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**Effects of exposure to sublethal concentrations of
methoxyfenozide on honey bee colony activity and
thermoregulation**

**William G. Meikle, Vanessa Corby-Harris, Mark J. Carroll, Milagra Weiss,
Lucy A. Snyder, Charlotte A.D. Meador, Eli Beren, and Nicholas Brown**

Carl Hayden Bee Research Center, USDA-ARS, Tucson, AZ, USA

19 **Abstract**

20 Methoxyfenozide is an insect growth regulator (IGR) commonly used in agricultural to simultaneously
21 control pests and preserve beneficial insect populations; however, its impact on honey bees is not fully
22 understood. We conducted field and laboratory experiments to investigate bee health in response to field-
23 relevant doses of this pesticide. Significant effects were observed in honey bee colony flight activity and
24 thermoregulation after being treated with methoxyfenozide. Data collected indicated that hives fed 500
25 ppb methoxyfenozide treated pollen patty had: 1) a significantly reduced rate of daily hive weight loss
26 due to forager departure at the start of the colony's daily activity; 2) the end of the colony's daily activity
27 delayed by 17-21 minutes compared to Control; and 3) higher temperature variability during the winter.
28 Colonies in the 125 ppb treatment group had fewer differences with the Control group, but did show a
29 delay in the foraging end time by 30-46 minutes compared to the Control. Bee colony metrics of adult bee
30 mass and brood surface area, and individual bee measurements of head weight, newly-emerged bee
31 weight, and hypopharyngeal gland size were not significantly affected by the methoxyfenozide exposure
32 levels of our experiments. An experiment conducted using the same treatment groups in the spring
33 resulted in fewer differences among groups than did the experiments conducted in the fall. Analyses of
34 methoxyfenozide concentrations in the treatment patty, wax, and bee bread showed that: 1) observed
35 methoxyfenozide concentrations were about 18-60% lower than the calculated concentrations; 2) no
36 residues were observed in wax in any treatment; and 3) methoxyfenozide was detected in stored bee bread
37 in the 500 ppb treatment, at concentrations about 1-2.5% of the observed concentration for that treatment.
38 These results suggest that there may be significant effects on honey bee colony behavior (and possibly
39 health) in the field that are difficult to detect through traditional hive inspections and individual metrics.

40

41 **Key words:** insect growth regulator, continuous hive monitoring, bee colony behavior, pollen patty,
42 pesticide concentration

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44

45 **Introduction**

46 Honey bee colonies are frequently exposed to agrochemicals, including many different classes of
47 insecticides, among them insect growth regulators (IGRs) [1]. Broadly speaking, IGRs interfere with
48 insect growth and development in their target pest species [2]. The IGR methoxyfenozide is an ecdysone
49 receptor agonist that binds the ecdysteroid receptor and activates the ecdysteroid signaling pathway [3, 4].
50 Unlike the ecdysteroid hormone that precisely controls larval development by binding to the receptor for
51 a defined period of time, ecdysteroid agonists bind irreversibly, disrupting the expression of genes
52 involved in cuticle development, sclerotization, and ecdysis [4, 5]. Methoxyfenozide specifically targets
53 lepidopterans and research shows that it has high affinity for the lepidopteran ecdysteroid receptor that is
54 not seen in other insect orders (reviewed in [3, 4]). Although the binding specificity of methoxyfenozide
55 to the *Apis mellifera* ecdysteroid receptor has not been determined, it is safe to assume that it should not
56 bind. Fewer non-target effects makes methoxyfenozide an attractive, targeted solution for controlling
57 pests while preserving beneficial insect populations. Methoxyfenozide is registered in more than 50
58 countries for use in a variety of crops, including those pollinated by honey bees. Methoxyfenozide use has
59 increased 15-fold between 2001 and 2015, from ~30,000 to ~450,000 pounds annually [6], mostly in
60 orchards, to control lepidopteran pests like the navel orangeworm, *Amyelois transitella* (Walker) [7].

61 Reported toxicity of methoxyfenozide to young worker bees is low, with an acute toxicity LD₅₀
62 greater than 100 ug/bee [8] and, when formulated with spinetoram, an oral toxicity LC₅₀ of about 712
63 mg/L (ppm) in young workers [9]. Nevertheless growers have been advised to avoid spraying during
64 bloom because the impact of methoxyfenozide on earlier life stages, older workers, or on colonies is not
65 fully understood [10]. Despite these spray recommendations, methoxyfenozide has been detected in
66 commercial colonies in bees (9-21 ppb) and hive materials including wax (80-495 ppb), pollen (35-128
67 ppb), and honey (3 ppb) [1, 11]. In a recent study, methoxyfenozide has been shown to decrease forager
68 survival at field-relevant doses [12].

69 Other IGRs have caused delayed sublethal effects in adult honey bees after larval exposure.
70 Larval bees exposed to the juvenile hormone analog pyriproxyfen in pollen patty at 321 ppb showed
71 increased deformities as adults, reduced adult survivorship and exposed adults had difficulties integrating
72 into the general adult population; fewer effects were observed at the lower concentration of 129 ppb [13].
73 The impact of methoxyfenozide on honey bee colonies is not fully understood. Although
74 methoxyfenozide is marketed as a safe pesticide due to its specificity to certain pests, it is nonetheless
75 possible that honey bees are negatively impacted simply due to the energetic cost of detoxifying an
76 exogenous chemical [14-16]. Pesticide exposure can lead to increased expression of stress response and
77 detoxification genes [17], but expression patterns vary [18]. It is unclear whether detoxification, *per se*, is
78 stressful to honey bees, although metabolomic analyses of bees exposed to plant secondary toxins suggest
79 that it is [14] and nurse-aged bees exposed to some classes of pesticides have reduced hypopharyngeal
80 glands [18, 19], a pattern observed with other stressors [20].

81 Effects that are difficult to detect on the level of the individual may be detected on the level of the
82 colony [21]. Continuously monitoring weight and internal hive temperature of honey bee hives has
83 provided information on bee colony growth and activity [22-24]. In those studies, continuous weight and
84 hive temperature data were detrended by subtracting the 25-hour running average from the raw data.
85 Running average weight data provided information on longer-term colony growth while the detrended
86 within-day data were modeled using sine curves to yield information on foraging activity and success.
87 Similarly, the average and detrended hive temperature data were related to capped brood levels [22].
88 These approaches were used to detect colony-level treatment effects of sublethal exposure of the
89 neonicotinoid pesticide imidacloprid on flight activity and internal hive temperature control (colony
90 thermoregulation) [25]. In another approach, continuous hive weight data were detrended by subtracting
91 the value at midnight, rather than the running average, from each of the values of that day until the next
92 midnight [26]. In that approach, a piecewise regression model was fitted to single-day datasets in the
93 original time scale using an R function [27] based on a bootstrapping method [28]. Several parameters of

94 the piecewise regression provide biological or behavioral interpretations, such as break points around
95 dawn and dusk reflecting the beginning and ending of the daily active periods for the hive [26].

96 To more fully address the possible consequences of methoxyfenozide exposure on honey bees,
97 we conducted replicated field and laboratory investigations of bee health in response to field-relevant
98 doses of this pesticide. In two different years, we mixed methoxyfenozide into pollen patty and fed it to
99 colonies. Colony parameters such as brood area, adult weight, foraging activity, and hive temperature
100 were measured. We find that methoxyfenozide does not exert a massively negative effect on honey bees,
101 but that small differences among hives in foraging activity and hive temperature regulation appear to be
102 exacerbated in the high-dose treatments compared to the control.

103

104 **Materials and methods**

105 **1. Preparation of pollen supplement**

106 A 50 mg/ml stock solution of methoxyfenozide was prepared in acetone. 18 kg of pollen patty supplement
107 was prepared in 1.5 kg batches in a stand mixer at a ratio of 1:1:1 corbicular pollen (Great Lakes Bee
108 Co.):granulated sugar:drivert sugar (Domino Foods). For the 500 ppb treatment, 15 μ L of the stock
109 solution of methoxyfenozide was added to 235 μ L of acetone yielding a 3 mg/ml solution. 250 μ L of this
110 3 mg/ml solution was added to 130 ml of water and then thoroughly mixed with 1.5 kg of patty.

111 Similarly, the 125 ppb treatment patty was prepared by adding 3.75 ul of the stock solution to 246.25 ul
112 acetone yielding a 0.75 mg/ml solution. 250 μ L of this 0.75 mg/ml solution was added to 130 mL water
113 and thoroughly mixed into 1.5 kg of patty. All of the diet was prepared at the same time for the Fall 2016
114 experiment. The diet was divided into 100 g patties and stored at -20°C until the supplement was fed.

115 Pollen supplement for Fall 2017 and Spring 2018 experiments was prepared in a similar manner
116 except that the methoxyfenozide stock solution was prepared to 6 mg/ml in acetone. 125 μ L of that
117 solution was then mixed with 100 mL water for a concentration of 0.75 mg/100 mL and applied to 1.5 kg
118 supplement for a concentration of 500 ppb. Likewise, 31 μ L of stock solution was diluted with 93 μ L

119 acetone and mixed with 100 mL water per 1.5 kg diet for a concentration of 125 ppb. Control patties were
120 treated with 125 μ L pure acetone and 100 mL water. Patties were divided and stored in the same manner.
121 Patty samples (3 g) were submitted to the Laboratory Approval and Testing Division, Agricultural
122 Marketing Service, USDA (LATD), Gastonia, NC, to determine methoxyfenozide concentrations.

123

124 **2. Fall 2016 field experiment.**

125 In August 2016, eighteen honey bee colonies were selected from apiaries at or near the Carl Hayden Bee
126 Research Laboratory, USDA-ARS, Tucson, AZ (32°16'30.17"N, 110°56'28.52"W) and the Santa Rita
127 Experimental Range (SRER) (31°46'38.08"N, 110°51'47.39"W) and moved to a single site at SRER in
128 August, 2016. Colonies had been stocked with Cordovan-Italian queens (C.F. Koehnen & Sons, Glenn,
129 CA) and housed in painted, 10-frame wooden Langstroth deep boxes fitted with migratory wooden lids
130 (Mann Lake Ltd, Hackensack, MN). Colonies were 6-18 months old at the start of the experiment. The
131 SRER apiary was provided with a permanent water source and hives were spaced 1-3 m apart. Hives were
132 placed on stainless steel electronic scales (TEKFA® model B-2418 and Avery Weigh-Tronix model
133 BSAO1824-200) (max. capacity 100 kg) connected to 12-bit dataloggers (Hobo® U-12, Onset Computer
134 Corporation) that were set to record weight every 5 minutes. The system had an overall precision of
135 approximately ± 20 g. On the same day, a hive temperature sensor (iButton Thermochron, precision ± 0.06
136 °C) enclosed in plastic tissue embedding cassettes (Thermo Fisher Scientific, Waltham, MA) was stapled
137 to the center of the top bar on the 5th frame in each hive and set to record every 30 min.

138 On 29-30 August, 2016, hives were given a full pre-treatment assessment (see [24, 29]). Briefly,
139 the hive was opened after the application of smoke, and each frame was lifted out, gently shaken to
140 dislodge adult bees, photographed using a 16.3 megapixel digital camera (Canon Rebel SL1, Canon USA,
141 Inc., Melville, NY), weighed on a portable scale (model EC15, Ohaus, 15 kg max. cap.), and replaced in
142 the hive. Frames were removed and replaced sequentially. During this first assessment (but not
143 subsequent assessments), all hive components (i.e. lid, inner cover, box, bottom board, frames, entrance

144 reducer, internal feeder) were also shaken free of bees and weighed to yield an initial mass of all hive
145 components. At the initial inspection, 3-5 g of wax were collected from each hive into 50 ml centrifuge
146 tubes and stored at -80°C; samples collected in September, prior to treatment, were pooled and subjected
147 to a full panel analysis for residues of 192 pesticides and fungicides, from all major classes, by LATD.
148 Samples from later assessments were pooled within treatment group and subjected only to
149 methoxyfenozide residue analysis.

150 The total weight of the adult bee population (whole colony adult bee mass) was calculated by
151 subtracting the combined weights of hive components obtained in the pre-treatment assessment from the
152 total weight of bees and hive materials recorded the midnight prior to the inspection. The area of sealed
153 brood per frame was estimated from the photographs using ImageJ version 1.47 software (W. Rasband,
154 National Institutes of Health, USA). The 18 colonies were randomly assigned to three treatment groups,
155 to be given the Low (125 ppb), High (500 ppb) and Control treatments. Care was taken to ensure the
156 average adult bee mass per treatment group varied by no more than 100 g among all treatments to
157 minimize pre-existing differences (Table 1). Each hive was given 100g pollen patty twice a week with the
158 appropriate concentration starting on 2 September, continuing weekly for 9 weeks until 28 October (1800
159 g treatment pollen patty total per hive). Consumption of the patty was measured by weighing any patty
160 (wet weight) that remained after one week. Hives were assessed on 2 November, 2016 (first post-
161 treatment assessment, 61 d after initial treatment). Each colony was then given 3 kg of 1:1 sugar syrup
162 and frame of capped honey on 10 November because of low food stores. Colonies were assessed and
163 sampled for the final time on 30 January, 2017, (2nd post-treatment assessment) to determine the long-
164 term effects of methoxyfenozide exposure on overwintering.

165

166 **3. Fall 2017 field experiment**

167 The experiment described above was repeated in 2017. In April, 2017 twenty colonies were
168 started from packages from the same supplier as the previous year and installed in SRER at a site about 2

169 km away from the Fall 2016 site. Colonies were in single deep hive boxes and maintained with new
170 Cordovan-Italian queens from the same queen supplier as the first year. Colonies were fed supplemental
171 pollen patty in the spring, and 12 kg 1:1 sugar syrup between May and early September. On 13 September
172 2017, full pre-treatment hive assessments were conducted on all colonies. Second deep boxes were added
173 to 6 of the colonies because of their size. Colonies were divided into 3 groups with similar adult bee
174 masses (<100 g difference), and each colony was given 100 g treatment pollen patty twice per week
175 beginning on 22 September and continuing for 8 weeks until 9 November (48 d after initial treatment,
176 1600 g treatment pollen patty total). Three supplemental feedings of 3 kg of 1:1 sugar syrup were
177 provided to each colony during the treatment period. The 1st post-treatment hive assessments were
178 conducted on 15 November. Smaller colonies were reduced to single boxes for overwintering on 16
179 November 2017. The 2nd post-treatment hive assessments were conducted on 13 February 2018. From 1
180 to 2 g of bee bread was collected from each hive at each assessment. As with the wax samples for the Fall
181 2016 experiment, samples collected in September, prior to treatment, were pooled and subjected to a full
182 panel of residue analyses while samples from later assessments were pooled within treatment group and
183 subjected only to methoxyfenozide residue analysis. Samples of protein patty from each treatment were
184 also analyzed for methoxyfenozide concentration. In addition, newly-emerged bees (NEBs) were also
185 sampled by pressing an 8 cm x 8 cm x 2 cm mesh queen cage into a section of capped brood, then
186 returning the following day to collect NEBs that had emerged within the cage over the previous 24 h. The
187 NEBs were then placed in a 50 mL centrifuge tube, frozen on dry ice, and stored at -80°C. At the
188 laboratory, 5 bees per hive per assessment date were placed in Eppendorf tubes, weighed, dried for 72 h at
189 60°C, then re-weighed to determine average wet and dry weight per bee.

190

191 **4. Spring 2018 field experiment**

192 The hives used in the Fall 2017 experiment were retained in the same treatment groups and
193 treated in the same manner as in the previous fall, using pollen patties with the same concentrations of

194 methoxyfenozide. Treatments consisting of 100 g pollen patty were started on 8 March and continued
195 twice weekly for 6 weeks until 13 April (42 d after initial treatment, 1200 g treatment pollen patty total).
196 All hives were given 4 L supplemental sugar syrup on 30 March. Hives were evaluated on 18 April (1st
197 post-treatment assessment) and again on 24 May (2nd post-treatment assessment). Bee bread and NEBs
198 were sampled at each assessment.

199

200 **5. Adult bee head weights from Fall 2017 experiment**

201 Head weight was measured on samples of nurse-aged bees collected from all colonies after the
202 November 2017 post-treatment hive evaluation. Nurse-aged bees visiting cells containing larvae for a
203 period of ≥ 5 seconds were collected. The captured bees were immediately flash frozen in the field and
204 were maintained at -80°C until their heads were weighed. To obtain head weights, each head was thawed
205 and weighed using a Sartorius CP2P microbalance at a resolution of 0.01 mg (see [30]).

206

207 **6. Hypopharyngeal gland size of nurse-aged workers fed methoxyfenozide**

208 In order to assess whether oral methoxyfenozide exposure affected nurse-aged bee health, HPG
209 size was measured in caged nurse-aged bees fed either 1000 ppb of methoxyfenozide in 30% sugar syrup
210 or a control dose of methoxyfenozide-free sugar syrup. A comparatively high dose was chosen to increase
211 the probability of a measurable effect. NEBs emerged overnight from brood frames taken from three
212 colonies in a temperature-controlled dark room ($32\text{--}34^{\circ}\text{C}$, 30–40% relative humidity). The next morning,
213 the NEBs were distributed among eight cages to a density of 100 bees per cage. Cage dimensions were
214 11.5 x 7.5 x 16.5 cm, with narrow sides, top and base made of Plexiglas® and the broad sides and floor
215 made of 3mm aperture galvanized steel mesh. A plastic 50 mL bottle for distilled water and a plastic
216 30mL bottle for syrup, each with a small hole in the lid, were inverted and placed over holes on the top of
217 each cage. Four cages were provided with 1000 ppb methoxyfenozide in 30% sucrose syrup and four
218 cages with 30% sucrose syrup without the pesticide as controls. All cages were provided with pollen

219 patty, 1:1:1 sucrose : drivert sugar : natural pollen (Great Lakes Pollen, Bulkfoods.com) *ad libitum*. The
220 caged bees were maintained at 34°C and 30–40% relative humidity. At 8d after emergence, 10 bees per
221 cage were flash frozen and maintained at -80°C until their HPGs were dissected and measured (see [31]).
222 Between 10 and 12 HPG acini were measured for each gland to obtain an average HPG acinus size for
223 each bee.

224

225 **7. Data analysis: Hive assessment, patty consumption, NEB weights, nurse bee** 226 **head weights and HPG size**

227 Adult bee masses, capped brood surface areas and NEB dry weights were compared among
228 treatments across sampling occasions using repeated-measures MANOVA (SAS version 9.4) with
229 treatment, experiment and day as main factors, all 2-way interactions, and with pre-treatment adult bee
230 mass as a covariate to control for pre-existing colony differences. Per colony patty consumption and nurse
231 bee head weights were analyzed using ANOVA, in which treatment was the main effect and hive was a
232 random effect. Caged-bee data were analyzed using ANOVA, with treatment as a fixed effect and cage
233 replicate as a random effect. Post hoc contrasts with the Bonferroni correction for multiple comparisons
234 were reported for significant treatment effects.

235

236 **8. Data analysis: Hive weight**

237 Continuous hive weight data were considered with respect to average daily weight and within-day
238 changes. Weight data were detrended for each day by subtracting the hive weight estimate at midnight (or
239 closest time thereafter) from each subsequent weight value over the next 24 h (see [25]). The resulting
240 within-day weight datasets were modeled using the “segmented” function in R which fits a segmented
241 line derived from a linear or generalized linear model to a dependent variable using a bootstrapping
242 procedure [27]. Bee colonies outside of a nectar flow during winter tend to lose weight and exhibit
243 consistent daily patterns (Fig 1). Piecewise regressions with 4 breakpoints were fit to the data, which

244 yielded estimates for 10 parameters: 4 break point values, 5 slope values and the adjusted r^2 . Because the
245 data were detrended by subtracting the raw data value at midnight, daily datasets were mathematically
246 independent. A repeated measures MANOVA was conducted on these daily parameter values of interest
247 (letters refer to Fig. 1):

- 248 1. Beginning of initial forager departure (usually the 1st break point, Point B);
- 249 2. End of initial forager departure (usually 2nd break point, Point C);
- 250 3. Time of final forager return (usually the 4th break point, Point E);
- 251 4. Average slopes of the 1st and 5th segments, which are weight changes at night when colonies are
252 not foraging (Segments AB and EF);
- 253 5. Slope of the first segment after initial forager departure, usually the 2nd segment (usually
254 Segment BC); and
- 255 6. Slope of the last segment before dusk, usually the 4th segment (usually Segment DE).

256 For statistical analysis, if the 1st break point occurred before 4AM, the 2nd break point was used as the
257 time of initial forager departure (with no restrictions placed on that second estimate) and the slope of the
258 3rd, rather than 2nd, segment was used as the rate of weight loss due to forager departure. Likewise, if the
259 4th break point occurred after 8PM then the 3rd break point was taken as the time of final forager return
260 (with no restrictions placed on that second estimate). For all analyses involving slopes, the pre-treatment
261 adult bee mass was included as a covariate to control for pre-existing differences in forager populations.

262

263 **Figure 1.** Graph showing data and fitted piecewise regression to data from a single hive in the Control
264 group for a single day (24 November 2017) during a nectar and pollen dearth.

- 265 1. **Point A:** First weight measure at midnight or shortly thereafter;
- 266 2. **Segment AB:** Inactive period in the early morning; hive weight change is likely due to bee
267 respiration and changes in the moisture content of nectar, pollen and wooden hive parts;
- 268 3. **Point B:** Bee departure at beginning of active period (usually close to dawn);

- 269 4. **Segment BC:** Active period usually showing hive weight loss due to far greater numbers of
270 departing bees compared to returning bees;
- 271 5. **Point C:** Point at which mass of returning bees increases relative to mass of departing bees;
- 272 6. **Segment CD:** Continued increased return of foragers;
- 273 7. **Point D:** Point at which mass of returning foragers, including nectar and pollen as well as bee
274 weight loss due to respiration, exceeds the mass of departing bees plus weight loss due to drying
275 and respiration in the colony;
- 276 8. **Segment DE:** Inactive period with hive weight change driven mainly by respiration and changes
277 in ambient humidity – usually close to parallel with segment AB;
- 278 9. **Point E:** Return of bees to the hive around dusk is completed;
- 279 10. **Segment EF:** Inactive period in the late evening; similar dynamics with moisture content of hive,
280 bee respiration and ambient humidity;
- 281 11. **Point F:** Last weight measure just before midnight.

282

283 **9. Data analysis: Hive temperature**

284 Internal hive temperature data were divided into daily average values and within-day detrended
285 data. Detrended data were calculated as the difference between the 25 hour running average and the raw
286 data [24]. Sine curves were fit to 3-day subsamples of detrended data taken sequentially by day, and curve
287 amplitudes, representing estimates of daily hive temperature variability, were used as response variables.
288 For hives with two boxes, only temperature data from the lower (brood) box were analyzed. Repeated
289 measures MANOVA (Proc Glimmix, SAS Inc. 2002) was used to evaluate the effects of treatment, day,
290 and their interaction, with the pre-treatment whole colony adult bee mass as a covariate on both the
291 average daily hive temperature and the amplitudes of the fit sine curves. Temperature amplitude datasets
292 were reduced to one value per hive point every 3 d for repeated measures analysis to ensure no overlap
293 between subsamples.

294

295 **Results**

296 **1. Hive assessment**

297 Methoxyfenozide treatment did not have a measurable impact on either whole colony adult bee
298 mass ($P=0.73$) or brood surface area ($P=0.43$) in the Fall 2016 or Fall 2017 experiments , nor were the
299 two experiments different from each other with respect to these metrics ($P=0.36$ and $P=0.26$, respectively)
300 (Table 1). Considered separately, neither adult bee mass nor brood surface area in the Spring 2018
301 experiment was affected by treatment ($P=0.14$ and $P=0.37$, respectively). During the Spring 2018
302 experiment, two colonies in the 125 ppb treatment group and one colony in the 500 ppb treatment group
303 died, in all cases around 1 May. Pre-treatment adult bee mass was significantly correlated with adult bee
304 mass at the 2nd post treatment assessment for the Fall 2016, Fall 2017 and Spring 2018 field experiments
305 (adjusted $r^2= 0.27, 0.44$ and 0.31 , respectively). Two colonies in the 500 ppb treatment group and one in
306 the 125 ppb treatment group died in the Fall 2016 experiment, and one colony in the 500 ppb treatment
307 group and two in the 125 ppb treatment group died in the Spring 2018 experiment.

308

309

310 **Table 1.** Adult bee masses and brood surface areas for the Fall 2016, Fall 2017 and Spring 2018 field

311 experiments.

312

Dates	Adult bee mass (g)			Capped brood surface area (cm ²)		
	125 ppb	500 ppb	control	125 ppb	500 ppb	control
Fall 2016:						
30 Aug. 2016	2166 ±141	2254 ±234	2202 ±204	2120.6 ±220.7	1972.4 ±436.0	2475.5 ±291.7
2 Nov. 2016	1550 ±153	1498 ±330	1633 ±139	583.2 ±53.1	498.0 ±134.6	546.6 ±97.7
1 Feb. 2017	782 ±224	777 ±296	1352 ±107	9.1 ±8.0	3.7 ±0.6	11.6 ±7.9
Fall 2017:						
13 Sep. 2017	1641 ±251	1591 ±208	1595 ±141	426.4 ±112.5	630.1 ±158.3	756.7 ±98.3
15 Nov. 2017	1563 ±170	1496 ±224	1434 ±83	274.9 ±38.4	210.6 ±33.1	269.4 ±53.0
13 Feb. 2018	1010 ±230	920 ±150	810 ±90	412.8 ±96.2	320.1 ±125.4	496.7 ±153.6
Spring 2018:						
18 Apr. 2018	1211 ±422	1453 ±416	780 ±136	1607.2 ±413.6	2613.8 ±640.1	1055.6 ±245.7
23 May 2018	1054 ±351	1553 ±491	534 ±162	1110.6 ±169.9	1292.7 ±259.2	477.3 ±167.7

313

314 **2. Patty consumption**

315

Bee colonies in all groups consumed all the pollen feed in the Fall 2016 and Fall 2017

316

experiments. In the Spring 2018 experiment some colonies did not consume all 1100 g patty but average

317

values (wet weight) among treatment groups were not significantly different when average adult bee mass

318 during feeding period was used as a covariate ($P=0.72$): the 500 ppb group consumed 900 ± 97 g, the 125
319 ppb group consumed 868 ± 154 g, and the control group consumed 871 ± 94 g. Total consumption was
320 related to colony size: the average adult bee mass of the colonies that consumed all the patty was $1.61 \pm$
321 0.28 kg while that for the colonies that did not was 0.74 ± 0.08 kg, and among colonies that did not finish
322 the patty, consumption was directly proportional to adult bee mass ($F_{1,10}=13.21$, $P=0.0046$, adj. $r^2=0.53$).

323

324 **3. NEB weights and nurse bee head weights**

325 Dry weights of NEBs were not significantly different among treatments in the Fall 2017
326 experiment ($P=0.45$). Methoxyfenozide application did not influence head weight of bees collected in the
327 Fall 2017 experiment. Average head weights (\pm s.e.) for bees in hives treated with the 500 ppb
328 (12.51 ± 0.37 mg), 125 ppb (12.29 ± 0.31 mg) and control (12.32 ± 0.19 mg) treatments did not differ
329 significantly ($P=1.0$).

330

331 **4. HPG size of caged bees**

332 Oral exposure to methoxyfenozide during young adult development did not impact the
333 hypopharyngeal gland sizes of nurse-aged workers ($P=0.31$). The average (\pm s.e.) acinus size of bees
334 exposed to 1000 ppb methoxyfenozide in syrup was 0.021 ± 0.007 mm², while those fed the control
335 treatment had glands that were 0.025 ± 0.007 mm².

336

337 **5. Pesticide analyses**

338 Wax samples taken during the Fall 2016 experiment were analyzed for methoxyfenozide residues
339 and none were found (Limit of Detection [LOD]=1 ppb) with the exception of trace amounts detected in
340 the 500 ppb treatment in early November, just after the end of the treatment period. For the Fall 2017
341 experiment, bee bread samples were analyzed rather than wax. In the initial sample, analyzed with respect
342 to a full panel of 192 compounds, only trace amounts of diphenylamine (LOD=2 ppb) and 118 ppb of

343 thymol were detected. No methoxyfenozide was detected in the bee bread except for samples collected
344 from hives in the 500 ppb treatment in the November, February and April hive assessments. Those
345 samples contained 5, 2 and 2 ppb methoxyfenozide, respectively. The protein patty samples for the 500
346 ppb, 125 ppb and control treatments were found to have 199, 103 and 0 ppb methoxyfenozide,
347 respectively.

348

349 **6. Hive weight for Fall 2016 and Fall 2017 experiments**

350 Piecewise regression curves fit the data well on average: average (\pm s.e.) adj. r^2 values for the 125
351 ppb, 500 ppb and Control groups were 0.94 ± 0.03 , 0.92 ± 0.05 and 0.91 ± 0.05 , respectively, for the Fall
352 2016 experiment and 0.96 ± 0.02 , 0.93 ± 0.04 and 0.94 ± 0.03 , respectively, for the Fall 2017 experiment.
353 Dusk break point was significantly affected by treatment (Fig 2). Post hoc contrasts showed that all
354 treatment groups were significantly different from each other. Lower values in the Control group indicate
355 that the dusk break point occurred significantly earlier in the day than for either of the other treatment
356 groups, and earlier for the 500 ppb group than for the 125 ppb group. The two fall field experiments were
357 themselves significantly different with respect to dusk break point. In the Fall 2016 experiment, the
358 Control treatment group average dusk break point was about 4:37 PM, with the 500 ppb treatment group
359 21 minutes later and the 125 ppb treatment group 30 minutes later than Control. In the Fall 2017
360 experiment, the Control treatment group average dusk break point was about 4:59 PM, with the 500 ppb
361 treatment group 17 minutes later and the 125 treatment group 46 minutes later than Control.

362

363 **Figure 2.** Break points associated with the end of the daily active period (dusk) for piecewise regression
364 curves on within-day hive weight changes for the Fall 2017 experiment. Dashed black vertical line shows
365 the end of the treatment period. Solid black horizontal line shows calculated sunset time.

366

367 Slopes of the segments associated with forager departure after the dawn break point were
368 significantly affected by treatment (Fig 3, S1 Table). Post hoc contrasts showed that slopes in the 500 ppb
369 treatment were significantly shallower (lower in absolute value) than those in the Control treatment
370 ($P=0.0043$), indicating a lower rate of forager departure. Neither the night segment slopes nor the dawn
371 break point were significantly affected by treatment ($P=0.51$ and 0.34 , respectively).

372

373 **Figure 3.** Segment slopes associated with departing foragers for the Fall 2017 experiment. Dashed black
374 vertical line shows the end of the treatment period. Solid gray horizon line shows slope=0.

375

376 Average piecewise regression curves were calculated for each treatment group by averaging slope
377 and break point values across all hives and sample days (Fig 4). Because the post-treatment data were
378 collected in the late fall and winter, with few foraging opportunities, all hives lost weight as the bees
379 consumed food stores. Average slopes at night tended to be positive, probably due to higher ambient
380 relative humidity (the woodenware of the hives can gain weight, as well as any open food cells within the
381 hive). In both years, average curves for the 500 ppb treatment group were shallower than the control,
382 probably indicating a lower foraging effort.

383

384 **Figure 4.** Piecewise regression curves calculated using average slope and break point values calculated
385 from all colonies within each treatment group and across all days from the end of treatment until the final
386 hive assessment. A) Fall 2016 experiment; B) Fall 2017 experiment. Graphs have the same scales.

387

388

389 **Table 2.** Results of post hoc comparisons for three response variables with significant treatment effects
390 for three field experiments on the effects of sublethal methoxyfenozide exposure to honey bee colonies
391 conducted in southern Arizona.

392

Year	Concentration	Response variables		
		Dusk break point	Departing slopes	Temperature amplitude
Fall 2016	500 ppb	a	a	a
&	125 ppb	b	ab	ab
Fall 2017	0 ppb	c	b	b
Spring 2018	500 ppb	-	a	-
	125 ppb	-	ab	-
	0 ppb	-	b	-

393 Treatment groups within the same time period with no letters in common are significantly different at
394 $\alpha=0.05$ with a Bonferroni comparison for multiple groups. Bold indicates which group or groups had the
395 highest value within each time period and response variable group. Dashes indicate no significant main
396 effect.

397

398 **7. Hive weight for Spring 2018 experiment**

399 Hive weight consistently decreased during the fall experiments, but average hive weight in all
400 treatment groups increased every day from the end of treatment on 20 April until 6 May. On 7 May
401 average hive in each treatment group started to decrease, indicating the end of a nectar flow, and most

402 hives lost weight after that day until the end of the experiment on 23 May. Within-day hive weight
403 patterns differ depending on whether there is a nectar flow [25] so those two periods (20 April-6 May and
404 7-23 May) were considered separately. During the nectar flow, only the departing slopes were
405 significantly different among treatments (S2 Table, Table 2). Average slope in the 500 ppb treatment
406 group was significantly higher than the slopes for either of the other treatment groups. In this case, while
407 average $r^2 \pm s.e.$ values for model fit were high: 0.96 ± 0.03 , 0.97 ± 0.03 , and 0.91 ± 0.06 for the 125 ppb, 500
408 ppb, and Control treatment groups, respectively, visual inspection of the data showed that breakpoints
409 near dawn were not being detected by the algorithm, causing inaccuracies in other segment parameters
410 such as slope values. Increasing the number of break points from 4 to 5 did not change the overall
411 goodness of fit (average $r^2 = 0.92$, 0.97 and 0.96 , respectively) and likewise did not improve detection of
412 the break points (Fig. 5). These results, therefore, should be subject to future verification. No parameters
413 were significantly affected by treatment during the 16 d after the end of the nectar flow.

414

415 **Figure 5.** Example of an unsatisfactory curve fit. Raw data, collected on 29 April 2018 from a hive in the
416 500 ppb treatment group, shown with two piecewise regression curve fits: a curve with 4 break points and
417 a curve with 5 break points. Arrow indicates expected “dawn” (1st) break point. Note that the 2nd segment,
418 typically associated with forager departure and therefore negative, is in this case positive.

419

420 **8. Hive temperature for Fall 2016 and Fall 2017 experiments**

421 Internal hive temperature was considered with respect to two response variables: 1) average daily
422 temperature; and 2) average temperature variability measured as amplitudes of curves fit to 3-day
423 datasets. Hive temperature has been positively correlated with total adult bee mass, so pre-treatment total
424 adult bee mass was included as a covariate in all analyses. No significant treatment effects were observed
425 either with respect to average temperature or to temperature amplitudes when all post treatment data
426 (from end of treatment until final hive assessment), but the low P values (0.06 and 0.07, respectively)

427 suggested there may be trends to explore by sharpening the focus of the analysis. Given that hive
428 temperature is a function of both the bee colony and external ambient conditions, treatment effects may
429 be more likely to be observed when a colony is challenged to manage its temperature. At the beginning of
430 November in both 2016 and 2017 average ambient daily temperatures at the study site were about 21.7 to
431 22.8°C. Thirty days later, however, average ambient daily temperatures had dropped to 9.4-12.8°C. Thus,
432 ambient temperatures 30 d after the end of treatment were considered more challenging to the bee
433 colonies and thus more likely to show an effect. Considering only the data from 30 d after the end of
434 treatment until the final hive assessment, treatment effects on average temperature remained not
435 significant ($P=0.06$) but significant treatment effects were observed with respect to temperature
436 amplitudes (variability) (S3 Table, Table 2, Fig 6). Temperature amplitudes were lower in the Control
437 treatment group (about 1.72°C lower in the Fall 2016 experiment and 2.40°C in the Fall 2017 experiment)
438 than the 500 ppb treatment group (about 2.55°C in the Fall 2016 experiment and 3.64°C in the Fall 2017
439 experiment). Neither the Control group nor the 500 ppb group was different from the 125 ppb group in
440 either experiment.

441

442 **Figure 6.** Internal hive temperatures. A) Average temperature for the Fall 2017 experiment; B) Average
443 amplitudes of sine curves fit to internal hive temperature data (temperature variability). Gray zones in A)
444 show the periods during which experimental treatments were applied.

445

446 **9. Hive temperature data for Spring 2018 experiment**

447 Neither average internal hive temperatures nor their daily variability were different among
448 treatments ($P=0.22$ and 0.32 , respectively) when the February adult bee mass was used as a covariate.
449 February adult bee mass was significantly correlated with both temperature variables.

450

451 **Discussion**

452 Methoxyfenozide is preferred as a pesticide for many crops pollinated by honey bees. One
453 important question regarding sublethal exposure of methoxyfenozide is how the treatment effects
454 manifest themselves, such as by reducing brood production or adult bee survivorship, changing colony
455 behavior, or affecting the growth or physiology of individual bees. Methoxyfenozide, with a reported
456 acute oral toxicity of more than 5.0 g per kg for humans, is considered mildly toxic for many vertebrates
457 and crustaceans; with respect to honey bees, it has been reported as “moderately toxic” [9] and “relatively
458 nontoxic” [32]. Most studies on sublethal pesticide exposure rely on invasive hive assessments to estimate
459 adult bee and brood populations. To avoid antagonizing the bees or losing the queen, these inspections are
460 typically carried out about every 4-6 weeks, as was done here. No effects of methoxyfenozide were
461 observed with respect to hive assessment data (total adult bee mass or brood surface area) in this study.
462 These data are important, but they provide little information on daily or hourly changes in colony-level
463 behavior such as foraging and thermoregulation. In this study we monitored colony behavior using hive
464 scales, to observe hourly weight changes associated with, for example, foraging activity [26], and using
465 temperature sensors to measure thermoregulation. Both types of sensors have been successfully used to
466 detect the effects of sublethal exposure of honey bees to a neonicotinoid, imidacloprid [24].

467 Hive weight data, after being detrended by removing the value at midnight from subsequent
468 values for the next 24 h, have common patterns [26]. For example, from midnight until the start of the
469 active period, usually just after dawn, bee flight activity is minimal and hive weight changes largely
470 involve water gain or loss depending on the amount of open nectar and ambient relative humidity. After
471 dawn, weight changes usually become sharply negative, as the daily active period for the colony begins
472 and foragers and other bees leave the hive. Whether hives gain or lose weight during the day depends
473 largely on factors such as the success of the foraging bees. In the late afternoon, hives tend to gain weight
474 as flying bees return to the colony; their return is usually complete about dusk, marking the end of the
475 active period. After that point, hive weight changes are once again largely due to the amounts of
476 hydrophilic materials present and to internal and ambient humidity. Hive scale data revealed treatment
477 effects with respect to the slope of the segment associated with forager departure, and the dusk break

478 point. Because these parameters resulted from the fit of single piecewise regression curves, they should
479 not be considered entirely independent from each other but rather reflect fundamental differences in
480 overall daily curve shape. Average curve shapes differed from year to year, showing the effects of yearly
481 variability in temperature and rainfall, but clear differences in curve patterns were also evident within
482 each year. Colonies fed 500 ppb methoxyfenozide in the fall in general had shallower curves, indicating
483 lower activity than colonies in the Control group, and the 500 ppb colonies also had delays in the dusk
484 break point of 17-21 minutes on average, consistent between the two fall experiments. Analyses of break
485 points other than dawn, or of slopes other than the initial forager slope, were not included here because
486 the meaning of any differences was not clear. Although the curves usually have similar shapes among
487 colonies and over time, not all parts of the curves have a clear interpretation. This is particularly true
488 during a forage dearth, when forage-related environmental signals are absent.

489 The Fall 2017 field experiment was continued through the following spring, in order to observe
490 longer-term effects of methoxyfenozide exposure. The spring environment differed from that of the fall in
491 two crucial respects: 1) rising temperatures, longer days and increasingly available forage promote colony
492 growth rather than stasis or decrease as observed in the fall; and 2) a nectar flow was under way during
493 the post-treatment period in the spring but not in either fall experiment. That the colonies grew rapidly in
494 the spring in spite of reduced pollen patty consumption suggests that alternative food sources played a
495 large role. Interestingly, few significant colony-level effects were observed in the spring. Three colonies
496 died after the first hive assessment in the spring, one in the 500 ppb treatment and the other two in the 125
497 ppb treatment. As in the fall, hive assessment data were not significantly different among groups.
498 Continuous weight data segment slopes associated with departing bees were significantly affected, with
499 the 500 ppb hives showing higher slope values (indicating fewer departing bees) than other treatment
500 groups, but this result did not conform with visual inspection of the data, reducing our confidence in that
501 result. Regression models with 5 breaks, rather than 4, were fit to the data but the quality of the fit was not
502 improved and merit further analysis. While the expectation was that hives may be more sensitive to IGR

503 in the spring, IGR effects may have been diluted by the pollen and nectar flows (see [23]) or the bees,
504 either on the individual or colony level, were more effective at detoxification.

505 Hive internal temperature has been correlated with adult bee mass and brood levels [23]. While
506 adult bee and brood levels were not significantly affected by treatment, internal hive temperature was
507 significantly more variable in the 500 ppb treatment group when November data were excluded (mild
508 temperatures in November would have resulted in lower variability values in any case; removing those
509 data put the focus on December through the final hive assessment in February). It may be that the
510 temperature data were more sensitive to brood levels than the inspection data, if only because of the larger
511 amount of temperature data, and as a consequence the temperature data revealed smaller differences.

512 Methoxyfenozide concentration was measured in wax samples during the Fall 2016 experiment,
513 and none was detected in any sample. Wax was sampled to determine if the lipophilic nature of the
514 compound facilitated its spread throughout the hive and that none was detected suggests that such
515 spreading occurs at very small quantities, if at all. For the Fall 2017 and Spring 2018 experiments, bee
516 bread was sampled. Low concentrations of methoxyfenozide were detected in bee bread from the 500 ppb
517 treatment group and none in the other groups. The concentrations were apparently stable over time, from
518 the end of the fall treatment in November until the end of the spring treatment the following April.
519 Treatment patties were also sampled for methoxyfenozide concentration. Observed concentrations in the
520 treatment patties were 17.6% lower than calculated concentrations in the 125 ppb treatment and 60.2%
521 lower than calculated concentrations in the 500 ppb treatment. Disparities at about that magnitude
522 between calculated and observed concentrations of pesticides mixed in pollen patties have been reported
523 elsewhere [34]. Such disparities may be due to several factors, including insufficient mixing in a
524 heterogeneous material (pollen patty), or a breakdown of the compound due to chemical reactions or
525 biological activity in the patty environment. Using the observed concentration values, the bee bread
526 results suggest that 1-2.5% of the bee bread sampled was treatment patty.

527 While significant effects of methoxyfenozide exposure on colony behavior were observed, no
528 differences were detected with respect to hive assessment data (adult bee mass and brood surface area) or

529 on the level of the individual bee (gland size, NEB dry weight and head mass of nurse bees). The kinds of
530 colony-level behaviors that were measured, i.e. foraging activity and thermoregulation, can be considered
531 functions of either colony size and age structure, or individual adult bee behavior, or both. Greater
532 foraging activity by one colony compared to another may result from a larger adult bee population with a
533 similar proportion of foragers or from a similar adult bee population with a higher proportion of foragers.
534 Reduced variability in internal hive temperatures may result from the better insulation properties of a
535 tighter bee cluster, from more heat, on average, per bee, or simply from more bees. Although brood levels
536 were not significantly different, the 500 ppb treatment group ranked last for each post-treatment hive
537 assessment in both fall experiments, suggesting that thermoregulation differences were likely (but not
538 definitively) linked to brood levels. The significance of the observed delay in the end of foraging activity
539 (“dusk”) is not clear, but longer-term effects of sublethal pesticide exposure are not always well
540 understood. Six of the 36 colonies involved in this study died and none was from the control group.
541 Further work is needed to link observed changes in colony behavior to longer-term effects on colony
542 performance and survivorship.

543

544 **Conclusions**

- 545 • Exposure of honey bee colonies to methoxyfenozide in supplement patty at a concentration of
546 500 ppb in the fall reduced the colony foraging population, delayed the end of the daily activity
547 period by 17-21 minutes, and was associated with higher internal hive temperature variability
548 (poorer thermoregulation).
- 549 • Exposure to the colonies did not have a measurable affect on the total adult bee mass, the amount
550 of brood, average newly-emerged bee body mass or head weight, and caged bees fed 1000 ppb
551 methoxyfenozide in sugar syrup showed no differences in hypopharyngeal gland size.
- 552 • Colonies treated for a 2nd consecutive time the following spring showed fewer differences than in
553 the fall.

- 554 • Continuous weight and temperature monitoring methods showed significant effects on colony-
- 555 level behavior whereas periodic colony assessments and sampling did not show effects.
- 556
- 557

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563

564

565

566 **S1 Table.** Effects of treatment on dawn break point for piecewise regressions fit to continuous weight
567 data for the Fall 2016 and Fall 2017 field experiments (PDF).

568

569 **S2 Table.** Effects of methoxyfenozide exposure on departing slopes for piecewise regressions fit to
570 continuous weight data for the Spring 2018 field experiment (PDF).

571

572 **S3 Table.** The effects of methoxyfenozide exposure on log amplitudes of sine waves fit to detrended
573 continuous temperature data (daily internal hive temperature variation) for the Fall 2016 and Fall 2017
574 field experiments 30 d after the end of treatment until the final assessment (PDF).

575

576 **S1 File.** Experimental data (XLSX).

577

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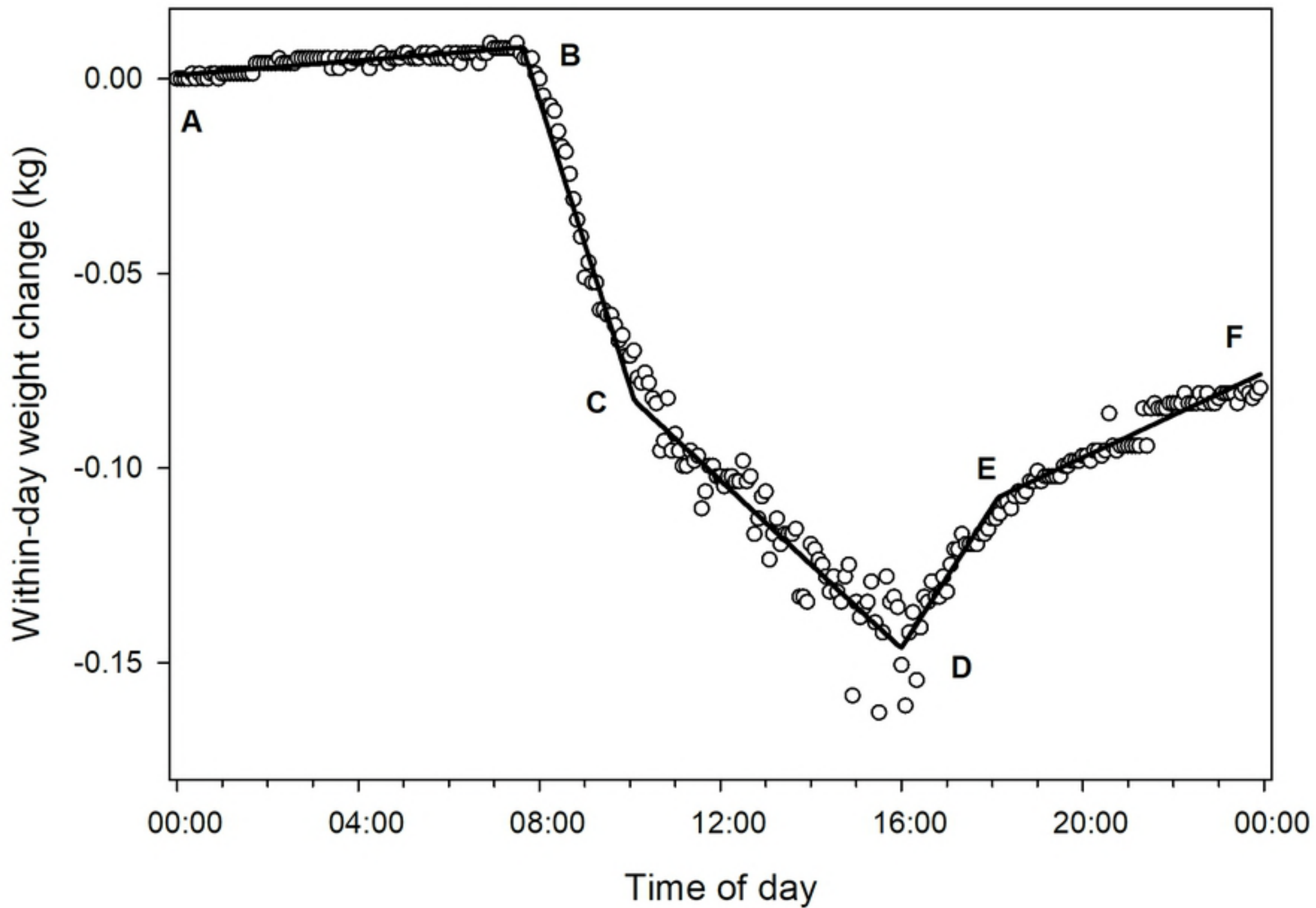
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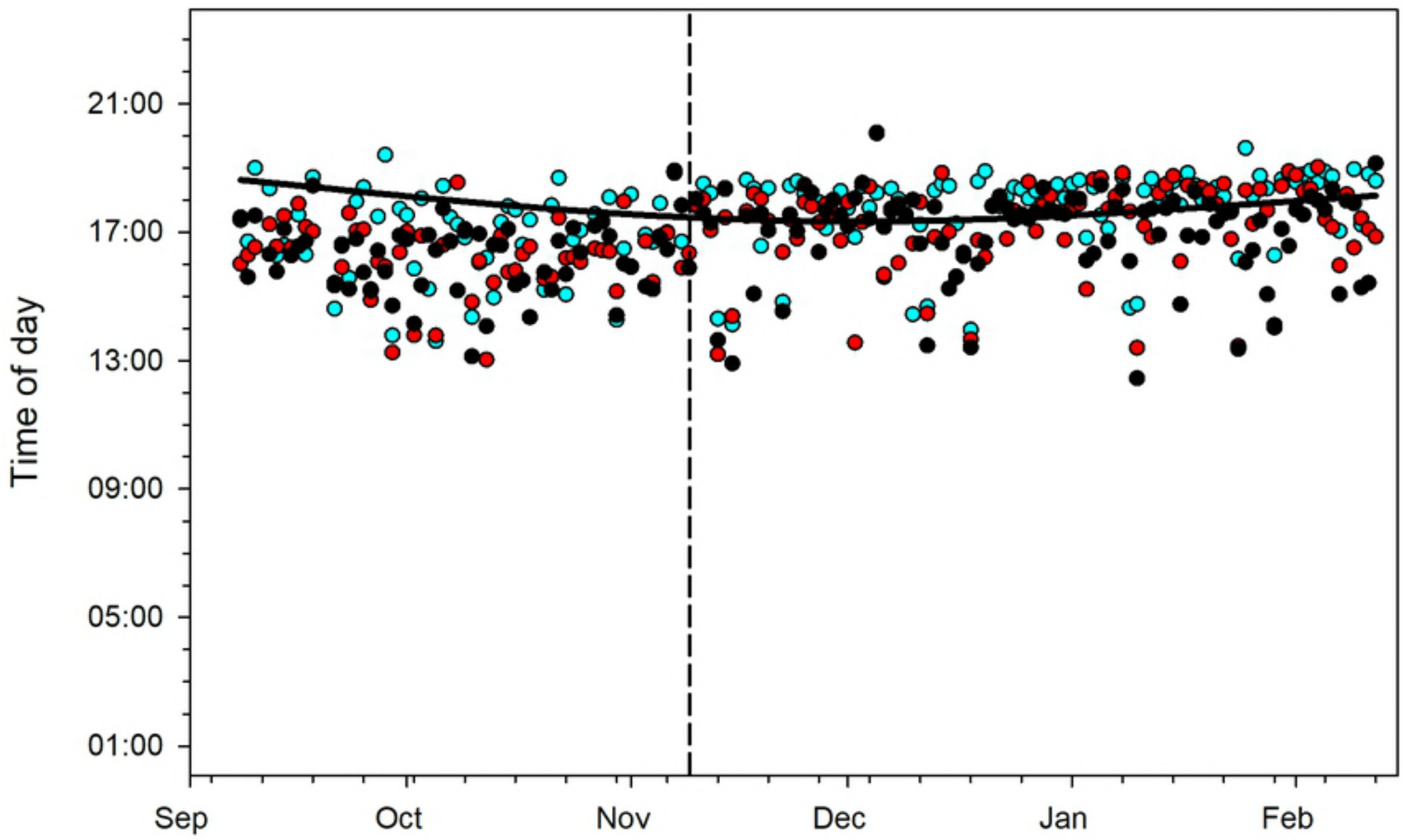
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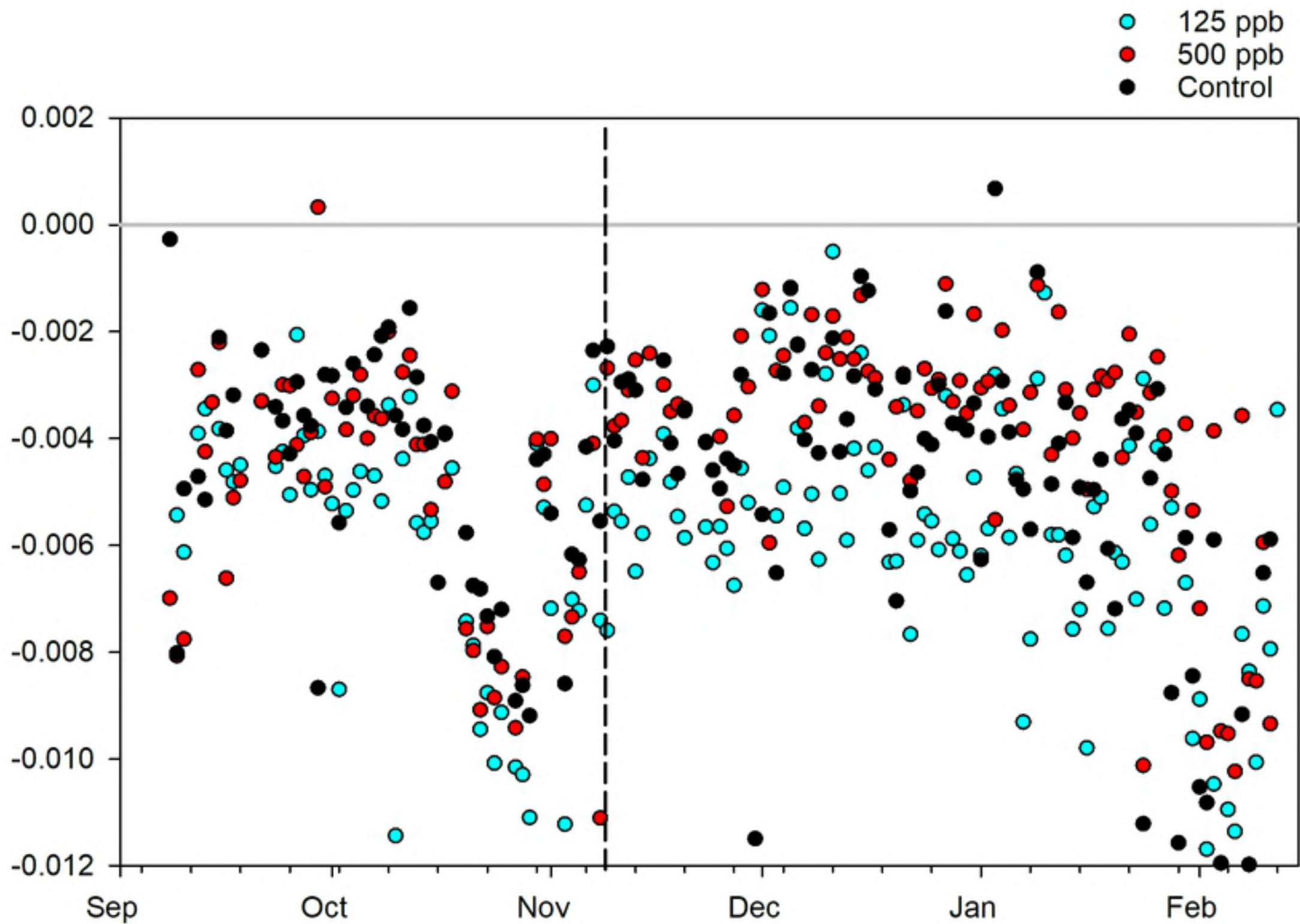
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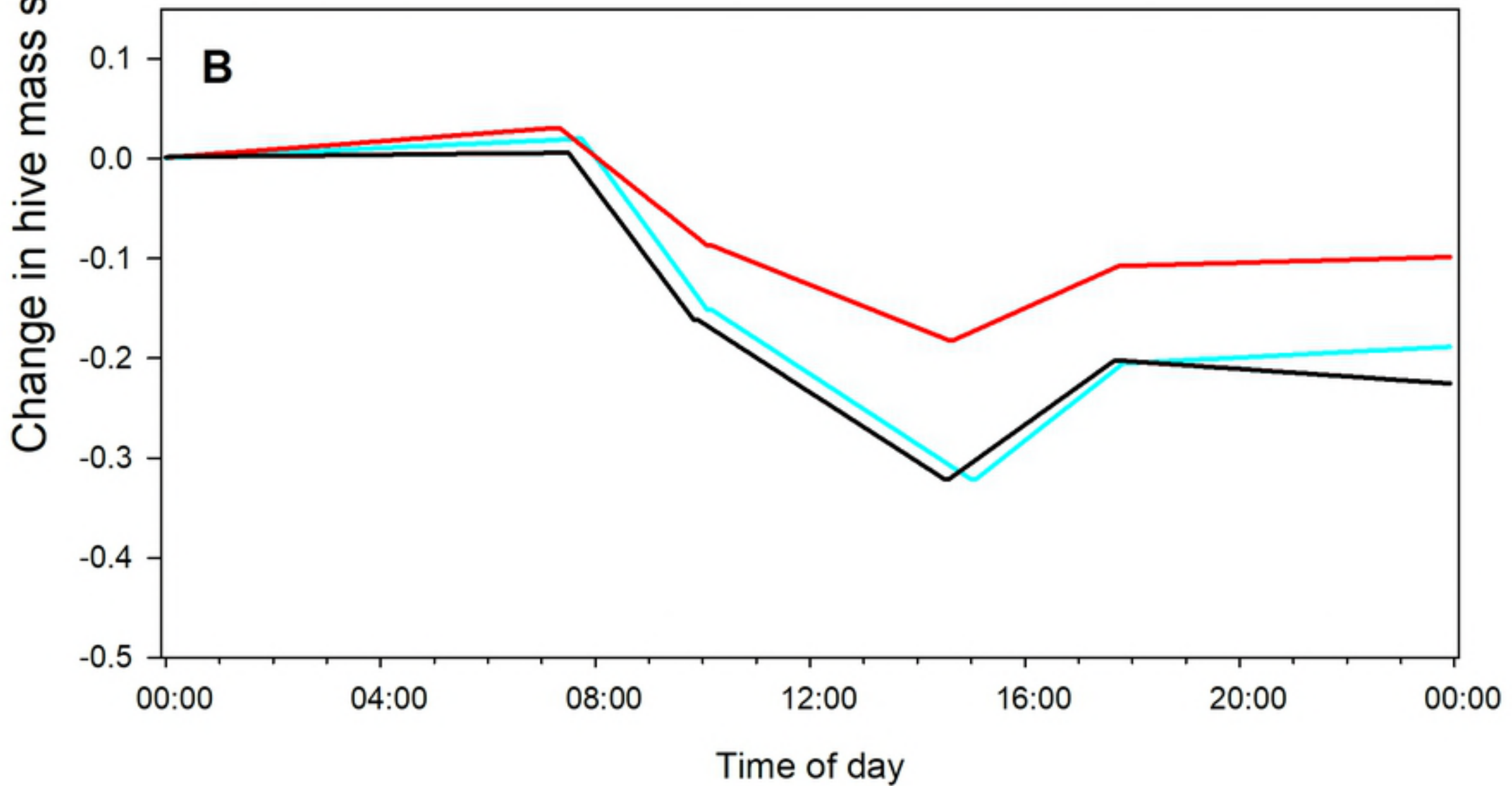
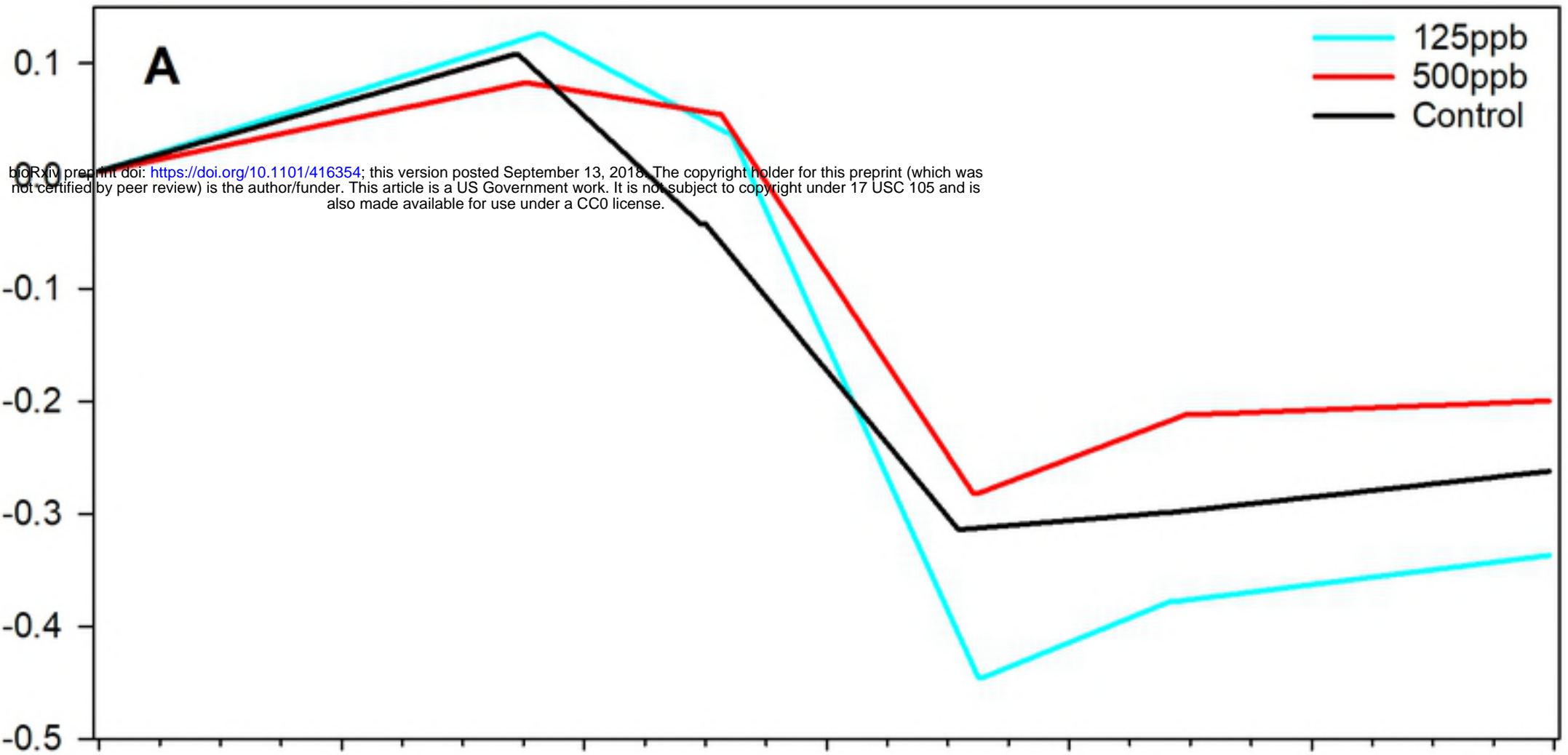


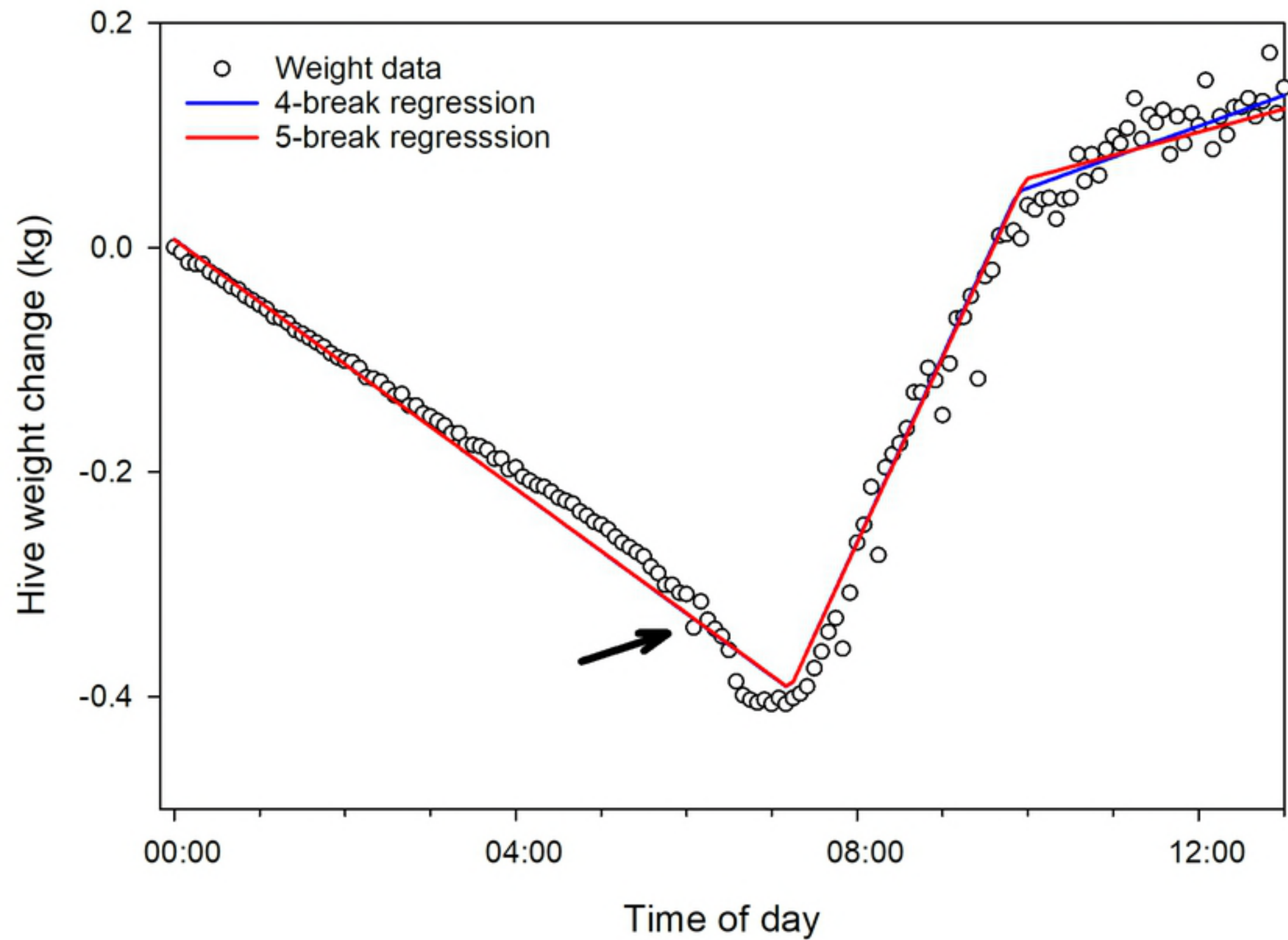
- 125 ppb
- 500 ppb
- Control



Slope of forager departure segment (g per 5 min)







Control
High concentration (500 ppb)
Low concentration (125 ppb)

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