

1 **Estimating the value of quarantine: eradicating invasive cane toads from tropical islands**

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3 Adam S Smart^{1*}, Reid Tingley¹ and Ben L Phillips¹

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5 ¹School of Biosciences, University of Melbourne, Melbourne, VIC, 3010, Australia

6 *Corresponding author: Adam Smart, email: asmart1@student.unimelb.edu.au

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14 **Abstract**

15 Islands are increasingly used to protect endangered populations from the negative impacts of invasive species.
16 Quarantine efforts on islands are typically undervalued, however. Using a field-based removal experiment, we
17 estimate the economic value of quarantine efforts aimed at keeping invasive cane toads (*Rhinella marina*) off
18 Australian islands. We estimate a mean density of 3444 [2744, 4386] individual toads per km² and a mean per-night
19 detection probability of 0.1 [0.07, 0.13]. Using a removal model and estimates of economic costs incurred during
20 toad removal, we estimate that eradicating cane toads would cost AUD\$96,556 per km². Across islands that have
21 been prioritized for conservation benefit across the toads predicted range, we estimate the remaining value of toad
22 quarantine to be more than \$1.3 billion. The value of a proposed waterless barrier on the mainland to prevent the
23 spread of toads into the Pilbara was in excess of \$26 billion. We conclude that quarantine of toads across Australia
24 provides substantial value in prevented eradication costs.

25 **Keywords:** Cane Toad, density, detection probability, eradication, islands, preventative management

26

27 **Introduction**

28 It is a truth universally acknowledged that an ounce of prevention is worth a pound of cure. Despite this truism,
29 conservation managers rarely ascribe value to preventative management. Without such valuation, we risk falling
30 prey to cognitive biases (e.g., immediacy bias), and so routinely commit substantially more money and effort to
31 tactical, “cure” type approaches, compared to strategic “prevention”. Quarantine against invasive species is a case in
32 point; vastly more resources are spent controlling the spread and impact of invaders than are spent on preventing
33 their arrival and establishment.

34 Quarantine is particularly likely to be undervalued in circumstances in which a failure incurs non-economic costs
35 (e.g., biodiversity loss) (Leung et al. 2002). One way to place a value on such quarantine efforts is to calculate the
36 cost of restoring the system to its former state. In the case of an invasive species with primarily non-economic
37 impacts, we can calculate the ongoing value of quarantine as the expense of a subsequent eradication program. Such
38 a valuation is a lower bound on the value of quarantine for two reasons. First, the same quarantine effort typically
39 protects against many potential invasive species. Second, any impact that the invasive species has before it is
40 eradicated (e.g., local extinction of a native species) must be added to the cost of restoration (Hoffmann &
41 Broadhurst 2016, Jardine & Sanchirico 2018). Thus, the cost of eradication of a single invader is a very conservative
42 estimate of the true value of quarantine efforts.

43 Islands are important resources for conservation quarantine because they offer a natural barrier to the spread of
44 invasive species. Conservationists routinely exploit this property of islands, not only to protect species that naturally
45 occur on islands, but also to provide refuge for species under threat on the mainland (Thomas 2011). In Australia
46 alone, 47 conservation translocations to islands have been carried out to date (Department of the Environment,
47 Water, Heritage and the Arts, 2009). In these circumstances – where the conservation value of an island has been
48 artificially bolstered – the arrival of invasive species can have a larger impact than they otherwise would.

49 Typically, island quarantine is used by conservation managers to protect native species from invasive predators (e.g.,
50 foxes, cats, weasels, rats). In Australia, however, islands are also used to mitigate the impact of cane toads (*Rhinella*
51 *marina*) on native predators (Moro *et al.* 2018). Cane toads were introduced to northeastern Australia in the 1930s
52 and, in northern Australia, continue to spread westerly at a rate of ~50 km per year (Phillips et al. 2010). This
53 invasion has had major impacts on populations of native predators, many of which have no resistance to the toad’s
54 toxin (Nelson et al. 2010; Greenlees et al. 2010; Llewelyn et al. 2014). In response to declines of multiple predator
55 species (e.g., dasyurids, monitors, snakes) the Australian government implemented the Cane Toad Threat Abatement
56 Plan (2011), which aimed to identify, and where possible reduce, the impact of cane toads on native species
57 (Shanmuganathan et al. 2010). A lack of viable methods for broad-scale control, however, has since led the
58 Australian government to place an increased emphasis on containment (on the mainland) and on quarantine (on
59 offshore islands) to mitigate the biodiversity impacts of cane toads.

60 While quarantine is currently the best available strategy, it is not a panacea: cane toads have already established on
61 at least 48 islands across northern Australia (McKinney et al. 2018 unpub data), with potential for further natural

62 and anthropogenic introductions. Thus, execution of the strategy outlined in the Cane Toad Threat Abatement Plan
63 requires ongoing quarantine and containment efforts. Here we estimate the monetary value of these ongoing efforts,
64 by quantifying the cost of eradicating cane toads from an island in northern Australia. We approach this problem by
65 estimating the density and detection probability of toads on an island, and use these estimates to calculate the
66 amount of time and money it would take to remove enough toads to ensure eradication.

67 **Methods**

68 **Study Area**

69 This study was carried out at Horan Island on Lake Argyle, Western Australia. Lake Argyle is Western Australia's
70 largest man-made reservoir and is located within the East Kimberly region. The study site is comprised of exposed
71 hilltops and savannah woodland. Freshwater is available year-round, with the lake contracting from May–
72 November. Toads are thought to have colonized islands on the lake via a flooding event in 2010/2011.

73 **Field sampling**

74 Cane toad surveys occurred over six nights, denoted, $t = \{0, 1, \dots, 5\}$, during November 2017. Surveys commenced
75 at 1830 each evening and lasted 3–5 hours. Temperatures ranged from 26–35°C. The entire island was
76 circumambulated each night by two people using headtorches; one individual focused on the higher part of the
77 shoreline, the other on the lower shoreline. Every toad encountered was collected and humanely killed on site in
78 accordance with The University of Melbourne animal ethics protocol (1714277.1) and State laws regarding handling
79 of non-native species. Each night, we recorded the number of individuals collected, c_t . Only post-metamorphic age
80 classes were encountered during sampling.

81 **Statistical analysis**

82 We assume that we do not encounter every individual on a given night, and so incorporate imperfect detection. We
83 aim to estimate two parameters: N_0 , the true number of toads on the island at the commencement of surveys, and p ,
84 the mean per-individual detection probability. The number of individuals collected each night, c_t , can be considered
85 a draw from a binomial distribution with:

$$86 \quad c_t \sim \text{Binom}(N_t, p).$$

87 Where N_0 , the pre-sampling population size, is a latent variable with a mean and variance equal to λ , such that:

$$88 \quad N_0 \sim \text{Poiss}(\lambda).$$

89 For $t > 0$:

$$90 \quad N_t = N_0 - \sum_0^{t-1} c_t$$

91 The model was fit with Markov chain Monte Carlo (MCMC) in JAGS v.4.6.0, run through R v3.4.1 via the package
92 rjags v4.6.0 (Plummer & Martyn 2013). Three model chains were run for 30,000 iterations, with the first 10,000
93 iterations discarded as a burn-in, which was sufficient for the MCMC chains to converge. Convergence was checked
94 using the Gelman-Rubin diagnostic (Gelman & Rubin 1992); all chains produced potential scale reduction factors <
95 1.1, indicating convergence of chains. The remaining samples were thinned by a factor of 2, resulting in 10,000
96 samples per chain for post-processing. Minimally informative prior distributions for p and λ were specified as
97 uniform between 0-1 and 0-10,000, respectively.

98 We denote a successful eradication to have occurred when only a single toad remains (i.e., no further breeding pairs
99 remain). As we assume that removal efforts take place on consecutive nights until completion, we disregard
100 breeding and immigration.

101 **Cost analysis**

102 We estimate the cost of eradicating toads on Horan Island based on consumable, personnel, and travel costs incurred
103 during toad collection (Table 1). Relative to most islands in northern Australia, Horan Island is readily accessible,
104 thus our travel costs are modest. We assume that eradication is conducted by a fully-equipped organization; hence
105 we do not include vehicle/boat purchase or hire (i.e., set-up costs). Travel costs include travelling to and from Horan
106 Island from Darwin, NT. Removal efforts are carried out on subsequent nights until eradication is reached; therefore,
107 the cost associated with travel to and from our site is incurred only once. Travel costs include a \$85/hour consultant
108 rate plus the additional costs of fuel, insurance, and vehicle maintenance (an extra \$36/hour based on a per km
109 charge). Thus, total travel costs are \$111/hour.

110 **Cost Scenarios**

111 We use our estimate of toad removal on Horan Island to highlight the potential value of quarantine efforts on high
112 priority islands, based on a list of the top 100 islands that the Australian Commonwealth has prioritized for
113 conservation due to their biodiversity value and presence of species listed under the Environment Protection and
114 Biodiversity Conservation Act (Department of the Environment and Energy [DEE], 1999). We refined this list to
115 include only islands that are >2 km from the Australian mainland, and occur within the predicted potential
116 distribution of cane toads (Kearney et al. 2008). All islands were cross-checked for the presence of cane toads
117 with the 'Feral Animals on Offshore Islands' database (DEE, 2016).

118 In addition to the islands derived from this report, we explore the value of a potential toad containment strategy
119 outlined in a revised version of the Cane Toad Threat Abatement Plan (unpub data). This strategy aims to develop a
120 'waterless barrier' on the Australian mainland by excluding cane toads from artificial water bodies on cattle stations
121 between Broome and Port Hedland in Western Australia. If implemented successfully, this strategy could keep toads
122 out of the Pilbara (and subsequent regions) –an effective quarantine of 268,000km² of the Australian mainland (see
123 Florance et al. 2011; Tingley et al. 2013; Southwell et al. 2017 for further information).

124 **Results**

125 The number of cane toads removed from Horan Island declined over time (Figure 1). Over the duration of our
126 survey, we captured and removed 1251 cane toads. The estimated probability of detecting an individual toad on a
127 given night was low (mean p [95% credible interval] = 0.1 [0.07, 0.13]; Figure 2). Given our survey effort and
128 estimated detection probability, the number of toads present on Horan Island at the initiation of our surveys (N_0) was
129 estimated to be 2638 [2140, 3421] (Figure 3). Horan Islands is 0.78km^2 , so this translates to a cane toad density of
130 3444 per km^2 [2744, 4386].

131 Using our estimates of p and λ , we examine the total survey effort (in days) required to eradicate toads from Horan
132 Island. Given our best estimate of N_0 , leaving a single individual is equivalent to leaving $r_{crit} = \frac{1}{N_0} = 0.000379$ of
133 the original number of individuals. The time to reach this point is given by $\ln r_{crit} / \ln(1 - p) = 72$ days. Multiplying
134 this estimate of the number of days required to achieve eradication by our incurred daily removal costs suggests that
135 \$77,670 (2017 AUD) would be required to eradicate toads from Horan island, or \$96,55 per km^2 of toad eradication.

136 **Value of quarantine**

137 Given our estimated per-area cost of cane toad eradication, we can explore the economic value of quarantine efforts
138 across Australian islands. We present the estimated value of island quarantine by State (Table 2), as well as the cost
139 required to eradicate toads from islands they have already colonized. We estimate it would cost \$415,151,032 to
140 remove cane toads from all islands across Australia on which they are currently known to be present. The remaining
141 value of quarantine across all toad-free islands in Australia is estimated at \$1,376,345,130, or a staggering
142 \$26,416,746,616 if the waterless barrier is implemented in Western Australia.

143 **Discussion**

144 Our results demonstrate the immense value of toad quarantine across Northern Australia. Using costs derived from
145 our removal efforts, we conservatively estimate that it costs \$98,569 per km^2 to eradicate cane toads. There is only
146 one instance in which cane toads have been successfully eradicated (Nonsuch Island in Bermuda: Wingate 2010).
147 This effort was the culmination of \$10,000 USD, six years, and countless hours of volunteer effort - all to remove
148 toads from an island roughly 0.065km^2 . Given the high monetary cost of toad eradication, and the susceptibility of
149 Australian fauna to multiple introduced species (e.g., cats, foxes, toads), island quarantine has significant value for
150 protecting declining populations from the detrimental impacts of invasive species (Ringma et al. 2018).

151 Our analysis of the feasibility and cost of cane toad eradication is timely, given renewed emphasis on Australia's
152 offshore islands as safe-havens to buffer biodiversity against cane toad impacts (Tingley 2017). Sixty-two offshore
153 islands designated as 'high conservation status' fall within the cane toads predicted distribution; 21 of those have
154 already been colonized by toads. Given these numbers, we estimate the remaining value of toad quarantine across
155 toad-free islands in northern Australia to be \$1.37B. This value is conservative for two reasons. First, it is a

156 reasonable expectation that as islands become home to increasing numbers of insurance populations or endangered
157 species, their value (measured as the cost of restoration) will increase. Second, as further islands are colonized by
158 toads, the value we place on remaining islands should increase. These questions are outside the scope of our analysis
159 but require careful consideration.

160 The vanguard of the cane toad invasion is currently sweeping across Western Australia at ~ 50 km per annum, but,
161 recent research suggests that a waterless barrier between the Kimberley and Pilbara could halt the toad invasion
162 (Florance et al. 2011; Tingley et al. 2013; Southwell et al. 2017). Applying our results to this management strategy
163 revealed that the value of quarantine over such an area (\$26.5B) is more than an order of magnitude higher than the
164 value of quarantine across all offshore islands combined (\$1.3B). Our results therefore hint at the immense value of
165 such a strategy, though questions surrounding implementation remain (Southwell et al. 2017).

166 Estimating the feasibility and cost of cane toad eradication enabled us to estimate a per-individual detection
167 probability and a density estimate. Despite substantial community and research effort into toad removal via trapping
168 and hand capture, we are unaware of a published detection estimate for this species. Our detection estimate is, of
169 course, specific to the details of our survey. Nonetheless, it is surprisingly low: individual toads in our closed system
170 had roughly a 10% chance of being seen on any given night. Individual toads are relatively easy to see when they
171 are active, but our results suggest that this gives a misleading impression of detectability.

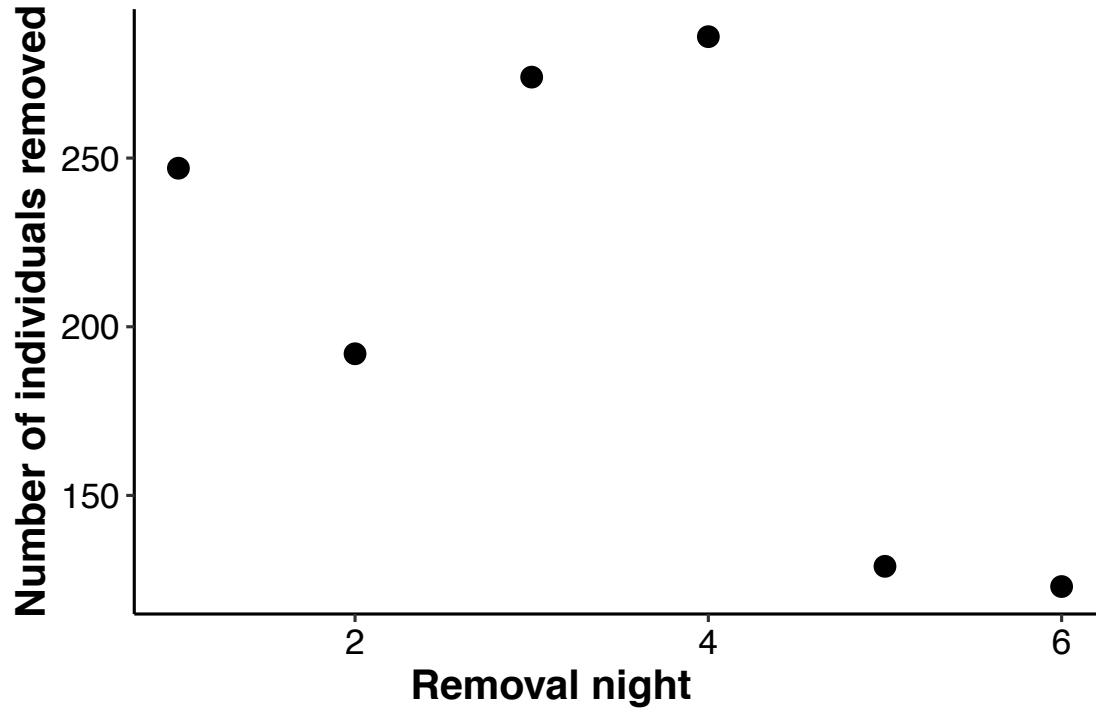
172 Our results suggest that cane toads can achieve a density of 3440 individuals per km². This estimate is similar to
173 density estimates derived from previous studies of invasive cane toads in the Solomon Islands archipelago (1035
174 individuals per km²; Pikacha et al. 2015) and the islands of Papua New Guinea (3000 individuals/km²; Zugg et al.
175 1975; Freeland et al. 1986). Studies conducted on the Australian mainland have reported densities as high as
176 213,400 individuals per km², but these higher estimates include metamorph toads, which often occur at very high
177 densities prior to dispersal. Metamorphs are strongly constrained to the edges of water bodies (Child et al. 2008),
178 and typically suffer high mortality from predation and desiccation before reaching maturity (Ward-Fear et al. 2010).
179 It is important to note, however, that our toad density estimate is derived from a population that is not limited by
180 access to freshwater. Cane toads are inherently limited in their ability to persist across the landscape in areas where
181 near-constant hydration is not possible (Florance et al. 2011). Our estimate is likely to be applicable solely to areas
182 in which water bodies persist through the dry season, such as natural catchments or structures engineered to house
183 water year-round (e.g., cattle watering points).

184 Placing a monetary value on preventative management is critical as conservation actions increasingly rely on
185 offshore islands and fenced areas as cost-effective avenues to protect biodiversity from the impacts of invasive
186 species. Quarantine measures often protect against multiple potential invaders but our results suggest that even when
187 considering a single species, the monetary value of quarantine can be substantial. Prevention, it seems, is worth
188 more than we might naively guess, even with aphorisms to remind us.

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193 research was supported by the Australian Research Council (FT160100198; DE170100601).

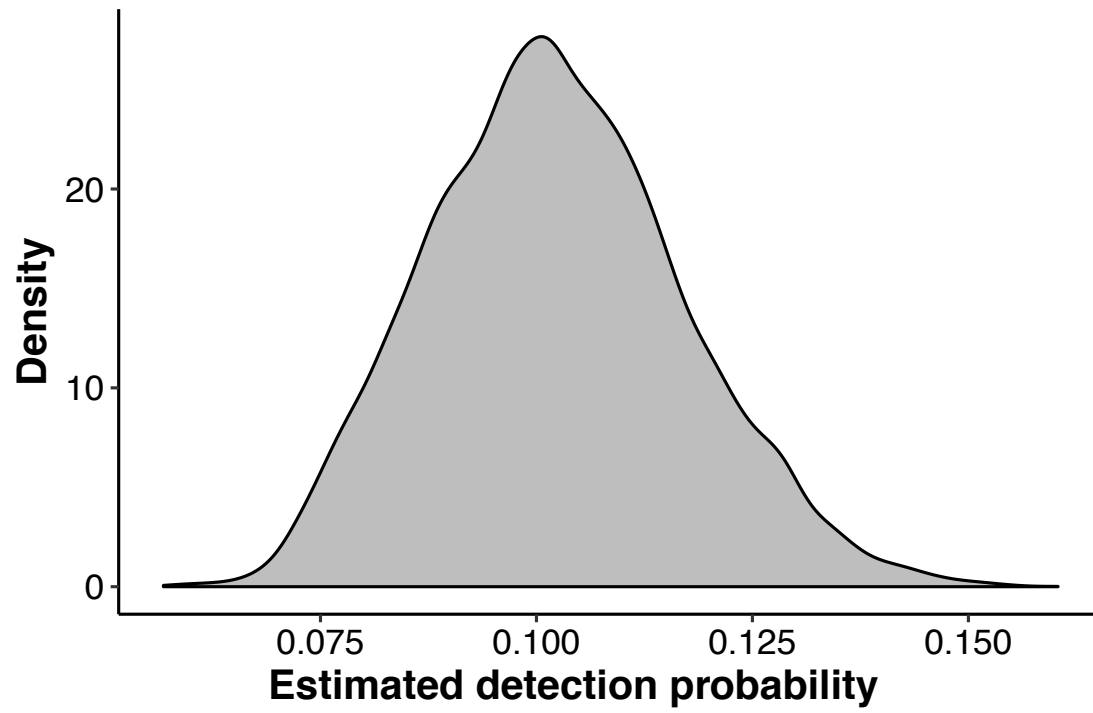
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196 Figure 1. The number of individual cane toads removed from Horan Island, NT over six consecutive collection
197 nights.

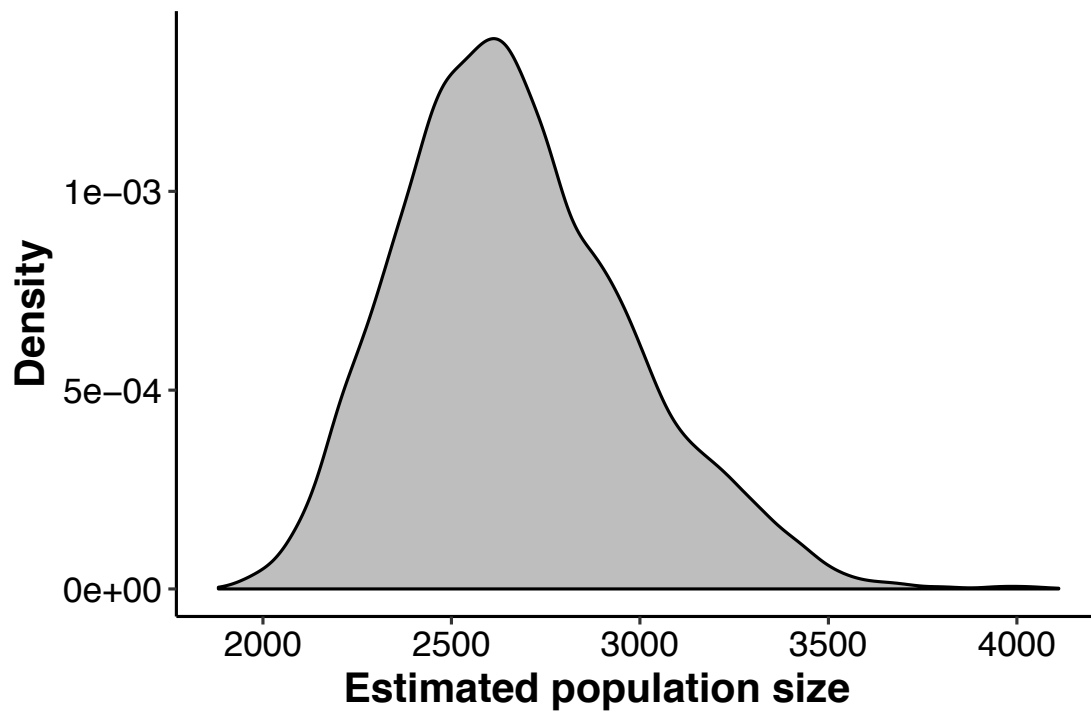
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200 Figure 2. Distribution of estimated detection probability of individual cane toads.

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203 Figure 3. Distribution of cane toad population size on Horan Island, NT.

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205 Table 1: Estimated costs (AUD) of cane toad eradication on Horan Island, NT.

| Cost category | Cost (AU) |
|---------------------------------|------------------|
| Consumable cost | \$5,611 |
| Personnel cost | \$61,200 |
| <i>Equipment preparation</i> | \$6,120 |
| <i>Survey time</i> | \$48,960 |
| <i>Waste disposal</i> | \$6,120 |
| Travel cost | \$9,768 |
| <i>Travel to site</i> | \$1,776 |
| <i>Motorized travel in site</i> | \$7,992 |

206

207 Table 2: Estimated cost (AUD) to eradicate cane toads across states and territories in Australia.

| Region within Australia | Cost to eradicate toads |
|---|--------------------------------|
| Cost to eradicate toads per km ² | \$98,569 |
| Horan Island | \$76,579 |
| WA islands (toad free) | \$112,578,720 |
| NT islands (toad present) | \$1,037,843,083 |
| NT islands (toad free) | \$121,443,116 |
| QLD islands (toad present) | \$224,148,083 |
| QLD islands (toad free) | \$293,707,917 |
| Waterless barrier (WA) | \$26,416,748,616 |

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209 **References**

- 210 1. Child, T., Phillips, B. L., Brown, G. P. and Shine, R. 2008. The spatial ecology of cane toads (*Bufo*
211 *marinus*) in tropical Australia: Why do metamorph toads stay near water? *Austral Ecology* **33**: 630-640.
- 212 2. Department of the Environment and Energy (2011). Environment Protection and Biodiversity Conservation
213 Act 1999 (EPBC Act). Retrieved from <http://www.environment.gov.au/epbc>
- 214 3. Department of Environment and Energy (2011). The biological effects, including lethal toxic ingestion,
215 caused by Cane Toads (*Bufo marinus*). Retrieved from
216 [http://www.environment.gov.au/system/files/resources/2dab3eb9-8b44-45e5-b249-651096ce31f4/files/tap-](http://www.environment.gov.au/system/files/resources/2dab3eb9-8b44-45e5-b249-651096ce31f4/files/tap-cane-toads.pdf)
217 [cane-toads.pdf](http://www.environment.gov.au/system/files/resources/2dab3eb9-8b44-45e5-b249-651096ce31f4/files/tap-cane-toads.pdf)
- 218 4. Department of the Environment, Water, Heritage and the Arts (2009). Prioritization of high conservation
219 status offshore islands. Retrieved from [https://www.environment.gov.au/system/files/resources/5325cdf1-](https://www.environment.gov.au/system/files/resources/5325cdf1-b56f-43b3-8bef-052d740d93fd/files/offshore-islands.pdf)
220 [b56f-43b3-8bef-052d740d93fd/files/offshore-islands.pdf](https://www.environment.gov.au/system/files/resources/5325cdf1-b56f-43b3-8bef-052d740d93fd/files/offshore-islands.pdf)
- 221 5. Department of the Environment and Energy (2016). Feral Animals on Offshore Islands Database. Retrieved
222 from <http://www.environment.gov.au/biodiversity/invasive-species/feral-animals-australia/offshore-islands>
- 223 6. Freeland, W. 1986. Populations of cane toad *Bufo marinus* in relation to time since colonization. *Wildlife*
224 *Research* **13**: 321–330. doi:10.1071/WR9860321
- 225 7. Florance, D., Webb, J. K., Dempster, T., Kearney, M. R., Worthing, A. and Letnic, M. 2011. “Excluding
226 Access to Invasion Hubs Can Contain the Spread of an Invasive Vertebrate.” *Proceedings of the Royal*
227 *Society B: Biological Sciences* **278**: 2900–2908. doi:10.1098/rspb.2011.0032.
- 228 8. Gelman, A., and Rubin, D. B. 1992. Inference from Iterative Simulation Using Multiple Sequences.
229 *Statistical Science* **7**: 457–511.
- 230 9. Greenlees, M. J., Phillips, B. L. and Shine, R. 2010. Adjusting to a Toxic Invader: Native Australian Frogs
231 Learn Not to Prey on Cane Toads. *Behavioral Ecology* **21**: 966–71. doi:10.1093/beheco/arq095.
- 232 10. Hoffmann, B D. and Broadhurst, L. M. 2016. The Economic Cost of Managing Invasive Species in
233 Australia. *NeoBiota* **31**: 1–18. doi:10.3897/neobiota.31.6960.
- 234 11. Jardine, S. L. and Sanchirico, J. N. 2018. Estimating the Cost of Invasive Species Control. *Journal of*
235 *Environmental Economics and Management* **87**: 242–57. doi:10.1016/j.jeem.2017.07.004.
- 236 12. Leung, B., Lodge, D. M., Finnoff, D., Shogren, J. F., Lewis, M. A. and Lamberti, G. 2002. An Ounce of
237 Prevention or a Pound of Cure: Bioeconomic Risk Analysis of Invasive Species. *Proceedings of the Royal*
238 *Society B: Biological Sciences* **269** (1508): 2407–13. doi:10.1098/rspb.2002.2179.
- 239 13. Llewelyn, J., Schwarzkopf, L., Phillips, B. L. and Shine, R. 2014. After the Crash: How Do Predators
240 Adjust Following the Invasion of a Novel Toxic Prey Type?: Adjusting to a Novel Toxic Prey Type.
241 *Austral Ecology* **39** (2): 190–97. doi:10.1111/aec.12058.
- 242 14. Nelson, D. W. M., Crossland, M. R. and Shine, R. 2010. Indirect Ecological Impacts of an Invasive Toad
243 on Predator–prey Interactions Among Native Species. *Biological Invasions* **12** (9): 3363–9.
244 doi:10.1007/s10530-010-9729-4.
- 245 15. Phillips, B. L., Brown, G. P. and Shine, R. 2010. Evolutionarily Accelerated Invasions: The Rate of
246 Dispersal Evolves Upwards During the Range Advance of Cane Toads: Dispersal Evolution During Range
247 Advance. *Journal of Evolutionary Biology* **23** (12): 2595–2601. doi:10.1111/j.1420-9101.2010.02118.x.
- 248 16. Pikacha, P., Lavery, T. and Leung, L. K. P. 2015. What Factors Affect the Density of Cane Toads (*Rhinella*
249 *Marina*) in the Solomon Islands? *Pacific Conservation Biology* **21** (3): 200. doi:10.1071/PC14918.

- 250 17. Plummer, Martyn. 2013. rjags: Bayesian graphical models using MCMC. R package version 3-10. URL:
251 <http://CRAN.R-project.org/package=rjags>
- 252 18. Ringma, J., Legge, S., Woinarski, J., Radford, J., Wintle, B. and Bode, M. 2018. Australia's mammal fauna
253 requires a strategic and enhanced network of predator-free havens. *Nature Ecology & Evolution* **2**: 410-411.
- 254 19. Shanmuganathan, T., Pallister, J., Doody, S., McCallum, H., Robinson, T., Sheppard, A., Hardy, C.,
255 Halliday, D., Venables, D., Voysey, R., Strive, T., Hinds, L. and Hyatt, A. 2010. Biological Control of the
256 Cane Toad in Australia: A Review: Biological Control of Cane Toad. *Animal Conservation* **13**: 16–23.
257 doi:10.1111/j.1469-1795.2009.00319.x.
- 258 20. Southwell, D., Tingley, R., Bode, M., Nicholson, E. and Phillips, B. L 2017. Cost and Feasibility of a
259 Barrier to Halt the Spread of Invasive Cane Toads in Arid Australia: Incorporating Expert Knowledge into
260 Model-Based Decision-Making. *Journal of Applied Ecology* **54** (1): 216–24. doi:10.1111/1365-
261 2664.12744.
- 262 21. Thomas, C. D. 2011. Translocation of Species, Climate Change, and the End of Trying to Recreate Past
263 Ecological Communities. *Trends in Ecology & Evolution* **26** (5): 216–21. doi:10.1016/j.tree.2011.02.006.
- 264 22. Tingley, R., Ward-Fear, G., Schwarzkopf, L., Greenlees, M. J., Phillips, B. L., Brown, G., Clulow, S.,
265 Webb, J., Capon, R., Sheppard, A., Strive, T., Tizard, M. and Shine, R. 2017. New weapons in the toad
266 toolkit a review of methods to control and mitigate the biodiversity impact of invasive cane toad (*Rhinella*
267 *Marina*). *The Quarterly Reviews of Biology* **92**: 123-149.
- 268 23. Tingley, R., Phillips, B. L., Letnic, M., Brown, G. P., Shine, R. and Bair9d, S. J. E. 2013. Identifying
269 Optimal Barriers to Halt the Invasion of Cane Toads *Rhinella Marina* in Arid Australia. *Journal of Applied*
270 *Ecology* **50** (1): 129–37. doi:10.1111/1365-2664.12021.
- 271 24. Ward-Fear, G., Brown, G. P. and Shine, R. 2010. Using a Native Predator (the Meat Ant, *Iridomyrmex*
272 *Reburus*) to Reduce the Abundance of an Invasive Species (the Cane Toad, *Bufo Marinus*) in Tropical
273 Australia. *Journal of Applied Ecology* **47** (2): 273–80. doi:10.1111/j.1365-2664.2010.01773.x.
- 274 25. Wingate, D. B. 2011. The successful elimination of Cane Toad, *Bufo marinus*, from an island with breeding
275 habitat off Bermuda. *Biological Invasions* **13**: 1487-1492.
- 276 26. Zug, G., Lindgren, E. and Pippet, J. 1975. Distribution and ecology of the marine toad, *Bufo marinus*, in
277 Papua New Guinea. *Pacific Science* **29**:31–50.
- 278