

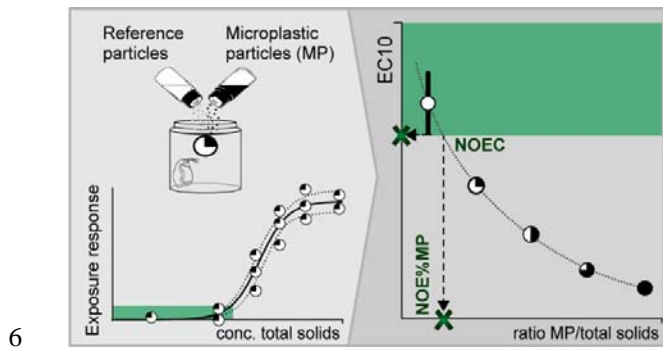
1 **A serial dilution method for assessment of microplastic toxicity in suspension**

2 Zandra Gerdes, Markus Hermann, Martin Ogonowski and Elena Gorokhova

3 Department of Environmental Science and Analytical Chemistry, Stockholm

4 University, Svante Arrhenius väg 8, SE-11418 Stockholm, Sweden.

5 **TOC**



7 **Abstract**

8 The occurrence of microplastic (MP) in the environment is of global concern. MP risk
9 assessment, however, is currently hampered by lacking ecotoxicological methods due
10 to conceptual and practical problems with particle exposure. Natural particles of
11 similar size as MP, e.g., clay and cellulose, occur abundantly in the environment. For
12 MP risk assessment and regulation it must be established whether the addition of MP
13 to these particles represents an additional hazard. We present a novel approach
14 employing a serial dilution of MP and reference particles, in mixtures, which allows
15 the differentiation of MP effects from other particulates. We demonstrate the
16 applicability of the method using an immobilisation test with *Daphnia magna*
17 exposed to polyethylene terephthalate (MP) and kaolin clay (reference material). In
18 the concentration range of 0.1 to 10000 mg L⁻¹ of total suspended solids (TSS), with
19 MP contributing 0-100 %, the LC₅₀ values for MP-kaolin mixtures were significantly
20 lower compared to the pure kaolin suspension. MP particles were thus more harmful
21 to daphnids than the reference material. The estimated threshold for %MP
22 contribution above which higher mortality was observed was 1 % MP at 36 mg TSS
23 L⁻¹. This approach has a potential for standardisation of MP ecotoxicological testing
24 as well as other particulate material of anthropogenic origin.

25 **Introduction**

26 The increasing environmental pollution with plastic waste is of global concern. What
27 is more, this debris eventually breaks down to small fragments collectively termed
28 microplastics (MP) that are omnipresent in aquatic environments, including alpine
29 lakes, rivers, oceans and arctic ice.¹⁻⁴ The amounts of the plastic debris in general, and
30 MP, in particular, are expected to increase because of increased production,
31 continuous discharge, and fragmentation.⁵ Research on the hazard assessment of solid
32 polymer particulates is in high demand due to public and scientific concerns.
33 Nevertheless, scientists disagree on the immediacy of the MP pollution problem,⁶⁻⁹
34 and it remains largely unclear whether MP are harmful to biota and what the impact
35 mechanisms are. The continuing uncertainty is, at least partly, related to the fact that
36 MP are a new type of environmental contaminant with yet unsettled methodology for
37 hazard testing.

38 The first experimental MP effect studies included a wide range of animal species
39 focusing mainly on feeding-related impacts in filter-feeders, such as bivalves¹⁰⁻¹² and
40 zooplankton^{13,14} Filter-feeders continue to be among the commonly used test
41 organisms in MP effect studies because they are susceptible to MP exposure via
42 ingestion. Since MP particles are nutritionally inert, their ingestion decreases the
43 energy intake. In other words, the ingestion of refractory material and alterations in
44 feeding (a primary response) leads to lower growth and reproduction (secondary
45 responses) as a result of the decreased caloric intake.¹⁵

46 All these processes occur not only with MP but also with any other refractory material
47 present in natural seston. Both mineral¹⁵⁻¹⁸ and MP^{13,14,19,20} particles have been
48 reported to alter feeding activity and reduce growth. Natural processes, such as wind

49 and resuspension, primarily affect the presence of nutritionally inert particles in the
50 water; whereas, human activities, like, dredging and stormwater runoff, may also
51 elevate their concentrations. High concentrations of total suspended solids (TSS) have
52 been found to reduce primary production,²¹ suppress population growth of
53 zooplankton²² and alter feeding behaviour in fish.²³ Therefore, to protect wildlife,
54 water quality standards are implemented for TSS concentrations or allowable TSS
55 levels in, e.g. stormwater effluents,²⁴ lakes and streams²⁵.

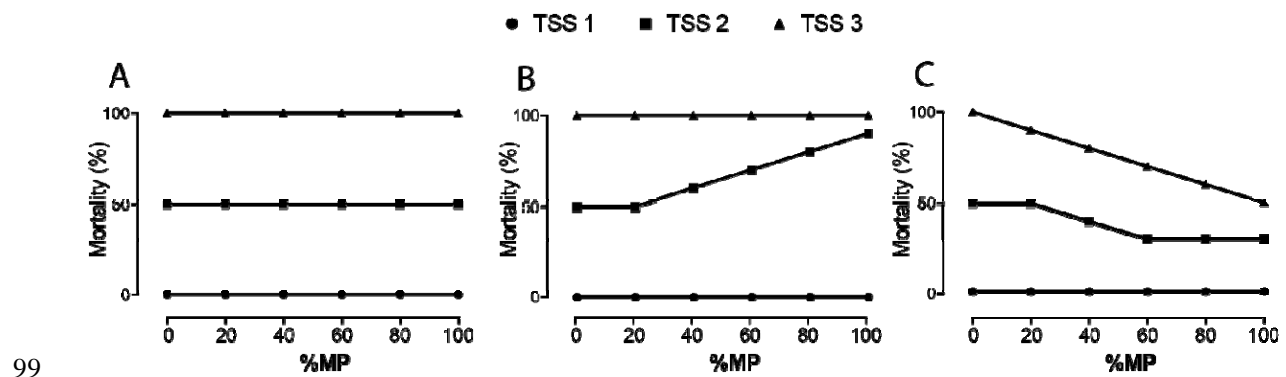
56 Regulatory efforts to set allowable MP levels are calling for adequate methodological
57 approaches for hazard assessment, relevant model species, and exposure scenarios. A
58 step towards quantifying hazardous properties of synthetic polymer microparticles is
59 to develop and apply standardised practices and experimental designs that will be able
60 to provide threshold values of these effects. However, given the presence of various
61 particulates and the hazardous effects of high TSS concentrations, such designs
62 should include the MP in question together with environmentally relevant reference
63 material(s). Particular attention should be paid to the similarity of basic physical
64 properties that are important for biological responses, e.g., size distribution and shape,
65 between the reference particles and the MP.^{13,26} Also, to maintain the experimental
66 reproducibility and stable encounter rates in a pelagic exposure scenario, it is
67 important that all particles be kept in suspension during the incubation.

68 A recent comparison of the effects exerted by MP and mineral particulates suggests
69 some similarity in responses across different levels of biological organisation, albeit
70 with an indication of a greater hazard by MP.²⁷ Since natural particles are more
71 abundant than MP in aquatic environments,⁷ the hazardous levels of MP should rather
72 be presented as a relative contribution of MP to TSS and not the absolute
73 concentrations.

74 To date, there is no standard approach for MP effect assessment,²⁸ despite a rapidly
75 rising number of reports on MP effects under laboratory conditions. This is partly
76 because it is challenging to design exposure experiments with environmentally
77 relevant concentrations of MP based on the commonly reported levels (<10 particles
78 m³)²⁹⁻³¹. Moreover, the existing approaches do not explicitly test the effects of MP *per*
79 *se* but those of nutrition-free particulates. To move the field of MP ecotoxicology
80 forward, we need to use test methods that (1) are appropriate for delineating effects of
81 different particulate materials in mixtures, (2) provide estimation of the critical
82 concentrations of MP in different environments, (3) allow high-throughput testing,
83 and (4) support read-across and categorical assessment of solid polymer particles.
84 Here, we propose a new approach employing a linear serial dilution of MP and
85 reference particles in mixtures, to identify MP-specific toxicity while controlling for
86 the total concentration of suspended matter in the experimental system. Further, we
87 demonstrate the applicability of this approach using the 96-h exposure of the
88 cladoceran *Daphnia magna* to a mixture of polyethylene terephthalate (PET) as a test
89 MP and kaolin as a reference particle.

90 METHOD

91 The MP Ratio Test was designed to examine whether a particulate material (test particle) is
92 harmful when co-occurring in a mixture with naturally present particulates (reference
93 particle) across a range of TSS concentrations. The rationale is as follows: if the test particle
94 is more harmful than the reference particle, then decreasing its contribution to a mixture with
95 the reference particles should decrease the overall toxicity, assuming additivity of the effects
96 (fig. 1). When the test (MP) and reference (mineral) particles are provided at varying
97 proportions for each TSS concentration, then, by using a range of TSS concentrations, a dose-
98 response relationship can be established for each mixture.



100 **Figure 1.** Possible outcomes of the MP Ratio Test, shown as a response (in this case
101 Mortality%) to the relative MP contribution (%MP) to the TSS in the test system, including
102 three test concentrations of TSS. The TSS concentrations denoted as 1, 2 and 3 represent
103 increasing levels of TSS in the system. Scenario A shows no MP effect because mortality is
104 only responding to the increasing TSS level. Scenario B shows an additive effect of MP
105 because Mortality% is positively affected %MP above a critical threshold. Scenario C shows
106 an ameliorating effect of %MP on TSS toxicity. Based on existing reports for TSS effects on
107 *Daphnia*, we expect no effect at low TSS concentrations.

108

109 **Test organism.** The freshwater cladoceran *Daphnia magna* was used as the test species.
110 These microcrustaceans are the most common model organisms in aquatic ecotoxicology and
111 have been used extensively in studies assessing both TSS^{13,22,32–34} and MP^{13,33,35} effects. All
112 experimental animals originated from the same clone (Clone 5; The Federal Environment
113 Agency, Berlin, Germany) cultured in M7 media at a density ~10 ind. L⁻¹ and fed *ad libitum*
114 with a mixture of *Pseudokirchneriella subcapitata* and *Scenedesmus spicatus*.

115 **The reference and test particles.** Kaolin (Sigma-Aldrich) was used as the reference particle;
116 it occurs globally in suspended particulates and has previously been used in tests with
117 daphnids, both as a reference particle when assessing MP effects¹³ and as a test particle when
118 assessing effects of TSS³². As the test MP, we used polyethylene terephthalate (PET
119 Goodfellow), to represent a plastic that is commonly found in the environment. The PET was
120 obtained as 3-5 mm-sized pellets from the manufacturer and milled to a powder by Messer
121 group GmbH, Germany. The powder was first mixed with milliQ water containing 0.01 %
122 v/v of a non-ionic surfactant (Tween 80, Sigma-Aldrich), and sequentially wet-sieved to
123 produce a size fraction similar to that of kaolin (for particle size distributions and details on
124 preparation of the test particles see Fig. S1-2, Supporting Information).

125 **Test suspensions.** Particle stocks were prepared by suspending weighed kaolin and MP with
126 M7 media (reconstituted lake water)³⁶; the volumes were adjusted to produce equal mass-
127 based concentrations (0.1, 1, 10, 100, 1000, 10000 mg L⁻¹). The test suspensions were
128 prepared in batches with 0 %, 20 %, 40 %, 60 %, 80 % and 100 % of MP contribution to
129 TSS. These test suspensions were then transferred to 50-mL polypropylene centrifuge tubes
130 and used in the exposure system.

131 **Experimental setup and procedures.** To demonstrate the application of the MP Ratio Test,
132 we conducted the *Daphnia* sp. acute immobilisation test (OECD 202), with some

133 modifications. The standard *Daphnia* immobilisation test assesses 48-h mortality
134 (immobilisation) in individuals exposed to a range of test concentrations. Based on our pilot
135 experiments with kaolin and the reported data for MP effects on *D. magna* mortality³⁷, we
136 prolonged the test duration to 96 h. Ten daphnids (<24 h old) were placed in each test tube
137 with the exposure media; 24 treatments (%MP × TSS concentration) were used with four
138 replicates per treatment and two particle-free controls per run. Three TSS concentrations
139 were tested for the 20-80 %MP mixtures and six concentrations with the single particle
140 exposure. When sealing the test tubes, care was taken to avoid trapping air bubbles inside the
141 tube. The tubes were mounted on a plankton wheel in a thermo-constant room at 21°C with a
142 light: dark cycle of 16:8 h and the test was terminated after 96 h by counting live and dead
143 animals. The daphnids were considered dead if they did not move for 30 s after being
144 agitated.

145

146 **Statistical analyses**

147 For each mixture, the LC₅₀ values and the corresponding 95 % confidence intervals were
148 calculated using a dose-response curve fitted with three-parameter logistic regression
149 (GraphPad Prism, v. 7.0; GraphPad Software, La Jolla California USA). The median
150 mortality in the particle-free control was used as the bottom constraint and ≤100 % mortality
151 as the top constraint. The LC₅₀ values were compared across the test mixtures to evaluate the
152 effect of %MP in the mixture; the non-overlapping confidence intervals were used as
153 evidence of the significant difference between the treatments (%MP). Further, we estimated
154 the corresponding LC₁₀ values and used them as a surrogate for NOEC.³⁸ One-phase
155 exponential decay function was fitted to describe the relationship between the LC₁₀ values
156 and %MP in the mixture. As NOEC, we used the interpolated LC₁₀ value corresponding to

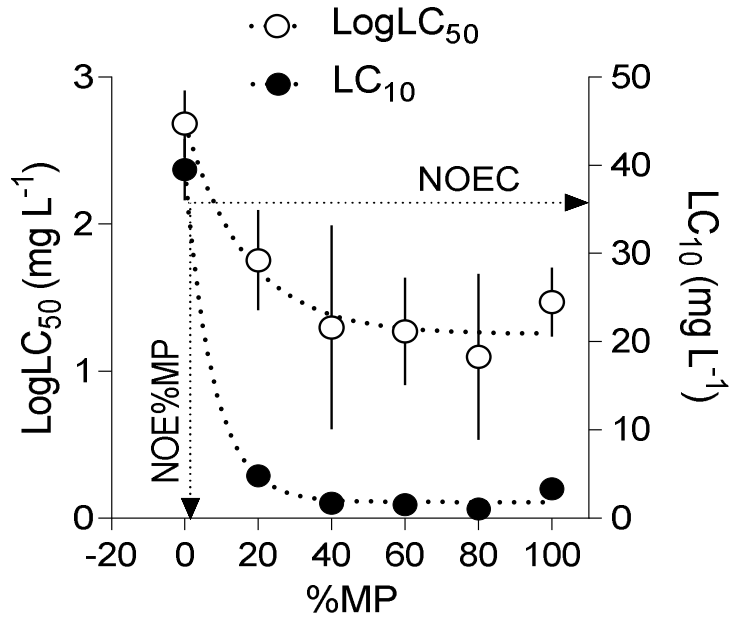
157 the lower bound of the 95 %-confidence interval for LC_{10} in pure kaolin. The critical
158 threshold for %MP in the mixture representing no effect level of %MP in the test system was
159 termed NOE%MP (No Effect Percentage of Microplastics).

160

161 RESULTS

162 Survival in the particle-free controls was high (average 95.4 %), and mortality in relation to
163 TSS followed the expected concentration-dependent response (Fig. S4, Supporting
164 Information). The LC_{50} values in the treatments with 20-100 % MP were significantly lower
165 than in the 100 % kaolin treatment, 13-43 $mg L^{-1}$ and 482 $mg L^{-1}$, respectively (Fig. 2), with
166 the non-overlapping confidence intervals between the treatments with pure kaolin and the
167 mixtures. The highest mortality was observed in the 80 % MP treatment; however, due to the
168 broad confidence intervals for the LC_{50} values, particularly in the mixtures, the differences
169 across all treatments with MP were not statistically significant.

170 The one-phase exponential decay function provided an adequate fit for the relationship
171 between the LC values and the %MP in the mixture (LC_{50} : $R^2 = 0.95$ and LC_{10} : $R^2 = 0.99$).
172 The curves suggested that at %MP exceeding 40 %, the LC values levelled off at 18 and 2 mg
173 L^{-1} for LC_{50} and LC_{10} , respectively. The fit for the LC_{10} values was used to derive the %MP
174 threshold (NOE%MP) above which significantly higher mortality was observed. This value
175 was determined as 1 % MP in the mixture with 36 $mg L^{-1}$ TSS.



176

177 **Figure 2.** Relationships between the estimated LogLC₅₀ values (mean and 95 % confidence
178 interval; left Y-axis) and the corresponding LC₁₀ values (right Y-axis) for *D. magna* and the
179 mass contribution of MP (%MP, 0 % to 100 %) in the test mixtures of MP and kaolin. One-
180 phase exponential decay was used to find the LC₁₀ for the MP-kaolin mixture corresponding
181 to the lower bound of the confidence interval for the kaolin treatment (NOEC) and the
182 NOE%MP representing no effect level of %MP in the test system.

183

184 DISCUSSION

185 **Test performance**

186 Using the MP Ratio Test, we delineated effects of MP from those of inorganic particles of
187 similar size and shape in the mixtures mimicking natural suspended solids in the aquatic
188 environment. Moreover, the test design allowed estimating critical levels for the suspended
189 MP that can be considered as hazardous. The hazardous level of MP contribution to TSS is an
190 important outcome of our test approach because the effects of particulate contaminants, such
191 as MP, depend not only on the absolute concentration but also on their relative contribution to
192 the suspended matter or sediment; the latter was also recently shown by Redondo-
193 Hasselerharm and co-workers.³⁹

194 Ecotoxicological data describing effect thresholds in ecologically meaningful settings are
195 needed in microplastic research to support the hazard assessment of solid polymer particles.
196 The MP Ratio Test can be used as a tool for screening a variety of polymer materials and
197 particle types as well as for selecting suitable test organisms and endpoints. It is conceptually
198 similar to both the already standard bioassay approach for testing toxic effluents and
199 sediments by serial dilutions⁴⁰ and evaluating algal toxicity in mixtures with varying
200 proportion of the test species.⁴¹ Furthermore, the need for well-characterised reference
201 materials when evaluating, for example, particle size effects, is also recognised in
202 nanomaterial toxicity assessment.^{42,43} However, to our knowledge, our study is the first to
203 assess the effects of suspended MP in mixtures with natural particles using a dose-response
204 approach.

205

206 **PET toxicity estimates**

207 We found PET MP to be more hazardous than kaolin for *D. magna*. The addition of PET
208 powder to the kaolin suspension increased *Daphnia* mortality, with LC₅₀ values dropping
209 more than 8-fold in mixtures with >20 % of MP. Moreover, TSS containing 1 % PET was
210 predicted to have significantly lower LC₁₀ than pure kaolin suspension. The corresponding
211 concentration for 1 % PET-kaolin mixture was 36 mg L⁻¹, which would represent NOEC of
212 TSS containing plastics (Fig. 2).

213 The reported effect concentrations of MP are highly variable and span orders of magnitude
214 even within the same level of biological organisation.²⁷ Unfortunately, only a few reports
215 provide dose-response data for MP-exposed microcrustaceans,^{33,37,44} and reference particles
216 are rarely employed.^{13,33} Although truly comparable published data on PET toxicity for
217 daphnids are not available; some reports are still relevant. For example, a 6-day static
218 exposure of the copepod *Parvocalanus crassirostris* to 14 mg L⁻¹ of PET (<11 µm; assuming
219 that particles are 10-µm and spherical, with a density of 1.38 g cm³) was found to decrease
220 population size.⁴⁵ This concentration is four times higher compared our LC₁₀ for 100 % MP.
221 The difference, at least partially, can be explained by the fact that our animals were starved
222 during the exposure. Moreover, the suspended amount of PET in the study with *P.*
223 *crassirostris*⁴⁵ is uncertain, because of the static exposure and the lack of information on the
224 MP mass in the system. For other planktonic filter-feeders and various MP, the reported LC₅₀
225 values are similar to what we have found for PET using the MP Ratio Test. For example, a
226 96-h LC₅₀ of 57.43 mg L⁻¹ was reported for *D. magna* neonates exposed to 1-µm
227 polyethylene (PE) particles, although some mortality was observed already at 12.5 mg L⁻¹.³⁷
228 Also with *D. magna*, exposure to PE fragments (10-75 µm) produced a 48-h LC₅₀ of 65 mg
229 L⁻¹.⁴⁴ These findings are comparable with the 96-h LC₅₀ observed in the 100 % MP treatment
230 (30 mg L⁻¹) as well as in the mixtures (13-43 mg L⁻¹), although our values are consistently
231 lower, which could be related to possible variations in MP aggregation and settling during the

232 exposure as well as in specific properties of polymers and particle size.⁴⁶ The differences in
233 the experimental setup highlight the importance of both particle characterisation in MP
234 research²⁸ and keeping the test particle in suspension when using pelagic feeders. Still, based
235 on either modelled data or some of the highest levels reported in the ocean,^{7,47} the levels of
236 MP accessible for zooplankton in nature are much – approximately two to four orders of
237 magnitude – lower than the experimentally determined MP levels for the observed effects.

238

239 **Important issues in experimental design**

240 The straightforward logistics of our approach makes it possible to examine a large number of
241 treatments (e.g., TSS concentrations, plastic material, particle size and shape) with reasonable
242 effort. When testing the PET effects on the daphnid survivorship, one person was able to
243 handle 50 experimental units per day routinely. Further developments should include a higher
244 number of treatments with lower %MP to provide more ecologically meaningful test
245 suspensions and improve the threshold estimates for the hazardous MP levels. Furthermore,
246 when more information of the environmentally realistic exposure becomes available, future
247 experimental designs can focus on narrowing concentration ranges, as well as including
248 endpoints that are more sensitive. When testing PET, for example, a higher resolution of the
249 %MP at the lower range (<20 %) would have provided a more precise estimate of NOEC and
250 the corresponding %MP in the mixture. Also, various reference particles can be used
251 depending on the research context, both natural and anthropogenic.

252 The selection of the reference particles is not a trivial task. One possible criterion is size
253 because size spectra for MP and many naturally occurring particles overlap (clay: <2 µm, silt:
254 2–50 µm, and sand: 50 µm –2 mm).⁴⁸ Another recently advocated option is to benchmark the
255 test particles to the reference MP using reference plastics supplied by commercial vendors,²⁸

256 which have a narrow size range and reliable certificates of analysis. Such reference materials
257 would add credibility to any adverse effects exerted by unknown materials. Here, we used
258 kaolin, because it has been relatively well studied in hazard assessment of suspended solids.⁴⁹
259 Several studies also suggest its low toxicity for daphnids,^{13,50} which was supported by our 96-
260 h LC₅₀ of 482 mg L⁻¹. However, in chronic tests with *D. magna*, Robinson and colleagues
261 observed a 7-d LC₅₀ of 74.5 mg L⁻¹,³² which can be related to both delayed effects and
262 possible difference in kaolin composition and aggregation during the test. Kaolin powder is a
263 commercially available standard product, but depending on the vendor, it might contain
264 impurities and vary in particle size distribution.

265 When testing effects of suspensions on planktonic organisms, it is essential to prevent that
266 particles sediment or float, because it will affect the encounter rate and intake by the animals.
267 Moreover, particles with different specific gravity will settle with different rates; hence
268 performing tests under static conditions would not provide stable exposure levels. The
269 exposure of planktonic organisms should preferably be conducted using a plankton wheel that
270 keeps particles in suspension, thus, ensuring stable exposure conditions.^{51,52} In plankton
271 ecology, the use of a plankton wheel is a standard procedure when conducting grazing
272 experiments because it minimises the sedimentation of algae. Although less common in
273 ecotoxicological testing, the plankton wheel has been used to assess the effects of suspended
274 clay and other particulate materials on planktonic filtrators.⁵³ Even though this method
275 requires some additional effort compared to static exposures commonly employed in OECD-
276 tests for soluble chemicals, it is a necessity for standardising the exposure conditions.

277 Another area of concern with respect to the standardisation of the testing procedure is particle
278 behaviour, such as aggregation and settling, during the exposure. Since suspended particles
279 may have complex behaviour, their dispersion stability during exposure should be controlled.
280 The extent of aggregation, i.e., aggregate size, composition and cohesion, is dependent on the

281 complex interplay between various components, such as particle material, size, concentration,
282 media ion composition, and organic material present. Aggregation may have affected the
283 amount and size spectra of PET and kaolin and could potentially explain the non-linear
284 response of LC to %MP. The relationship between the aggregate formation and the measured
285 response is particularly relevant at high TSS concentrations. On the other hand, aggregation,
286 at least to some extent, is preventable by, for example, dispersants and sonication, which may
287 be sufficient in short-term experiments. These methods are employed in nanomaterial effect
288 assessments⁴² and, in some cases, microplastic studies.⁵⁴

289 The exposure duration is yet another issue that needs to be considered in acute tests. Daphnid
290 energy expenditure may increase in the presence of non-food particles since the induced
291 filtering activity is similar as for food particles,⁵⁵ while the cost of cleaning appendages and
292 egestion through postabdominal rejections increase.¹⁵ Sensitivity to the refractory material
293 would increase with starvation, and the higher energy expenditure may decrease survival. In
294 *D. magna* neonates, the critical exposure time is 96 h, which marks the depletion of their fat
295 reserves.⁵⁶ Earlier studies using the *D. magna* acute immobilisation test for particle
296 suspensions have also shown that it is suitable to extend the exposure period to 96 h from the
297 standard 48 h to increase the sensitivity of the test.³⁷ However, if a test organism other than
298 *Daphnia* neonates is used, the duration of the exposure must be adjusted depending on its
299 capacity to withstand starvation.

300

301 **Implications for regulatory measures and concluding remarks**

302 Many suspension- and filter-feeders frequently face turbid environments with high
303 concentrations of refractory materials generated by natural processes, such as terrestrial
304 runoff, currents, and weather-induced bottom sediment resuspension,⁵⁷ but also by

305 anthropogenic activities, such as dredging and capping of contaminated sediments using
306 various materials. In aquatic ecology, conditions with elevated TSS are acknowledged as
307 stressful⁵⁸ and regulated by water quality standards.^{24,59} The quality standards vary across
308 regions and types of aquatic systems, e.g., lotic and lentic, and systems with different natural
309 levels of suspended solids. For example, the Alaskan state standard for clear-water lakes is a
310 maximum increase of TSS, above background levels, equivalent to 25 mg L^{-1} , whereas an
311 increase of 100 mg L^{-1} is acceptable for streams.²⁵ Similarly, hazard assessment of MP in
312 different systems would eventually require that effect-thresholds are established for the
313 critical MP concentration, such as NOE%MP, relative to the background TSS levels and their
314 combined interactions with biota.

315 By using reference particles – natural minerals or standardised plastics – with predictable
316 effects on the test organisms, one may identify and account for the general responses
317 anticipated from exposure to suspended solids. The relative importance of MP addition to the
318 suspension could thereby be assessed. Interestingly, the hazard level we found for MP
319 contribution to TSS using a planktonic organism is virtually identical to the EC₁₀ of 1 % MP
320 per sediment dry weight, reported for a benthic macroinvertebrate.³⁹ This could further
321 support the possibility to benchmark MP effects against the lower 95 % confidence bound of
322 the reference material and using the corresponding contribution of MP in the mixture to
323 estimate the hazard level of MP (Fig. 2).

324 The hazard assessment and the regulatory framework for MP contaminants in aquatic systems
325 require integration with an assessment of particulate matter pollution at large because the
326 approaches required to establish toxicity are similar. Moreover, raising levels of black carbon
327 in the atmosphere implies increased inputs of these particles in the aquatic systems, where
328 their environmental effects are also a matter of concern.⁶⁰ Addressing all types of particulate

329 pollution and focusing on physicochemical properties of these particles would provide a
330 translational value when developing testing and regulatory practices.

331

332 ACKNOWLEDGEMENT

333 This research was funded by projects WEATHER-MIC, irPLAST and MICROPOLL, which
334 are supported through the Joint Programming Initiative Healthy and Productive Seas and
335 Oceans (JPI-Oceans), Swedish Research Council for Environment, Agricultural Sciences and
336 Spatial Planning (FORMAS), the joint Baltic Sea research and development programme
337 (BONUS) and the Swedish Innovation Agency VINNOVA.

338 **Supporting Information.** Additional text and four figures describing particle preparation
339 and size distributions analyses, confirmed particle ingestion and separate dose-response
340 curves for all the tested mixtures of and kaolin and PET.

341

342 REFERENCES

- 343 (1) Imhof, H. K.; Ivleva, N. P.; Schmid, J.; Niessner, R.; Laforsch, C. Contamination of Beach
344 Sediments of a Subalpine Lake with Microplastic Particles. *Curr. Biol.* **2013**, *23* (19), R867–
345 R868.
- 346 (2) Hurley, R.; Woodward, J.; Rothwell, J. J. Microplastic Contamination of River Beds
347 Significantly Reduced by Catchment-Wide Flooding. *Nat. Geosci.* **2018**, *11* (4), 251–257.
- 348 (3) Cózar, A.; Echevarría, F.; González-Gordillo, J. I.; Irigoien, X.; Úbeda, B.; Hernández-León,
349 S.; Palma, Á. T.; Navarro, S.; García-de-Lomas, J.; Ruiz, A.; et al. Plastic Debris in the Open
350 Ocean. *Proc. Natl. Acad. Sci.* **2014**, *111* (28), 10239–10244.
- 351 (4) Peeken, I.; Primpke, S.; Beyer, B.; Gütermann, J.; Katlein, C.; Krumpfen, T.; Bergmann, M.;
352 Hehemann, L.; Gerds, G. Arctic Sea Ice Is an Important Temporal Sink and Means of
353 Transport for Microplastic. *Nat. Commun.* **2018**, *9* (1), 1505.
- 354 (5) Jambeck, J. R.; Geyer, R.; Wilcox, C.; Siegler, T. R.; Perryman, M.; Andrady, A.; Narayan, R.;
355 Law, K. L. Plastic Waste Inputs from Land into the Ocean. *Science* **2015**, *347* (6223), 768–
356 771.
- 357 (6) Backhaus, T.; Wagner, M. *Microplastics in the Environment: Much Ado about Nothing? A*
358 *Debate*; e26507v1; PeerJ Preprints, 2018.
- 359 (7) Lenz, R.; Enders, K.; Nielsen, T. G. Microplastic Exposure Studies Should Be
360 Environmentally Realistic. *Proc. Natl. Acad. Sci.* **2016**, 201606615.
- 361 (8) Burton, G. A. Stressor Exposures Determine Risk: So, Why Do Fellow Scientists Continue To
362 Focus on Superficial Microplastics Risk? *Environ. Sci. Technol.* **2017**, *51* (23), 13515–13516.
- 363 (9) Hale, R. C. Are the Risks from Microplastics Truly Trivial? *Environ. Sci. Technol.* **2018**, *52*
364 (3), 931–931.
- 365 (10) Wegner, A.; Besseling, E.; Foekema, E. m.; Kamermans, P.; Koelmans, A. a. Effects of
366 Nanopolystyrene on the Feeding Behavior of the Blue Mussel (*Mytilus Edulis L.*). *Environ.*
367 *Toxicol. Chem.* **2012**, *31* (11), 2490–2497.

- 368 (11) Cole, M.; Galloway, T. S. Ingestion of Nanoplastics and Microplastics by Pacific Oyster
369 Larvae. *Environ. Sci. Technol.* **2015**, *49* (24), 14625–14632.
- 370 (12) Sussarellu, R.; Suquet, M.; Thomas, Y.; Lambert, C.; Fabioux, C.; Pernet, M. E. J.; Le Goic,
371 N.; Quillien, V.; Mingant, C.; Epelboin, Y.; et al. Oyster Reproduction Is Affected by
372 Exposure to Polystyrene Microplastics. *Proc. Natl. Acad. Sci. U. S. A.* **2016**, *113* (9), 2430–
373 2435.
- 374 (13) Ogonowski, M.; Schür, C.; Jarsén, Å.; Gorokhova, E. The Effects of Natural and
375 Anthropogenic Microparticles on Individual Fitness in *Daphnia magna*. *PLOS ONE* **2016**, *11*
376 (5), e0155063.
- 377 (14) Cole, M.; Lindeque, P.; Fileman, E.; Halsband, C.; Galloway, T. S. The Impact of Polystyrene
378 Microplastics on Feeding, Function and Fecundity in the Marine Copepod *Calanus*
379 *helgolandicus*. *Environ. Sci. Technol.* **2015**, *49* (2), 1130–1137.
- 380 (15) Kirk, K. L. Effects of Suspended Clay on *Daphnia* Body Growth and Fitness. *Freshw. Biol.*
381 **1992**, *28* (1), 103–109.
- 382 (16) Kirk, K. L. Suspended Clay Reduces *Daphnia* Feeding Rate. *Freshw. Biol.* **1991**, *25* (2), 357–
383 365.
- 384 (17) Cranford, P. J.; Gordon, D. C. The Influence of Dilute Clay Suspensions on Sea Scallop
385 (*Placopecten Magellanicus*) Feeding Activity and Tissue Growth. *Neth. J. Sea Res.* **1992**, *30*,
386 107–120.
- 387 (18) Wilson, W. J. The Effects of Concentration and Particle Size of Suspended Materials on
388 Growth and Condition of the Pacific Oyster (*Crassostrea Gigas*). **1973**.
- 389 (19) Kaposi, K. L.; Mos, B.; Kelaher, B. P.; Dworjanyn, S. A. Ingestion of Microplastic Has
390 Limited Impact on a Marine Larva. *Environ. Sci. Technol.* **2014**, *48* (3), 1638–1645.
- 391 (20) Martínez-Gómez, C.; León, V. M.; Calles, S.; Gomáriz-Olcina, M.; Vethaak, A. D. The
392 Adverse Effects of Virgin Microplastics on the Fertilization and Larval Development of Sea
393 Urchins. *Mar. Environ. Res.* **2017**.
- 394 (21) Lloyd, D. S. Turbidity as a Water Quality Standard for Salmonid Habitats in Alaska. *North*
395 *Am. J. Fish. Manag.* **1987**, *7* (1), 34–45.

- 396 (22) McCabe, G. D.; O'Brien, W. J. The Effects of Suspended Silt on Feeding and Reproduction of
397 *Daphnia pulex*. *Am. Midl. Nat.* **1983**, *110* (2), 324–337.
- 398 (23) Robertson, M. J.; Scruton, D. A.; Clarke, K. D. Seasonal Effects of Suspended Sediment on the
399 Behavior of Juvenile Atlantic Salmon. *Trans. Am. Fish. Soc.* **2007**, *136* (3), 822–828.
- 400 (24) Multi-Sector General Permit for Stormwater Discharges Associated with Industrial Activity
401 (MSGP). United States Environmental Protection Agency (EPA), National Pollutant Discharge
402 Elimination System (NPDES) April 6, 2015.
- 403 (25) Lloyd, D. S. *Turbidity in Freshwater Habitats of Alaska – a Review of Published and*
404 *Unpublished Literature Relevant to the Use of Turbidity as a Water Quality Standard*; 85–1;
405 Alaska Department of Fish and Game - Habitat Division, 1985.
- 406 (26) Gerritsen, J.; Porter, K. G. The Role of Surface Chemistry in Filter Feeding by Zooplankton.
407 *Science* **1982**, *216* (4551), 1225–1227.
- 408 (27) Ogonowski, M.; Gerdes, Z.; Gorokhova, E. What We Know and What We Think We Know
409 about Microplastic Effects – A Critical Perspective. *Curr. Opin. Environ. Sci. Health* **2018**, *1*,
410 41–46.
- 411 (28) Connors, K. A.; Dyer, S. D.; Belanger, S. E. Advancing the Quality of Environmental
412 Microplastic Research. *Environ. Toxicol. Chem.* **2017**, *36* (7), 1697–1703.
- 413 (29) Zhao, S.; Zhu, L.; Li, D. Microplastic in Three Urban Estuaries, China. *Environ. Pollut.* **2015**,
414 *206*, 597–604.
- 415 (30) Cole, M.; Webb, H.; Lindeque, P. K.; Fileman, E. S.; Halsband, C.; Galloway, T. S. Isolation
416 of Microplastics in Biota-Rich Seawater Samples and Marine Organisms. *Sci. Rep.* **2014**, *4*.
- 417 (31) Fischer, E. K.; Paglialonga, L.; Czech, E.; Tamminga, M. Microplastic Pollution in Lakes and
418 Lake Shoreline Sediments – A Case Study on Lake Bolsena and Lake Chiusi (Central Italy).
419 *Environ. Pollut.* **2016**, *213*, 648–657.
- 420 (32) Robinson, S. E.; Capper, N. A.; Klaine, S. J. The Effects of Continuous and Pulsed Exposures
421 of Suspended Clay on the Survival, Growth, and Reproduction of *Daphnia magna*. *Environ.*
422 *Toxicol. Chem.* **2009**, *29* (1), 168–175.

- 423 (33) Casado, M. P.; Macken, A.; Byrne, H. J. Ecotoxicological Assessment of Silica and
424 Polystyrene Nanoparticles Assessed by a Multitrophic Test Battery. *Environ. Int.* **2013**, *51*, 97–
425 105.
- 426 (34) Rellstab, C.; Spaak, P. Starving with a Full Gut? Effect of Suspended Particles on the Fitness
427 of *Daphnia hyalina*. *Hydrobiologia* **2007**, *594* (1), 131–139.
- 428 (35) Besseling, E.; Wang, B.; Lürling, M.; Koelmans, A. A. Nanoplastic Affects Growth of *S.*
429 *Obliquus* and Reproduction of *D. Magna*. *Environ. Sci. Technol.* **2014**, *48* (20), 12336–12343.
- 430 (36) OECD. *Test No. 202: Daphnia sp. Acute Immobilisation Test*; Organisation for Economic Co-
431 operation and Development: Paris, 2004.
- 432 (37) Rehse, S.; Kloas, W.; Zarfl, C. Short-Term Exposure with High Concentrations of Pristine
433 Microplastic Particles Leads to Immobilisation of *Daphnia magna*. *Chemosphere* **2016**, *153*,
434 91–99.
- 435 (38) Beasley, A.; Belanger, S. E.; Brill, J. L.; Otter, R. R. Evaluation and Comparison of the
436 Relationship between NOEC and EC10 or EC20 Values in Chronic *Daphnia* Toxicity Testing.
437 *Environ. Toxicol. Chem.* **2015**, *34* (10), 2378–2384.
- 438 (39) Redondo-Hasselerharm, P. E.; Falahudin, D.; Peeters, E. T. H. M.; Koelmans, A. A.
439 Microplastic Effect Thresholds for Freshwater Benthic Macroinvertebrates. *Environ. Sci.*
440 *Technol.* **2018**, *52* (4), 2278–2286.
- 441 (40) Ferraz, M. A.; Alves, A. V.; de Cássia Muniz, C.; Pusceddu, F. H.; Gusso-Choueri, P. K.;
442 Santos, A. R.; Choueri, R. B. Sediment Toxicity Identification Evaluation (TIE Phases I and II)
443 Based on Microscale Bioassays for Diagnosing Causes of Toxicity in Coastal Areas Affected
444 by Domestic Sewage. *Environ. Toxicol. Chem.* **2017**, *36* (7), 1820–1832.
- 445 (41) Jónasdóttir, S. H.; Kjørboe, T.; Tang, K. W.; St. John, M.; Visser, A. W.; Saiz, E.; Dam, H. G.
446 Role of Diatoms in Copepod Production: Good, Harmless or Toxic? *Mar. Ecol. Prog. Ser.*
447 **1998**, *172*, 305–308.
- 448 (42) Handy, R. D.; Brink, N. van den; Chappell, M.; Mühlhling, M.; Behra, R.; Dušinská, M.;
449 Simpson, P.; Ahtiainen, J.; Jha, A. N.; Seiter, J.; et al. Practical Considerations for Conducting

- 450 Ecotoxicity Test Methods with Manufactured Nanomaterials: What Have We Learnt so Far?
451 *Ecotoxicology* **2012**, *21* (4), 933–972.
- 452 (43) Stone, V.; Nowack, B.; Baun, A.; van den Brink, N.; von der Kammer, F.; Dusinska, M.;
453 Handy, R.; Hankin, S.; Hassellöv, M.; Joner, E.; et al. Nanomaterials for Environmental
454 Studies: Classification, Reference Material Issues, and Strategies for Physico-Chemical
455 Characterisation. *Sci. Total Environ.* **2010**, *408* (7), 1745–1754.
- 456 (44) Frydkjær, C. K.; Iversen, N.; Roslev, P. Ingestion and Egestion of Microplastics by the
457 Cladoceran *Daphnia magna*: Effects of Regular and Irregular Shaped Plastic and Sorbed
458 Phenanthrene. *Bull. Environ. Contam. Toxicol.* **2017**, *99* (6), 655–661.
- 459 (45) Heindler, F. M.; Alajmi, F.; Huerlimann, R.; Zeng, C.; Newman, S. J.; Vamvounis, G.; van
460 Herwerden, L. Toxic Effects of Polyethylene Terephthalate Microparticles and Di(2-
461 Ethylhexyl)Phthalate on the Calanoid Copepod, *Parvocalanus Crassirostris*. *Ecotoxicol.*
462 *Environ. Saf.* **2017**, *141*, 298–305.
- 463 (46) Rist, S.; Baun, A.; Hartmann, N. B. Ingestion of Micro- and Nanoplastics in *Daphnia magna* –
464 Quantification of Body Burdens and Assessment of Feeding Rates and Reproduction. *Environ.*
465 *Pollut.* **2017**, *228*, 398–407.
- 466 (47) Enders, K.; Lenz, R.; Stedmon, C. A.; Nielsen, T. G. Abundance, Size and Polymer
467 Composition of Marine Microplastics ≥ 10 μ m in the Atlantic Ocean and Their Modelled
468 Vertical Distribution. *Mar. Pollut. Bull.* **2015**, *100* (1), 70–81.
- 469 (48) Handy, R. D.; Brink, N. van den; Chappell, M.; Mühlhling, M.; Behra, R.; Dušinská, M.;
470 Simpson, P.; Ahtiainen, J.; Jha, A. N.; Seiter, J.; et al. Practical Considerations for Conducting
471 Ecotoxicity Test Methods with Manufactured Nanomaterials: What Have We Learnt so Far?
472 *Ecotoxicology* **2012**, *21* (4), 933–972.
- 473 (49) McFarland, V. A.; Peddicord, R. K. Lethality of a Suspended Clay to a Diverse Selection of
474 Marine and Estuarine Macrofauna. *Arch. Environ. Contam. Toxicol.* **1980**, *9* (6), 733–741.
- 475 (50) Weltens, R.; Goossens, R.; Puymbroeck, S. V. Ecotoxicity of Contaminated Suspended Solids
476 for Filter Feeders *Daphnia magna*. *Arch. Environ. Contam. Toxicol.* **2000**, *39* (3), 315–323.

- 477 (51) Robins, D. B.; Bellan, I. E. A Controlled-Temperature Plankton Wheel. *Mar. Biol.* **1986**, *92*
478 (4), 587–593.
- 479 (52) Anraku, M. Some Technical Problems Encountered in Quantitative Studies of Grazing and
480 Predation by Marine Planktonic Copepods. *J. Oceanogr. Soc. Jpn.* **1964**, *20* (5), 221–231.
- 481 (53) Kirk, K. L.; Gilbert, J. J. Suspended Clay and the Population Dynamics of Planktonic Rotifers
482 and Cladocerans. *Ecology* **1990**, *71* (5), 1741–1755.
- 483 (54) Hüffer, T.; Praetorius, A.; Wagner, S.; von der Kammer, F.; Hofmann, T. Microplastic
484 Exposure Assessment in Aquatic Environments: Learning from Similarities and Differences to
485 Engineered Nanoparticles. *Environ. Sci. Technol.* **2017**, *51* (5), 2499–2507.
- 486 (55) Wiedner, C.; Vareschi, E. Evaluation of a Fluorescent Microparticle Technique for Measuring
487 Filtering Rates of *Daphnia*. *Hydrobiologia* **1995**, *302* (2), 89–96.
- 488 (56) Tessier, A. J.; Henry, L. L.; Goulden, C. E.; Durand, M. W. Starvation in *Daphnia*: Energy
489 Reserves and Reproductive Allocation. *Limnol. Oceanogr.* **1983**, *28* (4), 667–676.
- 490 (57) Gulati, R.; Demott, W. The Role of Food Quality for Zooplankton: Remarks on the State-of-
491 the-Art, Perspectives and Priorities. *Freshw. Biol.* **1997**, *38* (3), 753–768.
- 492 (58) Pelletier, M.; Ho, K.; Cantwell, M.; Perron, M.; Rocha, K.; Burgess, R. M.; Johnson, R.; Perez,
493 K.; Cardin, J.; Charpentier, M. A. Diagnosis of Potential Stressors Adversely Affecting
494 Benthic Invertebrate Communities in Greenwich Bay, Rhode Island, USA. *Environ. Toxicol.*
495 *Chem.* **2016**, *36* (2), 449–462.
- 496 (59) Canadian Council of Ministers of the Environment. 2002. *Canadian Water Quality Guidelines*
497 *for the Protection of Aquatic Life: Total Particulate Matter*; Canadian environmental quality
498 guidelines; Winnipeg, 1999; pp 1–13.
- 499 (60) Shrestha, G.; Traina, S. J.; Swanston, C. W. Black Carbon's Properties and Role in the
500 Environment: A Comprehensive Review. *Sustainability* **2010**, *2* (1), 294–320.