1	
2	
3	Using eDNA to biomonitor the fish community in a tropical oligotrophic lake
4	
5	
6	Martha Valdez-Moreno ^{1¶} , Natalia V. Ivanova ^{2*¶} , Manuel Elías-Gutiérrez ¹ , Stephanie L.
7	Pedersen ² , Kyrylo Bessonov ² , Paul D.N. Hebert ²
8	
9	¹ – El Colegio de la Frontera Sur, Unidad Chetumal, Av. Centenario Km 5.5, Chetumal,
10	Quintana Roo, Mexico
11	² – Centre for Biodiversity Genomics, Biodiversity Institute of Ontario, University of Guelph,
12	Guelph, ON N1G 2W1, Canada
13	
14	*Corresponding author:
15	E-mail: <u>nivanova@uoguelph.ca</u> (NVI)
16	
17	[¶] These authors contributed equally to this work.
18	

19 Abstract

Environmental DNA (eDNA) is an effective approach for detecting vertebrates and plants, 20 especially in aquatic ecosystems, but prior studies have largely examined eDNA in cool 21 22 temperate settings. By contrast, this study employs eDNA to survey the fish fauna in tropical 23 Lake Bacalar (Mexico) with the additional goal of assessing the possible presence of invasive fishes, such as Amazon sailfin catfish. Sediment and water samples were collected from eight 24 25 stations in Lake Bacalar on three occasions over a 4-month interval. Each sample was stored in the presence or absence of lysis buffer to compare eDNA recovery. Short fragments (184-187 26 27 bp) of the cytochrome c oxidase I (COI) gene were amplified using fusion primers and then 28 sequenced on Ion Torrent PGM and S5 before their source species were determined using a 29 custom reference sequence database constructed on BOLD. In total, eDNA sequences were recovered from 75 species of vertebrates including 47 fishes, 15 birds, 7 mammals, 5 reptiles, 30 and 1 amphibian. Although all species are known from this region, 6 fish species represent new 31 records for the study area, while 2 require verification. Sequences for five species (2 birds, 2 32 33 mammals, 1 reptile) were only detected from sediments, while sequences from 52 species were only recovered from water. Because DNA from the Amazon sailfin catfish was not detected, we 34 used a mock eDNA experiment to confirm our methods were appropriate for its detection. We 35 36 developed protocols that enabled the recovery of eDNA from tropical oligotrophic aquatic 37 ecosystems, and confirmed their effectiveness in detecting diverse species of vertebrates 38 including an invasive species of Amazon catfish.

39 Introduction

Environmental DNA (eDNA) has gained popularity for biomonitoring, especially for the 40 41 detection of invasive species and for baseline surveys of animal and plant communities [1,2]. This 42 eDNA derives from cells shed into the environment as mucus, urine, feces, or gametes [2–7]. More than 120 articles have now considered eDNA, including a special issue on the topic [8]. 43 The rapid adoption of eDNA-based bioassessments reflects two factors: improved access to the 44 reference sequences [9,10] required to link the short reads from eDNA to their source species, 45 and the availability of high-throughput sequencers that can generate large volumes of data at 46 47 modest cost. Conventional biomonitoring programs require extensive effort with sampling repeated over time 48 49 and space [11]. Because these methods utilize nets and electrofishing, they are invasive and can damage habitats [3,12,13]. Furthermore, they rarely reveal all species in an environment, and 50 51 each specimen that is collected must be identified to a species, requiring access to taxonomic 52 experts [14,15]. Given these limitations, eDNA-based methods are an attractive approach for the 53 quantification of biodiversity required for conservation and management [8,16]. 54 Most prior eDNA studies have focused on the detection of invasive fish species [2,3,17] or rare/ 55 endangered aquatic species [18,19] although they have also been used to examine the abundance 56 and biomass of fish species [20,21]. Early results revealed complexities in data acquisition and

interpretation that have provoked studies to improve sampling and DNA extraction while also
minimizing methodological biases [16,22,23].

Varied protocols have been used to preserve eDNA, but they usually involve the immediate
storage of samples at low temperature [21], precipitation [24], the addition of lysis buffers to

61	filters [25], and the inclusion of surfactants [26]. While eDNA has traditionally been assumed to
62	rapidly degrade, this is not always the case. For example, Deiner et al. [27] employed long-range
63	PCR (>16 kb) to recover whole mitochondrial genomes from the six fish species in a mock
64	community, and from 10 of 12 species represented in water samples from a stream. Prior studies
65	have also demonstrated that eDNA metabarcoding is an effective alternative to traditional
66	biomonitoring as it has revealed the species composition of fish communities in various
67	ecosystems including a species-rich coastal sea in Japan [28] and a river in Indiana, USA [29]. In
68	fact, in both cases, it revealed species that were overlooked by other sampling methods.
69	Although most past eDNA studies have examined temperate ecosystems, there is increasing
70	interest in the use of this method in the tropics [30–32]. The present study sought to develop an
71	eDNA protocol for biomonitoring the fish fauna of oligotrophic tropical Lake Bacalar and
72	adjacent habitats, and to verify its capacity to detect an invasive fish, the Amazon sailfin catfish,
73	Pterygoplichthys sp., which represents a serious threat in this region.

74 Material and Methods

75 Study site

Lake Bacalar, the largest freshwater habitat in the Yucatan Peninsula, is renowned for its striking blue color, the clarity of its water, and for the world's largest occurrence of microbialites. With a length of more than 10 km [33–35], it occupies a larger fault basin [36] (50 km long, 2 to 3 km wide). With sediments derived from karst limestone, it represents the world's largest freshwater lens system [37]. Lake Bacalar is not directly connected to the sea, but it shows a high rate of groundwater flow [38]. The northern part is connected to a complex system of lagoons including

82	Laguna Chile Verde and Laguna Guerrero. The southern part of the lake has an indirect
83	connection to the sea via the Hondo River that enters Chetumal Bay. There are four cenotes
84	(sinkholes) in the lake whose depths range from 50-90 m.
85	Water temperatures in Lake Bacalar range from 25-28 °C, while the Secchi transparency
86	averages 10.3 m. The lake is unstratified [33,39] with a water temperature of 27°C or higher
87	throughout the year. Its surface waters are slightly alkaline (pH=7.8) with a conductivity of 1220
88	μ S/cm [39] and HCO ₃ values higher than in marine environments [33].
89	All samples were taken from the littoral zone at a maximum depth of 0.5 m, except La Unión (2
69	An samples were taken from the intoral zone at a maximum depth of 0.5 m, except La Omon (2
90	m) and Xul-Ha (4 m). Eight sampling sites were examined along the main axis of Lake Bacalar
91	and associated systems in the Hondo River (Fig 1). Samples were collected in December 2015,

92 January 2016, and April 2016.

93 Fig 1. Sampling locations.

94 Sampling

Eight localities (one with three points) (Fig 1) were sampled and most were examined on more

96 than one occasion, producing a total of 14 sampling events. Three sites (Cocalitos, Huay Pix and

97 Xul-Ha) were sampled in December 2015, four sites in January 2016 (Alvaro Obregón Viejo,

Huay Pix, La Unión and Pedro A. Santos), and seven sites in April 2016 (Cacao, Cocalitos, Huay

99 Pix, Juan Sarabia, La Unión, Pedro A. Santos and Xul-Ha).

100 All sampling bottles and equipment were handled with gloves to minimize contamination with

101 human DNA. At each site, three replicate samples of water and sediment were taken. Each 1L

102 water sample was placed in a sterilized CIVEQ® bottle, while each sediment sample was placed

in a separate (100 cm³) Ziplock bag. Each water and sediment sample was collected using a new
turkey baster. Sediment samples were obtained from the upper 1 cm. All samples were
immediately placed on ice, and then transported to the lab in Chetumal for processing.

106 **Processing water samples**

107 Water samples were filtered within 7 hours of collection. Prior to processing, all lab surfaces and

108 materials were sterilized with 10% bleach, followed by 70% ethanol; gloves were changed

between each sample. Each sample was split into two 0.5 L subsamples that were filtered

110 through separate 0.22 µm filters. One of these filters was stored with 1 ml of PW1 solution with

111 grinding media while the other filter was placed with grinding media in a dry tube covered by

aluminum foil. This approach resulted in six samples per locality for each collection event (Fig

113 2). All filters were stored at -18°C before being transported on ice from Chetumal to Guelph

114 where DNA extraction was undertaken. The interval between filtration and DNA extraction was

always less than 48 hours.

116 Fig 2. Experimental design for the water and sediment samples.

117 **Processing sediment samples**

During the first collection trip, three replicate sediment samples (20 cm³ each) were collected at each site. Each sample was split into two 10 cm³ subsamples; 2 ml of 5M GuSCN buffer was either used as a preservation buffer or was added during lysis. For the second and third collection events, one pooled sample (60 cm³) was collected at each site and split into two replicates of 30 cm³; 6 ml of 5M GuSCN buffer [40] was used as a preservation buffer or was added during lysis step (Fig 2). Sediments were kept at -18°C before transportation on ice to Guelph.

124 **DNA extraction from water samples**

Prior to DNA extraction, all lab surfaces and pipettors were sterilized using 10% bleach,
followed by 70% ethanol. Pipettors were repeatedly cleaned with 10% bleach, followed by 70%
ethanol during extraction and gloves were frequently changed. Centrifuge rotors, adapters, and
tube racks were washed with diluted ELIMINase (Decon Labs) (1:10) and rinsed with deionized
water.

130 DNA was extracted from the filters from the water samples using a PowerWater DNA extraction

131 kit (MoBio) following a modified lysis protocol (tubes were incubated at 65°C prior to the

132 grinding stage) with minor modifications to avoid cross-contamination between samples; all

steps requiring 650 μ l were reduced to 625 μ l. Columns were incubated at 56°C for 15 min prior

to DNA elution in 100 μl of PW6 buffer and quantified using Qubit 2.0 and HS dsDNA kit.

DNA extraction from sediment samples

136 All lab surfaces and equipment were decontaminated as described above. The DNA extraction protocol employed ProK digestion in the presence of chaotropic salts followed by the use of 137 138 magnetic beads to bind and then release DNA. Sera-Mag SpeedBead Carboxylate-Modified Magnetic Particles (Hydrophilic) from GE Healthcare were employed to bind/release DNA. 139 140 They were prepared as described in [41] in a polyethylene glycol/NaCl buffer (0.1% Sera 141 Magnetic Particles (w/v) in 18% PEG-8000, 1M NaCl, 10 mM Tris-HCl, pH 8.0, 1 mM EDTA 142 pH 8.0, 0.05% Tween 20). Extraction was initiated by adding a 1/5 volume of 5M GuSCN lysis buffer [40] to non-treated samples and a 1/10 volume of ProK (20 mg/ml). Each tube was 143 vortexed and incubated for 2 hours at 56°C on an orbital shaker, then held for 30 min at 65° C 144 145 before being centrifuged at 2000 g for 2 min. DNA was extracted from 500 µl of the resultant

146 lysate after its transfer to a 2 ml Eppendorf LoBind tube. 500 μ l of prepared magnetic beads was 147 added to each lysate, and a pipettor was used to mix the beads and lysate before incubation at 148 room temperature for 10 min. Tubes were then placed on a magnet (DynaMag) and incubated 149 until the supernatant was clear. Beads were thoroughly washed once with 1 ml of PWB buffer 150 [42] and twice with 1 ml of 70% ethanol. Magnetic beads were then dried at room temperature 151 for 10-15 min before DNA was eluted into 100 μ l of 10 mM Tris-HCl, pH 8.0 and quantified 152 using Qubit 2.0 and HS dsDNA kit.

PCR amplification

Prior to PCR amplification, each DNA extract from sediments was diluted 1:10 with moleculargrade water while extracts from water were used without dilution.

156 In order to increase species recovery and overcome possible PCR bias introduced by UMI-

157 labelled fusion primers [43] we followed a two-step PCR approach [43] with the first round

158 employing conventional primers, where the diluted PCR product serves as a template for a

second round of PCR with fusion primers containing sequencing adapters and UMI-tags.

160 A 184-187 bp segment of the barcode region of COI was amplified with two primer sets

161 (AquaF2/C_FishR1, AquaF3/C_FishR1) known to be effective for vertebrates (Table 1). The

162 PCR reactions employed the master mix described in [44] and Platinum Taq. The first round of

163 PCR employed the following thermal regime: 94°C for 2 min, 40 cycles of 94°C for 40 s,

annealing at 51°C for AquaF2 or 50°C for AquaF3 for 1 min, and 72°C for 1 min, with a final

165 extension at 72°C for 5 min. Following the first round, PCR products were diluted 2-fold with

166 molecular grade water, and 2 μ l was transferred to a new tube for a second round of PCR to

167 create amplicon libraries tagged with unique molecular identifiers (UMIs) with fusion primers

- 168 containing Ion XpressTM MID tags and Ion Adapters (Table 4). The same UMI-tag was applied
- to all three replicates for each treatment from a particular collection event (Fig 6). The PCR
- regime for the second round consisted of 94°C for 2 min, 20 cycles of 94°C for 40 s, annealing at
- 171 51°C for 1 min, and 72°C for 1 min, with a final extension at 72°C for 5 min. PCR products were
- 172 visualized on an E-Gel96 (Invitrogen, Thermo Fisher Scientific).

173 **Table 1. Primers used for eDNA amplification**

Region/Primer name	Direction	Primer sequence	Reference
COX1			
AquaF2	F	ATCACRACCATCATYAAYATRAARCC	[44]
AquaF3	F	CCAGCCATTTCNCARTACCARACRCC	this study
C_FishR1		Cocktail primers (FR1d: FishR2; 1:1)	[45]
FR1d	R	ACCTCAGGGTGTCCGAARAAYCARAA	
FishR2	R	ACTTCAGGGTGACCGAAGAATCAGAA	
NGS-fusion			
IonA	CCATCTCATCCCTGCGTGTCTCC[GACT] on A F [IonExpress-MID][specific sequence]		Ion Torrent, Thermo Fisher Scientific
trP1	R	CCTCTCTATGGGCAGTCGGTGAT [specific sequence]	Ion Torrent, Thermo Fisher Scientific

175	Libraries for AquaF2/C_FishR1 and AquaF3/C_FishR1 were purified separately using magnetic
176	beads prepared as described in [41], quantified using Qubit 2.0 and HS dsDNA kit, and
177	normalized to 13 pM prior to sequencing. Most amplicon libraries were sequenced on an Ion
178	PGM using the 400 Template and Sequencing kits following manufacturer's instructions with
179	either a 316 or 318 chip. Water samples from April 2016 were also analyzed on an Ion S5 using
180	the Chef for templating and an Ion 520 and 530 kit with a 530 chip. Fig 7 summarizes the
181	treatment for each water or sediment sample. HTS data (FASTQ files) are available in the
182	European Nucleotide Archive (ENA) under submission number: PRJEB25950.

Detection of an invasive fish

We conducted a mock eDNA experiment to ensure our methods could detect species of *Pterigoplichthys*. Two of these fishes were kept in a 10 L aquarium for 10 days at 25°C. One
liter of water was collected and two 0.5 L subsamples were filtered through separate 0.22 µm
filters; filters were stored with buffer and no buffer as described above.

188 Bioinformatics workflow

We created a reference dataset (DS-EBACALAR) on BOLD with sequences from 3,534
Mexican fishes representing 576 BINs, including 219 records from Lake Bacalar and
surrounding habitats (Hondo River, and adjacent lakes – Huay Pix, Xul-Ha). This reference
barcode library included sequences for 53 of the 57 fish species (Table 2) known from these
habitats (see Fig 1). As well, four other datasets with all public data on BOLD for Amphibia,
Mammalia, Aves, and Reptilia were constructed. These five datasets were employed as a basis
for assigning a source species to each OTU recovered from the present eDNA analyses.

196	The following procedure was used to process the raw NGS reads: Cutadapt (v1.8.1) was used to
197	trim primer sequences; Sickle (v1.33) was used for size filtering (sequences from 100-205 bp
198	were retained), while Uclust (v1.2.22q) was employed to recognize OTUs based on the >98%
199	identity and a minimum read depth of 2 thresholds. The Local Blast 2.2.29+ algorithm was then
200	used to compare each OTU to the reference sequences in five datasets: DS-EBACALAR Bacalar
201	Fish Dataset for eDNA Detection (3,534 sequences), and public BOLD data for Amphibia, Aves,
202	Mammalia, and Reptilia represented by the following datasets: DS-EBACAMPH (11,018
203	sequences), DS-EBACAVES (28,914 sequences), DS-EBACMAMM (39,890 sequences), DS-
204	EBACREPT (5,424 sequences). Raw Blast output results were parsed using a custom-built
205	python script OTUBlastParser.py, and concatenated using ConcatenatorResults.py available at
206	https://gitlab.com/kbessonov/OTUExtractorFromBLASToutput; processed results in tab-
207	delimited format were imported to MS Excel, filtered by min score of 250, and 97-100% percent
208	identity range. Filtered results werefurther processed and visualized in Tableau software 10.2.
209	All human reads were excluded from the final results and species counts.
210	Species accumulation curves were calculated in EstimateS software available at
211	http://purl.oclc.org/estimates [46], with the following settings: 1000 randomizations; do not
212	extrapolate rarefaction curves; classic formula for Chao 1 and Chao 2. Results were exported to
213	Microsoft Excel and visualized in Tableau software 10.4.
214	Sequences with 97-100% identity to fish species represented in the DS-EBACALAR dataset

- were mapped onto a neighbor-joining Taxon ID tree generated in BOLD for these taxa, and
- visualized in iTOL [47]. Results

- 217 The eDNA analyses recovered sequences from 75 species of vertebrates including 47 fishes
- (Table 2), 15 birds, 7 mammals, 5 reptiles and 1 amphibian, most previously known from this
- region. Although substantially more species were recovered from the water than from the
- sediments (Table 3, Fig 3), five species were only recovered from sediments racoon (Procyon
- 221 *lotor*), pig (*Sus scrofa*), meso-american slider turtle (*Trachemys venusta*), brown jay (*Psilorhinus*
- 222 *morio*), and great blue heron (*Ardea herodias*). However, all fish species detected in sediment
- samples were also recovered from the water samples (Table 3, S1 Table).
- Fig 3. Number of species detected versus the number of samples analyzed for water,
- sediments, and combined. Species accumulation curves were calculated in EstimateS software
- with 1000 randomizations and classic formula for Chao 1 and Chao 2.

227 Table 2. Fishes that were detected in eDNA, known from the literature and those with

228 barcodes.

Species list	Reference	Sequence	Barcode
	known from the	recovered from	sequence
	study area	eDNA	available
Achirus lineatus	[48]		+
Anchoa parva	[49,50]		N/A
Anchovia clupeoides	[49]		+
Astyanax aeneus	[50–53]	*	+
Astyanax mexicanus		*	+
Atherinella alvarezi	[49,52]		+
Atherinella sp.	[48,49]	*√	+

Atherinomorus stipes [50] Bagre marinus [48] Bardiella ronchus *√	+ +
	+
Dandiella nonehus *1	
	+
Bathygobius curacao [49]	+
Bathygobius soporator [50] *	+
Belonesox belizanus [48,50,51,54] *	+
Bramocharax caballeroi *	+
Caranx hippos [48]	+
Centropomus undecimalis [48]	+
Chriodorus atherinoides [49,52]	+
Cribroheros robertsoni [48,54] *	+
Cryptoheros chetumalensis [48,53–55] *	+
Ctenogobius fasciatus *✓	+
<i>Cyprinodon artifrons</i> [49–52,54]	+
Cyprinodon beltrani /simus *?	+
Dormitator maculatus [49,50]	+
Dorosoma anale *✓	+
Dorosoma petenense [48,50,54] *	+
Engraulidae *1	+
<i>Eugerres plumieri</i> [48,54] *	+
Floridichthys polyommus[49,50,52,54,56]	+
Gambusia sexradiata [48,49,51,52,54] *	+
Gambusia yucatana [48,49,51,54] *	+

Gobiomorus dormitor	[48,49,52,54]	*	+
Gobiosoma sp.		*√	+
Heterandria bimaculata	[48–50,54]	*	+
Hyphessobrycon	[48,49,54]	*	+
compressus			
Hyporhamphus roberti	[49,52]		N/A
Ictalurus meridionalis	[48]		+
Jordanella pulchra	[48,49,54,56]	*	+
Lachnolaimus maximus		*	+
Lophogobius cyprinoides	[49,53,54,56]	*	+
Lutjanus griseus	[48]		+
Mayaheros urophthalmus	[53–55]	*	+
Megalops atlanticus	[48,49,54]	*	+
Mugil cephalus	[49]		+
Oligoplites saurus		*√	+
Ophisternon sp.	[48,49,54]	*	+
Oreochromis mossambicus	[48,50]	*	+
Oreochromis niloticus	[48]	*	+
Parachromis friedrichsthalii	[48,51,52,54]	*	+
Petenia splendida	[48,53–55]	*	+
Phallichthys fairweatheri	[48,49,52,54]	*	+
Poecilia mexicana	[48,51,54]	*	+
Poecilia orri	[48,52,54]		+

Poecilia petenensis	[48,49,52,54]	*	+
Pterygoplichthys sp.		X	+
Rhamdia guatemalensis	[48,50]	*	+
Rhamdia laticauda	[48,49,54]	*	+
Rocio octofasciata	[48,51,54]	*	+
Sciades assimilis	[48,49,51,52,56]	*	+
Sphoeroides testudineus	[48,54]		+
Strongylura notata	[49,52,54,56]	*	+
Strongylura timucu	[48]	*	+
Thorichthys aureus	[48]		N/A
Thorichthys aff. meeki	[49,56]		+
Thorichthys meeki	[48,51–54]	*	+
Trichromis salvini	[48,52,53,55]	*	+
Urolophus jamaicensis	[49,50]		N/A
Vieja fenestrata		*?	+
Vieja melanura	[48]	*	+
Xiphophorus maculatus	[48,49,51]	*	+

Number of t	axa	57		48	65	
			 		 	_
	1 D M A 1		DOID 1			

229

+ – Species with DNA barcode present in the BOLD dataset; * – detected in eDNA; \checkmark –

230 potentially new record or species for the study area; ? – questionable presence; X – detected in

231 mock experiment; N/A – no barcode sequence available.

	Total	Sediments only	Sediments + water	Water only
Actinopterygii	47	0	14	33
Amphibia	1	0	1	0
Reptilia	5	1	1	3
Mammalia	7	2	1	4
Aves	15	2	1	12
Total:	75	5	18	52

Table 3. Number of species recovered from the sediment and water samples.

234

The three fish orders with highest number of detected species were Cichliformes (12 species),

236 Cyprinodontiformes (11 species), and Gobiiformes (5 species) (S2 Table). Although

237 Pterygoplichthys sp. was not recovered from water and sediment samples, it was recovered from

the mock eDNA study (S1 Table and S2 Table).

In total, 272,610 reads obtained by the PGM from the water and sediment samples showed a 239 240 close match to a vertebrate species in one of the five reference libraries. The three sites with the 241 most fish species recovered were Alvaro Obregón Viejo, Cocalitos and Xul-Ha (Fig 4, Table 4), 242 with 24-25 fish species per site. On average, the sediment and water samples from a particular 243 site on a particular date generated eDNA sequences for 45% (mean =20.9; range= 15-25) of the 244 47 fish species detected in the survey (Table 4). Although nearly 2/3 of these reads derived from sediments (177,077 reads versus 95,533 from water), more species of vertebrates were recovered 245 246 from the water samples (Fig 4). In fact, no species were recovered from the sediment samples 247 taken at three sites (Cocalitos, Huay Pix, Xul-Ha) in December 2015 or from those preserved 248 without buffer from Alvaro Obregón Viejo in January 2016. Although water samples revealed

- 249 more species, there was evidence of temporal shifts in species recovery and DNA concentration
- with the lowest value in December 2015 (Fig 4, 5A). Best results were obtained with filtered
- 251 water rather than sediments for both treated and untreated samples (Fig 3, 4, 5A).
- **Table 4. Number of vertebrate species detected at each of the eight sampling sites**

Sampling site No	Sampling site	Actinopterygii	Amphibia	Aves	Mammalia	Reptilia
1	Pedro A. Santos / 18.920052°; -88.170100°	18		2	1	
2	Cocalitos / 18.651469°; -88.408948°	24	1	5		1
3	Xul-Ha / 18.560084°; -88.446216°	25		1		2
4 a	Huay Pix / 18.514639°; -88.426750°	19		3		
4 b	Huay Pix / 18.528561° -88.429242°	15		1		
4 c	Huay Pix / 18.516144°; -88.436876°	19		1	1	1
5	Juan Sarabia/ 18.493471°; -88.478857°	19		1	1	
6	Alvaro Obregón Viejo / 18.299259°; -88.597250°	24		6	2	1
7	Cacao / 18.186488°;-88.694636°	23		1	2	
8	La Unión / 17.894676°; -88.870059°	23		4	4	2

253

Fig 4. Number of vertebrate species recovered in eDNA from each sampling event at each

255 collection site.

257 Fig 5. Species recovery and DNA concentration for each sampling event and treatment. A –

number of vertebrate species recovered for each sampling event and treatment (only PGM data

- 259 for fishes are shown, circle size indicates average read count per sampling event). B DNA
- 260 concentration of sediment and water samples (one replicate per sampling event) measured on
- 261 Qubit. Whiskers correspond to data within 1.5x of interquartile range (IQR).
- 262 With two exceptions, negative PCR and DNA extraction controls did not produce any vertebrate
- reads aside from those deriving from humans (S2 Table). Two reads of *Trichromis salvini* were
- recovered from a negative control on the S5, and two reads of *Mayaheros urophthalmus* were
- obtained from a negative control on the PGM. Because both cases involved a species common in
- samples from the same run, they likely reflect tag-switching due to a misread UMI [57,58] or
- cross-contamination. Given their rarity, it is unlikely such analytical errors significantly
- 268 impacted our overall conclusions.

Comparison of results from two sequencing platforms

Two sequencing platforms (PGM, S5) were used to analyze the same set of water samples from
April 2016 and from the mock experiment. The PGM generated 57,689 reads that matched
vertebrates while the S5 generated 1,106,574 vertebrate reads. Reflecting its 20-fold higher
sequence count, the S5 recovered more fish species than the PGM (41 vs. 34). However, large
shifts in the relative number of read counts for the component species were also detected. For
example, *Trichromis salvini* comprised 59.1% of the S5 reads, but just 5.9% of those from PGM
(Fig 6).

Fig 6. Percentage of total read coverage for each species recovered in water samples from
April 2016 and mock experiment, analyzed with Ion Torrent PGM and S5 instruments.

279	To verify the accuracy of the identifications assigned to the sequences recovered from eDNA, we
280	generated a Neighbor-Joining tree in BOLD for detected fish species and mapped corresponding
281	top-hit process IDs (Fig 7).
282	Fig 7. Top hit process IDs mapped on the Neighbor-Joining tree for fish taxa visualized in
283	iTOL. Blue – species previously reported from the Lake Bacalar region; red – species new to

Lake Bacalar); orange – species complexes lacking resolution with COI: 1 – *Bramocharax*-

285 Astyanax complex, 2 – Cyprinodon simus/beltrani; green – Pterygoplichthys from mock eDNA

286 experiment.

287 **Discussion**

Although eDNA has often been thought to degrade rapidly, factors such as temperature,

alkalinity, and trophic state [31,59] affect its stability. For example, cold temperatures, low UV-

B levels, and high pH slow eDNA degradation [59], while acidity promotes it [59,60]. The

overall probability of eDNA detection also depends on its production which may vary by

species, by season, by density, and diet, and its loss from the study system via water discharge ordiffusion [59].

Because techniques for eDNA analysis are still being optimized, some contradictory results have been reported. For example, a laboratory study showed that eDNA degradation increased with rising temperature, particularly in water samples from an oligotrophic lake [31]. By contrast, Robson et al [32] evaluated effects of high water temperature and fish density on the detection of invasive Mozambique tilapia in ponds via eDNA protocols and found that increased water

temperatures did not affect degradation rates. However, they did detect increased rates of eDNA
shedding at 35°C.

The present study represents the first time that the effectiveness of eDNA has been tested in an 301 302 oligotrophic tropical lake, where high temperatures ($>27^{\circ}C$ year round) should speed DNA 303 degradation. However, its water is slightly alkaline (pH 7.8) [39] which may aid eDNA preservation, perhaps explaining the better recovery of eDNA from water samples than 304 sediments (Fig 2, 5A). The recovery of species from sediments were low (0-8 species per site) 305 despite the higher DNA concentration in most samples treated with buffer (Fig 5B). The low 306 307 recovery may reflect the presence of inhibitors or nuclease activity although best practices were 308 followed to minimize DNA loss [61,62]. In particular, samples were transferred onto ice 309 immediately following collection, and DNA was promptly extracted from the samples. 310 Most prior eDNA studies have examined vertebrates in temperate freshwater streams [63], lakes 311 [64] or sea water [65,66]. However, Lopes et al. [30] recently demonstrated that eDNA surveys 312 were effective in revealing anurans in streams in the Brazilian Atlantic forest, while Robson et al. [32] developed eDNA protocols that employ qPCR to detect invasive fish in tropical 313 314 environments. Our results further validate the effectiveness of eDNA surveys in these settings. There are three major approaches for of eDNA detection: the use of qPCR to detect one or a 315 316 small number of target species [67,68], shotgun sequencing of aquatic metagenomes [69] and 317 PCR-based metabarcoding [28,29,63–66]. Metabarcoding studies cannot deliver information on species composition without a reliable reference database, such as that available for Mexican 318 freshwater fishes which includes records for 93% of the 70 species known from Lake Bacalar 319 320 and associated wetlands [48,51–54,70]. The present study has affirmed the effectiveness of

321 eDNA analysis as a tool for rapidly assessing the species composition of fish communities. The 322 extraction of eDNA from just 42 liters of water collected on three dates at 14 sampling points 323 revealed 41 of 57 of previously recorded species and added 6 potentially new species to the 324 fauna of the Lake Bacalar region. The present results mirror those obtained in a study which used eDNA to examine the composition of fish communities a coastal sea [28]. The analysis of 94 325 326 water samples obtained by sampling 47 sites in six hours revealed 40 of the 80 species known 327 from 14 years of underwater surveys, as well as 23 new species. The present study recovered sequences records from six fish species new to the Lake Bacalar 328 329 region (Bardiella ronchus, Ctenogobius fasciatus, Gobiosoma sp., Dorosoma anale, Oligoplites 330 saurus and one Engraulidae). Two of the new species (B. ronchus, O. saurus) detected in Juan

331 Sarabia and Alvaro Obregón Viejo are known from the adjacent Chetumal Bay, but have not

previously been reported from inland waters (Valdez-Moreno, pers. obs.). The new gobiid, *C*.

333 fasciatus, was detected in four localities (Xul-Ha, Pedro A. Santos, Cocalitos, Huay Pix), but was

overlooked in prior field surveys because of its morphological similarity to other gobids in Lake

Bacalar, especially *Lophogobius cyprinoides*. The other two new records require confirmation –

the presence of a *Gobiosoma* in Huay Pix (51 reads) and a member of the family Engraulidae in

337 Pedro A. Santos (>7000 reads).

A few sequences (13) of *Lachnolaimus maximus*, a commercially fished marine species, were recovered from Huay Pix, but the presence of its DNA was almost certainly mediated by human activity as fish were prepared for consumption.

As reported previously, species in two characid genera (*Astyanax* and *Bramocharax*) are difficult to discriminate using DNA barcodes as they show low divergences between species and genera

343	[51,71]. Moreover, in a study utilizing three mitochondrial genes (Cyt b, 16S, COI) and a single
344	nuclear gene (RAG1), Bramocharax was found to be polyphyletic, with species in this genus
345	being sisters to different clades of Astyanax, making Astyanax paraphyletic [72]. In fact, all
346	Bramocharax species grouped with sympatric Astyanax lineages (or even with allopatric
347	Astyanax populations), with less than 1% divergence. In our dataset, Astyanax aeneus, Astyanax
348	mexicanus, and Bramocharax caballeroi also formed an intermixed cluster in the Taxon ID tree
349	generated (see the dataset DS-EBACALAR in www.boldsystems.org). Among the species
350	detected with eDNA, Astyanax mexicanus was the least certain as it was only represented by 7
351	reads from the S5 with an average identity of 0.97 (S2 Table), which may be low quality reads.
352	Although Dorosoma anale was only represented by 13 reads from water at La Unión station, this
353	species is known to occur at sites in northern Belize close to the Hondo River (Valdez-Moreno
354	pers. obs.).
355	The presence of two other species (Cyprinodon beltrani (two PGM reads) and Vieja fenestrata
356	(five S5 reads) was less certain. C. beltrani is native to Chichancanab lagoon, and cannot be
357	
	distinguished from C. simus with DNA barcodes (see DS-EBACALAR in
358	
	distinguished from C. simus with DNA barcodes (see DS-EBACALAR in
358	distinguished from <i>C. simus</i> with DNA barcodes (see DS-EBACALAR in <u>www.boldsystems.org</u>), while <i>V. fenestrata</i> is native to the Papaloapan River [52] so its presence
358 359	distinguished from <i>C. simus</i> with DNA barcodes (see DS-EBACALAR in www.boldsystems.org), while <i>V. fenestrata</i> is native to the Papaloapan River [52] so its presence in the study area is unlikely.
358 359 360	 distinguished from <i>C. simus</i> with DNA barcodes (see DS-EBACALAR in www.boldsystems.org), while <i>V. fenestrata</i> is native to the Papaloapan River [52] so its presence in the study area is unlikely. Our results also revealed good eDNA recovery for other vertebrates, including rare species, such
358 359 360 361	 distinguished from <i>C. simus</i> with DNA barcodes (see DS-EBACALAR in www.boldsystems.org), while <i>V. fenestrata</i> is native to the Papaloapan River [52] so its presence in the study area is unlikely. Our results also revealed good eDNA recovery for other vertebrates, including rare species, such as <i>Tamandua mexicana</i>, which was represented in both PGM and S5 sequences from a river

364 similar to other species.

365 Biodiversity in freshwater ecosystems in undergoing losses as a result of landscape 366 transformation, pollution, and biological invasions. In fact intentional or accidental introduction 367 of invasive alien species is the third leading cause of global biodiversity loss [73]. In 2008 it was 368 estimated that nearly 40% of the freshwater fish species in North America, including Mexico, 369 were threatened by invasive species [74]. The Lake Bacalar water basin has been recognized as one of the hydrological basins with high priority for conservation by the Mexican National 370 371 Commission for the Use and Knowledge of the Biodiversity (CONABIO) [75]. Fortunately, most 372 wetlands and lakes in this region are relatively pristine, excepting the Hondo Riveron the border 373 between Mexico and Belize, which has been heavily impacted by the discharge of organic waste and pesticides, by vegetation clearing, and by the introduction of invasive such as tilapia [76], 374 and the Amazon sailfin catfish (*Pterygoplichthys pardalis*), which was recently detected [77,78] 375 376 near La Unión [79]. The impact of this declining water quality and rising incidence of invasive species on the native fish fauna needs to be carefully monitored and eDNA-based studies could 377 378 provide a cost-effective way to meet this need

379 **Conclusions**

We developed field sampling protocols and a HTS pipeline which enabled the efficient recovery of eDNA from several tropical aquatic ecosystems. Water samples consistently revealed more vertebrate species than sediment samples although about 10% of the species were only recovered from sediments.

eDNA sequences were recovered from 75 species of vertebrates including 47 species of fishes,
with six new records for the Lake Bacalar region and two other species whose detection is likely
to be due to human activity. Sequences were also detected from another 28 vertebrate species

including 15 birds, 7 mammals, 5 reptiles, and 1 amphibian, all species known from thewatershed.

This study indicates that eDNA can aid conservation and monitoring programs in tropical areas by improving our capacity to map occurrence records for resident and invasive species. There remains a need to convince both regulatory agencies and the public that this approach can provide the detailed information on species composition needed to underpin conservation policy for tropical aquatic ecosystems.

394 Acknowledgments

395 We thank Jose Angel Cohuo, Miguel Valadez, and Rosaura Castro from the Instituto

396 Tecnológico de Chetumal and Fernando Cortés Carrasco from El Colegio de la Frontera Sur for

help with collecting and processing samples. DNA extraction and NGS were carried out at the

398 Centre for Biodiversity Genomics (CBG), University of Guelph through funding to PDNH from

Ann McCain Evans and Chris Evans. We thank Thomas Braukmann for designing the

400 bioinformatics pipeline for processing the raw data. This paper represents a contribution from the

401 Chetumal node of the Mexican Barcode of Life (MEXBOL) network. MVM and MEG thank

402 CONACYT for support through the Sabbatical Stays Program (Grants 261790 and 262267).

403 Author contributions

404 Conceived and designed the experiments: MVM, MEG.

405 Coordinated the collection of material and performed field experiments: MVM, MEG.

406 Performed the lab experiments: NVI, SLP.

- 407 Analyzed the data: NVI, MVM, MEG, SLP.
- 408 Designed illustrations: NVI, MVM.
- 409 Designed *OTUBlastParser.py* and *ConcatenatorResults.py* Python scripts: KB.
- 410 Wrote the paper: MVM, NVI, MEG, SLP, PDNH.
- 411 All authors have read and approved the final manuscript.

412 **Data accessibility for BOLD datasets:**

- 413 dx.doi.org/10.5883/DS-EBACALAR
- 414 dx.doi.org/10.5883/DS-EBACAMPH
- 415 dx.doi.org/10.5883/DS-EBACAVES
- 416 dx.doi.org/10.5883/DS-EBACREPT
- 417 dx.doi.org/10.5883/DS-EBACMAMM

418 **References**

419	1.	Deiner K, Bik HM, Mächler E, Seymour M, Lacoursière-Roussel A, Altermatt F, et al.
-----	----	--

- 420 Environmental DNA metabarcoding: Transforming how we survey animal and plant
- 421 communities. Mol Ecol. 2017;26: 5872–5895. doi:10.1111/mec.14350
- 422 2. Barnes MA, Turner CR. The ecology of environmental DNA and implications for
- 423 conservation genetics. Conserv Genet. 2015;17: 1–17. doi:10.1007/s10592-015-0775-4
- 424 3. Thomsen PF, Willerslev E. Environmental DNA an emerging tool in conservation for

425 monite	oring past and prese	nt biodiversity. E	Biol Conserv. 2014;
------------	----------------------	--------------------	---------------------

- 426 doi:10.1016/j.biocon.2014.11.019
- 427 4. Mächler E, Deiner K, Spahn F, Altermatt F. Fishing in the water: effect of sampled water
- 428 volume on environmental DNA-based detection of macroinvertebrates. Environ Sci
- 429 Technol. 2016;50: 305–312. doi:10.1021/acs.est.5b04188
- 430 5. Taberlet P, Coissac E, Hajibabaei M, Rieseberg LH. Environmental DNA. Mol Ecol.
 431 2012;21: 1789–1793. doi:10.1111/j.1365-294X.2012.05542.x
- 432 6. Dejean T, Valentini A, Duparc A, Pellier-Cuit S, Pompanon F, Taberlet P, et al.
- 433 Persistence of environmental DNA in freshwater ecosystems. PLoS One. 2011;6: 8–11.
- 434 doi:10.1371/journal.pone.0023398
- 435 7. Schmelzle MC, Kinziger AP. Using occupancy modelling to compare environmental
- 436 DNA to traditional field methods for regional-scale monitoring of an endangered aquatic
- 437 species. Mol Ecol Resour. 2016;16: 895–908. doi:10.1111/1755-0998.12501
- 438 8. Goldberg CS, Strickler KM, Pilliod DS. Moving environmental DNA methods from
- 439 concept to practice for monitoring aquatic macroorganisms. Biol Conserv. 2014;
- 440 doi:10.1016/j.biocon.2014.11.040
- Hebert PDN, Cywinska A, Ball SL, deWaard JR. Biological identifications through DNA
 barcodes. Proc Biol Sci. 2003;270: 313–21. doi:10.1098/rspb.2002.2218
- 10. Ratnasingham S, Hebert PDN. BOLD: The Barcode of Life Data System
- 444 (www.barcodinglife.org). Mol Ecol Notes. 2007;7: 355–364.
- 445 11. Hajibabaei M. The golden age of DNA metasystematics. Trends Genet. 2012;28: 535–

446 537. doi:10.1016/j.tig.2012.08.001

447	12.	Lodge DM, Turner CR, Jerde CL, Barnes MA, Chadderton L, Egan SP, et al.
448		Conservation in a cup of water: estimating biodiversity and population abundance from
449		environmental DNA. Mol Ecol. 2012;21: 2555-2558. doi:10.1111/j.1365-
450		294X.2012.05600.x
451	13.	Goldberg CS, Turner CR, Deiner K, Klymus KE, Thomsen PF, Murphy MA, et al.
452		Critical considerations for the application of environmental DNA methods to detect
453		aquatic species. Methods Ecol Evol. 2016;7: 1299–1307. doi:10.1111/2041-210X.12595
454	14.	Sigsgaard EE, Carl H, Møller PR, Thomsen PF. Monitoring the near-extinct European
455		weather loach in Denmark based on environmental DNA from water samples. Biol
456		Conserv. 2014;183: 46–52. doi:10.1016/j.biocon.2014.11.023
457	15.	de Carvalho MR, Bockmann FA, Amorim DS, Brandão CRF, de Vivo M, de Figueiredo
458		JL, et al. Taxonomic impediment or impediment to taxonomy? A commentary on
459		systematics and the cybertaxonomic-automation paradigm. Evol Biol. 2007;34: 140–143.
460		doi:10.1007/s11692-007-9011-6
461	16.	Rees HC, Maddison BC, Middleditch DJ, Patmore JRM, Gough KC. The detection of
462		aquatic animal species using environmental DNA – a review of eDNA as a survey tool in
463		ecology. Journal of Applied Ecology. 2014. doi:10.1111/1365-2664.12306
464	17.	Comtet T, Sandionigi A, Viard F, Casiraghi M. DNA (meta)barcoding of biological
465		invasions: a powerful tool to elucidate invasion processes and help managing aliens. Biol
466		Invasions. 2015;17. doi:10.1007/s10530-015-0854-y

467	18.	Bergman PS, Schumer G, Blankenship S, Campbell E. Detection of adult green sturgeon
468		using environmental DNA analysis. PLoS One. 2016;11: 1-8.
469		doi:10.1371/journal.pone.0153500
470	19.	Eva B, Harmony P, Thomas G, Francois G, Alice V, Claude M, et al. Trails of river
471		monsters: Detecting critically endangered Mekong giant catfish Pangasianodon gigas
472		using environmental DNA. Glob Ecol Conserv. 2016;7: 148-156.
473		doi:10.1016/j.gecco.2016.06.007
474	20.	Doi H, Inui R, Akamatsu Y, Kanno K, Yamanaka H, Takahara T, et al. Environmental
475		DNA analysis for estimating the abundance and biomass of stream fish. Freshw Biol.
476		2016; 30–39. doi:10.1111/fwb.12846
477	21.	Takahara T, Minamoto T, Yamanaka H, Doi H, Kawabata Z. Estimation of fish biomass
478		using environmental DNA. PLoS One. 2012;7: 3-10. doi:10.1371/journal.pone.0035868
479	22.	Deiner K, Walser J-C, Mächler E, Altermatt F. Choice of capture and extraction methods
480		affect detection of freshwater biodiversity from environmental DNA. Biol Conserv. 2014;
481		doi:10.1016/j.biocon.2014.11.018
482	23.	Turner CR, Barnes MA, Xu CCY, Jones SE, Jerde CL, Lodge DM. Particle size
483		distribution and optimal capture of aqueous macrobial eDNA. Methods Ecol Evol. 2014;5:
484		676–684. doi:10.1111/2041-210X.12206
485	24.	Ficetola GF, Miaud C, Pompanon F, Taberlet P. Species detection using environmental
486		DNA from water samples. Biol Lett. 2008;4: 423-5. doi:10.1098/rsbl.2008.0118
487	25.	Wegleitner BJ, Jerde CL, Tucker A, Chadderton WL, Mahon AR. Long duration, room

488		temperature preservation of filtered eDNA samples. Conserv Genet Resour. 2015;7: 789-
489		791. doi:10.1007/s12686-015-0483-x
490	26.	Yamanaka H, Minamoto T, Matsuura J, Sakurai S, Tsuji S, Motozawa H, et al. A simple
491		method for preserving environmental DNA in water samples at ambient temperature by
492		addition of cationic surfactant. Limnology. 2017;18: 233-241. doi:10.1007/s10201-016-
493		0508-5
494	27.	Deiner K, Renshaw MA, Li Y, Olds BP, Lodge DM, Pfrender ME. Long-range PCR
495		allows sequencing of mitochondrial genomes from environmental DNA. Methods Ecol
496		Evol. 2017; doi:10.1111/2041-210X.12836
497	28.	Yamamoto S, Masuda R, Sato Y, Sado T, Araki H, Kondoh M, et al. Environmental DNA
498		metabarcoding reveals local fish communities in a species-rich coastal sea. Sci Rep.
499		2017;7: 40368. doi:10.1038/srep40368
500	29.	Olds BP, Jerde CL, Renshaw MA, Li Y, Evans NT, Turner CR, et al. Estimating species
501		richness using environmental DNA. Ecol Evol. 2016;6: 4214-4226.
502		doi:10.1002/ece3.2186
	20	
503	30.	Lopes CM, Sasso T, Valentini A, Dejean T, Martins M, Zamudio KR, et al. eDNA
504		metabarcoding: a promising method for anuran surveys in highly diverse tropical forests.
505		Mol Ecol Resour. 2017;17: 904–914. doi:10.1111/1755-0998.12643

- 506 31. Eichmiller J, Best SE, Sorensen PW. Effects of temperature and trophic state on
- 507 degradation of environmental DNA in lake water. Environ Sci Technol. 2016;
- 508 acs.est.5b05672. doi:10.1021/acs.est.5b05672

509	32.	Robson HLA, Noble TH, Saunders RJ, Robson SKA, Burrows DW, Jerry DR. Fine-tuning
510		for the tropics: application of eDNA technology for invasive fish detection in tropical
511		freshwater ecosystems. Mol Ecol Resour. 2016;16: 922-932. doi:10.1111/1755-
512		0998.12505
513	33.	Gischler E, Gibson MA, Oschmann W. Giant Holocene freshwater microbialites, Laguna
514		Bacalar, Quintana Roo, Mexico. Sedimentology. 2008;55: 1293-1309.
515		doi:10.1111/j.1365-3091.2007.00946.x
516	34.	Castro-Contreras SI, Gingras MK, Pecoits E, Abert NR, Petrash D, Castro-Contreras SM,
517		et al. Textural and geochemical features of freshwater. Palaios. 2013;29: 192-209.
518		doi:10.2110/palo.2013.063
519	35.	Centeno CM, Legendre P, Beltrán Y, Alcántara-Hernández RJ, Lidström UE, Ashby MN,
520		et al. Microbialite genetic diversity and composition relate to environmental variables.
521		FEMS Microbiol Ecol. 2012;82: 724–735. doi:10.1111/j.1574-6941.2012.01447.x
522	36.	Bauer-Gottwein P, Gondwe BRN, Charvet G, Marín LE, Rebolledo-Vieyra M, Merediz-
523		Alonso G. Review: The Yucatán Peninsula karst aquifer, Mexico. Hydrogeol J. 2011;19:
524		507–524. doi:10.1007/s10040-010-0699-5
525	37.	Moore YH, Stoessel RK, Easley D. Fresh-water/sea-water relationship within a
526		groundwater system, northeastern coast of the Yucatán Peninsula. Gr Wat. 1992;30: 343-
527		350.
528	38.	Perry E, Velazquez-Oliman G, Marin LE. The hydrogeochemistry of the karst aquifer
529		system of the northern Yucatan Peninsula, Mexico. Int Geol Rev. 2002;44: 191-221.

530 doi:10.2747/0020-6814.44.3.191

531	39.	Pérez L, Bugja R, Lorenschat J, Brenner M, Curtis J, Hoelzmann P, et al. Aquatic
532		ecosystems of the Yucatan Peninsula (Mexico), Belize, and Guatemala. Hydrobiologia.
533		2011;661: 407-433. doi:10.1007/s10750-010-0552-9
534	40.	Ivanova NV, Fazekas AJ, Hebert PDN. Semi-automated, membrane-based protocol for
535		DNA isolation from plants. Plant Mol Biol Report. 2008;26: 186–198.
536	41.	Rohland N, Reich D, Rohland and Reich, Rohland N, Reich D. Cost-effective, high-
537		throughput DNA sequencing libraries for multiplexed target capture. Genome Res.
538		2012;22: 939–946. doi:10.1101/gr.128124.111
539	42.	Ivanova N V., Dewaard JR, Hebert PDN. An inexpensive, automation-friendly protocol
540		for recovering high-quality DNA. Mol Ecol Notes. 2006;6: 998–1002. doi:10.1111/j.1471-
541		8286.2006.01428.x
542	43.	Berry D, Ben Mahfoudh K, Wagner M, Loy A. Barcoded primers used in multiplex
543		amplicon pyrosequencing bias amplification. Appl Environ Microbiol. 2011;77: 7846-9.
544		doi:10.1128/AEM.05220-11
545	44.	Ivanova N V, Clare EL, Borisenko A V. DNA barcoding in mammals. In: Kress J,
546		Erickson D, editors. Analytical protocols In: DNA barcodes, methods in molecular
547		biology. Humana Press; 2012. pp. 153-182. doi:10.1007/978-1-61779-591-6
548	45.	Ivanova NV, Zemlak TS, Hanner RH, Hebert PDN. Universal primer cocktails for fish
549		DNA barcoding. Mol Ecol Notes. 2007;7: 544–548.
550	46.	Colwell RK. EstimateS: Statistical estimation of species richness and shared species from

551		samples. Version 9. User's guide and application. [Internet]. 2013. Available:
552		http://purl.oclc.org/estimates
553	47.	Letunic I, Bork P. Interactive tree of life (iTOL) v3: an online tool for the display and
554		annotation of phylogenetic and other trees. Nucleic Acids Res. 2016;44: W242-5.
555		doi:10.1093/nar/gkw290
556	48.	López-Vila JM, Valdéz-Moreno ME, Schmitter-Soto JJ, Mendoza-Carranza M, Herrera-
557		Pavón RL. Composition and structure of the ichthyofauna in the Hondo River, Mexico-
558		Belize, using a harpoon. Rev Mex Biodivers. 2014;85: 866–874. doi:10.7550/rmb.35806
559	49.	Schmitter-Soto JJ. Catalogo de los peces continentales de quintana roo [Internet]. 1996.
560		Available: http://www.nativefishlab.net/library/textpdf/20569.pdf
561	50.	Schmitter-Soto JJ. Distribution of continental fishes in northern Quintana Roo, Mexico.
562		Southwest Nat. 1999;44: 166–172.
563	51.	Valdez-Moreno M, Ivanova N V., Elías-Gutiérrez M, Contreras-Balderas S, Hebert PDN.
564		Probing diversity in freshwater fishes from Mexico and Guatemala with DNA barcodes. J
565		Fish Biol. 2009;74: 377–402. doi:10.1111/j.1095-8649.2008.02077.x
566	52.	Miller RR. Peces dulceacuícolas de México. México, D.F.: Comisión Nacional para el
567		Conocimiento y Uso de la Biodiversidad, Sociedad Ictiológica Mexicana A. C., El colegio
568		de la Frontera Sur y Consejo de los Peces del Desierto México-Estados Unidos; 2009.
569	53.	Neil SJ. Field studies of the behavioral ecology and agonistic behavior of Cichlasoma
570		meeki (Pisces: Cichlidae). Environ Biol Fishes. 1984;10: 59–68. doi:10.1007/BF00001662
571	54.	Schmitter-Soto JJ, Gamboa-Pérez HC. Composición y distribución de peces continentales

572		en el sur de Quintana Roo, Península de Yucatán, México. Rev Biol Trop. 1996;44: 199-
573		212.
574	55.	Gamboa-Pérez H and Schmitter-Soto J.J. Distribution of cichid fishes in the littoral of
575		Lake Bacalar, Yucatan Peninsula. Environ Biol Fishes. 1999;54: 35–43.
576	56.	Miller RR. Composition and derivation of the freshwater fish fauna of Mexico. An la Esc
577		Nac Ciencias Biológicas. 1986;30: 121–153.
578	57.	Palmer JM, Jusino MA, Banik MT, Lindner DL. Non-biological synthetic spike-in
579		controls and the AMPtk software pipeline improve mycobiome data. PeerJ. 2018;6:
580		e4925. doi:10.7717/peerj.4925
581	58.	Esling P, Lejzerowicz F, Pawlowski J. Accurate multiplexing and filtering for high-
582		throughput amplicon-sequencing. Nucleic Acids Res. 2015;43: 2513-24.
583		doi:10.1093/nar/gkv107
584	59.	Strickler KM, Fremier AK, Goldberg CS. Quantifying effects of UV-B, temperature, and
585		pH on eDNA degradation in aquatic microcosms. Biol Conserv. 2014;
586		doi:10.1016/j.biocon.2014.11.038
587	60.	Seymour M, Durance I, Cosby BJ, Ransom-Jones E, Deiner K, Ormerod SJ, et al. Acidity
588		promotes degradation of multi-species environmental DNA in lotic mesocosms. Commun
589		Biol. 2018;1: 4. doi:10.1038/s42003-017-0005-3
590	61.	Prosser S, Martínez-Arce A, Elías-Gutiérrez M. A new set of primers for COI
591		amplification from freshwater microcrustaceans. Mol Ecol Resour. 2013;13: 1151–1155.
592		doi:10.1111/1755-0998.12132

593	62.	Elías-Gutiérrez M, Valdez-Moreno M, Topan J, Young MR, Cohuo-Colli JA. Improved
594		protocols to accelerate the assembly of DNA barcode reference libraries for freshwater
595		zooplankton. Ecol Evol. 2018;8: 3002-3018. doi:10.1002/ece3.3742
596	63.	Vences M, Lyra ML, Perl RGB, Bletz MC, Stanković D, Lopes CM, et al. Freshwater
597		vertebrate metabarcoding on Illumina platforms using double-indexed primers of the
598		mitochondrial 16S rRNA gene. Conserv Genet Resour. 2016;8: 323-327.
599		doi:10.1007/s12686-016-0550-y
600	64.	Sato H, Sogo Y, Doi H, Yamanaka H. Usefulness and limitations of sample pooling for
601		environmental DNA metabarcoding of freshwater fish communities. Sci Rep. 2017;7:
602		14860. doi:10.1038/s41598-017-14978-6
603	65.	Port JA, O'Donnell JL, Romero-Maraccini OC, Leary PR, Litvin SY, Nickols KJ, et al.
604		Assessing vertebrate biodiversity in a kelp forest ecosystem using environmental DNA.
605		Mol Ecol. 2016;25: 527–541. doi:10.1111/mec.13481
606	66.	Andruszkiewicz EA, Starks HA, Chavez FP, Sassoubre LM, Block BA, Boehm AB.
607		Biomonitoring of marine vertebrates in Monterey Bay using eDNA metabarcoding. PLoS
608		One. 2017;12: e0176343. doi:10.1371/journal.pone.0176343
609	67.	Dougherty MM, Larson ER, Renshaw MA, Gantz CA, Egan SP, Erickson DM, et al.
610		Environmental DNA (eDNA) detects the invasive rusty crayfish Orconectes rusticus at
611		low abundances. Journal of Applied Ecology. 2016. doi:10.1111/1365-2664.12621
612	68.	Klymus KE, Richter CA, Chapman DC, Paukert C. Quantification of eDNA shedding
613		rates from invasive bighead carp Hypophthalmichthys nobilis and silver carp

614 *Hypophthalmichthys molitrix*. Biol Conserv. 2015;183: 77–84.

- 615 doi:10.1016/j.biocon.2014.11.020
- 616 69. Cowart DA, Murphy KR, Cheng C-HC. Metagenomic sequencing of environmental DNA
- 617 reveals marine faunal assemblages from the West Antarctic Peninsula. Mar Genomics.
- 618 2018;37: 148–160. doi:10.1016/J.MARGEN.2017.11.003
- 619 70. Espinoza Pérez H, Gaspar Dillanes MT, Fuentes Mata P. Listado faunísticos de México.
- 620 III. Los peces dulceaciucolas Mexicanos. México, D. F.: Universidad Nacional Autonoma
 621 de Mexico; 1993.
- 622 71. Rossini BC, Oliveira CAM, Melo FAG de, Bertaco V de A, Astarloa JMD de, Rosso JJ, et
- al. Highlighting *Astyanax* species diversity through DNA barcoding. PLoS One. 2016;11:

624 e0167203. doi:10.1371/journal.pone.0167203

- 625 72. Ornelas-García C, Domínguez-Domínguez O, Doadrio I. Evolutionary history of the fish
- 626 genus *Astyanax* Baird & Girard (1854) (Actinopterygii, Characidae) in Mesoamerica
- reveals multiple morphological homoplasies. BMC Evol Biol. 2008;8: 340.
- 628 doi:10.1186/1471-2148-8-340
- 629 73. Sarukhán J, Whyte A, editors. Millenium ecosystem assessment. Ecosystems and human
- 630 well-being: biodiversity synthesis [Internet]. Ecosystems. Washington, DC: Island Press;
- 631 2005. Available:
- https://www.millenniumassessment.org/documents/document.356.aspx.pdf
- 633 74. Jelks HL, Walsh SJ, Burkhead NM, Contreras-Balderas S, Diaz-Pardo E, Hendrickson
- DA, et al. Conservation status of imperiled North American freshwater and diadromous

	635	fishes. Fisheries. 2008;33: 372–407. doi:10.1577/1548-8446-33.8.372
--	-----	---

636	75.	Arriaga Cabrera	L, Aguilar Sierra	V. Alcocer Durán J	J, Jiménez Rosenberg R,	Muñoz

- 637 López E, Vázquez Domínguez E, editors. Regiones hidrológicas prioritarias: fichas
- técnicas y mapa (escala 1:4,000,000) [Internet]. Mexico: Comisión Nacional para el
- 639 Conocimiento y Uso de la Biodiversidad; 1998. Available:
- 640 https://books.google.ca/books?id=b116YgEACAAJ
- 641 76. Schmitter-Soto JJ, Ruiz-Cauich LE, Herrera RL, González-Solís D. An index of biotic
- 642 integrity for shallow streams of the Hondo River basin, Yucatan Peninsula. Sci Total

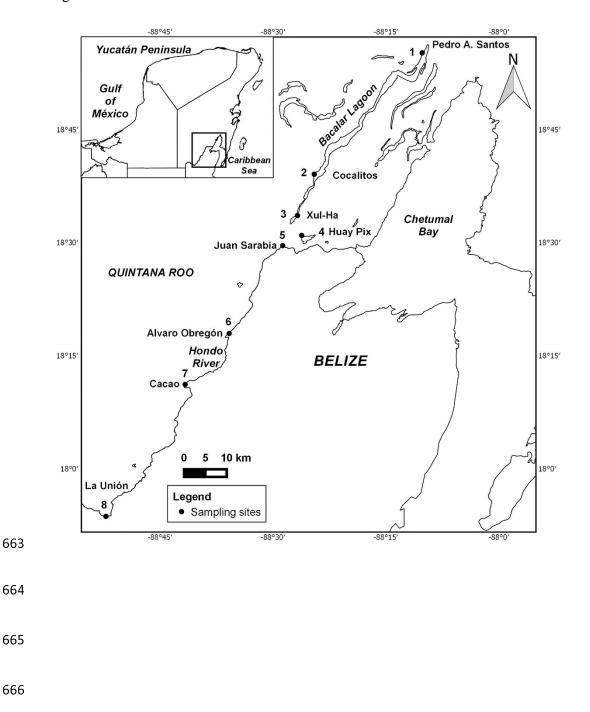
643 Environ. 2011;409: 844–852. doi:10.1016/j.scitotenv.2010.11.017

- Wakida-Kusunoki AT, Luis Enrique Amador-del Ángel. Aspectos biológicos del pleco
 invasor *Pterygoplichthys pardalis* (Teleostei: Loricariidae) en el río Palizada, Campeche,
 México. Rev Mex Biodivers. 2011;82: 870–878.
- 647 78. Alfaro REM, Fisher JP, Courtenay W, Ramírez Martínez C, Orbe-Mendoza A, Escalera
- 648 Gallardo C, et al. Armored catfish (Loricariidae) trinational risk assessment guidlines for
- 649 aquatic alien invasive species. Test cases for the snakeheads (Channidae) and armored
- 650 catfishes (Loricariidae) in North American inland waters. Montreal, Canadá: Commission
- 651 for Environmental Cooperation; 2009. pp. 25–49.
- 652 79. Schmitter-Soto JJ, Quintana R, Valdéz-Moreno ME, Herrera-Pavón RL, Esselman PC.
- Armoured catfish (*Pterygoplichthys pardalis*) in the Hondo River basin, Mexico-Belize.
- 654 Mesoamericana. 2015;19: 9–19.

655 Supporting Information

- 656 S1 Table. Supplementary Table 1. Summary of coverage for vertebrate reads for all collection
- events and for positive controls.
- 658 S2 Table. Supplementary Table 2. Taxonomy, top hit Process IDs, average overlap, average
- score, coverage, and average identity for vertebrate species detected by eDNA in environmental
- samples, positive and negative controls.

662 Fig 1.



667

669 Fig 2.

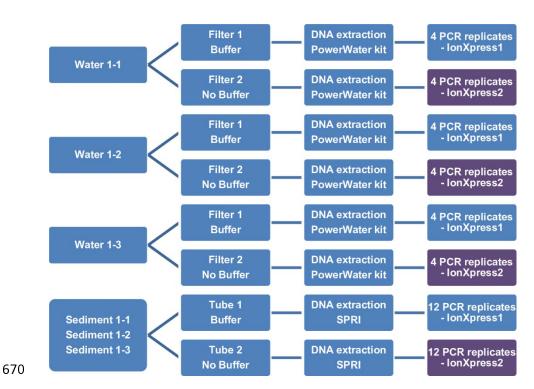


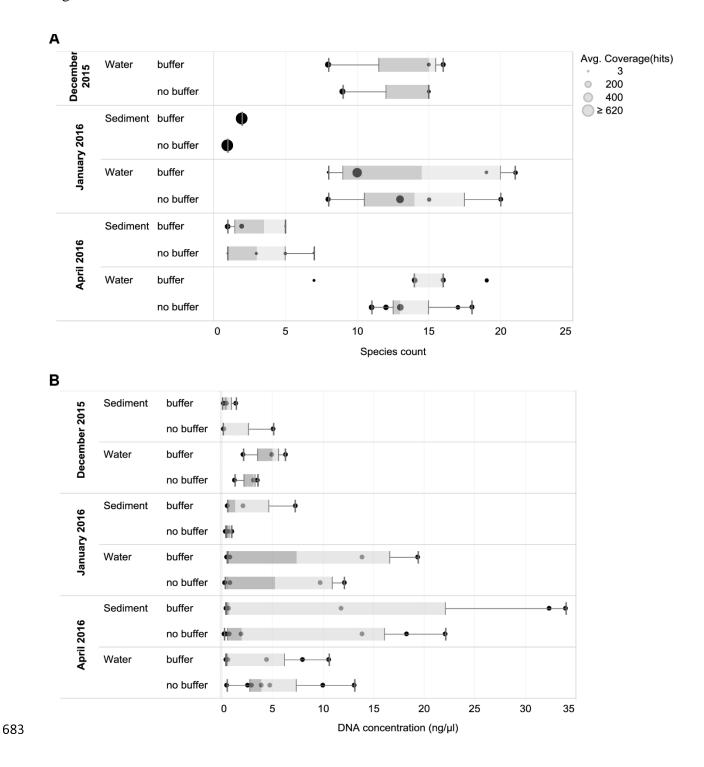
Fig 3.

Water and sediments Water Sediments Species(est) Sampling event

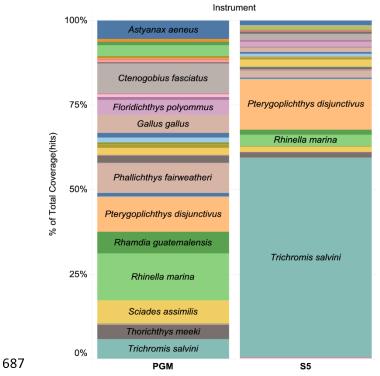
673 Fig 4.

Month of Sampling date	Site	Sample type								Amphibia
December 2015	Cocalitos	Water								Mammalia
	Huay Pix	Water								Actinopter
	Xul-Ha	Water								
January 2016	Alvaro Obregón Viejo									
		Water								
	Huay Pix	Water								
	La Unión	Sediment						_		
		Water								
	Pedro A. Santos									
April 2016	Cacao	Water								
	Cocalitos	Sediment							_	
		Water								
	Huay Pix	Sediment					_			
	huan Qanakia	Water								
	Juan Sarabia	Sediment					_			
	La Unión	Water Sediment								
	La Union	Water						_		
	Pedro A. Santos									
	Feuro A. Santos	Water								
	Xul-Ha	Sediment								
	Xuinu	Water								
			0	5	10	15 Species	20 count	25	30	

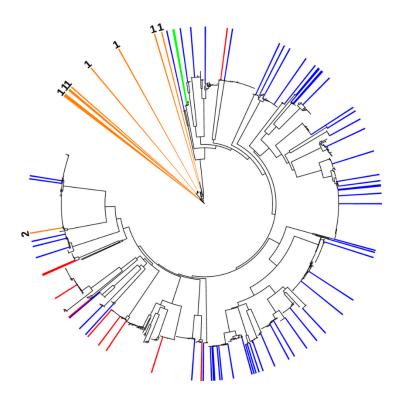


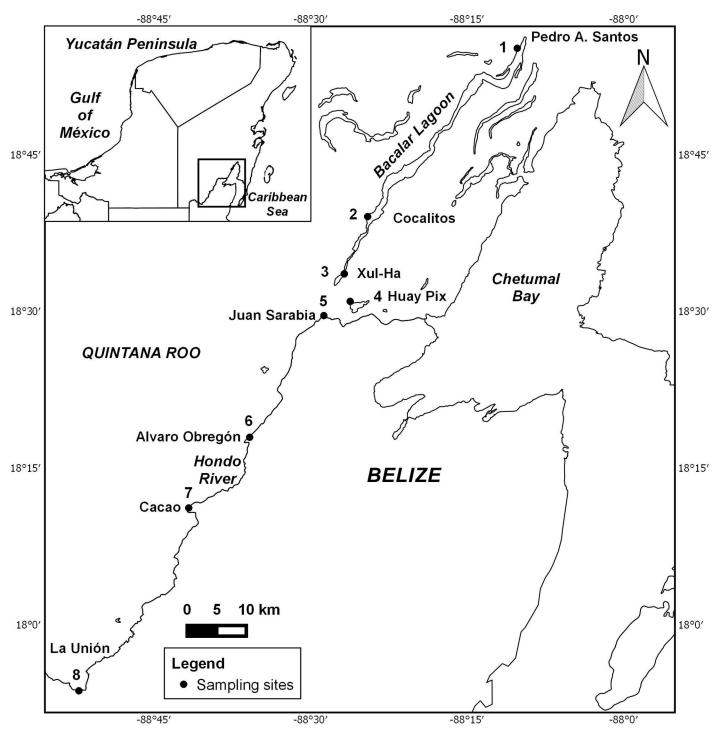


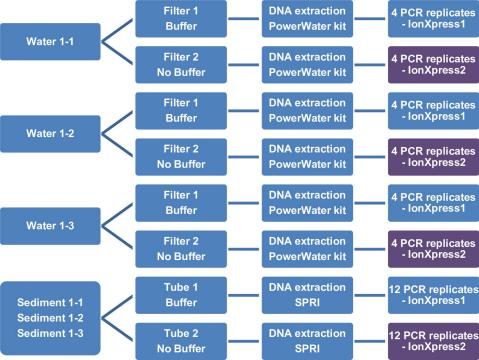
686 Fig 6.

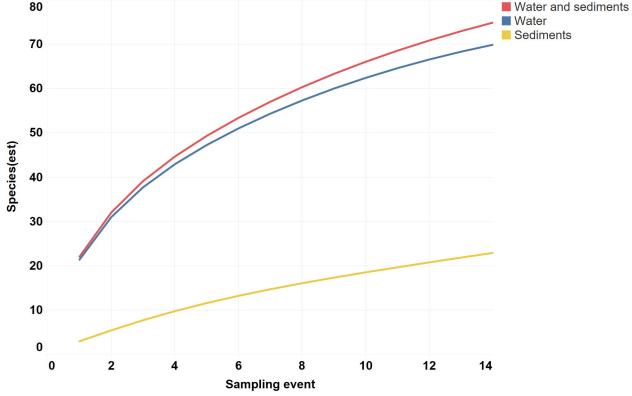


688 Fig 7.

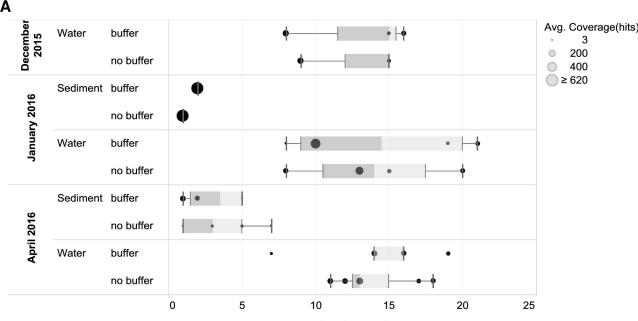






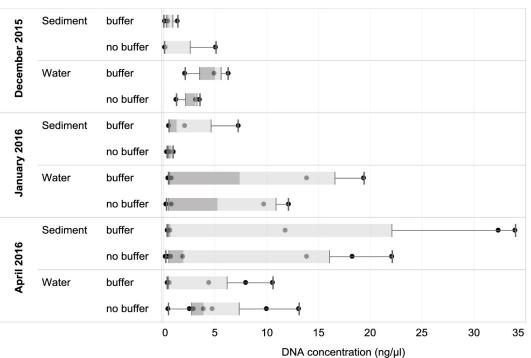


Month of Sampling date	Site	Sample type								Amphibia Aves
December 2015	Cocalitos	Water								Mammalia Reptilia
	Huay Pix	Water								Actinopterygi
	Xul-Ha	Water								
January 2016	Alvaro Obregón Viejo	Sediment								
		Water								
	Huay Pix	Water								
	La Unión	Sediment								
		Water								
	Pedro A. Santos	Water								
April 2016	Cacao	Water								
	Cocalitos	Sediment								
		Water								
	Huay Pix	Sediment								
		Water								
	Juan Sarabia	Sediment								
		Water								
	La Unión	Sediment								
		Water								
	Pedro A. Santos	Sediment								
		Water								
	Xul-Ha	Sediment								
		Water								
			0	5	10 s	15 Species co	20 ount	25	30	



Species count

В



Instrument

