

Trait-based approach to bacterial growth efficiency

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ABSTRACT

14 Bacterial growth efficiency (BGE) is the proportion of assimilated carbon that is converted into
biomass and reflects the balance between growth and energetic demands. Often measured as an
16 aggregate property of the community, BGE is highly variable within and across ecosystems. To
understand this variation, we used a trait-based approach with 20 bacterial isolates enriched from
18 lake communities to determine how consumer and resource identity affect BGE. We used
phenotypic and genomic approaches to characterize the metabolic physiology of each isolate and
20 test for predicted trade-offs between growth rate and efficiency. Across resource types, 20 % of
the variation in BGE could be attributed to the coarse-scale taxonomic resolution of the isolate,
22 while 58 % of the variation could be explained by isolate identity. Resource identity explained a
relatively small amount of variation (7 %) in BGE across isolates but accounted for > 60 % of
24 the variation within an isolate alone. Metabolic trade-offs and genomic features associated with
BGE suggest that BGE is a species trait, which regardless of resource environment, contributes
26 to variation in BGE. Genomic and phylogenetic information from microbiomes may help predict
aggregate community functions such as BGE to better understand the fate of organic matter
28 resources in ecosystems.

INTRODUCTION

32 In most ecosystems, heterotrophic bacteria play a pivotal role in determining whether organic
carbon is respired and thus lost as CO₂ or is instead converted into biomass and retained in food
34 webs (Pomeroy et al. 1998; Ducklow 2008). Many factors control how bacteria process carbon,
but perhaps the most important is reflected in measurements of bacterial growth efficiency
36 (BGE). BGE is the proportion of assimilated organic carbon that is converted into bacterial
biomass (del Giorgio and Cole 1998). When BGE is high, more carbon is turned into biomass
38 where it can be retained for longer periods of time while also serving as a source of energy for
other members of the food web. In contrast, when BGE is low, microbially assimilated carbon
40 has a shorter residence time and is released back to the environment as CO₂.

When measured at the community scale, BGE is notoriously variable among habitats and
42 has proven difficult to predict (del Giorgio and Cole 1998). While it can be influenced by a range
of chemical and physical properties (Apple and del Giorgio 2007; Hall and Cotner 2007; del
44 Giorgio and Newell 2012; Sinsabaugh et al. 2013), variation in BGE may also reflect taxon-
specific differences in microbial metabolism. BGE may vary among microbial taxa for a number
46 of reasons. For example, BGE results from the physiological balance between cellular growth
and energetic demands. As such, bacterial growth strategy is predicted to constrain BGE via
48 physiological trade-offs (Litchman et al. 2015). For example, it has been hypothesized that
oligotrophs have higher maximum growth efficiency than copiotrophs (Roller and Schmidt
50 2015). In addition, consumers that specialize on only a few resources are predicted to be more
efficient at using those resources than more generalist consumers (Dykhuizen and Davies 1980;
52 Glasser 1984). As such, consumer properties such as maximum growth rate and the number of
resources used (i.e., niche breadth) could underlie species-specific differences in BGE.

54 Variation in BGE may also be influenced by the resources used to meet energetic and
growth demands. For example, different resources can affect ATP yield depending on the
56 metabolic pathways that are used by a bacterial population (Fuhrer et al. 2005; Flamholz et al.
2013). Also, it is important to recognize that the energy-producing catabolic processes and
58 biomass-producing anabolic processes are not independent (Russell and Cook 1995). For
example, cells have the potential to produce >30 ATP from a single glucose molecule if the
60 molecule is completely oxidized, but in this scenario, there would be no remaining glucose to
yield new biomass. Instead, to build new biomass, cells must use the intermediate products of
62 glycolysis to form proteins and other cellular material, which diminishes the maximum ATP
yield (Gottschalk 1986). Furthermore, biomass production requires materials (e.g., intermediates
64 of Krebs cycle) and energy. For example, the synthesis of proteins, which constitute ~70% of
cellular dry mass, requires amino acid building blocks and 4 ATP per peptide bond (Tempest and
66 Neijssel 1984; Gottschalk 1986). Therefore, because resources differ in their potential energy
yield and bacteria differ in their ability to extract energy and form biomass from a given
68 resource, BGE should vary based on the resources available to bacteria.

In this study we measured BGE in a set of bacterial isolates growing with only a single
70 carbon source. In addition to partitioning variation in BGE based on consumer and resource
identity, we tested for hypothesized trade-offs between BGE, growth rate, and niche breadth.
72 Using genomes of each isolate, we evaluated whether metabolic pathways could explain
differences in BGE among our diverse collection of aquatic bacteria. Last, to test if resources
74 have different effects on the metabolic traits that underlie BGE (production and respiration), we
tested if there were relationships between respiration and production rate when isolates were
76 grown on each resource and if the relationships were unique to each resource. Our trait-based

approach provides a framework for understanding linkages between community structure and
78 function due to the physiological constraints on BGE and suggest that large changes in
community composition or available resources may alter BGE in predictive ways.

80

METHODS

82 **Bacterial Strains** — Using a novel cultivation approach, we isolated 20 bacterial strains from
lakes in the Huron Mountain Research Preserve (Powell, MI, USA) by incubating inert carbon
84 beads (Bio-Sep Beads) in the water column for one week. Prior to the incubations, the beads
were saturated with a sterile complex carbon substrate, i.e., Super Hume (CropMaster, United
86 Agricultural Services of America, Lake Panasoffkee, Florida, USA). Super Hume is a lignin-rich
resource comprising 17 % humic and 13 % fulvic acids, and has been shown to be an analog of
88 terrestrial DOC in aquatic ecosystems that can be used by diverse bacteria (Lennon et al. 2013).
We used this enrichment technique to select for bacteria with a range of metabolic potential
90 (Ghosh et al. 2009). After the incubation, beads were rolled on R2 agar plates (BD Difco, Sparks
Maryland, USA) and incubated at 25 °C. We picked random colonies from plates and serially
92 transferred until axenic. All strains were preserved in 25 % glycerol at -80 °C.

We identified each bacterial strain by direct sequencing the 16S rRNA gene. We obtained
94 genomic DNA from log phase cultures using the FastPrep DNA extraction kit according to the
manufacturer's specifications (MP Biomedical). We used 10 ng of genomic DNA to amplify the
96 16S rRNA gene using the 27F and 1492R bacterial primers (See Supplemental for primer
sequences and PCR conditions). We sequenced the PCR products at the Indiana Molecular
98 Biology Institute (IMBI) at Indiana University (Bloomington, Indiana, USA). Raw sequence

reads were quality-trimmed based on a Phred quality score of 25. Forward and reverse reads
100 were manually merged after aligning sequences to the Ribosomal Database Project reference
alignment (Cole et al. 2009) using mothur (Schloss et al. 2009). After merging into full length
102 16S rRNA sequences, alignments were checked using ARB (Ludwig et al. 2004) and sequences
were compared to the Silva All-Species Living Tree Project database (Yilmaz et al. 2014) for
104 taxonomic identification (Fig. 1).

106 **Bacterial Growth Efficiency** — To test for differences in bacterial growth efficiency (BGE)
across our isolates, we measured BGE for each isolate when cultured on one of three different
108 carbon substrates: glucose, succinate, or protocatechuate (Fig. 2). These carbon sources were
chosen based on differences in their bioavailability and structure but also the required pathways
110 for metabolism (see Fig. 2).

To determine BGE we measured bacterial respiration and production rates and then
112 calculated BGE as $BP/(BP + BR)$, where BP is bacterial productivity and BR is bacterial
respiration (del Giorgio and Cole 1998). BP and BR were measured using triplicate cultures of
114 each isolate. Cultures of each isolate were grown in R2 broth (BD Difco, Sparks Maryland,
USA) until mid-log phase. We then transferred 100 μ L of culture into 10 mL of M9 with the
116 appropriate carbon source (25 mM C) and allowed 24 h for the cultures to acclimate. We then
transferred 100 μ L of culture into 10 mL of fresh carbon-amended M9 and incubated 1-3 h to
118 replenish nutrients. Using these transfers, we were able to establish populations of each isolate at
target cell densities between 10^4 and 10^5 cells mL^{-1} . We used the populations to measure BP and
120 BR, which were normalized to cell density using plate counts of colony forming units. We

measured BP using the ^3H -Leucine assay (Smith and Azam 1992) with 1.5 mL of culture. We
122 added ^3H -Leucine to a final concentration of 50 mM and incubated for 1 h. Following
incubation, we terminated production with trichloroacetic acid (final concentration 3 mM) and
124 measured leucine incorporation using a liquid scintillation counter. We measured BR using an
automated O_2 measurement system (PreSens Sensor Dish System, PreSens, Regensburg,
126 Germany) on 5 mL of culture based on the rate O_2 consumption during three-hour incubations.
We estimated BR as the slope of O_2 concentration during the incubation using linear regression.
128 We used theoretical respiratory quotients for each resource to convert O_2 depletion into C
respiration assuming aerobic growth.

130

Taxonomic and Phylogenetic Relationships — We compared differences in BGE across
132 isolates and resources using linear models. First, we used a taxonomic framework to compare
BGE between isolates (Lennon et al. 2012). Isolates were classified into taxonomic groups based
134 on the species tree constructed in ARB. We then used mixed linear models to compare BGE
across taxonomic groups and resources. To test the hypothesis that taxonomy (i.e., at the class
136 level) affects BGE, we nested resources within isolate. To test the hypothesis that resource
identity affects BGE, we nested isolates within resource. We identified the best statistical models
138 based on the variation explained (R^2) and AIC values. Second, we tested if phylogenetic
relationships between isolates explained differences in BGE across isolates. We created a
140 phylogenetic tree based on the full-length 16S rRNA gene sequences. We aligned sequences
using the SINA aligner (Pruesse et al. 2012) and checked alignments using ARB. We generated a
142 phylogenetic tree using the CIPRES science gateway (Miller et al. 2010). The phylogenetic tree
was created using RAxML-HPC v.8 on XSEDE based on the GTRCAT DNA model (Stamatakis

144 2006; Stamatakis et al. 2008). We created a consensus tree based on 1,000 bootstraps using
CONSENSE (Felsenstein 2008). We used Blomberg's K to compare trait variation across the
146 tree and test if phylogenetic relationships between isolates could explain differences in traits
(Blomberg et al. 2003). Blomberg's K is a test for phylogenetic signal that determines if trait
148 variation is better explained by phylogenetic relationships or Brownian motion. Last, to
determine if the distribution of BGE across isolates was unimodal, we used Hartigan's dip test
150 for unimodality (Hartigan and Hartigan 1985). Hartigan's dip test is used to determine if a
distribution is unimodal by testing the null hypothesis that there is a dip in the distribution. A
152 significant Hartigan's dip test would suggest that the distribution is unimodal. Alternatively, the
distribution has an internal "dip" (reported as D). All statistical tests were conducted in the R
154 statistical environment (R Core Team 2012). We used the nlme package (Pinheiro et al. 2016)
for the mixed linear models, the picante package (Kembel et al. 2015) for the phylogenetic
156 methods, and the diptest package (Maechler 2015) for Hartigan's dip test.

158 **Phenotypic Comparisons and Trade-offs** — To test the hypothesis that phenotypic differences
and physiological trade-offs underlie BGE variation, we compared the maximum growth rate
160 (μ_{max}) and niche breadth of each isolate. First, to test whether BGE was affected by growth
strategy (i.e., copiotrophs vs. oligotrophs), we measured the maximum growth rate of each
162 isolate. Bacterial growth rates were measured based on changes in optical density during 18-h
incubations. Bacterial strains were grown in R2 broth in 48-well plates. We incubated plates with
164 continuous shaking and measured optical density every 15 min using a plate reader (BioTek
MX). Growth curves were analyzed by fitting a modified Gompertz growth model (Zwietering et

166 al. 1990; Lennon 2007) to the observed growth curves using maximum likelihood fitting. We
used the model fit as our estimate of μ_{\max} .

168 Second, to test whether BGE was affected by niche breadth, we generated carbon usage
profiles using BioLog EcoPlates™ (Garland and Mills 1991). The EcoPlate is a phenotypic
170 profiling tool consisting of 31 unique carbon sources. In addition to the carbon source, each well
contains a tetrazolium dye, which in the presence of NADH will be reduced and change color.
172 We used this colorimetric assay to generate carbon usage profiles for each strain. We
standardized profiles for each strain by subtracting water blanks (average water blank + 1 SD),
174 and relativizing across substrates. Using these data, we calculated resource niche breadth using
Levin's Index (Colwell and Futuyma 1971).

176

Genomic Comparisons — To test the hypothesis that variation in metabolic pathways could
178 explain differences in BGE, we compared the genomes of each isolate. First, we determined the
metabolic pathways found in the genome of each isolate. We characterized each isolate using
180 whole genome sequencing. Genomic DNA libraries for each isolate were prepared using the
Illumina TruSeq DNA sample prep kit using an insert size of 250 base pairs (bp). Libraries were
182 sequenced on an Illumina HiSeq 2500 (Illumina, San Diego, GA) using 100-bp paired-end reads
at the Michigan State University Research Technology Support Facility. We processed raw
184 sequence reads (FASTQ) by removing the Illumina TruSeq adaptors using Cutadapt (Martin
2011), interleaving reads using Khmer (McDonald and Brown 2013), and quality-filtering based
186 on an average Phred score of 30 using the FASTX-toolkit (Hannon Lab 2010). Finally, we
normalized coverage to 25 based on a k-mer size of 25 using Khmer. We assembled the genomes

188 using Velvet (Zerbino and Birney 2008) after optimizing assembly parameters for each isolate
with Velvet Optimizer (Gladman and Seemann 2012). We annotated contigs larger than 200 bp
190 using Prokka (Seemann 2014), and predicted metabolic and physiological functions using
MAPLE with bidirectional best-hit matches (Takami et al. 2012). We identified functional
192 pathway based on the presence of intermediate genes within a pathway. We scored pathways as
functional if more than 80 % of the intermediate genes were recovered in the genomes.

194 To test the hypothesis that metabolic pathways affect BGE, we used multivariate methods
to compare the pathways of each isolate. First, we used PERMANOVA to determine if there
196 were differences in pathways associated with different levels of BGE. When significant
differences were found, we used indicator species analysis (Dufrene and Legendre 1997) to
198 determine which metabolic pathways contributed to group differences in BGE. Next, to
determine if metabolic pathways could explain differences in BGE within a group, we used
200 distance-based redundancy analysis (dbRDA) which is a multivariate technique that tests if a
quantitative predictor can explain differences in multivariate datasets (Legendre and Legendre
202 2012). Because we scored pathways as present or absent, metabolic distances between isolates
were calculated using the Jaccard Index. We tested for significance using a permutation test. If
204 the dbRDA model was significant, we used Spearman's rank-order correlation to test for
correlations between BGE and individual metabolic pathways. We used the vegan R package
206 (Oksanen et al. 2013) for multivariate analyses.

208 **Resource Effects** — To test the hypothesis that resources have different effects on components
of metabolism that affect BGE, we used a linear model to test for a relationship between BR and

210 BP. Because BP required energy through respiration, we used production rate as the dependent
variable and respiration rate as the independent variable. We used an indicator variable linear
212 regression to test for changes in BP rate due to BR. We included resource identity and group
(high- versus low-BGE) as the categorical predictors and BR as the continuous predictor
214 (Lennon and Pfaff 2005). In addition, we included all interactions terms. Respiration and
production rates were \log_{10} -transformed to meet model assumptions. Last, to determine if the
216 relationship between BR and BP rates was isometric (proportional scaling, slope = one) or
allometric (disproportional scaling, slope \neq one), we used a one-sample t-test to determine if the
218 slope was different from one. All statistical tests were conducted in the R statistical environment.

220

RESULTS

Bacterial Growth Efficiency — Using measures of bacterial productivity (BP) and respiration
222 (BR), we calculated bacterial growth efficiency (BGE) on 20 isolates each on three resources:
glucose, succinate, and protocatechuate (Fig. 3). All isolates belonged to the Proteobacteria
224 phylum with representatives from the Alpha-, Beta-, and Gamma- Proteobacteria subphyla (Fig.
1). Based on mixed-effects linear models we found that isolate and resource identity explained a
226 substantial amount of variation in BGE. Across resources, isolate identity explained 58 % of the
variation in BGE, and 67 % of the variation within resource (AIC = -48), and the taxonomic
228 order of each isolate explained 20 % of the variation in BGE across resources, and 28 % of the
variation within resource (AIC = -94). Across isolates, resource identity only explained 7 % of
230 the variation in BGE across all isolates, but 63% of variation within isolate (AIC = -117).

Based on Blomberg's K, there was weak phylogenetic signal of BGE among our isolates.

232 We did not detect phylogenetic signal for BGE when isolates used succinate ($K < 0.001$, $p =$
0.316) or protocatechuate ($K < 0.001$, $p = 0.257$), but there was a significant phylogenetic signal
234 when isolates used glucose ($K < 0.001$, $p = 0.04$). Based on Hartigan's dip test, we found that
there was a bimodal distribution of BGE among our isolates ($D = 0.07$, $p = 0.58$, Supp. Fig. 1).
236 Using this distribution, we split isolates into two groups, which we define as the "high BGE",
and "low BGE" groups.

238

Phenotypic Comparisons — Using linear models, we found phenotypic differences between
240 isolates that were related to BGE. While there was no relationship between BGE and μ_{\max} ($F_{1,7}$
 $= 0.51$, $r^2 = 0.06$, $p = 0.50$) in the low-BGE group of bacteria, we did identify a significant
242 negative relationship between BGE and μ_{\max} for the high-BGE group of bacteria ($F_{1,7} = 9.52$, r^2
 $= 0.54$, $p = 0.015$), and BGE decreased 2.58 % for each per minute increase in μ_{\max} . In contrast
244 to our predictions, there was no relationship between niche breadth (Levin's Index) and BGE for
the low-BGE group ($F_{1,7} = 1.47$, $r^2 = 0.17$, $p = 0.27$) or high-BGE group ($F_{1,7} = 0.92$, $r^2 = 0.11$,
246 $p = 0.37$).

248 **Genomic Comparisons** — We found genomic differences between isolates that were related to
BGE. First, isolates in the high-BGE group had 13 % more metabolic pathways than isolates in
250 the low BGE-group ($m_{\text{high}} = 72$, $m_{\text{low}} = 64$, t-test: $t_{18} = -2.36$, $p = 0.03$). Second, the genomes of
the low-BGE group and high-BGE group contained different metabolic pathways
252 (PERMANOVA, $R^2 = 0.14$, $p = 0.008$). Although we were unable to account for variation in the

254 metabolic pathways using BGE in the high BGE isolates (dbRDA: $F_{1,7} = 1.05$, $R^2 = 0.13$, $p =$
256 0.39), there was a marginally significant relationship between metabolic pathway composition
and BGE in isolates belonging to the low-BGE group ($F_{1,7} = 2.15$, $R^2 = 0.18$, $p = 0.08$) despite
the fact that correlations between individual metabolic pathways and BGE were weak (all rho
values $< |0.7|$).

258

Resource Effects — Indicator variable linear regression revealed a positive relationship between
260 respiration and production rates (Fig. 5, $F_{9,42} = 4.92$, $R^2 = 0.51$, $p < 0.001$) with there being a
higher y-intercept for the high-BGE group of isolates (Table 1). Resource identity had no effect
262 on the BR-BP relationship, but we did find a significant interaction between the resource
protocatechuate and group (Table 1). Last, we did not find evidence that the slope of the BR-BP
264 relationship was different from one ($t_{42} = -0.26$, $p = 0.79$) suggesting that the two measures of
bacterial metabolism scale proportionately with one another.

266

DISCUSSION

268 We measured bacterial growth efficiency (BGE) in 20 environmental bacterial isolates on three
resources that varied in their bioavailability, structure, and pathways required for metabolism.
270 While BGE varied among strains, phylogenetic relatedness did not statistically explain
differences in efficiency. Instead, a substantial amount (20 %) of the variation in BGE could be
272 explained by an isolate's taxonomic order while a much smaller amount of the variation (7 %)
could be attributed to resource identity (Fig. 3). We found evidence for a predicted trade-off
274 between maximum growth rate and BGE, but only on the most labile resource (glucose) (Fig. 3).

Even though we found that resource identity explained 63% of the variation in BGE within an
276 isolate, resource identity did not alter the relationship between respiration and production rate
observed across isolates (Fig. 5) suggesting that resource identity has a stronger effect on BGE
278 within a species than across species. Together, we propose that growth efficiency is a
physiological trait independent of resource identity, but resource characteristics may modify
280 species-specific physiological performances. We propose that taxonomic groups of bacteria may
have fundamentally different growth efficiencies such that changes in community composition
282 may alter the fate of carbon resources (i.e., biomass versus CO₂) within the ecosystem.

284 **Bacterial Growth Efficiency as a Trait**

Our results indicate that there are species-specific properties regulating BGE, which may be
286 conserved at higher taxonomic levels. This conclusion is consistent with the view that BGE
represents a complex bacterial trait (i.e., aggregate property of numerous cellular functions) with
288 ecological significance, and that different groups of bacteria have fundamentally different
strategies for carbon allocation. Our phylogenetic analyses suggest that BGE may be an over-
290 dispersed trait (at least with glucose) such that the efficiency of closely related bacteria may be
less similar than expected. One potential explanation for this pattern of over-dispersion is that
292 our culture collection lacked phylogenetic resolution within some of our taxonomic groups (e.g.,
Betaproteobacteria) or that the variation in BGE within a taxonomic group (e.g., order) may not
294 be the same across taxonomic groups. Alternatively, BGE may not be a phylogenetically
conserved trait. Though some traits such as phosphorus acquisition, photosynthesis, and
296 methanogenesis are phylogenetically conserved deep in the microbial tree of life (Martiny et al.

2006, 2013), others such as complex carbon metabolism are not (Zimmerman et al. 2013).

298 Therefore, it is possible that BGE may be similar to traits such as complex carbon metabolism
that are not deeply conserved, which appears to be common among complex traits (Martiny et al.
300 2015). Regardless, our data reveal that BGE is a complex bacterial trait that is influenced by
taxonomic affiliation. As such, it may be possible to make predictions about BGE and other
302 ecosystem functions given information about composition of resident microbiomes (Goberna and
Verdú 2016).

304

Bacterial Growth Efficiency on Different Resources

306 Differences in resource complexity and the metabolic pathways required for degradation may
explain species-specific differences in BGE due to resource identity. Within an isolate, resource
308 identity accounted for 63 % of the variation in BGE. Given that different resources are processed
via different metabolic pathways, resource-based variation in BGE within a species is expected.
310 For example, BGE was higher when isolates were supplied with glucose compared to when they
were supplied with protocatechuate. Glucose is a simple sugar that is able to be metabolized by
312 numerous pathways and converted to acetyl-CoA (White et al. 2012). Protocatechuate, on the
other hand, is a complex aromatic compound that requires a specific metabolic pathway to be
314 converted to acetyl-CoA. Furthermore, because protocatechuate is chemically more complex, it
requires more energy (i.e., ATP) to be degraded than more labile resources such as glucose
316 (Harwood and Parales 1996). Therefore, resource complexity and the metabolic pathways
required may explain the within-isolate variation in BGE. However, across isolates, we did not
318 find resource-specific differences in the relationship between respiration and production rate (Fig

5, Table 1). Such findings suggest that energetic demands required to use different resources may
320 also be species-specific trait. That is, the energetic demands for individual species may be highly
constrained and therefore not change much when growing on different resources. Together, these
322 findings suggest that the effect of resources on the efficiency of entire microbiomes may depend
on the composition of bacteria consuming those resources.

324

Bacterial Growth Efficiency Groups

326 Across all isolates, we found a bimodal distribution of BGE suggesting that there were two
distinct groups with contrasting efficiencies. One group had low BGE (<5 %) across all
328 treatments, and the other group ranged in BGE from 7-30 % (Fig. 4 & 5). Although, the range of
BGE measured across isolates is similar to the range observed in many ecosystems (del Giorgio
330 and Cole 1998), our results suggest that some species of bacteria grow relatively inefficiently,
irrespective of resource quality. One explanation is that the minimum cellular energetic demand
332 (i.e., cellular maintenance costs) is higher in some bacteria than others (Russell and Cook 1995).
Furthermore, energetic demand may be higher when bacteria are grown in minimal media where
334 they must produce all cellular components from a single carbon resource (Tao et al. 1999).
Alternatively, nutrient concentrations (e.g., phosphorus) and other physical properties (e.g.,
336 temperature) may regulate efficiency (Smith and Prairie 2004) and the effects of these properties
may be species-specific. As such, it is possible that maintenance costs, resource imbalances, and
338 the physical growth conditions affected BGE of our isolates. Differences in low-BGE and high-
BGE isolates was also reflection in genomic content, including the number and presence-absence
340 of metabolic pathways. Together, these findings suggest that there are fundamental differences

between bacterial species that determine BGE, which can be predicted based on genomic
342 content.

344 **Physiological Trade-Offs**

We found evidence to support a predicted trade-off between maximum growth rate and BGE
346 (Fig. 4), which is predicted across microbial and non-microbial systems (Glasser 1984; Roller
and Schmidt 2015). Theoretical models of microbial communities predict a rate-efficiency trade-
348 off (Allison 2014), and this trade-off has been observed across microbial taxa (Lipson 2015).
Physiologically, the trade-off is based on allocation constraints imposed by the balance between
350 energy requirements and biomass yield: organisms with higher maximum growth rates may have
more energetic requirements and thus lower BGE (Russell and Baldwin 1979; Russell and Cook
352 1995). Furthermore, processes that limit respiration, such as oxygen availability, have been
shown to suppress bacterial growth rate (Meyenburg and Andersen 1980). Therefore, respiration
354 rate is likely a major control on biomass production and BGE. Consistent with this, we observed
a power-law relationship between respiration and production rates (Fig. 5; Table 1).
356 Furthermore, this relationship between respiration and production and the non-zero intercept
suggest that there is a minimum respiration rate required before any biomass can be produced,
358 which is commonly interpreted as the cellular maintenance requirement. Therefore, it is possible
that the maintenance energy demand of bacterial species explains the physiological trade-off
360 between maximum growth rate and growth efficiency.

Theory also predicts a trade-off between resource niche-breadth and growth efficiency
362 (Glasser 1984). This trade-off is based on the assumption that there is an energetic cost to

maintaining numerous metabolic pathways (Johnson et al. 2012). As such, species with more
364 metabolic pathways should have more energetic requirements and thus lower BGE; although, the
effects of genome reduction has been debated (Giovannoni et al. 2005; Livermore et al. 2014). In
366 this study, we did not find evidence of a trade-off between resource niche breadth and BGE (Fig.
4). One possible explanation is that the resources used in our phenotypic assay (i.e. Ecolog
368 plates) did not reflect the full metabolic potential of our isolates. Alternatively, there may not be
a strong trade-off between niche breadth and efficiency, but further experiments with additional
370 isolates and resources would be required to test this prediction more rigorously.

372 **Genomic Signatures**

In addition to the physiological differences documented among our isolates, we found genomic
374 evidence of metabolic pathways that are associated with BGE. Specifically, we found genomic
differences between isolates that belong to low-BGE and high-BGE groups. We discovered that
376 isolates in the high-BGE group had 13 % more metabolic pathways than the low-BGE group.
Furthermore, we identified particular pathways that were unique to each group (Table 2).
378 Together, our findings suggest that there are genomic features that may contribute to or regulate
BGE.

380 In general, the genomic composition of BGE groups appear to reflect differences in
cellular biosynthesis. It is possible that species with particular biosynthesis pathways may
382 generate essential cellular components with less energetic demand. For example, the low-BGE
isolates lacked some metabolic pathways, including pyridoxal biosynthesis and histidine
384 degradation, which were present in the high-BGE group. The pyridoxal biosynthesis pathway

386 produces vitamin B₆ from erythrose-4-phosphate (Mukherjee et al. 2011). Because vitamin B₆ is
essential for growth, the isolates lacking the pyridoxal pathway will use alternatives such as
uptake from the environment if they are auxotrophic (i.e., unable to synthesize) or other
388 synthesis pathways such as the Deoxyxylulose-5-phosphate synthase (DXS) pathway
(Mukherjee et al. 2011). However, the DXS pathway requires pyruvate (a precursor for Krebs
390 cycle) and thus may limit central metabolism and possibly lead to lower BGE. Likewise, the
histidine degradation pathway is used to breakdown histidine into ammonium and glutamate
392 (Bender 2012). Alternatively, glutamate can be synthesized from α -ketoglutarate; however,
because α -ketoglutarate is an intermediate component of Krebs cycle this may limit central
394 metabolism and possibly lead to reduced BGE.

396 **Conclusion**

At the cellular level, BGE reflects the balancing energetic and cellular growth demands. We find
398 evidence of this based on physiological trade-offs (i.e., maximum growth rate) as well as
metabolic pathways. As such, changes in community composition and resource availability have
400 to potential to alter food web and ecosystem function due to changes in BGE. For example,
communities dominated by species with low BGE should yield a net release of CO₂ from the
402 ecosystem. Alternatively, communities comprised of individuals with high BGE should yield a
net increase in ecosystem productivity. However, variation in BGE can arise within a species due
404 to the ways in which it processes different resources. Therefore, changes in the resource supply
will alter the performance of individual taxa, but we predict that these changes will not be as
406 strong as changes in BGE that arise owing to differences in community composition. A trait-

based approach can be used to provide a mechanistic link between the structure and function of

408 bacterial communities.

410

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REFERENCES

- 416
Allison, S. D. 2014. Modeling adaptation of carbon use efficiency in microbial communities.
418 *Frontiers in Microbiology* 5:1–9.
- Apple, J. K., and P. A. del Giorgio. 2007. Organic substrate quality as the link between
420 bacterioplankton carbon demand and growth efficiency in a temperate salt-marsh estuary. *The*
ISME Journal 1:729–742.
- 422 Bender, R. A. 2012. Regulation of the histidine utilization (Hut) system in bacteria.
Microbiology and Molecular Biology Reviews 76:565–584.
- 424 Blomberg, S. P., T. Garland, and A. R. Ives. 2003. Testing for phylogenetic signal in
comparative data: behavioral traits are more labile. *Evolution* 57:717–745.
- 426 Buchan, A., L. S. Collier, and E. L. Neidle. 2000. Key aromatic-ring-cleaving enzyme,
protocatechuate 3,4-dioxygenase, in the ecologically important marine *Roseobacter* lineage.
428 *Applied and Environmental Microbiology* 66:4662–4672.
- Cole, J. R., Q. Wang, E. Cardenas, J. Fish, B. Chai, R. J. Farris, A. S. Kulam-Syed-Mohideen, et
430 al. 2009. The Ribosomal Database Project: improved alignments and new tools for rRNA
analysis. *Nucleic Acids Research* 37:D141–D145.

- 432 Colwell, R. K., and D. J. Futuyma. 1971. On the Measurement of Niche Breadth and Overlap.
Ecology 52:567–576.
- 434 del Giorgio, P. A., and J. J. Cole. 1998. Bacterial growth efficiency in natural aquatic systems.
Annual Review of Ecology and Systematics 29:503–541.
- 436 del Giorgio, P. A., and R. E. I. Newell. 2012. Phosphorus and DOC availability influence the
partitioning between bacterioplankton production and respiration in tidal marsh ecosystems.
438 Environmental Microbiology 14:1296–307.
- Ducklow, H. W. 2008. Bacterial Production and Biomass in the Oceans. Pages 1–47 in Microbial
440 Ecology of the Oceans.
- Dufrene, M., and P. Legendre. 1997. Species assemblages and indicator species: the need for a
442 flexible asymmetrical approach. Ecological Monographs 67:345–366.
- Dykhuizen, D., and M. Davies. 1980. An Experimental Model: Bacterial Specialists and
444 Generalists Competing in Chemostats. Ecology 61:1213–1227.
- Felsenstein, J. 2008. CONSENSE: Consensus tree program.
- 446 Flamholz, A., E. Noor, A. Bar-Even, W. Liebermeister, and R. Milo. 2013. Glycolytic strategy as
a tradeoff between energy yield and protein cost. Proceedings of the National Academy of
448 Sciences 110:10039–44.
- Fuhrer, T., E. Fischer, and U. Sauer. 2005. Experimental identification and quantification of
450 glucose metabolism in seven bacterial species. Journal of Bacteriology 187:1581–1590.
- Garland, J. L., and A. L. Mills. 1991. Classification and characterization of heterotrophic

- 452 microbial communities on the basis of patterns of community-level sole-carbon-source
utilization. *Applied and Environmental Microbiology* 57:2351–9.
- 454 Ghosh, D., K. Roy, V. Srinivasan, T. Mueller, O. H. Tuovinen, K. Sublette, A. Peacock, et al.
2009. In-situ enrichment and analysis of atrazine-degrading microbial communities using
456 atrazine-containing porous beads. *Soil Biology and Biochemistry* 41:1331–1334.
- Giovannoni, S. J., H. J. Tripp, S. Givan, M. Podar, K. L. Vergin, D. Baptista, L. Bibbs, et al.
458 2005. Genome streamlining in a cosmopolitan oceanic bacterium. *Science* 309:1242–5.
- Gladman, S., and T. Seemann. 2012. Velvet Optimizer.
- 460 Glasser, J. W. J. 1984. Evolution of efficiencies and strategies of resource exploitation. *Ecology*
65:1570–1578.
- 462 Goberna, M., and M. Verdú. 2016. Predicting microbial traits with phylogenies. *The ISME*
Journal 10:959–967.
- 464 Gottschalk, G. 1986. *Bacterial Metabolism* (2nd ed.). Springer-Verlag, New York, NY.
- Hall, E. K., and J. B. Cotner. 2007. Interactive effect of temperature and resources on carbon
466 cycling by freshwater bacterioplankton communities. *Aquatic Microbial Ecology* 49:35–45.
- Hannon Lab. 2010. FASTX Toolkit.
- 468 Hartigan, J. A., and P. M. Hartigan. 1985. The dip test of unimodality. *The Annals of Statistics*
13:70–84.
- 470 Harwood, C. S., and R. E. Parales. 1996. The beta-ketoadipate pathway and the biology of self-
identity. *Annual Review of Microbiology* 50:553–90.

- 472 Johnson, D. R., F. Goldschmidt, E. E. Lilja, and M. Ackermann. 2012. Metabolic specialization
and the assembly of microbial communities. *The ISME Journal* 6:1985–1991.
- 474 Kembel, S. W., D. D. Ackerly, S. P. Blomberg, W. K. Cornwell, M. R. Helmus, M. Helene, and
C. O. Webb. 2015. *Picante: R tools for integrating phylogenies and ecology*.
- 476 Legendre, P., and L. F. J. Legendre. 2012. *Numerical Ecology* (Vol. 24). Elsevier.
- Lennon, J. T. 2007. Diversity and metabolism of marine bacteria cultivated on dissolved DNA.
478 *Applied and Environmental Microbiology* 73:2799–2805.
- Lennon, J. T., Z. T. Aanderud, B. K. Lehmkuhl, and D. R. J. Schoolmaster. 2012. Mapping the
480 niche space of soil microorganisms using taxonomy and traits. *Ecology* 93:1867–1879.
- Lennon, J. T., S. K. Hamilton, M. E. Muscarella, A. S. Grandy, K. Wickings, and S. E. Jones.
482 2013. A source of terrestrial organic carbon to investigate the browning of aquatic ecosystems.
PloS One 8:e75771.
- 484 Lennon, J. T., and L. E. Pfaff. 2005. Source and supply of terrestrial organic matter affects
aquatic microbial metabolism. *Aquatic Microbial Ecology* 39:107–119.
- 486 Lipson, D. A. 2015. The complex relationship between microbial growth rate and yield and its
implications for ecosystem processes. *Frontiers in Microbiology* 6:1–5.
- 488 Litchman, E., K. F. Edwards, and C. A. Klausmeier. 2015. Microbial resource utilization traits
and trade-offs: implications for community structure, functioning, and biogeochemical impacts at
490 present and in the future. *Frontiers in Microbiology* 06:1–10.
- Livermore, J. A., S. J. Emrich, J. Tan, and S. E. Jones. 2014. Freshwater bacterial lifestyles

- 492 inferred from comparative genomics. *Environmental Microbiology* 16:746–58.
- Ludwig, W., O. Strunk, R. Westram, L. Richter, H. Meier, Yadhukumar, A. Buchner, et al. 2004.
- 494 ARB: a software environment for sequence data. *Nucleic Acids Research* 32:1363–71.
- Maechler, M. 2015. diptest: Hartigan’s dip test statistic for unimodality.
- 496 Martin, M. 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads. *EMBnet.journal* 17:10.
- 498 Martiny, A. C., M. L. Coleman, and S. W. Chisholm. 2006. Phosphate acquisition genes in *Prochlorococcus* ecotypes: evidence for genome-wide adaptation. *Proceedings of the National*
- 500 *Academy of Sciences* 103:12552–7.
- Martiny, A. C., K. Treseder, and G. Pusch. 2013. Phylogenetic conservatism of functional traits
- 502 in microorganisms. *The ISME Journal* 7:830–838.
- Martiny, J. B. H., S. E. Jones, J. T. Lennon, and A. C. Martiny. 2015. Microbiomes in light of
- 504 traits: A phylogenetic perspective. *Science (New York, N.Y.)* 350:aac9323-aac9323.
- McDonald, E., and C. T. Brown. 2013. Khmer: working with big data in bioinformatics. *arXiv*
- 506 1303.2223:1–18.
- Meyenburg, K. V. O. N., and K. B. Andersen. 1980. Are growth rates of *Escherichia coli* in
- 508 batch cultures limited by respiration? *Journal of Bacteriology* 144:114–123.
- Miller, M. A., W. Pfeiffer, and T. Schwartz. 2010. Creating the CIPRES Science Gateway for
- 510 inference of large phylogenetic trees. 2010 Gateway Computing Environments Workshop.
- Mukherjee, T., J. Hanes, I. Tews, S. E. Ealick, and T. P. Begley. 2011. Pyridoxal phosphate:

- 512 biosynthesis and catabolism. *Biochimica et Biophysica Acta - Proteins and Proteomics*
1814:1585–1596.
- 514 Oksanen, J., F. G. Blanchet, R. Kindt, P. Legendre, P. R. Minchin, R. B. O’Hara, G. L. Simpson,
et al. 2013. *Vegan: community ecology package*.
- 516 Pinheiro, J., D. Bates, D. Saikat, D. Sarkar, S. Heiskterkamp, B. Van Willigen, and R.-C. D.
Team. 2016. *nlme: linear and nonlinear mixed effects models*.
- 518 Pomeroy, L. R., P. J. Williams, F. Azam, and J. E. Hobbie. 1998. The microbial loop.
Oceanography 20:28–33.
- 520 Pruesse, E., J. Peplies, and F. O. Glöckner. 2012. SINA: accurate high-throughput multiple
sequence alignment of ribosomal RNA genes. *Bioinformatics* 28:1823–1829.
- 522 R Core Team. 2012. *R: A language and environment for statistical computing*.
- Roller, B. R., and T. M. Schmidt. 2015. The physiology and ecological implications of efficient
524 growth. *The ISME Journal* 9:1481–1487.
- Russell, J. B., and R. L. Baldwin. 1979. Comparison of maintenance energy expenditures and
526 growth yields among several rumen bacteria grown on continuous culture. *Applied and
Environmental Microbiology* 37:537–543.
- 528 Russell, J. B., and G. M. Cook. 1995. Energetics of bacterial growth: balance of anabolic and
catabolic reactions. *Microbiological Reviews* 59:48–62.
- 530 Schloss, P. D., S. L. Westcott, T. Ryabin, J. R. Hall, M. Hartmann, E. B. Hollister, R. A.
Lesniewski, et al. 2009. Introducing mothur: open-source, platform-independent, community-

- 532 supported software for describing and comparing microbial communities. *Applied and Environmental Microbiology* 75:7537–7541.
- 534 Seemann, T. 2014. Prokka: rapid prokaryotic genome annotation. *Bioinformatics* 30:2068–9.
- Sinsabaugh, R. L., S. Manzoni, D. L. Moorhead, and A. Richter. 2013. Carbon use efficiency of
536 microbial communities: stoichiometry, methodology and modelling. (J. Elser, ed.) *Ecology Letters* 16:930–939.
- 538 Smith, D. C., and F. Azam. 1992. A simple, economical method for measuring bacterial protein synthesis rates in seawater using 3H-leucine. *Marine Microbial Food Webs* 6:107–114.
- 540 Smith, E. M., and Y. T. Prairie. 2004. Bacterial metabolism and growth efficiency in lakes: the importance of phosphorus availability. *Limnology and Oceanography* 49:137–147.
- 542 Stamatakis, A. 2006. RAxML-VI-HPC: maximum likelihood-based phylogenetic analyses with thousands of taxa and mixed models. *Bioinformatics* 22:2688–2690.
- 544 Stamatakis, A., P. Hoover, and J. Rougemont. 2008. A rapid bootstrap algorithm for the RAxML web servers. *Systematic Biology* 57:758–71.
- 546 Takami, H., T. Taniguchi, Y. Moriya, T. Kuwahara, M. Kanehisa, and S. Goto. 2012. Evaluation method for the potential functionome harbored in the genome and metagenome. *BMC Genomics*
548 13:699.
- Tao, H., C. Bausch, C. Richmond, F. R. Blattner, and T. Conway. 1999. Functional genomics: expression analysis of *Escherichia coli* growing on minimal and rich media. *Journal of Bacteriology* 181:6425–6440.
- 550

- 552 Tempest, D. W., and O. M. Neijssel. 1984. The status of YATP and maintenance energy as
biologically interpretable phenomena. *Annual Reviews in Microbiology* 459–486.
- 554 White, D., J. Drummond, and C. Fuqua. 2012. *The Physiology and Biochemistry of Prokaryotes*
(4th ed.). Oxford University Press, New York, NY.
- 556 Yilmaz, P., L. W. Parfrey, P. Yarza, J. Gerken, E. Pruesse, C. Quast, T. Schweer, et al. 2014. The
SILVA and “All-species Living Tree Project (LTP)” taxonomic frameworks. *Nucleic Acids*
558 *Research* 42:643–648.
- Zerbino, D. R., and E. Birney. 2008. Velvet: algorithms for de novo short read assembly using de
560 Bruijn graphs. *Genome Research* 18:821–9.
- Zimmerman, A. E., A. C. Martiny, and S. D. Allison. 2013. Microdiversity of extracellular
562 enzyme genes among sequenced prokaryotic genomes. *The ISME Journal* 7:1187–1199.
- Zwietering, M. H., I. Jongenburger, F. M. Rombouts, and K. van’t Riet. 1990. Modeling of the
564 bacterial growth curve. *Applied and Environmental Microbiology* 56:1875–1881.
- 566

TABLES

568 **Table 1:** Indicator variable linear regression coefficients for model testing if respiration rate (i.e.,
BR), resource identity, and group (high BGE or low BGE) explain differences in production rate
570 (i.e. BP). Glucose and low BGE are used as baseline factors. Resp = respiration rate, Group =
high BGE group. Suc = succinate. Pro = protocatechuate. Model includes all two-way interaction
572 terms.

	Estimate	Std. Error	<i>t</i>-value	<i>p</i>-value
Intercept	-2.62	0.63	-4.13	0.0002
Resp	0.90	0.37	2.43	0.02
Suc	1.43	0.85	1.68	0.10
Pro	2.04	1.32	1.54	0.13
Group	1.86	0.77	2.42	0.02
Resp * Suc	-0.78	0.50	-1.57	0.12
Resp * Pro	0.52	0.62	-0.84	0.41
Resp * Group	0.18	0.49	0.36	0.72
Suc * Group	-0.45	0.70	-0.63	0.53
Pro * Group	-2.16	0.86	-2.51	0.02

576 **Table 2:** Genetic pathways unique to the high BGE isolates. Pathways are functional metabolic
pathways identified from genome sequencing and predicted using Maple. Prob. = probability
578 statistic from indicator species analysis: the probability that the “species” (i.e., pathway), is not
unique to the group.

Group	Prob.	Pathway	Reference Function
	0.01	M00124	Pyridoxal biosynthesis, erythrose-4P → pyridoxal-5P
High BGE	0.03	M00045	Histidine degradation, histidine → N-formiminoglutamate → glutamate
	0.03	M00565	Trehalose biosynthesis, D-glucose-1P → trehalose

580

582

FIGURE LEGENDS

584 **Fig. 1:** Maximum likelihood phylogenetic tree of lake bacterial isolates used to study BGE
variation. Nearest relatives and other type-strains are included as taxonomic references. Isolates
586 are organized by and labeled with taxonomic class. The outgroup (*Aquifex*) is included as the tree
root. Scale bar represents 0.01 base substitutions.

588 **Fig. 2:** Carbon resources used to study BGE variation in environmental isolates. **A:** Glucose –
the baseline resource used to compare BGE across isolates. Glucose can be degraded by the
590 Embden-Meyerhof-Parnas, pentose phosphate, or Entner-Doudoroff pathway. Ultimately, these
pathways produce pyruvate (and then acetyl-CoA), which enters Krebs cycle and is used to
592 produce energy and intermediate for biomass synthesis, when cells are grown aerobically.
Alternatively, glucose can be fermented into organic acids (e.g., lactate), but these reactions yield
594 less energy (Gottschalk 1986). **B:** Succinate – is a simple organic acid. Succinate is an
intermediate of Krebs cycle and thus it does not require previous degradation. Additionally,
596 succinate can be used directly to produce energy via succinate dehydrogenase (White et al.
2012). **C:** Protocatechuate – is a complex resource with a aromatic core. Typically, it is degraded
598 to acetyl-CoA and Succinyl-CoA via the β -keto adipate pathway (Harwood and Parales 1996).
Protocatechuate is commonly used to study aromatic resource degradation in ecosystems, and the
600 β -keto adipate pathway is commonly found in bacteria across the phylum Proteobacteria (Buchan
et al. 2000).

602

604 **Fig. 3:** Bacterial growth efficiency (BGE) of each isolate for each resource. BGE was calculated
based on measured production (BP) and respiration (BR) rates using the following equation:
606 $BGE = BP / (BP + BR)$. Cladogram is based on the consensus phylogeny. Taxonomic class and
order are included based on the ribosomal database taxonomy: α = Alphaproteobacteria, β =
608 Betaproteobacteria, γ = Gammaproteobacteria, *Xan.* = Xanthomonadales, *Aero.* = Aeromonadales,
Pseudo. = Pseudomonadales.

610

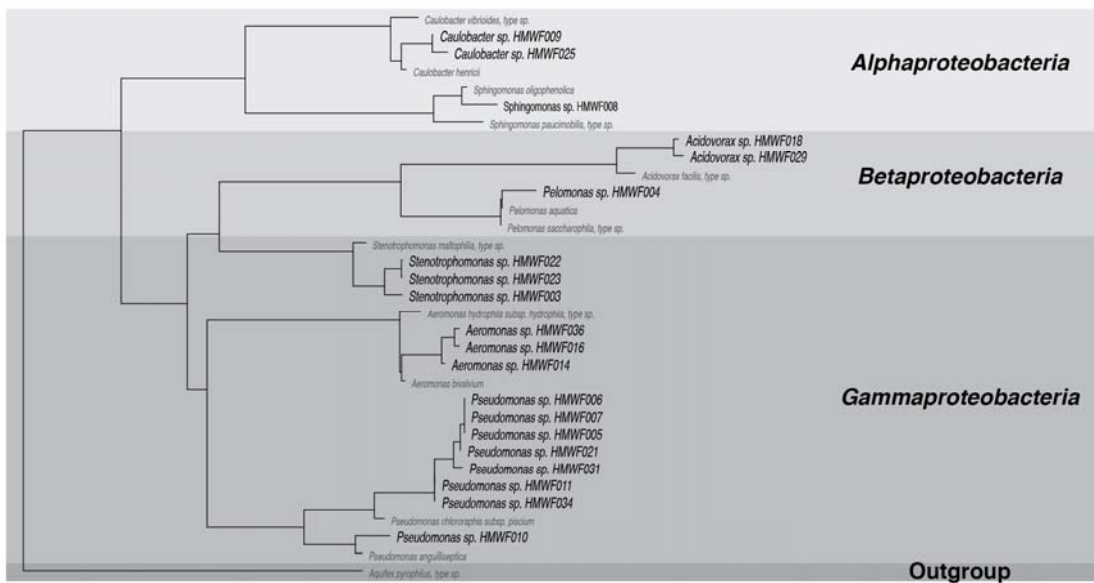
Fig. 4: Phenotypic traits associated with BGE. A: Maximum growth rate, a measure of growth
612 strategy, demonstrates a trade-off (negative relationship) with BGE in the high BGE group ($F_{1,7}$
 $= 9.52$, $r^2 = 0.54$, $p = 0.015$), but not the low BGE group ($F_{1,7} = 0.51$, $r^2 = 0.06$, $p = 0.50$). B:
614 Levin's Index, a measure of niche breadth, does not demonstrate a trade-off with BGE in either
the high or low BGE groups (high: $F_{1,7} = 0.92$, $r^2 = 0.11$, $p = 0.37$; low: $F_{1,7} = 1.47$, $r^2 = 0.17$, $p =$
616 0.27). High and low BGE groups were determined based on bimodal distribution of BGE.

618 **Fig. 5:** Relationship between respiration and production rates. Respiration and production rates
were compared using an indicator variable linear regression ($F_{9,42} = 4.92$, $R^2 = 0.51$, $p < 0.001$).
620 According to the regression model, production rate increases with respiration rate proportionally
(i.e., slope not significantly different from one, $t_{42} = -0.26$, $p = 0.79$). In addition, group (high vs.
622 low BGE) was a significant factor and isolates in the high BGE group had a greater y-intercept
($p = 0.02$). Symbols indicates isolate group (high and low BGE), and color indicates the resource
624 being used. Symbol size is scaled by growth efficiency.

626

FIGURES

Fig. 1:

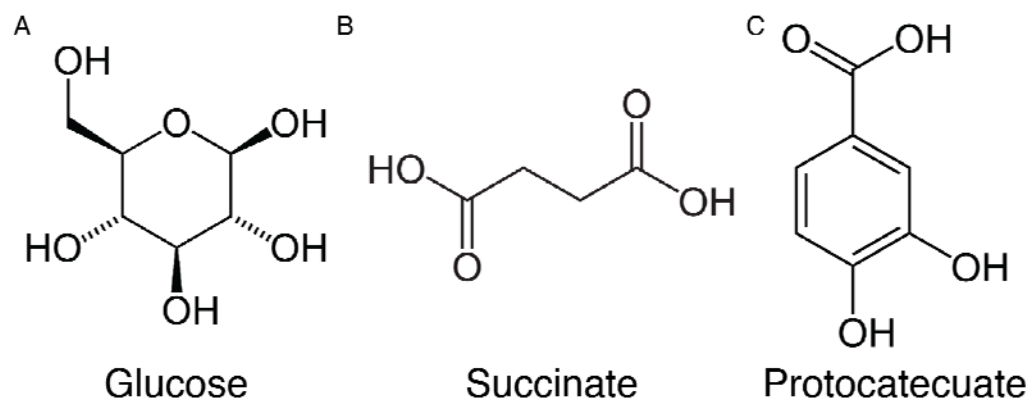


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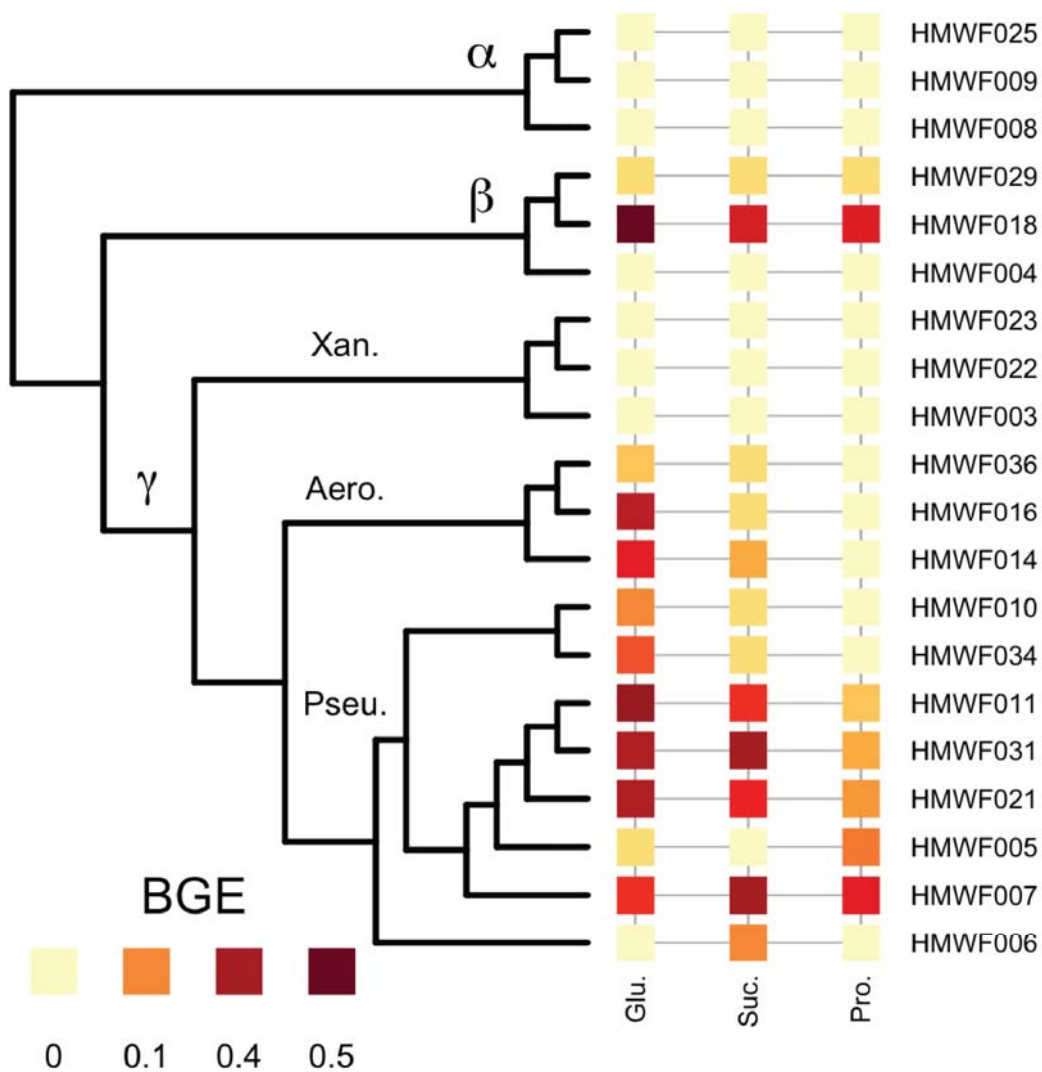
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Fig. 2:



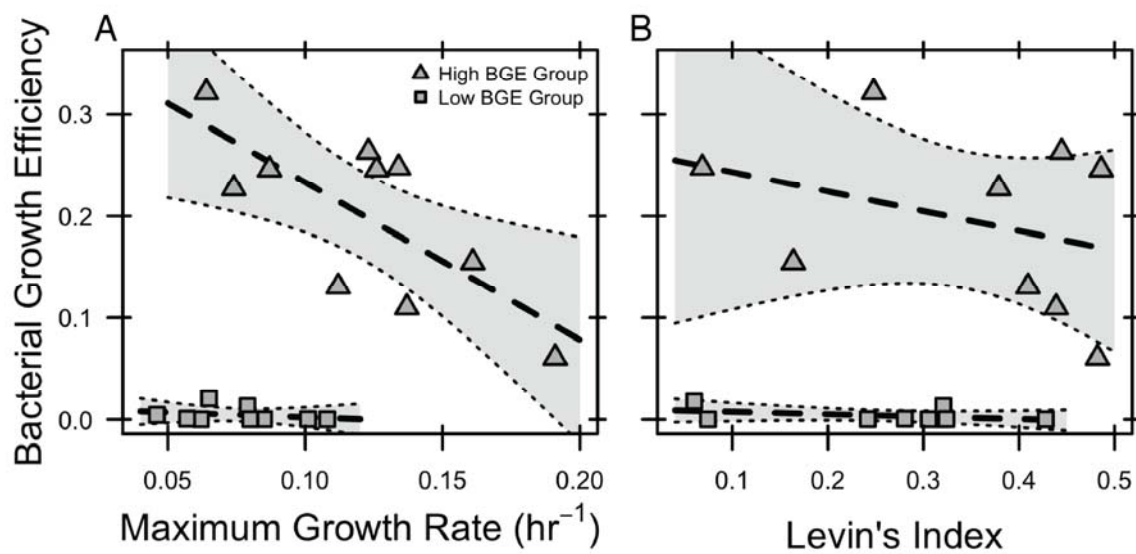
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Fig. 3:



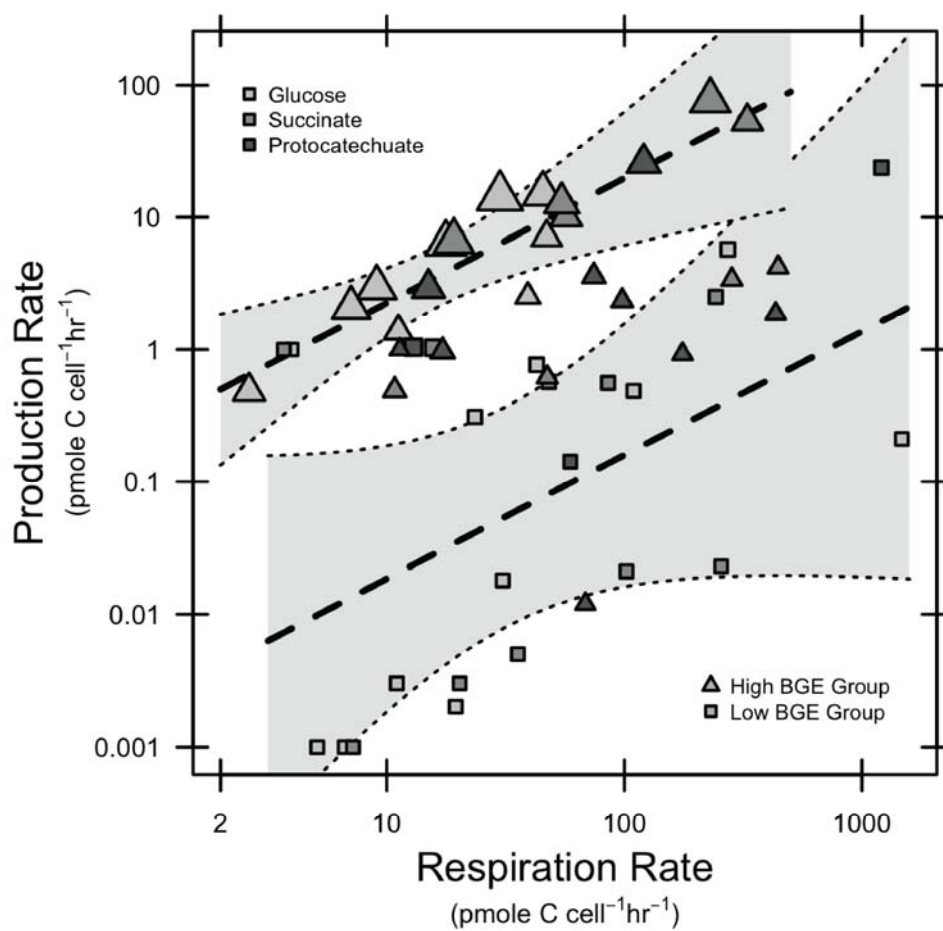
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638 **Fig. 4:**



640

Fig. 5:



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