Interspecific plant competition mediates the metabolic and ecological signature of a plant-herbivore interaction under warming and elevated CO2

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Abstract

1. Biotic interactions shape community evolution, but we lack mechanistic insights on how metabolic and ecological processes under climate change are altered by biotic interactions.

2. We used a two-trophic model community consisting of the aphid *Dysaphis plantaginea* feeding on the forb *Plantago lanceolata,* and a grass competitor *Lolium perenne* that does not experience herbivory by the aphid. Monocultures and mixtures were exposed to the herbivory treatment and to three relevant simulated environmental changes as prevalent under current climate change (increased temperature, CO_2, and increased temperature and CO_2)

3. Elevated CO_2 reduced the nitrogen content of *P. lanceolata*, while simultaneous increases of CO_2 and temperature modified the plant metabolic component and the magnitude of these responses in different directions. Elevated CO_2 enhanced defence systems in *P. lanceolata*, but these effects were not altered by warming. Interspecific plant competition did, however, neutralise these responses. There were no indications for indirect effects of climate on aphid population growth as mediated by changes in plant defence, nutritional quality or biomass.

4. We thus demonstrate interactions between abiotic and biotic processes on plant metabolite profiles, but more importantly, that climate change effect on a selection of the metabolic pathways are altered by herbivory and competition. Our experiment under semi-natural conditions thus demonstrates the non-additive and often neutralizing effects of biotic interactions on plant metabolism and species performance under climate-associated environmental change.

Introduction

Interactions between insect herbivores and host plants are influenced by primary and secondary metabolites. The concentrations and types of these molecules in plants is likely affected by climate change (Bidart-Bouzat & Imeh-Nathaniel, 2008; Zvereva & Kozlov, 2006). For example, plants exposed to elevated CO_2 show higher concentrations of many organic molecules, including carbohydrates (fertilization effect), but generally lower or unaltered nitrogen levels, and consequently higher C:N ratios (Bezemer & Jones, 1998; Lincoln *et al.*, 1993; Robinson *et al.*, 2012; Stiling & Cornelissen, 2007). Lower nitrogen concentrations imply lower levels of leaf protein and amino acids and reduced nutritive value for herbivores (Lincoln *et al.*, 1986). Consequently, the performance of insects generally reduces under high CO_2 (Bezemer & Jones, 1998; Hunter, 2001; Lincoln *et al.*, 1993). Contrary to elevated CO_2, moderate warming has positive effects on insect herbivore performance by reducing development time (Bale *et al.*, 2002; Van De Velde *et al.*, 2016) and increasing fecundity (Meisner *et al.*, 2014). Climate change may thus paradoxically constrain rather than facilitate insect herbivore population growth by lowering plant quality (Bauerfeind & Fischer, 2013; Jamieson *et al.*, 2015). These resource-quality mediated effects are, however, anticipated to not fully counterbalance the direct effects of warming on development, rendering the overall herbivore response to warming positive (Bauerfeind & Fischer, 2013; Zvereva & Kozlov, 2006).

Secondary metabolites (e.g. lignin, tannins, phenolics and terpenoids), greatly affect tissue quality by determining the nutritive value, palatability, digestibility and/or toxicity of foliage (see Table 1 for a synthesis). Compared to primary metabolites, responses of secondary metabolites to elevated CO_2 and warming are more variable and less understood (Bidart-Bouzat & Imeh-Nathaniel, 2008; Robinson *et al.*, 2012). According to the carbon-nutrient

balance hypothesis (Bryant *et al.*, 1983), elevated CO_2 is expected to increase secondary

metabolite levels as a result of the 'excess' carbon (fertilizing). Yet, because of the relatively

lower N-availability, the increase of nitrogen-containing metabolites will be lower than

expected (Robinson *et al.*, 2012). On the other hand, trade-offs between growth and

secondary metabolisms are anticipated to promote primary metabolic activity under

warming and therefore favour photosynthesis and growth relative to defence. This

expectation corresponds to the growth-differentiation balance hypothesis (Herms & Mattson,

1992). The different components associated with climate change thus induce direct, and often

orthogonal responses in chemical defences (Bidart-Bouzat & Imeh-Nathaniel, 2008; Zavala *et*

al., 2013; Zvereva & Kozlov, 2006).

Biotic interactions have a strong impact on life histories and physiology of species, and these

interactions may shift responses of plants towards climate change (Van de Velde et al. 2016).

Competition for limited nutrients for example, enhances the availability of carbon in a focal

plant relative to its demand. This, in turn, can increase carbon-based defences and thus reduce

herbivory compared with plants that are less constrained by competitors (Bryant *et al.*, 1983).

Also the identity of a neighbouring plant may affect the resistance of a focal plant by a

different outcome of the plant-plant competition (Barton & Bowers, 2006; Broz *et al.*, 2010).

Conspecific competition, in particular, may result in stronger decline of plant growth and

increased defence compared to heterospecific competition. Indeed, Broz *et al.* (2010) for

instance showed that focal plants with conspecific neighbours allocated more resources

towards production of carbon-based defence molecules, whereas those grown with

heterospecific neighbours allocated more resources towards growth. Hence, elevated CO_2 as

well as warming is anticipated to indirectly affect the defence of a focal plant, through species

specific stimulation of plant growth and thus plant-plant competition.

Although atmospheric CO_2 and temperature increase concurrently, empirical studies

investigating their combined effect on multi-trophic communities are surprisingly scarce and

insufficient to effectively guide theory or synthesis (Cornelissen, 2011; Dyer er al. 2013;

Zvereva & Kozlov, 2006). The current study therefore investigates effects of elevated CO_2 as

well as combined effects of elevated CO_2 and warming on a simple model community

consisting of three species: rosy apple aphid *Dysaphis plantaginea* Passerini (*Hemiptera:*

Aphididae) feeding on plantain, *Plantago lanceolata* L*.,* and a heterospecific neighboring

plant species, perennial ryegrass, *Lolium perenne* L. The aphid does not feed on *L. perenne.*

The experimental design consisted of monocultures of *P. lanceolata* and mixtures of *L.*

perenne and *P. lanceolata* exposed to the two aforementioned simulated climate scenarios as

well as to a control scenario with the current climate. As the experiments were conducted

outdoors, we were also able to *post-hoc* study how the imposed climatic and competition

treatments affected mildew infections. Biotrophic parasites such as mildew extract nutrients

from living cells and have extended periods of physiological interaction with their hosts

(Agrios, 2005). Elevated CO_2 and warming have been shown to affect the severity of

pathogen infections (Mikkelsen *et al.*, 2015, Thomas & Blanford, 2003). We took advantage

of this unplanned infestation to study the impact of elevated CO_2 and warming on mildew

infestation and the putative additive effects of mildew infestation on plant-insect herbivore

interaction.

Based on the above, we anticipate that elevated CO_2 and warming will alter foliar nutrients

and defence molecules, that altered host quality and plant resistance will affect insect

herbivore performance, and that elevated CO_2 and warming should indirectly influence host

quality and plant resistance via effects on neighbouring plants. We predict investments in

defence to be higher under conspecific competition relative to heterospecific competition

because of increased resource competition. In addition, increased CO_2 is anticipated to

increase chemical defence by promoting the production of secondary metabolites, while

simultaneously increased temperature would in contrast promote plant growth. Depending on

the relative rates of these physiological changes, net impact on herbivore populations may

either remain constant, decline or increase if defence is respectively proportionally promoted

or constrained relative to metabolism.

We do, however, aim to move beyond this generic perspective and to acquire a deep

mechanistic understanding of the changing ecological interactions under elevated CO_2 and

warming. To this end, we selected metabolites and groups of metabolites as molecular

markers to evaluate defences and stress levels (see Table 1). In particular, we quantified an

indicator of oxidative cellular damage (lipid peroxidation). Oxidative stress in plants is

commonly induced under abiotic stress, but also by herbivory (Wu & Baldwin, 2010).

Elevated CO_2 on the other hand, often reduces abiotic stress impact (Zinta et al. 2018).

Herbivory damage also induces the production of secondary metabolites associated with the

salicylic acid and jasmonic acid pathway (Kuśnierczyk *et al.*, 2011; Morkunas & Gabryś,

2011). Therefore, as an indicator of plant-insect interaction, jasmonic acid levels were

determined. Simultaneously with the stress status, we evaluated the plant capacity to defend

itself by quantifying protective metabolites. Here we determined molecular antioxidants

(tocopherols, antioxidant capacity, carotenoids, proline) providing protection against

oxidative stress (Gómez-Ariza *et al.*, 2007, Peshev *et al.*, 2013). We also quantified several

groups of molecules that contribute to defence against insects (glycosides, tannins, lignin).

Finally, elevated CO_2 as well as elevated temperature are known to affect stomatal aperture

and therefore water transport. To evaluate the plant response to its water status we quantified

osmolytes (proline, soluble sugars). Increase in these molecules reduces the cellular water

potential, improving water uptake. Given the anticipated complex interplay between elevated

CO_2, temperature and competition, we performed a multivariate grouping analysis to

document concerted responses in the quantified metabolites.

Table 1: Overview of molecular markers determined in this study and their principal function.

METABOLITE	PRINCIPAL FUNCTION	FUNCTION DESCRIPTION	REF.
TOCOPHEROL (VITAMINE E)	Lipophilic antioxidant	Components of biological membranes; act as scavengers of oxygen radicals generated by stress, in particular singlet oxygen. Tocopherols also stabilize membrane structures.	Blokhina et al (2003)
CAROTENOIDS	Lipophilic antioxidant	Lipophilic antioxidants localized in plastids of photosynthetic and non-photosynthetic plant tissues. Carotenoids react with lipid peroxidation products to end the chain reactions and dissipate the excess excitation energy.	Bartley and Scolnlk (1995)
PHENOLICS/ FLAVONOIDS	Antioxidant	Polyphenols have free radical scavenging activity and are involved in hydrogen peroxide neutralization. Flavonoids are also secondary ROS scavenging molecules in plants under excess excitation.	Blokhina et al (2003)
TOTAL ANTIOXIDANT CAPACITY (TAC)		Total antioxidant capacity (TAC) is the measure of the gross capacity of a plant extract to scavenge reactive oxygen species.	Das and Roychoudhury (2014)
MALONDIALDE HYDE, (MDA)	Lipid peroxidation marker	Oxidative stress induces lipid peroxidation disrupting membrane functions. Lipid peroxidation products (MDA) indicate membrane oxidation levels.	Das and Roychoudhury (2014)
SOLUBLE SUGARS (GLUCOSE, SUCROSE, FRUCTOSE)	Osmolytes	Soluble sugars protect plant cells through osmoregulation, improving water uptake and retention, and may impact cellular membranes through interaction with the lipid bilayer.	Shao et al., (2006)
PROLINE	Osmolytes/ antioxidant	Proline is an osmolyte, contributing to water uptake, as well as an antioxidant.	Szabados and Savoure (2010).
GLYCOSIDES: CATAPOL, AUCUBIN	Biotic stress defence	Secondary metabolites that help repelling herbivores, protects plant from UV and contributes to control of competitive interactions between plants.	Bowers (1991).
TANNINS	Biotic stress defence	Tannins are astringent bitter polyphenols and act as feeding deterrents to many insect pests.	Barbehenn et al., (2011)
LIGNIN	Biotic stress defence	A complex phenolic polymer that plant cell wall rigidity and hydrophobic properties, and is an important barrier against pests and pathogens.	Liu et al., (2018)
SALICYLIC ACID	Biotic stress	Salicylic acid is a phytohormone involved in regulation of	War et al.,

Materials and methods

Experimental set-up (Fig S1.1 in supplementary material)

The study was performed at the Drie Eiken Campus, University of Antwerp, Wilrijk, Belgium

(51° 09' N, 04° 24'E) in 12 sunlit, south-facing, climate-controlled chambers (details in

Naudts *et al.* 2014). Three climate scenarios (four chambers per scenario) were simulated in

an additive design: (1) current atmospheric CO_2 concentration and temperature (C); (2) future

atmospheric CO_2 and current temperature (CO_2); and (3) future atmospheric CO_2 and

temperature (TCO_2). Climate scenarios with elevated CO_2 had a target CO_2 concentration of

620 µmol mol^{-1} and future temperature chambers simulated a continuous 3 °C warming above

fluctuating ambient temperatures. Climate manipulations were based on the IPCC-SRES B2-

scenario prediction of moderate change for the year 2100 (IPCC, 2001).

The CO_2 concentration in each chamber was continuously measured and maintained at the

target concentration with a CO_2 control group with an infrared analyser (WMA-4, PPSystems,

Hitchin, UK). Air temperature and relative humidity were monitored every 0.5 h with a

combined humidity–temperature sensor (Siemens QFA66, Erlangen, Germany), by averaging

instantaneous readings in half-hour mean values. During the experiment, the CO_2

concentration was 382 ± 55 µmol mol^{-1} (SD) in the current climate, while it was 615 ± 70

µmol mol^{-1} (SD) in the climate scenarios with future CO_2 concentration (CO_2 and TCO_2). The

monthly average air temperature in the C and CO_2 chambers was 16.2, 17.2 and 18.7 °C in

June, July and August, respectively. TCO_2 chambers were 2.9 ± 1.0 °C (SD) warmer than

current temperature chambers. Average vapour pressure deficit was 0.60 ± 0.34 and $0.64 \pm$

0.52 kPa (SD) in the climate treatments with ambient and warmed air, respectively. Irrigation

mimicked the rainfall outside (Naudts et al. 2014), with monthly totals equalling 64.4, 85.1

and 80.2 mm in June, July and August, respectively. Water freely drained while capillary rise

was prevented by a drainage system placed below the chambers. Future climate chambers

(TCO$_2$) received the same amount of water as current climate chambers, so that any increase

in water consumption would result in (aggravated) soil drought.

Plant and insect communities

We used two common co-occurring grassland species, *L. perenne* and *P. lanceolata*. *P.*

lanceolata is characterized by the presence of the iridoid glycosides aucubin and catalpol.

These compounds stimulate feeding and oviposition by specialist insects and can act as

defence against generalist herbivores (Bowers & Puttick, 1988; Puttick & Bowers, 1988).

Both plant species were sown at the end of March in a non-climate controlled greenhouse

with a time lag of one week to prevent size differences at the start of the experiment (Cotrufo

& Gorissen, 1997), and were watered twice a week. Four or five-week-old seedlings were

transplanted into PVC containers (24 cm inner diameter and 40 cm height), filled with sandy

soil (93.9% sand, 4.1% silt, 2.0% clay; pH 7.5; Kjeldahl-N 0.125 g kg^{-1}; 2.1% C in humus).

Each of the 12 chambers received 20 containers with two different compositions: (1) 10

monocultures of *P. lanceolata*, and (2) 10 mixtures of both plant species in a 50:50 ratio.

Each community contained 18 individuals planted in a hexagonal grid with 4.5 cm interspace.

Interspecific interactions were maximized by avoiding clumping. All communities were

fertilized with 10 g m^{-2} NH$_4$NO$_3$, 5 g m^{-2} P$_2$O$_5$, 10 g m^{-2} K$_2$O and micro-elements (Fe, Mn,

Zn, Cu, B, Mo), given dissolved in water in two equal amounts.

When the plants were three months old, *P. lanceolata* was involuntarily infested with

powdery mildew *Podosphaera plantaginis*. *P. plantaginis* is a biotrophic fungal pathogen

which means that it feeds on living plant tissue, but does not kill the infected host. The

powdery mildew was found in all climate scenarios and plant compositions but not all the

containers were equally affected. We took advantage of this unplanned infestation to study the

impact of elevated CO_2 and warming on mildew infestation and the putative additive effects

of mildew infestation on plant-insect herbivore interaction. Because these infections occurred

post-hoc, we do not have proper controls to quantify mildew effects on metabolite

composition.

The rosy apple aphid *D. plantaginea* was used as an insect herbivore. It overwinters as eggs

on apple trees, the primary host plant, and migrates in spring to the obligate alternate hosts, *P.*

major L. and *P. lanceolata* (Alford, 2014). On *Plantago* spp., they give birth to apterous

(wingless) morphs that reproduce by parthenogenesis (Blommers *et al.*, 2004). *L. perenne* is

not a host plant for *D. plantaginea*. The aphids were reared in small cages on *P. lanceolata*

under laboratory conditions of 22 ± 1 °C. They were introduced on *P. lanceolata* when the

plants were 20 weeks old. At this time, two adult, apterous aphids were placed with a dry

paintbrush on the apex of each *P. lanceolata* plant in monocultures and mixtures.

Consequently, at the start of the infestation each container was inoculated with 36

(monocultures) or 18 (mixtures) aphids. In each chamber, four monocultures of *P. lanceolata*

and four mixtures were randomly chosen for aphid infestation All containers were meshed,

and those not receiving aphids acted as controls. The meshes consisted of an 85-cm-tall

cylinder of lightweight netting to ensure aphids did not migrate between pots. The

infrastructure did not physically limit plant growth and did not cause photosynthetic stress

effects (F_v/F_m, the intrinsic efficiency of PSII in controls equalled 0.84 which is an optimal

value (Johnson *et al.*, 1993).

Data collection

In the fourth week after the aphid introduction we determined the mildew infestation. The

degree of powdery mildew on *P. lanceolata* was categorized by a rating system: 1) healthy

(no visible lesions); 2) 1% - 25% of the leaves damaged; 3) 26% - 50% of the leaves

damaged; 4) 51% - 75% of the leaves damaged; 5) more than 75% of the leaves damaged. At

the same time, aphid populations of four replicate communities per plant composition in each

chamber were collected (totalling 4 replicate communities × 2 plant compositions × 12

chambers). Aphids were brushed directly into 70% ethanol with a dry paintbrush. It was not

possible to collect all the aphids because their number per container was either too high or too

low (hard to find), therefore a subsample of the population per container was collected: we

searched for and collected aphids for 30 minutes. The total number of aphids per community

was divided by the number of *P. lanceolata* individuals in that community. At the same time,

we also harvested the plants of two replicate communities per treatment in each chamber

(totalling 2 replicate communities × 4 treatments × 12 chambers). Alive aboveground plant

biomass in each community was separated from dead biomass by species, dried at 70 °C for

48 h and weighed. These biomass values per species were likewise divided by the number of

the species' individuals in the community, providing primary data for the statistical analysis.

The alive biomass samples of *P. lanceolata* plants of the two harvested communities per

treatment in each chamber (2 replicate communities × 4 treatments × 12 chambers) were

ground in a mill. Three subsamples per community were analysed for nitrogen and carbon

content using a NC element analyser (NC-2100 element analyser, Carlo Erba Instruments,

Milano, Italy), which were averaged prior to data analysis.

A separate subsample of the milled live aboveground biomass of *P. lanceolata* of the two

harvested communities per treatment in each chamber (2 replicate communities × 4 treatments

× 12 chambers) was taken to quantify biochemical parameters. These parameters are

indicators of stress in the organism (membrane damage as malondialdehyde levels, MDA),

indicators of antioxidant defences i.e. total antioxidant capacity (TAC) and antioxidant

molecules (tocopherols, carotenes, proline, phenols); indicators of biotic stress and defence,

i.e. glycosides (catapol, aucubin), tannin, lignin and salicylic acid; and metabolites involved in

plant water-deficit defence (proline, soluble sugars). We refer to Supplementary material 2 for

a detailed description of the methods used to quantify these metabolites.

Data analysis

We focused on *P. lanceolata* because this is the host plant for the aphid and powdery mildew.

In a first step we examined covariation in the metabolites responses to the three treatments

(aphid infestation, climate scenario and plant composition) by using a hierarchical clustering

analysis. Metabolites were normalized by Z-transformation and subjected to a hierarchical

clustering analysis with an Euclidean distance metric, and visualized as a heat map using

Multi Experiment Viewer (Mev) version 4.8 (Saeed *et al.*, 2003). A distance cut off of 0.3

was applied. The clusters were used in the structural equation models (see below).

In a second step, we fitted two piecewise Structural Equation Models (SEM), which combine

information from multiple separate linear models into a single causal network (Shipley,

2009). The first SEM investigated the effect of aphid infestation, climate scenario, plant

composition and mildew infestation on the metabolites and alive aboveground biomass of *P.*

lanceolata. In a second SEM, we separated the metabolites of *P. lanceolata* in control pots

from those with aphids. This allowed us to test the effects of the metabolites on the aphid

population, as well as to quantify effects of the aphid population on these metabolites. The

response of the aphid population was measured as their number at the end of the infestation.

The four clusters obtained by the clustering analysis were used to divide the metabolites in

separate groups and hence to avoid redundancy in the analyses. We standardized the

metabolites, the alive aboveground biomass of *P. lanceolata* and the number of aphids by

converting to Z-scores to equalize variances. To reduce the number of mildew categories, we

rearranged degree of mildew in two categories: (1) no or mild mildew infection (category 1 -

2) (2) severe mildew infection (category 3 - 5).

Traditional SEM estimation methods assume that all variables follow a normal distribution

and all observations are independent (Grace, 2006). In our analyses, we used piecewise SEM

that allows fitting general linear mixed effect models that can incorporate random effects.

Each mixed effects model was fitted using the "lme" function in the "nlme" package (version

3.1-128) in R. For each model, we fitted a random effect of chamber. The overall path model

(the SEM) was fitted using the "piecewiseSEM" package (version 1.2.1) in R (Lefcheck,

2016). Goodness of fit was estimated using Shipley's test of d-separation, which yields a

Fisher's C statistic that is Chi-squared distributed (Shipley, 2009). If the resulting P-value >

0.05, then the SEM can be said to adequately reproduce the hypothesized causal network.

As a final confirmatory step, all data were analysed with General Linear Mixed models

(GLM) in SAS (version 9.2, SAS Institute Inc., Cary, NC) (Littell *et al.*, 1996) with chamber

as a random factor nested within climate scenario. Climate scenario, plant composition, aphid

infestation and two-way and three-way interactions between these predictors were included as

fixed factors. Because mildew infestation did not have a significant effect on biochemical

plant responses, this treatment was excluded from the GLM analysis. Non-significant factors

were backwards-excluded from the model. In case of significant effects, *a posteriori* means

comparisons using the Tukey test, corrected for multiple comparisons, were made. Effects

were considered significant at $P \leq 0.05$. Repeated univariate analyses are subject to error

proliferations. We deliberately decided not to correct results from these analyses for multiple

testing, but instead interpret them in a conservative way in congruence with the higher

described grouping analyses. Results and discussion of the series of univariate analyses are

therefore provided in Supplementary Material 3.

Results

Hierarchical clustering analysis of metabolites

The variation in primary and secondary metabolites of *P. lanceolata* was subjected to a hierarchical clustering analysis in order to understand their grouping according to their covariation in response to the imposed herbivory, climate and competition treatments. We obtained four major clusters of metabolites (clusters 1 - 4 in Fig. 1).

The first cluster containing metabolites consistently upregulated by the presence of aphids contains the phenol antioxidants, the lipophilic antioxidant tocopherol, the C:N ratio, tannins, and the non-structural carbohydrates soluble sugars (including glucose, fructose) and starch (Fig. 1). The second cluster contains another group of a.o., antioxidant and defence molecules that are downregulated by elevated CO_2 and more so by aphids in mono-cultures, i.e. the carotenoids, as well as the growth regulator, jasmonic acid, total proteins, phosphorous and lipids (Fig. 1).

The nutrient N which contributes solely to the third cluster, is downregulated under elevated CO_2 . Carbon concentration, the C:P and N:P ratio, and lignin:N were classified in the fourth cluster that is specifically upregulated by aphids in monocultures under elevated CO_2 (Fig. 1). Cluster 4 also contains metabolites that play a role in plant defences against herbivores, such as salicylic acid and the iridoid glycosides (catalpol and aucubin), and also contains TAC, lignin, cellulose, proline and MDA. Notably in this cluster, metabolites are strongly induced in CO_2 and TCO_2.

Fig. 1 *Heat map showing the metabolite levels in the leaves of P. lanceolata, normalized to Z-*
score for each metabolite (blue-white-red heat map). Blue and red colours indicate a low and
high metabolite level, respectively. Clustering was based on the Euclidean distance for
metabolites. Labels 1-4 and colours indicate the four prominent clusters. Labels C, CO_2 and
TCO_2 indicate current climate, elevated CO_2 and combined warming and elevated CO_2,
respectively. Plant communities consist of monocultures of P. lanceolata (mono) and mixtures
of Lolium perenne and P. lanceolata (mix) with (A) and without aphids (C).

 Causal changes as inferred by structural equation models and univariate analyses

SEMs quantify the strength of both the direct and indirect interactions between the

environmental scenarios related to climate change, herbivory and plant phenotype. The first

SEM presents the effect of aphid infestation (qualitative effect), climate scenario, plant

composition and mildew infestation on the chemical composition (the metabolites of *P.*

lanceolata were subdivided into the four groups obtained from the hierarchical clustering; see

Fig. 1) and the live aboveground biomass of *P. lanceolata* (Fig. 2A). The hypothesized

structural relationship with climate scenario affecting mildew infection severity, and climate

scenario, plant composition and aphid infestation interactively affecting metabolites and

biomass adequately fits the data ($\chi^2 = 42.53$, df = 32, p = 0.101). Fig. 2A and SEM statistics in

Table S2.1 (in supplementary material 2) provide us with three insights. Firstly, TCO_2

increased the mildew infestation compared to C, but the mildew infestation, in turn, did not

affect the metabolites and the alive aboveground biomass of *P. lanceolata*. Secondly, neither

climate scenario, nor aphid infestation or plant composition had a significant impact on the

live aboveground biomass of *P. lanceolata.* Univariate analyses showed that live biomass

tended to be higher in CO_2 than C and was significantly lower in TCO_2 than CO_2, which

combined lead to similar values in TCO_2 and C (Fig. 3; see also supplementary material 3.1).

Fig. 2 (a) Structural equation model showing how climate scenario (CO_2 and TCO_2), mildew infestation, plant composition and aphid infestation affect the chemical composition and the live aboveground biomass of P. lanceolata. (b) Structural equation model showing how climate scenario, plant composition and chemical composition of P. lanceolata affect aphid population and how aphid population, in turn, affects the chemical composition of P. lanceolata. The four clusters refer to those obtained by the hierarchical clustering analysis (see Fig. 1). Solid black, green and orange arrows represent significant relationships (P ≤0.05) and dashed grey lines significant interactions. Blue lines stand for significant effects of both CO_2 and TCO_2, green lines for significant effects of CO_2 and orange lines for significant effects of TCO_2. Light grey arrows represent non-significant relationships. Standardized path coefficients are shown next to pathways. For the effect of CO_2 and TCO_2, the average path coefficients are shown. The individual path coefficients of CO_2 and TCO_2 can be seen in Table S2.1 and Table S2.2 (see supplementary material section 2). Metabolites levels, live aboveground biomass of P. lanceolata and number of aphids were scaled before analysis.

Fig. 3 *Effect of climate scenario (C, CO_2 and TCO_2), aphid infestation and plant composition on the live (A)*
and dead (B) aboveground biomass of individual P. lanceolata plants. Bars represent means ± SE. Plant
communities consist of monocultures of P. lanceolata and mixtures of Lolium perenne and P. lanceolata.

Thirdly, climate scenario, aphid infestation and plant composition impacted the distinguished

clusters in a different manner. We noticed a significant interaction between climate scenario

and aphid infestation for metabolites associated with cluster 1 (see supplementary material

3.2; Table S3.1 and Fig. S3.1). Aphid infestation increased the concentrations of the

metabolites on average and pending on the environmental change scenario between 30 and

100% in cluster 1 and plant composition (except for C:N ratio; Fig. 4A-B). The

concentrations of glucose and soluble sugars increased at elevated CO_2 but only in

monocultures, while TCO_2 did not alter the concentrations of glucose relative to the CO_2

treatment but decreased the concentrations of soluble sugars in monocultures (Fig. 4C). In

general CO_2, compared to C, increased the concentration of tannin while TCO_2 did not impose

further differences. Elevated temperature and CO_2 reduced the concentration of starch (except

for mixtures with aphids) and the C:N ratio compared to CO_2.

Fig. 4 *Effect of climate scenario (C, CO_2 and TCO_2), aphid infestation and plant composition on a selection of*
relevant metabolites from the four distinguished clusters. Bars represent means ± SE. Plant communities
consist of monocultures of P. lanceolata and mixtures of Lolium perenne and P. lanceolata. As cluster 3 only
contained one metabolite, it is not depicted here.

Also the metabolites in cluster 2 (see supplementary material 3.2 – Table S3.2, Fig. S3.2)

differed according to climate scenario and aphid infestation. More specifically, univariate

analyses of the metabolites indicates that the concentration of jasmonic acid in C increased

with aphid infestation in monocultures and mixtures while aphid infestation in CO_2 and TCO_2

decreased it in monocultures (Fig. 4D). Interspecific interactions between *P. lanceolata* and *L.*

perenne mitigated the negative effect of aphid infestation in these climate scenarios, and

hence have a neutralizing effect. Aphid infestation reduced the concentration of P in

monocultures, but did not alter it in mixtures (Fig. 4E). Furthermore, also CO_2, compared to

C, reduced the concentration of P and the concentration was further reduced in TCO_2.

Furthermore, aphid infestation decreased the concentration of proteins in mixtures in C but

did not alter the concentrations in monocultures and mixtures in CO_2 and TCO_2 (Fig. 4F). In

general, CO_2 compared to C decreased the concentration of proteins irrespective of the plant

composition while aphid infestation and TCO_2, compared to CO_2, did not alter it. The

concentration of lipids decreased with aphid infestation in monocultures and mixtures in all

climate scenarios (except for mixtures in C). Also, CO_2, compared to C, decreased the

concentration of lipids in monocultures without aphids and mixtures with aphids, but in

TCO_2, compared to CO_2, this concentration increased again to similar levels as C in

monocultures without aphids. Carotenoid concentration increased in TCO_2, compared to C

and CO_2. Leaf nitrogen was not impacted by any of the treatments (cluster 3; see

supplementary material 3.2; Table S3.3).

Aphid infestation, CO_2 and TCO_2 increased the metabolites in cluster 4 (see supplementary

material 3.2; Table S3.4, Fig. S3.3 & Fig. S3.4) and specifically inhibited the expression of

salicylic acid under the two environmental change scenarios (Fig 4G). Besides these single

factor treatments effects, climate scenario, aphid infestation and plant composition also

interacted with each other in affecting the metabolites in cluster 4. CO_2 and TCO_2

strengthened (salicylic acid and lignin) or induced (proline) an effect of aphid infestation but

only in monocultures (Fig. 4G). TCO_2, compared to CO_2 and C increased the N:P and the C:P

ratio in monocultures (Fig. 4H,I). Cellulose levels were always higher in monocultures and

aphid infestation but the magnitude of increase was lowers under the two climate-related

environmental change scenarios relative to the current conditions (control).

The second SEM presents the effect of climate scenario, plant composition and chemical

composition of *P. lanceolata* on the aphid population and whether the aphid population, in

turn, altered the chemical composition of *P. lanceolata* (quantitative effect). The mildew

infestation was left out of this analysis since it is a parasite of the plants, not of the aphids.

The results of this SEM model show the following goodness of fit statistics: $\chi^2 = 77.52$, df =

80 and p = 0.558 (Table S2.2 and Fig. 2b), which provides us with two insights. Firstly,

climate scenario, plant composition and the chemical composition of *P. lanceolata* did not

affect the number of aphids. Secondly, the aphid population size in turn, did not change the

chemical composition of *P. lanceolata*. Bringing together the findings from both SEM

models, we conclude that the treatment aphid infestation altered the chemical composition of

P. lanceolata but the number of aphids had no effect on it.

Discussion

A future climate clearly altered leaf nutritional quality and the expression of defence

molecules, while aphid infestation overall impaired nutritional requirements for insect

herbivores. Despite the strong metabolic changes as induced by climate-related environmental

change, no feedbacks on aphid population performance and herbivory were detected. More

importantly, we show that interspecific plant competition neutralizes the positive effects of

elevated CO_2 on the production of defence molecules in *P. lanceolata*. Induced effects of

climate-related environmental change on plant performance can thus be altered and even

nullified by competitive and food web interactions. We first discuss the general herbivore-

induced responses followed by an in-depth discussion of their dependency on the imposed

climate and competition treatments.

Herbivore-induced responses

Our experiment, as conducted under unique semi-natural conditions confirmed in first

instance findings that aphid infestation reduces N, P, total proteins and lipids in *P. lanceolata*

leaves, thereby impairing nutritional requirements of insect herbivores (Schoonhoven *et al.*,

2005). In contrast to previous research (Walling, 2000), however, aphid infestation promoted

the salicylic acid and jasmonic acid pathway (Kuśnierczyk *et al.*, 2011; Morkunas & Gabryś,

2011) as parallelized by the increased production of catalpol, aucubin, tannin, lignin and

cellulose.

Aphid herbivory increased oxidative stress in plants as well (Wu & Baldwin, 2010). We

assessed the induction of oxidative stress by measuring MDA, which aphid herbivory

enhanced considerably. A common defence response of plants against oxidative stress

includes the increase of antioxidant metabolites. Indeed, TAC and the well-known antioxidant

molecules phenols and tocopherols increased due to aphid herbivory. Moreover, *P. lanceolata*

responded to herbivory by increasing soluble sugars. At higher concentrations, also sugars can

act as antioxidants and may play a signalling role in regulating stress and defence responses

(Gómez-Ariza *et al.*, 2007, Peshev *et al.*, 2013). The induced resistance against oxidative

stress associated with aphid feeding appeared sufficiently effective to constrain biomass loss

from herbivory.

Climate dependency of the plant phenotype

Elevated CO_2 and temperature did not change the impact of herbivory on *P. lanceolata*

biomass. This implies that warming and CO_2 did not affect the net interaction strength

between plants and herbivores under semi-natural conditions. In contrast to earlier studies

(Van De Velde *et al.*, 2016), our experiment allows a profound analysis of the underlying

biochemical processes.

First, we found elevated CO_2 to increase starch and soluble sugars (in monocultures) and to

lower leaf proteins, hereby increasing the C:N ratio. In addition, elevated CO_2 reduced P and

consequently increased the C:P ratio (in monocultures), thereby reducing the plant's nutritive

value to herbivores (Huberty & Denno, 2006; Lincoln *et al.*, 1986; Mattson, 1980;

Schoonhoven *et al.*, 2005). We demonstrate here that these effects can be countered when

elevated CO_2 is accompanied with a temperature rise. This contradicts findings based on a

limited set of primary metabolites (Murray *et al.* 2013), but urges to apply broader phenomic

approaches to fully understand plant responses to climate change.

Second, elevated CO_2 increased the defence molecules lignin and tannin, especially under

aphid infestation. The induction of plant chemical defences (Bidart-Bouzat & Imeh-Nathaniel,

2008; Bidart-Bouzat *et al.*, 2005) is thus not consistent for all defence molecules, and are

consequently anticipated to render prediction on future plant-insect herbivore interaction

highly complex (Agrawal, 1999). In our study, the higher C:N ratio reflected higher lignin

and tannin levels. While there is a consensus that the carbon-nutrient-balance hypothesis fails

as predictive framework (Hamilton *et al.*, 2001; Lindroth, 2012), alternative hypotheses have

been proposed stating that resource utilization for chemical defence is linked with

photosynthesis, hormone regulation and the control of gene expression (Zavala *et al.*, 2017).

We here provide evidence for this alternative mechanism as we found CO_2 to alter

phytohormones (i.e., salicylic acid and jasmonic acid) that play an important role in

promoting compounds responsible for herbivore defence (Wu & Baldwin, 2010). Since no

further quantitative changes were observed under higher temperature, we can confidently

conclude that climate-change-induced changes in metabolic profiles are more due to enhanced

CO_2 than to temperature increases acting in parallel.

Competition: a biotic interaction mediating climate-induced effects

The presence and identity of neighbouring plants can influence the quality of the host plant by

altering primary and secondary chemistry (Barton & Bowers, 2006; Broz *et al.*, 2010, Lankau,

2012; Thorpe *et al.*, 2011). We provide evidence that these biotic interactions are additionally

able to alter climate-induced phenomic changes, and by that further impacting ecological

interactions under climate-related environmental change. Energetic growth-defence-trade-offs

have usually been attributed to these competition effects (Herms & Mattson, 1992), but recent

studies demonstrated the role of light limitation for plant defence downregulation (Ballaré,

2014; Campos *et al.*, 2016). More specifically, a reduced red to far-red ratio downregulates

defences by a simultaneous inhibition of the jasmonic acid and the salicylic acid pathway

(Wit *et al.*, 2013). The lower levels of defence molecules and salicylic acid in *P. lanceolata*

support this hypothesis. Competition did, however, also counter the CO_2-induced changes in

primary and secondary metabolites, as well as those induced by herbivores, rendering an

understanding of plant defences far from trivial (de Vries *et al.*, 2017). It is well-understood

that bottom-up effects from plant quality to herbivore abundance cannot be generalized across

the feeding guilds (Bezemer & Jones, 1998; Robinson *et al.*, 2012; Stiling & Cornelissen,

2007), but phloem feeders such as aphids were at least expected to be facilitated. Despite the

complex, high-dimensional changes in plant metabolomes under competition and climate

change, the overall effects on aphid population sizes were small to non-existent.

We here thus demonstrate that the joint action of atmospheric factors associated with climate

change and biotic interactions are able to induce pronounced changes in plant metabolomes

and biomass, but that translation of these biochemical changes towards ecological responses

appears non-trivial. Climate-change-imposed changes are shown to be completely neutralized

when common biotic interactions are taken into consideration. Plant-plant interactions thus

add a layer of complexity in mechanistic studies of climate change effects on plant-enemy

ecological interaction.

Supplementary data:

Supplementary material 1: Schematic representation of the experimental setup

Supplementary material 2: Metabolite quantification methods and SEM statistics

Supplementary material 3: Results from the univariate GLM analyses (statistics and figures)

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References

References

Agrawal, AA. (1999). Induced responses to herbivory in wild radish: effects on several herbivores and

plant fitness. Ecology, **80**, 1713-1723.

Agrios, G.N. (2005). Plant pathology, Amsterdam, Netherland, Elsevier Academic Press.

Alford, D.V. (2014). Pests of fruit crops: a colour handbook, London, U.K, Manson Publishing.

Bale, J.S., Masters G.J., Hodkinson I.D. et al.. (2002). Herbivory in global climate change research:

direct effects of rising temperature on insect herbivores. Global change biology, **8**, 1-16.

Ballaré, CL. (2014). Light regulation of plant defense. Annual review of plant biology, **65**, 335-363.

Ballaré ,C.L., Scopel, A.L., Sanchez, R.A. (1990). Far-red radiation reflected from adjacent leaves: an

early signal of competition in plant canopies. Science, **247**, 329.

Bartley,G.E., Scolnlk,P.A.(1995).Plant carotenoids:Pigments for photoprotection, visualattraction, and

human health. Plant and Cell, **7**, 1027-1038.

Barton, K.E., Bowers, M.D. (2006). Neighbor species differentially alter resistance phenotypes in

Plantago. Oecologia, **150**, 442-452.

Bauerfeind, S.S., Fischer, K. (2013). Increased temperature reduces herbivore host-plant quality.

Global change biology, **19**, 3272-3282.

Bezemer, T.M., Jones, T.H. (1998). Plant-insect herbivore interactions in elevated atmospheric CO_2:

quantitative analyses and guild effects. Oikos, **82**, 212-222.

Bidart-Bouzat, M.G., Imeh-Nathaniel, A. (2008). Global change effects on plant chemical defenses

against insect herbivores. Journal of Integrative Plant Biology, **50**, 1339-1354.

Bidart-Bouzat, M.G., Mithen, R., Berenbaum, M.R. (2005). Elevated CO_2 influences herbivory-

induced defense responses of Arabidopsis thaliana. Oecologia, **145**, 415-424.

Blokhina O., Virolainen E., Fagerstedt K.V. (2003) Antioxidants, oxidative damage and oxygen

deprivation stress: a review, Annales of Botany, **91**, 179–194.

Blommers, L.H.M., Helsen, H.H.M., Shao, H.B., Liang, Z.S., Shao, M.A. (2006). Osmotic regulation

of 10 wheat (Triticum aestivum L.) genotypes at soil water deficits. Biointerphases, **47**, 132-139.

Bowers, M.D., Puttick, G.M.. (1988). Response of generalist and specialist insects to qualitative

allelochemical variation. Journal of chemical ecology, **14**, 319-334.

Bowers, M. D. (1991). Iridoid glycosides. In:G. A. Rosenthal and M. R. Berenbaum (eds.).

Herbivores: Their Interactions with Secondary Plant Metabolites, 297–325. Academic Press, New

York.

Broz, A.K., Broeckling, C.D., De-La-Peña, C. et al.. (2010). Plant neighbor identity influences plant

biochemistry and physiology related to defense. BMC Plant Biology, **10**, 1.

Bryant, J.P., Chapin Iii, F.S., Klein, D.R. (1983). Carbon/nutrient balance of boreal plants in relation

to vertebrate herbivory. Oikos **78**, 357-368.

Campos, M.L., Yoshida, Y., Major, I.T., et al.. (2016). Rewiring of jasmonate and phytochrome B

signalling uncouples plant growth-defense tradeoffs. Nature Communications, **7**, 12570.

Cornelissen, T. (2011). Climate change and its effects on terrestrial insects and herbivory patterns.

Neotropical Entomology, **40**, 155-163.

Cotrufo, M.F., Gorissen, A. (1997). Elevated CO2 enhances below-ground C allocation in three

perennial grass species at different levels of N availability. New Phytologist, **137**, 421-431.

Darrow, K, Bowers, M.D. (1999). Effects of herbivore damage and nutrient level on induction of

iridoid glycosides in Plantago lanceolata. Journal of chemical ecology, **25**, 1427-1440.

Das, K., Roychoudhury, A. (2014) Reactive oxygen species (ROS) and response of antioxidants as

ROS-scavengers during environmental stress in plants. Frontiers in Environmental Sciences, **2,** 00053

De Boeck, H.J, Lemmens, C., Zavalloni, C. et al.. 2008. Biomass production in experimental

grasslands of different species richness during three years of climate warming. Biogeosciences, **5**, 585-

594.

De Valpine, P., Harte, J.. (2001). Plant responses to experimental warming in a montane meadow.

Ecology, **82**, 637-648.

De Vries, J., Evers, J.B., Poelman, E.H. (2017). Dynamic plant–plant–herbivore interactions govern

plant growth–defence integration. Trends in plant science, **22**, 329 - 337.

Dieleman, W.I., Vicca S., Dijkstra, F.A. et al.. (2012). Simple additive effects are rare: a quantitative

review of plant biomass and soil process responses to combined manipulations of CO2 and

temperature. Global change biology, **18**, 2681-2693.

Gómez-Ariza, J., Campo, S., Rufat, M., Estopà M, Messeguer, J., Segundo, B.S., Coca, M..(2007).

Sucrose-mediated priming of plant defense responses and broad-spectrum disease resistance by

overexpression of the maize pathogenesis-related PRms protein in rice plants. Molecular Plant-

Microbe Interactions, **20**, 832-842.

Grace, J.B. (2006). Structural equation modeling and natural systems, Cambridge, UK, Cambridge

University Press.

Hamilton, J., Zangerl, A., Delucia, E., Berenbaum, M. (2001). The carbon–nutrient balance

hypothesis: its rise and fall. Ecology letters**, 4**, 86-95.

Herms, D.A., Mattson, W.J. (1992). The dilemma of plants: to grow or defend. Quarterly Review of

Biology, **67**, 283-335.

Huberty, A.F., Denno, R.F. (2006). Consequences of nitrogen and phosphorus limitation for the

performance of two planthoppers with divergent life-history strategies. Oecologia, **149**, 444-455.

Hunter, M.D. (2001). Effects of elevated atmospheric carbon dioxide on insect-plant interactions.

Agricultural and Forest Entomology, **3**, 153-159.

IPCC. (2001). Climate Change: the scientific basis. Contribution of Working Group I to the third

sssessment report of the Intergovernmental Panel on Climate Change. . eds Houghton J, Ding Y,

Griggs D, Noguer M, Van Der Linden P, Dai X, Maskell K, Ca. J., Cambridge, UK and New York,

NY, USA, Cambridge University Press.

Jamieson, M.A., Schwartzberg, E.G., Raffa, K.F., Reich, P.B., Lindroth, R.L. (2015). Experimental

climate warming alters aspen and birch phytochemistry and performance traits for an outbreak insect

herbivore. Global Change Biology, **21**, 2698-2710.

Johnson, G., Young, A., Scholes, J., Horton, P. (1993). The dissipation of excess excitation energy in

British plant species. Plant, Cell & Environment, **16**, 673-679.

Kaloshian, I, Walling, L.L. (2005). Hemipterans as plant pathogens. Annu. Rev. Phytopathol., **43**,

491-521.

Kuśnierczyk, A., Tran, D.H., Winge, P., Jørstad, T.S., Reese, J.C., Troczyńska, J., Bones, A.M.

(2011). Testing the importance of jasmonate signalling in induction of plant defences upon cabbage

aphid. Brevicoryne brassicae. attack. BMC genomics, **12**, 423.

Lankau, R.A. (2012). Coevolution between invasive and native plants driven by chemical competition

and soil biota. Proceedings of the National Academy of Sciences, **109**, 11240-11245.

Lefcheck, J.S. (2016). piecewiseSEM: Piecewise structural equation modelling in R for ecology,

evolution, and systematics. Methods in Ecology and Evolution, **7**, 573-579.

Lilley, J., Bolger, T., Gifford, R. (2001). Productivity of Trifolium subterraneum and Phalaris aquatica

under warmer, high CO2 conditions. New Phytologist, **150**, 371-383.

Lincoln, D., Couvet, D., Sionit, N. (1986). Response of an insect herbivore to host plants grown in

carbon dioxide enriched atmospheres. Oecologia, **69**, 556-560.

Lincoln, D.E., Fajer, E.D., Johnson, R.H. (1993). Plant insect herbivore interactions in elevated CO2

environments. Trends in Ecology & Evolution, **8**, 64-68.

Lindroth, R.L. (2012). Atmospheric change, plant secondary metabolites and ecological interactions.

In: The ecology of plant secondary metabolites: from genes to global processes.. eds Iason G, Dicke

M, Hartley S. pp Page. Cambridge, UK, Cambridge University Press.

Littell, R.C., Milliken, G.A., Stroup, W.W., Wolfinger, R.D. (1996). SAS system for mixed models,

Cary, North Carolina, SAS institute Inc.

Liu, Q., Luo, L., Zheng, L. (2018). Lignins: Biosynthesis and Biological Functions in

Plants. International journal of molecular sciences, **19,** 19020335

Long, S.P, Ainsworth, E.A., Rogers, A., Ort, D.R. (2004). Rising atmospheric carbon dioxide: plants

FACE the future. Annual review of plant biology, **55**, 591-628.

Mattson, W.J. (1980). Herbivory in relation to plant nitrogen content. Annual review of ecology and

systematics, **11**, 119-161.

Meisner, M.H, Harmon, J.P., Ives, A.R. (2014). Temperature effects on long-term population

dynamics in a parasitoid-host system. Ecological Monographs, **84**, 457-476.

Mikkelsen, B.L., Jørgensen, R.B., Lyngkjær, M.F. (2015). Complex interplay of future climate levels

of CO2, ozone and temperature on susceptibility to fungal diseases in barley. Plant Pathology, **64**,

319-327.

Morkunas, I, Gabryś, B. (2011). Phytohormonal signaling in plant responses to aphid feeding. Acta

Physiologiae Plantarum, **33**, 2057-2073.

Murray, T., Ellsworth, D., Tissue, D., Riegler, M. (2013). Interactive direct and plant-mediated effects

of elevated atmospheric [CO2] and temperature on a eucalypt-feeding insect herbivore. Global change

biology, **19**, 1407-1416.

Naudts, K, Van Den Berge, J., Farfan, E. et al.. 2014. Future climate alleviates stress impact on

grassland productivity through altered antioxidant capacity. Environmental and Experimental Botany,

99, 150-158.

Peshev, D., Vergauwen, R., Moglia, A., Hideg, É., Van Den Ende, W. (2013). Towards understanding

vacuolar antioxidant mechanisms: a role for fructans? Journal of Experimental Botany, **64**, 1025-1038.

Puttick, G.M., Bowers, M.D. (1988). Effect of qualitative and quantitative variation in allelochemicals

on a generalist insect: iridoid glycosides and the southern armyworm. Journal of chemical ecology, **14**,

335-351.

Rawson, H. (1992). Plant responses to temperature under conditions of elevated CO2. Australian

Journal of botany, **40**, 473-490.

Robinson, E.A, Ryan G.D., Newman, J.A. (2012). A meta-analytical review of the effects of elevated

CO2 on plant–arthropod interactions highlights the importance of interacting environmental and

biological variables. New Phytologist, **194**, 321-336.

Ryan, G.D., Sylvester, E.V., Shelp, B.J., Newman, J.A. (2015). Towards an understanding of how

phloem amino acid composition shapes elevated CO2-induced changes in aphid population dynamics.

Ecological Entomology, **40**, 247-257.

Saeed, A., Sharov, V., White, J. et al.. (2003). TM4: a free, open-source system for microarray data

management and analysis. Biotechniques, **34**, 374.

Schoonhoven, L.M., Van Loon, J.J., Dicke, M. (2005). Insect-plant biology, Oxford, Oxford

University Press.

Shipley, B. (2009). Confirmatory path analysis in a generalized multilevel context. Ecology, **90**, 363-

368.

Stiling, P, Cornelissen, T. (2007). How does elevated carbon dioxide. CO2. affect plant-herbivore

interactions? A field experiment and meta-analysis of CO2-mediated changes on plant chemistry and

herbivore performance. Global change biology, **13**, 1823-1842.

Szabados, L., Savoure, A. (2010). Proline: a multifunctional amino acid. Trends in Plant Science, **15**,

89–97.

Thomas, M.B., Blanford, S. (2003). Thermal biology in insect-parasite interactions. Trends in Ecology

& Evolution, **18**, 344-350.

Thorpe, A.S., Aschehoug, E.T., Atwater, D.Z., Callaway, R.M. (2011). Interactions among plants and

evolution. Journal of Ecology, **99**, 729-740.

Vaal, F. (2004). Life history data of the rosy apple aphid Dysaphis plantaginea. Pass.). Homopt.,

Aphididae. on plantain and as migrant to apple. Journal of Pest Science, **77**, 155-163.

Van De Velde, H., Nijs, I., Bonte, D. (2016). Warming affects different components of plant–

herbivore interaction in a simplified community but not net interaction strength. Oikos, **126**, 285-295.

Walling, L.L. (2000). The myriad plant responses to herbivores. Journal of Plant Growth Regulation,

19, 195-216.

War A. R., Paulraj M. G., Ahmad T., Buhroo A. A., Hussain B., Ignacimuthu S . (2012). Mechanisms

of plant defense against insect herbivores. Plant Signaling and Behaviour, **7**, 1306–1320.

Williams, A.L., Wills, K.E., Janes, J.K., Vander Schoor, J.K., Newton, P.C., Hovenden, M.J. (2007).

Warming and free-air CO_2 enrichment alter demographics in four co-occurring grassland species.

New Phytologist, **176**, 365-374.

Wit, M., Spoel, S.H., Sanchez-Perez, G.F., Gommers, C.M., Pieterse, C.M., Voesenek, L.A., Pierik, R.

(2013). Perception of low red: far-red ratio compromises both salicylic acid-and jasmonic

acid-dependent pathogen defences in Arabidopsis. The Plant Journal, **75**, 90-103.

Wu, J., Baldwin, I.T. (2010). New insights into plant responses to the attack from insect herbivores.

Annual review of genetics, **44**, 1-24.

Zavala, J., Gog, L., Giacometti, R. (2017). Anthropogenic increase in carbon dioxide modifies plant–

insect interactions. Annals of Applied Biology, **170**, 68-77.

Zavala, J.A., Nabity, P.D., Delucia, E.H. (2013). An emerging understanding of mechanisms

governing insect herbivory under elevated CO2. Annual review of entomology, **58**, 79-97.

Zinta, G., AbdElgawad, H., Peshev, D., Weedon, J.T., Van den Ende, W., Nijs, I., Janssens, I.A.,

Beemster, G.T.S., Asard, H. (2018). Dynamics of metabolic responses to periods of combined heat

and drought in *Arabidopsis thaliana* under ambient and elevated atmospheric CO_2 Journal of

Experimental Botany, **69**, 2159–2170

Zvereva, E.L., Kozlov, M.V. (2006). Consequences of simultaneous elevation of carbon dioxide and

temperature for plant-herbivore interactions: a metaanalysis. Global change biology, **12**, 27-41.

 Figure legends

 Fig. 1 Heat map showing the metabolite levels in the leaves of *P. lanceolata*, normalized to Z-

 score for each metabolite (blue-white-red heat map). Blue and red colours indicate a low and

 high metabolite level, respectively. Clustering was based on the Euclidean distance for

 metabolites. Labels 1-4 and colours indicate the four prominent clusters. Labels C, CO_2 and

 TCO_2 indicate current climate, elevated CO_2 and combined warming and elevated CO_2,

 respectively. Plant communities consist of monocultures of *P. lanceolata* (mono) and

 mixtures of *Lolium perenne* and *P. lanceolata* (mix) with (A) and without aphids (C).

 Fig. 2 (a) Structural equation model showing how climate scenario (CO_2 and TCO_2), mildew

 infestation, plant composition and aphid infestation affect the chemical composition and the

 live aboveground biomass of *P. lanceolata*. (b) Structural equation model showing how

 climate scenario, plant composition and chemical composition of *P. lanceolata* affect aphid

 population and how aphid population, in turn, affects the chemical composition of *P.*

 lanceolata. The four clusters refer to those obtained by the hierarchical clustering analysis

 (see Fig. 1).

 Solid black, green and orange arrows represent significant relationships (P $\leq$ 0.05) and dashed

 grey lines significant interactions. Blue lines stand for significant effects of both CO_2 and

 TCO_2, green lines for significant effects of CO_2 and orange lines for significant effects of

 TCO_2. Light grey arrows represent non-significant relationships. Standardized path

 coefficients are shown next to pathways. For the effect of CO_2 and TCO_2, the average path

 coefficients are shown. The individual path coefficients of CO_2 and TCO_2 can be seen in

 Table S1 and Table S2 (see supplementary material section 2). Metabolites levels, live

 aboveground biomass of *P. lanceolata* and number of aphids were scaled before analysis.

Fig. 3 Effect of climate scenario (C, CO_2 and TCO_2), aphid infestation and plant composition

on the live (A) and dead (B) aboveground biomass of *Plantago lanceolata*. Bars represent

means ± SE. Plant communities consist of monocultures of *P. lanceolata* and mixtures of

Lolium perenne and *P. lanceolata*.

Fig. 4 Effect of climate scenario (C, CO_2 and TCO_2), aphid infestation and plant composition

on the on a selection of relevant metaboloites from the four distinguished clusters. Bars

represent means ± SE. Plant communities consist of monocultures of *P. lanceolata* and

mixtures of *Lolium perenne* and *P. lanceolata*.