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Title: Species limits in butterflies (Lepidoptera: Nymphalidae): Reconciling classical taxonomy

with the multispecies coalescent

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Abstract.— Species delimitation is at the core of biological sciences. During the last decade, molecular-based approaches have advanced the field by providing additional sources of evidence to classical, morphology-based taxonomy. However, taxonomy has not yet fully embraced molecular species delimitation beyond threshold-based, single-gene approaches, and taxonomic knowledge is not commonly integrated to multi-locus species delimitation models. Here we aim to bridge empirical data (taxonomic and genetic) with recently developed coalescent-based species delimitation approaches. We use the multispecies coalescent model as implemented in two Bayesian methods (DISSECT/STACEY and BP&P) to infer species hypotheses. In both cases, we account for phylogenetic uncertainty (by not using any guide tree) and taxonomic uncertainty (by measuring the impact of using or not a priori taxonomic assignment to specimens). We focus on an entire Neotropical tribe of butterflies, the Haeterini (Nymphalidae: Satyrinae). We contrast divergent taxonomic opinion—splitting, lumping and misclassifying species—in the light of different phenotypic classifications proposed to date. Our results provide a solid background for the recognition of 22 species. The synergistic approach presented here overcomes limitations in both traditional taxonomy (e.g. by recognizing cryptic species) and molecular-based methods (e.g. by recognizing structured populations, and not raise them to species). Our framework provides a step forward towards standardization and increasing reproducibility of species delimitations.

KEYWORDS: Bayes factor; clearwing Satyrinae; Haeterini; Neotropics; speciation; systematics

INTRODUCTION

Traditionally, taxonomic delimitations have relied on diagnostic phenotypic characters to classify distinct populations into species and subspecies (hereafter, the 'traditional taxonomic approach'). More recently, coalescent-based methods that quantify reproductive isolation using genetic data have been proposed as a means to calculate the probability of speciation (hereafter, the 'coalescent approach'; e.g. Knowles & Carstens, 2007; Yang & Rannala, 2010; Fujita *et al.*, 2012). Both approaches do not necessarily agree in their species hypotheses because their scopes are centered on different sections of the speciation continuum; while traditional taxonomy primarily depends on the evolution of informative and consistent phenotypic characters, the coalescent approach is guided by any gene flow reduction that could be associated with the onset of reproductive isolation. As a consequence, phenotype-based delimitation may not identify cryptic and incipient species (i.e. genetically divergent lineages embarked in the process of speciation; Rosindell *et al.*, 2010) whereas coalescent-based delimitation may simply reveal population structure (i.e. subpopulations with a long non-breeding history) (Sukumaran & Knowles, 2017). Despite the paramount importance of delimiting species for multiple disciplines and practices in science and society (e.g. ecology, evolution, conservation biology, among others), it still remains unclear how to reconcile conflicts between traditional taxonomy and the coalescent approach while taking into account their respective benefits and limitations.

Taxonomists have not fully embraced the recent developments in molecular species delimitation beyond threshold-based, single-gene approaches including DNA barcoding based on COI proposed as a fast approach to delimit species (Hebert *et al.*, 2004), and delimitation methods based on single markers (Pons *et al.*, 2006; Puillandre *et al.*, 2012). End-users of the coalescent

approach, on the other hand, do not usually incorporate taxonomic knowledge to inform their

models (e.g. Leaché & Fujita, 2010; Olave *et al.*, 2014; but see Aydin *et al.*, 2014; Jones *et al.*,

2015). In a Bayesian framework, taxonomic information could be explicitly acknowledged in the

form of prior distributions, and thus alternative species hypotheses can be statistically weighed.

Indeed, the latest Bayesian implementations, which include the multispecies coalescent model

(MSC; Degnan & Rosenberg, 2009), can accommodate parameters that control the number of

species in a dataset, their divergence times and ancestral population sizes, all in a single

probabilistic framework. These properties among others, such as the recognition of genealogical

incongruence and incomplete lineage sorting, arguably make MSC-based methods more

biologically realistic than threshold-based molecular species delimitation (Knowles & Carstens,

2007; Fujita *et al.*, 2012). However, it remains unexplored how divergent taxonomic opinions

affect species delimitations, when these opinions (such as "splitters" *vs*. "lumpers") are translated

into prior distributions for molecular species delimitation analyses.

Here we aim to reconcile genetic and taxonomic data using species delimitation models that

estimate coalescence and species divergence in a fully Bayesian framework. We study the

butterflies classified in the tribe Haeterini (Nymphalidae: Satyrinae), insects that exclusively

inhabit tropical rainforests in southern Mexico, Central and South America. Transparent wings

are the most obvious characteristic of this group, an attribute shared by four out of five genera

within the tribe—*Pierella* being the exception by having full scale-cover (Fig. 1).

Taxonomic work on this charismatic group involves numerous researchers over the last 150

years (e.g. Herrich-Schäffer, 1864; Weymer, 1910; Miller, 1968; Constantino, 1995; Lamas,

1997; Penz *et al.*, 2014; Paluch *et al.*, 2015; Willmott, 2015; Zacca *et al.*, 2016). Haeterini is a

monophyletic tribe (Wahlberg *et al.*, 2009; Chazot *et al.*, 2019), and current taxonomic

understanding (Lamas, 2004; Penz *et al.*, 2014; Willmott, 2015; Zacca *et al.*, 2016) is that

Haeterini consists of 29 described species and 39 subspecies. However, these estimates remain

contentious because published taxonomic studies have mostly focused on smaller groups within

Haeterini and have relied on different morphological character systems (wing coloration,

genitalia shape or male androconial organs), and have not taken into account variation in DNA

sequence data. The taxonomic opinions among these authors reflect the well-recognized

subjectivity of species-level taxonomic work, even though Haeterini has a comprehensive

taxonomic knowledge as compared to most other tropical insects. This offers the opportunity to

quantify taxonomic opinion and definitions among authors, as a preliminary step for

standardization.

Specifically, we aim to test alternative species delimitation hypotheses and to evaluate the

following two interconnected expectations:

1. Lineage delimitation using the multispecies coalescent model (MSC) will recover the

taxonomic subspecies in Haeterini. Butterfly subspecies are commonly described based on

parapatric or allopatric geographical distributions, and reduced gene flow could facilitate

phenotypic divergence. If phenotype divergence resulted from genetic differentiation, then any

reproductive isolation among subspecies will be recovered by the MSC.

2. Taxonomic knowledge modeled as prior distributions will group specimens into MSC Clusters

(MSCC). However, regardless of prior distribution, the MSC will standardize species

delimitation because divergent taxonomic opinion among authors, if any, will be evident in the

phylogeny of Haeterini. Therefore, the MSC will inform on whether the use of different

morphological characters (wing coloration, genitalia shape or male androconial organs) might

favor narrower or broader species definitions in Haeterini butterflies.

MATERIALS AND METHODS

Taxon sampling and molecular dataset

Haeterini butterflies were collected by the authors and collaborators throughout most of the

geographical range of the tribe, including localities from northern Costa Rica to southeastern

Brazil. Specimens were identified to the species and subspecies level following published

taxonomic revisions (Constantino, 1995; Lamas, 1997; Penz *et al.*, 2014; Paluch *et al.*, 2015;

Willmott, 2015; Zacca *et al.*, 2016) and comparing wing morphology to type specimen

photographs at https://www.butterfliesofamerica.com (last accessed August 2018). We sampled

all five Haeterini genera, including 18 of 29 currently valid species, and 18 of 39 subspecies

(Lamas, 2004; Penz *et al.*, 2014; Paluch *et al.*, 2015; Willmott, 2015; Zacca *et al.*, 2016). Note

that several subspecies were not represented in our sample because they are only known from

type collections (hampering their access for genetic studies) or have rarely been collected in

recent years.

We followed standard lab protocols (Wahlberg & Wheat, 2008) to sequence 6 gene fragments

from 63 specimens: the mitochondrial locus COI (1,475 bp) and the nuclear loci CAD (850 bp),

EF1α (1,240 bp), GAPDH (691 bp), RpS5 (617 bp), and *wingless* (400 bp). Sanger sequencing

was conducted by the company Macrogen (South Korea), and sequence quality control and DNA

alignments were carried out using the program Geneious R7. We retrieved from GenBank the

DNA sequences of sixteen species classified in the subfamily Satyrinae in order to root the

phylogeny of Haeterini. All DNA sequences were deposited in GenBank (accession numbers

MH802134–MH802346, see Table 1) and the DNA alignments used in this study were archived

in TreeBASE (ID 23439). Detailed specimen voucher information can be found in Table 1.

Phylogenetic analyses

We inferred phylogenies using single-gene datasets partitioned by codon position to rule out any

tree topology conflict due to contamination. In addition, in order to test the monophyly of each

genus, we inferred a phylogeny using the concatenated multi-locus dataset consisting of 6

genetic markers that proved to be phylogenetically informative (Table 2), and 63 Haeterini

specimens and 16 outgroup taxa. We used PartitionFinder v2.1.1 (Lanfear *et al.*, 2017) to

estimate the best-fit partitioning strategy for the concatenated dataset using 18 data blocks (each

codon position separately for each gene region) and the following settings: branchlengths =

linked (higher likelihood than the unlinked option), models = mrbayes, model_selection = bic

and search = greedy. All phylogenetic analyses using single-gene and concatenated datasets were

carried out using MrBayes v3.2.3 (Ronquist *et al.*, 2012) through the CIPRES portal (Miller *et*

al., 2010). We used the reversible-jump Markov chain Monte Carlo approach (rjMCMC) to

allow moving across nucleotide substitution schemes (nst = mixed) with different rate variation

across sites (+I and +Γ). This approach integrated the uncertainty of substitution models that fit

the data with the inference of tree topology and branch lengths. Two independent analyses and

four chains, one cold and three heated, were run for 10 million cycles and sampled every 1,000

cycles, discarding the first 25% sampled parameters as burn-in. We evaluated convergence using

the average standard deviation of split frequencies (< 0.005), potential scale reduction factor ($\sim$

1.000), estimated sample sizes (ESS > 200), and by inspection of stationary distribution of log-

probabilities in both independent runs. We summarized the 7,500 sampled trees using the 50%

majority-rule consensus method.

Molecular species delimitation

Once we had confidence on the monophyly of each genus and a notion of the phylogenetic

relationships among the 63 specimens, we carried out molecular species delimitation under the

MSC framework. Although the inference of phylogenetic relationships is not a pre-requisite for

subsequent delimitation analyses, it becomes informative when large phylogenies need to be

divided into smaller well-supported subclades, and analyzed separately due to computational

limitation (see BP&P analyses below). We used a comprehensive taxon sampling of Haeterini, a

multi-locus dataset, and two popular Bayesian implementations of the MSC, namely STACEY

(Jones, 2017) and BP&P (Yang, 2015):

STACEY.—We used DISSECT (Jones *et al.*, 2015) which is a taxonomic assignment-free

Bayesian method for grouping individuals into multispecies coalescent clusters (MSCC). The

method is implemented in the STACEY v1.2.4 package (Jones, 2017) available in BEAST v2.4.7

(Bouckaert *et al.*, 2014). All gene markers sequenced for this study are likely unlinked in the

genome of the butterflies and thus gene trees, substitution and clock models were all treated as

unlinked in the analysis. We assigned to the mitochondrial COI locus a gene ploidy of 0.5 and to

the remaining nuclear loci a gene ploidy of 2.0 (diploid). Uncorrelated relaxed-clock models

were chosen for all loci, and we estimated nuclear clock rates relative to the COI mean clock rate

fixed to 1.0. The relative clock mean priors were all log normal (M = 0, S = 1). We used the

birth-death-collapse model following Jones *et al.* (2015) with GrowthRate prior as log normal

(M = 5, S = 2) and relative DeathRate as uniform in [0, 1], while the popMean prior was set to

log normal (M = −7, S = 2).

In STACEY, the discovery of MSCCs relies on two parameters that control node collapsing in

the phylogeny, the collapseHeight (ε) and collapseWeight (ω). The parameter ε distinguishes

very shallow species divergences (node heights) and should be assigned a small value (Jones *et*

al., 2015), thus, we set ε to $1e{-}4$. The parameter ω controls the number of MSCCs and can be

used as a proxy for prior taxonomic knowledge. The 1/X distribution of the prior for the number

of MSCCs has a mean of $1 + (n - 1) \times (1 - \omega)$, where *n* is the number of individuals in the

dataset. We set ω to 0.73 or 0.59, corresponding to 18 described species and 26 taxa (the 18

sampled subspecies elevated to species), respectively. In addition, we carried out a third analysis

that did not take into account prior taxonomic information on the number of species by using a

Beta distribution (α = 2, β= 2) as a prior for the parameter ω. The analyses were run four

independent times for 200 million cycles each, with parameters sampled every 20,000 cycles.

Sampled trees were combined after discarding the first 25% samples as burn-in and checking that

ESS values were > 200. MSCCs, their posterior probabilities and pairwise similarity

probabilities were obtained using SpeciesDelimitationAnalyser v1.8.0 (Jones *et al.*, 2015)

acknowledging $\varepsilon = 1e{-}4$. We chose the clustering with the highest posterior probability (counts)

as the working species hypothesis.

BP&P.—We used the multispecies coalescent model as implemented in BP&P v3.4 (Yang, 2015) to jointly infer species trees and delimit MSCCs, without using any guide tree (Yang & Rannala, 2014; Rannala & Yang, 2017) nor taxonomic assignment to specimens (Yang & Rannala, 2010; Rannala & Yang, 2013). The heredity scalar was set to 1 for all nuclear loci and to 0.25 for the mitochondrial locus. The prior for ancestral population sizes controlled by the parameter θ was assigned the inverse gamma distribution (IG[α, β]), and we evaluated three different scenarios: i) large ancestral population size (IG[3, 0.2]), ii) medium ancestral population size IG[3, 0.1], and iii) small ancestral population size IG[3, 0.02]. These three different scenarios are expected to impact the inferred number of MSCCs, so that a larger ancestral population prior would favor fewer species in the model. Leaché and Fujita (2010) showed that scenario (i) is a conservative combination of parameters, because it favors fewer species; as a counterpart, we have used scenario (iii) with prior mean an order of magnitude lower than in scenario (i). We separately analyzed two Haeterini subclades, namely i) *Pierella* and ii) the remaining genera, to overcome computational limitation in BP&P. We used the inverse gamma distribution for the divergence time of the root in the species trees, which is controlled by the parameter τ_0. The prior distribution of τ_0 was diffuse ($\alpha = 3$) and we specified β = 0.042 for the *Pierella* species tree and β = 0.098 for the remaining Haeterini genera. These values enforced a sequence divergence mean of 2.1% for *Pierella* and 4.9% for the remaining genera, which translate into absolute times of ~7 Ma for *Pierella*'s crown age and ~17 Ma for the remaining genera, assuming a butterfly mutation rate of 2.9×10^{-9} (Keightley *et al.*, 2015). The analyses were run two independent times using the rjMCMC algorithm 1 with gamma variable fine-tuning shape $\alpha = 2$ and mean $m = 1$, each for 500,000 cycles with sampling frequency of 50

and burn-in of the first 10,000 cycles. We chose the species delimitation model with the highest

posterior probability as the working species hypothesis.

Species hypothesis testing and divergence time estimation

We compared the statistical support (model adequacy) for eight species hypotheses in a fully

Bayesian framework:

i) Taxonomic species (18 lineages);

ii) Taxonomic subspecies elevated to species (26 lineages);

iii) STACEY's clusters under $\omega = 0.73$ (accounting for number of taxonomic species; 22

lineages);

iv) STACEY's clusters under $\omega = 0.59$ (accounting for number of subspecies elevated to species;

24 lineages);

v) STACEY's clusters under $\omega = $ Beta ($\alpha = 2$, $\beta = 2$) (non-informative prior on the number of

MSCCs; 63 lineages);

vi) BP&P's clusters under $\theta = $ IG[3, 0.2] (large ancestral population size prior; 21 lineages);

vii) BP&P's clusters under $\theta = $ IG[3, 0.1] (medium ancestral population size prior; 28 lineages);

viii) BP&P's clusters under $\theta = $ IG[3, 0.02] (small ancestral population size prior; 55 lineages).

In order to account for incomplete lineage sorting and to avoid any of the pitfalls of using

concatenated datasets (Edwards *et al.*, 2016; Bravo *et al.*, 2018), we inferred species tree

topology and divergence times using the Bayesian multispecies coalescent method implemented

in StarBEAST2 (Ogilvie *et al.*, 2017). We did not use the concatenated-based partition estimated

by PartitionFinder in the framework of the multispecies coalescent model because 1) gene trees

should be unlinked, 2) each linked codon position within a single locus should share the same

clock, and 3) mitochondrial and nuclear loci do not share the same ploidy level. Thus, we used

jModelTest v2.1.7 (Darriba *et al.*, 2012) to evaluate the substitution models available in

starBEAST2, including or not rate variation among sites (+I and +Γ), for each gene locus.

Nucleotide substitution models for each gene locus were chosen on the basis of the Bayesian

Information Criterion (BIC). We preliminarily evaluated the fit of two tree models, namely Yule

and birth-death, for all 8 species delimitation scenarios using 50 path-sampling steps under

thermodynamic integration (Lartillot & Philippe, 2006), each running for 60 million cycles to

ensure final ESS values > 200. The Yule tree model had the highest marginal likelihood estimate

in all cases and it was preferred over the birth-death model. We therefore report here the path-

sampling analyses based on the Yule tree model for the eight species delimitation hypotheses.

Other parameters, including gene ploidy, clock models and popMean prior, were set as in the

STACEY analyses. The support for each of the eight species delimitation hypotheses was

assessed via Bayes factors (*ln*Bf) (Kass & Raftery, 1995) calculated from the posterior tree

samples, and we considered $ln\mathrm{Bf} = 2$–10 to represent positive but not conclusive support and

*ln*Bf > 10 as decisive support for the species hypothesis with the highest marginal likelihood

estimated through path sampling.

We must rely on secondary calibrations to date the phylogeny of Haeterini because there are no

described fossils assigned to the tribe. Based on a densely sampled, fossil-calibrated butterfly

phylogeny (Chazot *et al*., 2019), we constrained the ages of six well-supported nodes that do not

belong to Haeterini. We followed a conservative approach by using uniform priors encompassing

the 95% highest posterior density (HPD) intervals from (Chazot *et al.*, 2019). The constrained

nodes included the divergences of:

i) Brassolini and Morphini to 32–58 Ma;

ii) Melaniti and Dirini to 23–47 Ma;

iii) Lethina, Parargina and Mycalesina to 25–44 Ma;

iv) The crown age of Satyrini to 32–53 Ma;

v) The crown age of the Satyrini's subclade encompassing the tribes Pronophilina, Euptychiina,

Satyrina, Erebiina, and other closely related subtribes, to 25–43 Ma;

vi) The crown age of Satyrinae to 41–67 Ma.

Time-calibrated species trees were inferred using BEAST v2.4.7 and the Yule tree prior. Eight

independent analyses were carried out to estimate divergence times for the eight species

hypotheses described above. Each analysis was run four independent times for 200 million

cycles each, with parameters sampled every 20,000 cycles. Trees were summarized in

TreeAnnotator (part of BEAST v2.4.7) using the maximum clade credibility method after

discarding the first 25% samples as burn-in and merging the four independent runs in

LogCombiner (part of BEAST v2.4.7). Convergence among runs was evaluated on the basis of

ESS values > 200. DNA alignments and the eight time-calibrated species trees were deposited at

Figshare (https://doi.org/10.6084/m9.figshare.7637021.v1). For qualitative evaluation we present

"cloudograms" (Figs 2, 3), which are phylogenetic diagrams that reflect topological uncertainty

of species trees. Cloudograms were recovered using DensiTree v2.2.5 (Bouckaert, 2010) based

on 500 trees from the posterior distribution.

RESULTS

Data compilation and potential phylogenetic biases

The single-gene and concatenated multi-locus tree topologies were congruent and showed no

evident signature of cross-contamination (Fig. S1). The inferred inter-generic relationships were

robust as indicated by high posterior probabilities (PP = 1.0 for the concatenated multi-locus

tree) and the posterior MSC trees (Figs 2, 3). All clearwing Haeterini form a monophyletic group

sister to the Haeterini butterflies having full scale-cover on wings. Therefore, in terms of

phylogenetic branching pattern, the genus *Pierella* diverged early in the evolution of Haeterini,

followed by the monotypic genus *Dulcedo*, the genus *Pseudohaetera*, and the divergence

between the genera *Haetera* and *Cithaerias*. Mixed node support for inter-specific relationships

were recovered in single-gene and in the concatenated multi-locus datasets, with PP ranging

from ~0.6 to 1.0. The six loci chosen for this study have been previously utilized in butterfly

species-level systematics, thus, we expected these loci to be phylogenetically informative (Table

2). Instead, we recovered low node support among certain Haeterini species (Fig. S1), such as

those that rapidly diverged early in the evolution of *Pierella*, or in the recent radiation (< 2.5 Ma)

of the genera *Pierella* and *Cithaerias*.

Species delimitation

Regardless of which prior was used to take into account taxonomic knowledge for the

collapseWeight parameter (ω), STACEY converged in similar MSCCs suggesting that either 22

(under parameter ω = 0.73) or 24 (under parameter ω = 0.59) lineages are the most adequate

representation of species in our dataset. Indeed the pairwise similarity matrices generated for the

two analyses are highly congruent, though the pairwise posterior probabilities seem to decrease

in the STACEY analysis using $\omega = 0.59$ (Fig. S2). On average, the posterior probability that two

or more specimens inferred by STACEY under parameter $\omega = 0.73$ belong to a single MSCC

ranged from 0.42 to 0.99, with a median of 0.85 and mode of 0.93. The lowest posterior

probability (0.42) was for the MSCC encompassing *Pierella lena lena*. In contrast, the STACEY

analysis that did not take into account any taxonomy information (parameter ω with Beta

distribution [$\alpha = 2$, $\beta = 2$]) did not group any specimen in the dataset into MSCCs.

The MSCCs inferred by BP&P showed high sensitivity to prior distributions for the parameter θ.

The scenario with small ancestral population size, i.e. $\theta = IG[3, 0.02]$, suggested that nearly

every specimen in the dataset represent a single divergent lineage. This extreme scenario

suggesting 55 MSCCs was considered in the hypothesis testing exercise using Bayes factors,

even though it significantly departs from current taxonomic understanding of Haeterini. The

most likely numbers of MSCCs recovered by the other two BP&P analyses were 21 (when $\theta =$

$IG[3, 0.2]$; prior mean of 0.1) or 28 (when $\theta = IG[3, 0.1]$; prior mean of 0.05). The two

independent runs for each analysis converged in the same MSCCs, and the posterior probabilities

of the delimited species for the analysis under $\theta = IG[3, 0.2]$ ranged from 0.44 to 1.00, with a

median of 0.96 and mode of 0.99. The two lowest posterior probabilities (0.44) were for the

MSCCs encompassing *Pierella lena lena* and for *Pierella hyceta hyceta*. Overall, the large

impact of the prior distribution of the parameter θ in BP&P, resulting in 21, 28 or 55 MSCCs,

suggests that more data, both genetic and taxon sampling, is needed to strongly inform the

molecular species delimitation. However, the resulting most-probable number of MSCCs favored

under the two scenarios for large and medium ancestral population sizes are congruent with the

taxonomic knowledge of the group, as well as with the species delimitation exercises carried out

in STACEY (Fig. 2).

The eight species delimitation hypotheses ranged from 18 to 63 species (or MSC clusters) (Fig.

2). The species hypotheses recovered by STACEY using ω = 0.73 (22 MSCCs) and BP&P using

θ = IG[3, 0.2] (21 MSCCs) were strongly supported compared to the remaining six species

delimitation hypotheses (*ln*Bf >> 10). The STACEY analysis recovering 22 MSCCs is

supported, but not conclusively (*ln*Bf = 9.69), over the BP&P analysis recovering 21 MSCCs

(Table 3). The differences between these two species hypotheses are: 1) the delimitation of two

lineages of *Pseudohaetera hypaesia*, one from central Peru (Chanchamayo Valley) and the other

from southeastern Ecuador (Morona Santiago and Zamora Chinchipe Provinces); BP&P

suggested a single MSCC for these two lineages whereas STACEY suggested two separate

MSCCs, 2) split of the subspecies *Pierella helvina ocreata* and *P. helvina incanescens* into two

separate species only by STACEY, and 3) split of *Pierella chalybaea* from Ecuador (Sucumbios)

from remaining *P. chalybaea* only by BP&P. On the other hand, the BP&P analyses under the

medium ancestral population size (θ = IG[3, 0.1]) recovered most of the described subspecies in

Haeterini, thus, we used this delimitation as a proxy for recognizing taxonomic subspecies in the

framework of the MSC despite it not being supported by the Bayes factor model comparison

(Fig. 2).

Absolute divergence times

We show in Figure 3 two time-calibrated species trees, one approximating species in Haeterini

(i.e. STACEY's clusters under ω = 0.73; which had the highest marginal likelihood estimate

among the eight species delimitation hypotheses) and the second approximating subspecies in

Haeterini (BP&P's clusters under θ = IG[3, 0.1]; which recovered most of the described

taxonomic subspecies). The remaining six species trees can be found in the Supporting

Information (Fig. S3). Species tree topologies and branch lengths in absolute time remained

virtually the same regardless of species delimitation hypothesis. The median crown age of extant

Haeterini was estimated at 27.2 Ma (95% HPD: 21.8 to 32.7 Ma; entire posterior range: 15.4 to

40.5 Ma). The rapid early radiation of *Pierella* happened in the late Miocene, at about 7.22 Ma

(95% HPD: 5.68 to 9.03 Ma; entire posterior range: 4.97 to 11.4 Ma), whereas the recent

radiation of *Pierella* and *Cithaerias* occurred in the Pleistocene, at about 1 to 2 Ma (95% HPD:

0.81 to 2.58 Ma; entire posterior ranges for the recent radiation of *Pierella*: 0.57 to 2.83 Ma and

for *Cithaerias*: 1.19 to 3.25 Ma). We estimate that these two diversification events gave rise to

about 70% of extant Haeterini species.

DISCUSSION

Regardless of the species concept advocated by different researchers, traditional taxonomy and

coalescent-based approaches can act in synergy to infer statistically robust species hypotheses.

Importantly, the multispecies coalescent (MSC) model and the Bayesian framework to delimit

MSC clusters (MSCCs) allow the quantification and testing of species boundaries informed by

taxonomic knowledge. Furthermore, divergent taxonomic opinion among authors working on

particular species groups has been statistically weighed by the approach followed in this study.

This is important as a first step towards reliable standardization of taxonomy and higher-level

systematics based on models that take into account the process of speciation (e.g. incomplete

lineage sorting) and not just arbitrary genetic distances as in threshold-based approaches (e.g.

based on COI barcoding). Here, we have not used any threshold-based method because it would

not contribute to the described probabilistic framework as these methods utilize single-locus

datasets and neglect tree topology uncertainty. However, future research can further address

comparisons between threshold-based approaches (e.g., single-threshold, multi-thresholds, multi-

rate processes) and the multispecies coalescent model.

More realistic models of speciation that include, for example, inter-specific gene flow might be

more accurate at estimating species histories (Müller *et al.*, 2018). For example, in Europe alone,

around 16% of butterfly species are known to hybridize (Descimon & Mallet, 2009), which is in

line with the idea of speciation as a continuum (De Queiroz, 2007) and that inter-specific gene

flow in animals might not be uncommon (Mallet *et al.*, 2007). Another source of information for

taxonomic conclusions is the development of new methods that jointly model phenotypic traits

and genetic data (Solís-Lemus *et al.*, 2015). Unfortunately, phenotypic and life history data, such

as larval host plants and habitat association, that could be used in species delimitation is scarce

for Haeterini butterflies. Taken together, these methodological advances and the generation of

biological data from various sources suggest that in the near future coalescent-based approaches

based on multi-locus data may be able to overcome many current shortcomings in delimiting

species.

Limitations and strengths

The dataset and approach that we have followed here are not exempt of limitations. First, the

amount of missing information, including taxonomic sampling and gapped molecular dataset,

may have reduced the power of our analyses. However, it has been noted previously that the

MSC might still be accurate with sampling schemes including fewer than five individuals per

lineage, as long as multi-locus datasets (in our case 6 unlinked loci) are utilized (Zhang *et al.*,

2011). Although the impact of missing data in the MSC framework needs to be further

examined, a recent simulation study suggested that coalescent-based species tree inference might

be highly accurate even with severely gapped multi-locus datasets (Nute *et al.*, 2018). Second,

inter-specific gene flow may impact current implementations of the MSC by obscuring lineages

divergence (Luo *et al.*, 2018). Nevertheless, unless there is a high level of gene flow and

hybridization, the MSC-based delimitation and species tree inference should be robust (Eckert &

Carstens, 2008; Zhang *et al.*, 2011). However, we note that this issue needs to be further studied

in the light of more data and using recently-developed approaches, such as the isolation-with-

migration model (Müller *et al.*, 2018). Third, the impact of priors in MSC-based species

delimitation might be high when the molecular dataset does not hold sufficient signal to

converge on the same MSCCs (Leaché & Fujita, 2010; Jones *et al.*, 2015). Our results showed

that the choice of priors heavily influenced the number of MSCCs estimated by MSC-based

methods. However, our priors rely on existing taxonomic knowledge and thus we have reduced

the clustering space of 63 specimens based on actual evidence coming from previous

morphology-based studies. The number of most likely MSCCs inferred here (from 21 to 28

clusters) remains highly congruent with the morphological diversity encountered in this group of

butterflies.

The strengths of our study rely on three key aspects. First, we have not used any a priori

taxonomic assignment of individuals to species, nor any guide species tree to delimit species.

This allowed us to include both tree topology and taxonomic assignment uncertainties explicitly

into the models, avoiding potential biases that may have precluded accurate estimation of

speciation probabilities (Leaché & Fujita, 2010). However, we note that the posterior

probabilities of certain species delimitations, such as *Pierella lena lena*, remain low, which may

be explained by non-sufficient signal of our molecular dataset. Adding more genetic data may

increase the posterior probabilities while holding the same delimitations, given that different

approaches converged in the same species hypotheses regardless of the low posterior

probabilities (e.g. *Pierella lena lena* was recovered as most probable by both STACEY and

BP&P, Fig. 2). However, this needs to be confirmed by future genome-wide studies on

Haeterini. Second, taxonomic knowledge has been formally taken into account in our

probabilistic scenarios. Previous studies have put forward the statistical comparison (e.g. via

Bayes factors) of alternate species delimitation models, but these mostly evaluated different

individual reassignments based only on node collapsing of sister lineages in a phylogeny

(Grummer *et al.*, 2014; Yu *et al.*, 2017). Our pipeline, on the contrary, has aided the exploration

of alternate MSCCs that are congruent as well with other sources of information in butterfly

taxonomy, such as morphology and geography. Third, the MSCCs found in this study are not

nested, meaning that specimens could be re-assigned to any combination of clusterings. The use

of Bayes factors as a selection tool is appropriate because of its flexibility in testing non-nested

models (Leaché *et al.*, 2014). We used path sampling under thermodynamic integration which

has been shown to be highly accurate in testing non-nested species delimitation models

(Grummer *et al.*, 2014).

Performance of methods

Our approach focused on using the multispecies coalescent model, informed by taxonomic

knowledge, to assign individuals to MSCCs based on different priors and methods implemented

in STACEY and BP&P. This is arguably a less arbitrary approach to reduce the space of all

possible clusterings for model testing using Bayes factors, compared to other approaches based

solely in taxonomic expertise (Leaché *et al.*, 2014) or in multi-locus networks (Grummer *et al.*,

2014) and population assignments (Olave *et al.*, 2014). Furthermore, the approach outlined here

is flexible, simple and fast because it avoids preliminary estimation of guide trees and *a priori*

taxonomic assignments of individuals. Taxonomic knowledge modeled as prior distribution has

the potential to ameliorate potential biases (e.g. population structure raised to species), which

have been observed when relying solely on genetic information and guide trees (Olave et al.,

2014). As Sukumaran & Knowles (2017) argue, the only way that the multispecies coalescent

can be used as a reliable species delimitation approach is to explicitly incorporate other sources

of information as *a priori* hypotheses. We showed that the number of inferred MSCCs heavily

depends on priors, and indeed, under the scenario of medium ancestral population sizes (θ =

IG[3, 0.1]) BP&P recovered traditionally viewed infra-specific diversity, and perhaps structured

populations. Note that subspecies in butterfly taxonomy are seen as populations with limited

gene flow due to allopatric distributions (Lamas, 2008). Overall, we highlight the importance of

taxonomically-informed molecular species delimitation and the use of Bayes factor model

comparison.

Both MSC-based methods used in this study, STACEY and BP&P, were fast and simple in their

implementation thanks to available guidelines and manuals (e.g. Jones *et al.*, 2015; Yang, 2015).

The most supported BP&P's delimitation followed a conservative prior distribution for the size

of ancestral population size, that is θ = IG[3, 0.2] with a prior mean = 0.1, which was pointed out

as the most appropriate prior mean (Leaché & Fujita, 2010). This species delimitation model is

highly similar to the most supported STACEY model, except for three species hypotheses. In

line with morphological differences, STACEY favored the split of *Pierella helvina incanescens*

and *P. helvina ocreata*, two subspecies that were originally considered separate species but

synonymized due to allopatric geographical distribution by Constantino (1995). On the other

hand, BP&P favored the split of one *Pierella chalybaea* specimen (Ecuador, Sucumbíos) from

other conspecific individuals, including others from Ecuador. STACEY did not favor such a

split, which again is in line with the absence of any clear morphological difference in *P.*

chalybaea from western Amazonia (Zacca *et al.*, 2016). Finally, STACEY delimited three

separate species within *Pseudohatera hypaesia* while BP&P delimited only two. However, there

are no clear wing coloration differences among populations of *P. hypaesia* throughout its range

from Colombia to Bolivia (Gerardo Lamas, *pers. comm.*). The supported STACEY's

delimitation here thus points out that large genetic divergences exist in the otherwise similar-

looking *P. hypaesia*, warranting a taxonomic revision of these montane butterflies from the

tropical Andes.

Taxonomic implications

The probabilistic framework applied in this study allows the statistical test of alternative species

hypotheses in a taxonomic group that has likely evolved for nearly 27 million years. The most

likely scenario among those tested here suggests that at least four divergent lineages should be

elevated to species by current taxonomic standards. It is clear that a more comprehensive

sampling and datasets, including morphological and molecular characters, are needed to robustly

delineate species boundaries in Haeterini. Nonetheless, the four divergent lineages found in this

study corroborate morphological differences that have been previously acknowledged in these

groups but that taxonomists conservatively considered conspecific variations (subspecies) mainly

because of allopatric distributions, a historical practice in butterfly taxonomy (Lamas, 2008). In

almost all MSC-based analyses, the following taxa were considered independent evolutionary

lineages: i) *Pierella hyceta hyceta* and *Pierella hyceta ceryce*, ii) *Pierella lena lena* and *Pierella*

lena brasiliensis, iii) *Pierella helvina incanescens* and *Pierella helvina ocreata*, iv) two

divergent sympatric populations in central Peru (Chanchamayo Valley) of *Pseudohaetera*

hypaesia. According to our results, these eight lineages are likely full species (a taxonomic

revision in preparation is addressing this).

Zacca *et al.* (2016) split *Pierella lamia* into seven species mainly based on allopatric

geographical distributions, and also on genitalia and androconial patches on male wings (despite

their similar wing coloration). We did not find support for the species status of *Pierella*

keithbrowni from southeastern Brazil proposed by these authors. In the case of *P. keithbrowni*,

the main diagnostic characters were the androconial patch shape, and ductus bursae in female

genitalia longer than in *P. nice*, albeit male genitalia in *Pierella lamia* complex are not

differentiated. In all MSC-based analyses, *P. keithbrowni* is not genetically different from other

populations in central Brazil and Ecuador. Our results therefore suggest that differences in the

aforementioned morphological characters may represent variation within a single evolutionary

lineage, and thus their usage needs to be complemented with other lines of evidence. Note that

androconial patch morphology has been widely used to diagnose butterfly species boundaries in

other satyrine groups (e.g. Núñez Aguila *et al.*, 2013; Penz *et al.*, 2017). Our results question the

reliability of this character system only in Haeterini lineages, thus, further research in other

butterfly groups can test the generality of our conclusions that different character systems favor

narrower or broader delimitations in butterflies. However, as noted previously, granting species

status to butterfly populations primarily based on geographical distribution (allopatric

populations) might be unjustified if other evidence fails to recognize clear divergence (Descimon

& Mallet, 2009). Therefore, we suggest that *P. keithbrowni* should be synonymized with *P.*

chalybaea (work in preparation will address this).

CONCLUSIONS

Here we provide evidence that for the butterfly tribe Haeterini, the multispecies coalescent model

generally recognizes traditionally viewed butterfly subspecies and species, with some exceptions

linked to the use of different morphological character systems. By using a probabilistic

framework, we show that divergent taxonomic opinion (concepts) were used by different

authors, including butterfly species that were over-split (*lamia* complex), lumped (at least 6

subspecies raised to species here), or misclassified (e.g. *Pierella lena browni* was previously

synonymized with *P. lena brasiliensis* (Lamas, 2004), despite *browni* being evolutionarily more

closely related to *P. lena lena* than to *brasiliensis*). Androconial patch morphology is commonly

used as informative character to diagnose species, but we show that at least in *Pierella* it alone

may not be well-suited to distinguish infra- and inter-specific diversity. Furthermore, taxonomic

knowledge informed as priors in MSC-based species delimitation using genetic data is a robust

approach to reduce the clustering space in model testing.

The low node support recovered among certain Haeterini species may be attributed to incomplete

lineage sorting due, for example, to ancient rapid radiation as in the crown node of *Pierella* or to

recent speciation as in *Pierella* and *Cithaerias*. Haeterini butterflies evolved for nearly 27 million

years but most extant species (*ca*. 80%) likely diverged rather recently—within the past 2 million

years. Future macroevolutionary studies using the revised species boundaries of Haeterini might

address this puzzling diversification history, whether it was the result of high tropical species

turnover over millions of years—with constant and high extinction and speciation rates, in line

with Antonelli *et al.* (2015) —or a Pleistocene major burst in diversification, in line with the

Quaternary diversification model, as characterized via simulated data (Matos-Maraví, 2016).

SUPPLEMENTARY MATERIAL

Figure S1: Consensus trees based on single-gene and concatenated loci datasets (pages 2–9).

Figure S2: Pairwise similarity matrices based on delimitation analyses with STACEY (pages 10–

14).

Figure S3: Time-calibrated Maximum Clade Credibility species trees of the eight species

delimitation hypotheses (pages 15–24).

DNA alignments and phylogenies of the single-gene and concatenated datasets were archived in

TreeBASE (ID: 23439). Accession URL:

http://purl.org/phylo/treebase/phylows/study/TB2:S23439.

Input files for species delimitation in BP&P and STACEY, and time-calibrated phylogenies:

FigShare, https://doi.org/10.6084/m9.figshare.7637021.v1.

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REFERENCES

Antonelli, A., Zizka, A., Silvestro, D., Scharn, R., Cascales-Miñana, B., & Bacon, C.D. (2015)

An engine for global plant diversity: Highest evolutionary turnover and emigration in the

American tropics. *Frontiers in Genetics*, **6**, 130.

Aydin, Z., Marcussen, T., Ertekin, A.S., & Oxelman, B. (2014) Marginal likelihood estimate

comparisons to obtain optimal species delimitations in *Silene* sect. *Cryptoneurae*

(Caryophyllaceae). *PLoS ONE*, **9**, e106990.

Bouckaert, R., Heled, J., Kühnert, D., Vaughan, T., Wu, C.-H., Xie, D., Suchard, M.A.,

Rambaut, A., & Drummond, A.J. (2014) BEAST 2: a software platform for Bayesian

evolutionary analysis. *PLoS Computational Biology*, **10**, e1003537.

Bouckaert, R.R. (2010) DensiTree: making sense of sets of phylogenetic trees. *Bioinformatics*,

26, 1372–1373.

Bravo, G.A., Antonelli, A., Bacon, C.D., Bartoszek, K., Huynh, S., Jones, G., Knowles, L.L.,

Lamichhaney, S., Marcussen, T., Nakhleh, L.K., Oxelman, B., Pfeil, B., Schliep, A.,

Werneck, F.P., Wiedenhoeft, J., Willows-Munro, S., & Edwards, S. V (2018) Embracing

heterogeneity: building the Tree of Life and the future of phylogenomics. *PeerJ PrePrints*,

6, e26449v3.

Chazot, N., Wahlberg, N., Freitas, A.V.L., Mitter, C., Labandeira, C., Sohn, J.-C., Sahoo, R.K.,

Seraphim, N., de Jong, R., & Heikkilä, M. (2019) The priors and posteriors in Bayesian

timing of divergence analyses: the age of butterflies revisited. *Systematic Biology*, In Press.

Constantino, L.M. (1995) Revisión de la tribu Haeterini Herrich-Schäffer, 1864 en Colombia

(Lepidoptera: Nymphalidae, Satyrinae). *SHILAP Revista de Lepidopterología*, **23**, 49–76.

Darriba, D., Taboada, G.L., Doallo, R., & Posada, D. (2012) jModelTest 2: more models, new

heuristics and parallel computing. *Nature Methods*, **9**, 772.

Degnan, J.H. & Rosenberg, N.A. (2009) Gene tree discordance, phylogenetic inference and the multispecies coalescent. *Trends in Ecology and Evolution*, **24**, 332–340.

Descimon, H. & Mallet, J. (2009) Bad species. *Ecology of Butterflies in Europe* (ed. by J. Settele, T. Shreeve, M. Konvička, and H. Van Dyck), pp. 219–249. Cambridge University Press, Cambridge, UK.

Eckert, A.J. & Carstens, B.C. (2008) Does gene flow destroy phylogenetic signal? The performance of three methods for estimating species phylogenies in the presence of gene flow. *Molecular Phylogenetics and Evolution*, **49**, 832–842.

Edwards, S. V., Xi, Z., Janke, A., Faircloth, B.C., McCormack, J.E., Glenn, T.C., Zhong, B., Wu, S., Lemmon, E.M., Lemmon, A.R., Leaché, A.D., Liu, L., & Davis, C.C. (2016) Implementing and testing the multispecies coalescent model: a valuable paradigm for phylogenomics. *Molecular Phylogenetics and Evolution*, **94**, 447–462.

Fujita, M.K., Leaché, A.D., Burbrink, F.T., McGuire, J.A., & Moritz, C. (2012) Coalescent-based species delimitation in an integrative taxonomy. *Trends in Ecology and Evolution*, **27**, 480–488.

Grummer, J.A., Bryson, R.W., & Reeder, T.W. (2014) Species delimitation using Bayes factors: simulations and application to the *Sceloporus scalaris* species group (Squamata: Phrynosomatidae). *Systematic Biology*, **63**, 119–133.

Hebert, P.D.N., Penton, E.H., Burns, J.M., Janzen, D.H., & Hallwachs, W. (2004) Ten species in one: DNA barcoding reveals cryptic species in the neotropical skipper butterfly *Astraptes fulgerator*. *Proceedings of the National Academy of Sciences of the United States of America*, **101**, 14812–14817.

Herrich-Schäffer, G.A.W. (1864) Prodromus systematis Lepidopterorum: versuch einer systematischen Anordnung der Schmetterlinge. *CorrespondenzBlatt des zoologisch-mineralogischen Vereines zu Regensburg*, **18(7/8)**, 89112.

Jones, G. (2017) Algorithmic improvements to species delimitation and phylogeny estimation under the multispecies coalescent. *Journal of Mathematical Biology*, **74**, 447–467.

Jones, G., Aydin, Z., & Oxelman, B. (2015) DISSECT: an assignment-free Bayesian discovery method for species delimitation under the multispecies coalescent. *Bioinformatics*, **31**, 991–998.

Kass, R.E. & Raftery, A.E. (1995) Bayes Factors. *Journal of the American Statistical Association*, **90**, 773–795.

Keightley, P.D., Pinharanda, A., Ness, R.W., Simpson, F., Dasmahapatra, K.K., Mallet, J., Davey, J.W., & Jiggins, C.D. (2015) Estimation of the spontaneous mutation rate in *Heliconius melpomene*. *Molecular Biology and Evolution*, **32**, 239–243.

Knowles, L.L. & Carstens, B.C. (2007) Delimiting species without monophyletic gene trees. *Systematic Biology*, **56**, 887–895.

Lamas, G. (1997) Lista sinonímica de los géneros *Cithaerias* Hübner y *Haetera* Fabricius (Lepidoptera: Nymphalidae: Satyrinae), con la descripción de dos subespecies nuevas del Perú. *Revista Peruana de Entomología*, **40**, 133–138.

Lamas, G. (2004) *Atlas of Neotropical Lepidoptera: Checklist: Part 4A. Hesperioidea-Papilionoidea.* Association for Tropical Lepidoptera, Gainesville, Florida.

Lamas, G. (2008) La sistemática sobre mariposas (Lepidoptera: Hesperioidea y Papilionoidea) en el mundo: estado actual y perspetivas futuras. *Contribuciones Taxonómicas en Órdenes de Insectos Hiperdiversos* (ed. by J. Llorente Bousquets and A. Lanteri), pp. 57–70. Las

Prensas de Ciencias, UNAM, México D.F.

Lanfear, R., Frandsen, P.B., Wright, A.M., Senfeld, T., & Calcott, B. (2017) Partitionfinder 2:

new methods for selecting partitioned models of evolution for molecular and morphological

phylogenetic analyses. *Molecular Biology and Evolution*, **34**, 772–773.

Lartillot, N. & Philippe, H. (2006) Computing Bayes factors using thermodynamic integration.

Systematic Biology, **55**, 195–207.

Leaché, A.D. & Fujita, M.K. (2010) Bayesian species delimitation in West African forest geckos

(*Hemidactylus fasciatus*). *Proceedings of the Royal Society B: Biological Sciences*, **277**,

3071–3077.

Leaché, A.D., Fujita, M.K., Minin, V.N., & Bouckaert, R.R. (2014) Species delimitation using

genome-wide SNP Data. *Systematic Biology*, **63**, 534–542.

Luo, A., Ling, C., Ho, S.Y.W., & Zhu, C.-D. (2018) Comparison of methods for molecular

species delimitation across a range of speciation scenarios. *Systematic Biology*, **67**, 830–

846.

Mallet, J., Beltrán, M., Neukirchen, W., & Linares, M. (2007) Natural hybridization in

heliconiine butterflies: the species boundary as a continuum. *BMC Evolutionary Biology*, **7**,

28.

Matos-Maraví, P. (2016) Investigating the timing of origin and evolutionary processes shaping

regional species diversity: Insights from simulated data and Neotropical butterfly

diversification rates. *Evolution*, **70**, 1638–1650.

Miller, L.D. (1968) The higher classification, phylogeny and zoogeography of the Satyridae

(Lepidoptera). *Memoirs of the American Entomological Society*, **24**, 1–174.

Miller, M.A., Pfeiffer, W., & Schwartz, T. (2010) Creating the CIPRES Science Gateway for

inference of large phylogenetic trees. *Proceedings of the Gateway Computing Environments Workshop (GCE), 14 Nov.*, 1–8.

Müller, N.F., Ogilvie, H.A., Zhang, C., Drummond, A., & Stadler, T. (2018) Inference of species histories in the presence of gene flow. *bioRxiv*, 348391.

Núñez Aguila, R., Matos-Maraví, P., & Wahlberg, N. (2013) New *Calisto* species from Cuba, with insights on the relationships of Cuban and Bahamian taxa (Lepidoptera, Nymphalidae, Satyrinae). *Zootaxa*, **3669**, 503–521.

Nute, M., Chou, J., Molloy, E.K., & Warnow, T. (2018) The performance of coalescent-based species tree estimation methods under models of missing data. *BMC Genomics*, **19**, 286.

Ogilvie, H.A., Bouckaert, R.R., & Drummond, A.J. (2017) StarBEAST2 brings faster species tree inference and accurate estimates of substitution rates. *Molecular Biology and Evolution*, **34**, 2101–2114.

Olave, M., Solà, E., & Knowles, L.L. (2014) Upstream analyses create problems with DNA-based species delimitation. *Systematic Biology*, **63**, 263–271.

Paluch, M., Zacca, T., Siewert, R.R., Mielke, O.H.H., & Casagrande, M.M. (2015) A new subspecies of *Pierella lena* (Lepidoptera: Nymphalidae: Satyrinae) in northeastern Brazil. *Florida Entomologist*, **98**, 916–920.

Penz, C.M., Alexander, L.G., & DeVries, P.J. (2014) Revised species definitions and nomenclature of the rose colored *Cithaerias* butterflies (Lepidoptera, Nymphalidae, Satyrinae). *Zootaxa*, **3873**, 541–559.

Penz, C.M., Casagrande, M.M., Devries, P., & Simonsen, T.J. (2017) Documenting diversity in the Amazonian butterfly genus *Bia* (Lepidoptera, Nymphalidae). *Zootaxa*, **4258**, 201–237.

Pons, J., Barraclough, T.G., Gomez-Zurita, J., Cardoso, A., Duran, D.P., Hazell, S., Kamoun, S.,

Sumlin, W.D., & Vogler, A.P. (2006) Sequence-based species delimitation for the DNA taxonomy of undescribed insects. *Systematic Biology*, **55**, 595–609.

Puillandre, N., Lambert, A., Brouillet, S., & Achaz, G. (2012) ABGD, Automatic Barcode Gap Discovery for primary species delimitation. *Molecular Ecology*, **21**, 1864–1877.

De Queiroz, K. (2007) Species concepts and species delimitation. *Systematic Biology*, **56**, 879–886.

Rannala, B. & Yang, Z. (2013) Improved reversible jump algorithms for Bayesian species delimitation. *Genetics*, **194**, 245–253.

Rannala, B. & Yang, Z. (2017) Efficient Bayesian species tree inference under the multispecies coalescent. *Systematic Biology*, **66**, 823–842.

Ronquist, F., Teslenko, M., van der Mark, P., Ayres, D.L., Darling, A., Höhna, S., Larget, B., Liu, L., Suchard, M.A., & Huelsenbeck, J.P. (2012) MrBayes 3.2: efficient Bayesian phylogenetic inference and model choice across a large model space. *Systematic Biology*, **61**, 539–542.

Rosindell, J., Cornell, S.J., Hubbell, S.P., & Etienne, R.S. (2010) Protracted speciation revitalizes the neutral theory of biodiversity. *Ecology Letters*, **13**, 716–727.

Solís-Lemus, C., Knowles, L.L., & Ané, C. (2015) Bayesian species delimitation combining multiple genes and traits in a unified framework. *Evolution*, **69**, 492–507.

Sukumaran, J. & Knowles, L.L. (2017) Multispecies coalescent delimits structure, not species. *Proceedings of the National Academy of Sciences*, **114**, 1607–1612.

Wahlberg, N., Leneveu, J., Kodandaramaiah, U., Peña, C., Nylin, S., Freitas, A.V.L., & Brower, A.V.Z. (2009) Nymphalid butterflies diversify following near demise at the Cretaceous/Tertiary boundary. *Proceedings of the Royal Society B: Biological Sciences*,

276, 4295–4302.

Wahlberg, N. & Wheat, C.W. (2008) Genomic outposts serve the phylogenomic pioneers: designing novel nuclear markers for genomic DNA extractions of Lepidoptera. *Systematic Biology*, **57**, 231–242.

Weymer, G. (1910) Familie: Satyridae. *Die Gross-Schmetterlinge der Erde. Vol. 5* (ed. by A. Seitz), pp. 173–283. Stuttgart, Alfred Kernen Verlag,

Willmott, K.R. (2015) A new subspecies of *Cithaerias pyropina* (Salvin & Godman, 1868) from the Cordillera de Cutucú in southeastern Ecuador (Nymphalidae: Satyrinae). *Tropical Lepidoptera Research*, **25**, 1–4.

Yang, Z. (2015) The BPP program for species tree estimation and species delimitation. *Current Zoology*, **61**, 854–865.

Yang, Z. & Rannala, B. (2010) Bayesian species delimitation using multilocus sequence data. *Proceedings of the National Academy of Sciences of the United States of America*, **107**, 9264–9269.

Yang, Z. & Rannala, B. (2014) Unguided species delimitation using DNA sequence data from multiple loci. *Molecular Biology and Evolution*, **31**, 3125–3135.

Yu, G., Rao, D., Matsui, M., & Yang, J. (2017) Coalescent-based delimitation outperforms distance-based methods for delineating less divergent species: the case of *Kurixalus odontotarsus* species group. *Scientific Reports*, **7**, 16124.

Zacca, T., Siewert, R.R., Casagrande, M.M., Mielke, O.H.H., & Paluch, M. (2016) Taxonomic revision of the "*Pierella lamia* species group" (Lepidoptera: Nymphalidae: Satyrinae) with descriptions of four new species from Brazil. *Zootaxa*, **4078**, 366–386.

Zhang, C., Zhang, D.X., Zhu, T., & Yang, Z. (2011) Evaluation of a Bayesian coalescent method

744  of species delimitation. *Systematic Biology*, **60**, 747–761.

 FIGURES AND TABLES

Figure 1: Representatives of the tribe Haeterini. a) *Pierella nereis* (Brazil, Minas Gerais, Santa

Barbara; Milwaukee Public Museum), b) *Dulcedo polita* (Costa Rica, Sarapiqui, Tirimbina

Biological Station; Phil DeVries Collection), c) *Pseudohaetera mimica* (Peru, Junin, Satipo;

Natural History Museum of Los Angeles County), d) *Haetera piera* (Ecuador, Napo, Garza

Cocha; Phil DeVries Collection), e) *Cithaerias cliftoni* (Ecuador, Oriente; Natural History

Museum of Los Angeles County), f) *Cithaerias aurora tambopata* (Peru, Madre de Dios,

Pakitza, Manu National Reserve; Smithsonian Institution).

Figure 2: Evaluated species delimitation hypotheses using Bayes factors. The eight scenarios

were: Taxonomic species (*spp*, 18 lineages) or subspecies raised to species (*sspp*, 26 lineages),

STACEY's delimited species under prior accounting for number of taxonomic species (*spp*

prior, 22 lineages) or number of subspecies raised to species (*sspp prior*, 24 lineages), as well as

with prior not informed by taxonomy (*no prior*, 63 lineages), and BP&P's delimited species

under prior for ancestral population size as large (21 lineages), medium (28 lineages), or small

(53 lineages). The "cloudogram", which is a diagram representing phylogenetic uncertainty of

the 63 Haeterini specimens, was generated based on 500 posterior trees from STACEY analysis

(thicker blue line represents the consensus phylogeny). The delimitation model STACEY under

prior accounting for taxonomic species (outlined by a surrounding box) received significant

support based on Bayes factors over all other models, and thus is the classification that we

propose here.

Figure 3: Time-calibrated phylogenetic hypotheses of models that best approximate species and

subspecies in Haeterini. A) "Cloudogram" of the best-fit species delimitation model based on

Bayes factors, STACEY under prior accounting for taxonomic species. B) "Cloudogram" of the

delimitation model that best approximated described subspecies, BP&P under prior for medium

ancestral population size. Time axis in both panels is scaled to million years. *The species status

of *Pierella helvina ocreata* and *P. helvina incanescens* may change with the inclusion of *P.*

helvina helvina, but it is likely that *P. helvina ocreata* and *P. helvina incanescens* are separate

species. Inset butterfly: *Pierella hyceta hyceta*; Peru, Pasco, Cañón de Huancabamba, 1200 masl,

29.vii.2017. Photo: Markéta Aubrechtová.

A: Haeterini species tree, STACEY, spp prior

B: Haeterini (sub)species tree, BP&P, medium pop size

784   Table 1: Voucher locality information and associated genetic data. All Haeterini specimens were

785   identified to the species and subspecies level based on the most recent taxonomic revisions

786   (Constantino 1995; Lamas 1997; Penz et al. 2014; Paluch et al. 2015; Willmott 2015; Zacca et al.

787   2016). GenBank accession numbers for each of the sequenced locus are presented for every

788   specimen.

Code	Genus	Species	Subspecies	Country	Locality	CAD	COI	EF1a	GAPDH	RpS5	*wingless*
EW1-1	Pararge	aegeria		FRANCE	Carcassonne, 2 km S Trassonel	EU141293	DQ176379	DQ338913	EU141476	EU141372	DQ338620
NW121-17	Lethe	minerva		INDONESIA	Bali	EU141309	DQ338768	DQ338909	EU141492	EU141387	DQ338616
EW10-5	Bicyclus	anynana		ZIMBABWE	Harare	EU141295	AY218238	AY218258	EU141478	EU141374	AY218276
CP06-89	Oressinoma	sorata		PERU	Pasco, Oxapampa	MH802140	GQ357209	GQ357278	GQ357440	GQ357570	GQ357342
EW25-17	Orsotriaena	medus		BANGLADESH	Sylhet Div., Lowacherra Forest	-	DQ338766	DQ338906	EU528405	EU528453	DQ338633
NW108-3	Taygetis	virgilia		BRAZIL	SP, Campinas, Santa Genebra	EU141305	DQ338812	DQ338958	EU141487	EU141383	DQ338683
PM07-05	Calisto	smintheus		CUBA	Alrededores de La Platica	JN881807	JN881904	JN881778	JN881827	JN881845	JN881870
CP03-63	Manerebia	cyclopina		PERU	Junín, Quebrada Siete Jeringas	-	DQ338785	DQ338928	EU528397	EU528443	GQ864477
NW162-21	Satyrus	actaea		FRANCE	Aude, Villegly	GQ864709	GQ864807	GQ864901	GQ865030	GQ865494	GQ864495
EW24-7	Erebia	oeme		FRANCE	Languedoc, Ariege 09, Ustou	EU141296	DQ338780	DQ338923	EU141479	EU141375	DQ338640
CP13-01	Aeropetes	tulbaghia		S. AFRICA	Mpumalanga Verloren Valei	-	DQ338579	DQ338907	EU528381	EU528419	DQ338634
NW66-6	Melanitis	leda		AUSTRALIA	Queensland, Cairns	EU141330	AY090207	AY090173	EU141508	EU141408	AY090140
NW122-21	Brassolis	sophorae		BRAZIL	SP, Campinas	-	EU528314	EU528291	GQ357384	EU528425	EU528270
NW66-5	Morpho	helenor		-	London Pupae Supplies	EU141329	AY090210	AY090176	EU141507	EU141407	AY090143
NW121-20	Elymnias	casiphone		INDONESIA	Bali	EU141310	DQ338760	DQ338900	-	EU141388	DQ338627
NW101-2	Zeuxidia	dorhni		INDONESIA	Java, Bandung	-	DQ338752	DQ338892	EU528417	EU528471	DQ338609
PM14-21	Cithaerias	andromeda	bandusia	BRAZIL	MT, Alta Floresta	KJ145993	KJ145988	KJ145996	KJ146001	KJ146004	KJ146008
NW92-1	Cithaerias	aurora	aurora	PERU	Loreto, Río Paiwa	-	MH802187	-	MH802251	MH802286	MH802332
NW93-1	Cithaerias	aurora	aurora	PERU	Loreto, Río Paiwa	-	DQ338756	DQ338896	MH802252	MH802287	DQ338613
CP01-52	Cithaerias	aurora	tambopata	PERU	Tambopata Research Center	MH802134	MH802157	MH802205	MH802239	MH802258	MH802304
LEP-08919	Cithaerias	cliftoni		ECUADOR	Pastaza, Kapawi lodge	MH802148	MH802174	MH802218	MH802246	MH802276	MH802320
LEP-10406	Cithaerias	cliftoni		ECUADOR	Sucumbios, Cuyabeno lodge	KM013123	KM012939	KM012990	KM013275	KM013174	KM013055
PM14-17	Cithaerias	cliftoni		ECUADOR	Sucumbios, Garza Cocha	-	MH802196	-	-	MH802295	MH802341
PM24-06	Cithaerias	pireta	pireta	COSTA RICA	Sarapiqui, Agrícola Sofía, nr. Tirimbina	-	-	MH802238	-	MH802299	MH802345
LEP-09925	Cithaerias	pireta	pireta	ECUADOR	Carchi, Finca San Francisco	MH802154	MH802182	MH802225	MH802249	MH802282	MH802327
LEP-09926	Cithaerias	pireta	pireta	ECUADOR	Pichincha, km. 20 Pacto-Guayabillas Rd.	KM013135	KM012940	KM013007	KM013282	KM013181	KM013069
BCI86376	Cithaerias	pireta	pireta	PANAMA	BCI	-	HM406591	MH802203	-	MH802256	MH802302
LEP-14256	Cithaerias	pyropina	julia	ECUADOR	Morona Santiago	-	MH802183	MH802226	-	MH802283	MH802328
LEP-14257	Cithaerias	pyropina	julia	ECUADOR	Morona Santiago	-	MH802184	MH802227	-	MH802284	MH802329
PM0095	Cithaerias	pyropina	pyropina	PERU	Pasco, Cañón de Huancabamba	-	MH802192	MH802232	-	MH802291	MH802337

Code	Genus	species	subspecies	Country	Locality						
PM11-05	Dulcedo	polita		COSTA RICA	Heredia, Tirimbina	-	KJ145990	KJ145998	KJ146002	KJ146006	KJ146010
LEP-14258	Dulcedo	polita		ECUADOR	Esmeraldas, Río Chuchubi	-	MH802185	MH802228	MH802250	-	MH802330
FLMNH284117	Haetera	macleannania		ECUADOR	Esmeraldas, Finca Cypris	-	MH802165	MH802209	MH802242	MH802267	MH802311
LEP-04409	Haetera	piera	negra	ECUADOR	Morona Santiago, nr. Yaupi	MH802142	MH802167	MH802211	MH802244	MH802269	MH802313
LEP-08916	Haetera	piera	negra	ECUADOR	Pastaza, Kapawi lodge	MH802147	MH802173	MH802217	MH802245	MH802275	MH802319
PM0091	Haetera	piera	negra	PERU	Pasco, Cañón de Huancabamba	-	MH802191	MH802231	MH802253	MH802290	MH802336
CP01-84	Haetera	piera	pakitza	PERU	Tambopata Research Center	EU141292	DQ018959	DQ018926	EU141475	EU141371	DQ018897
PM14-20	Pierella	chalybaea		BRAZIL	MT, Alta Floresta	KM013148	KM012978	KM013025	-	KM013197	KM013087
LEP-04597	Pierella	chalybaea		ECUADOR	Orellana, Napo Wildlife Center	-	MH802168	MH802212	-	MH802270	MH802314
LEP-08934	Pierella	chalybaea		ECUADOR	Pastaza, Kapawi lodge	MH802151	MH802178	MH802221	-	MH802279	MH802323
PM14-11	Pierella	chalybaea		ECUADOR	Sucumbios, Garza Cocha	MH802155	MH802193	MH802233	-	MH802292	MH802338
PM24-01	Pierella	helvina	incanescens	COSTA RICA	Sarapiqui, Finca Starke	-	MH802197	-	-	-	-
PM24-02	Pierella	helvina	incanescens	COSTA RICA	Heredia, Tirimbina	-	MH802198	MH802235	-	MH802296	MH802342
PM24-03	Pierella	helvina	incanescens	COSTA RICA	Heredia, Tirimbina	-	MH802199	MH802236	MH802254	MH802297	MH802343
FLMNH146191	Pierella	helvina	ocreata	ECUADOR	Esmeraldas, Río Chuchubi	-	MH802164	MH802208	-	-	MH802310
LEP55828	Pierella	helvina	ocreata	ECUADOR	Esmeraldas, San Francisco Ridge	-	MH802166	MH802210	MH802243	MH802268	MH802312
BCI80049	Pierella	helvina	ocreata	PANAMA	BCI	-	MH802156	MH802202	-	MH802255	MH802301
CP01-69	Pierella	hortona	albofasciata	PERU	Tambopata Research Center	MH802135	-	MH802206	-	MH802259	-
LEP-08928	Pierella	hortona	hortona	ECUADOR	Pastaza, Kapawi lodge	MH802150	MH802177	MH802220	-	MH802278	MH802322
CP05-34	Pierella	hortona	hortona	PERU	Paraíso, Río Momón	MH802139	-	-	-	MH802265	MH802308
CP02-85	Pierella	hyceta	ceryce	PERU	Junín, Aldea	MH802136	MH802158	-	-	MH802260	MH802305
CP04-47	Pierella	hyceta	ceryce	PERU	Junín, Río Colorado, Quebrada Perla	MH802138	MH802160	-	-	MH802263	MH802307
LEP-08925	Pierella	hyceta	hyceta	ECUADOR	Zamora-Chinchipe, San Roque, Ridge E	KJ145994	KJ145989	KJ145997	-	KJ146005	KJ146009
LEP-08926	Pierella	hyceta	hyceta	ECUADOR		-	MH802176	-	-	-	-
CP03-60	Pierella	hyceta	hyceta	PERU	Yanachaga-Chemillén	MH802137	MH802159	-	-	MH802261	MH802306
PM0089	Pierella	hyceta	hyceta	PERU	Pasco, Cañón de Huancabamba	-	MH802189	MH802230	-	MH802289	MH802334
PM0090	Pierella	hyceta	hyceta	PERU	Pasco, Cañón de Huancabamba	-	MH802190	-	-	-	MH802335
SnPe05	Pierella	hyceta	hyceta	PERU	Cuzco, San Pedro	-	MH802201	-	-	MH802300	MH802346
NW108-2	Pierella	keithbrowni		BRAZIL	SP, Ilha do Cardoso, Cananéia	-	MH802186	MH802229	-	MH802285	MH802331
LEP-08908	Pierella	lena	brasiliensis (browni)	ECUADOR	Pastaza, Kapawi lodge	MH802144	MH802170	MH802214	-	MH802272	MH802316
LEP-08911	Pierella	lena	brasiliensis	ECUADOR	Morona-Santiago, Wachirpas airfield	MH802145	MH802171	MH802215	-	MH802273	MH802317
LEP-08912	Pierella	lena	brasiliensis	ECUADOR	Pastaza, Kapawi lodge	MH802146	MH802172	MH802216	-	MH802274	MH802318

NW93-2	Pierella	lena	brasiliensis	PERU	Loreto, Río Paiwa	-	DQ338757	DQ338897	-	MH802288	DQ338614
PM14-19	Pierella	lena	lena	BRAZIL	MT, Alta Floresta	KM013139	KM012979	KM013011	-	KM013185	KM013073
PM14-13	Pierella	lena	lena	ECUADOR	Sucumbios, Garza Cocha	-	MH802195	-	-	MH802294	MH802340
LEP-04613	Pierella	lucia		ECUADOR	Orellana, Napo Wildlife Center	MH802143	MH802169	MH802213	-	MH802271	MH802315
LEP-08935	Pierella	lucia		ECUADOR	Pastaza, Kapawi lodge	MH802152	MH802179	MH802222	-	MH802280	MH802324
PM14-12	Pierella	lucia		ECUADOR	Sucumbios, Garza Cocha	-	MH802194	MH802234	-	MH802293	MH802339
CP05-11	Pierella	lucia		PERU	Cordillera del Cóndor	-	MH802162	-	-	-	-
LEP-09797	Pierella	luna	lesbia	ECUADOR	Manabi, above Camarones, Pedernales-Jama Rd.	MH802153	MH802181	MH802224	-	MH802281	MH802326
LEP-09921	Pierella	luna	lesbia	ECUADOR	Esmeraldas, Tundaloma lodge	KM013149	KM012980	KM013026	-	KM013198	KM013088
PM24-04	Pierella	luna	luna	COSTA RICA	Heredia, Tirimbina	-	MH802200	MH802237	-	MH802298	MH802344
BCI90713	Pierella	luna	luna	PANAMA	Argos Plantations	-	HM406585	MH802204	-	MH802257	MH802303
CP14-08	Pierella	nereis		BRAZIL	SP, São Luiz do Paraitingo	MH802141	MH802163	MH802207	-	MH802266	MH802309
LEP-08924	Pseudohaetera	hypaesia		ECUADOR	Zamora-Chinchipe, Quebrada Guayzimi (close to La Merced)	MH802149	MH802175	MH802219	MH802247	MH802277	MH802321
LEP-09562	Pseudohaetera	hypaesia		ECUADOR	Morona Santiago, El Boliche	-	MH802180	MH802223	MH802248	-	MH802325
CP03-99	Pseudohaetera	hypaesia		PERU	Junín, Quebrada Siete Jeringas	-	DQ338758	DQ338898	MH802240	MH802262	DQ338625
CP04-57	Pseudohaetera	hypaesia		PERU	Junín, Mina Pichita	-	MH802161	-	MH802241	MH802264	-
PM0062	Pseudohaetera	hypaesia		PERU	Junín, San Ramón, Nueva Italia	-	MH802188	-	-	-	MH802333

789   Table 2: Characteristics of the molecular dataset used in this study, including gene and specimen

790   coverage, GC content and the number of variable sites.

Genes	Specimens	Length (bp)	Variable sites	Missing data (%)	GC content (%)
CAD	29 (46%)	850	157	25.8	33.5
COI	60 (95%)	1475	474	26.5	29.5
EF1α	49 (78%)	1240	230	17.5	48.6
GAPDH	21 (33%)	691	129	2.6	45.4
RpS5	55 (87%)	617	124	1.9	45.2
wingless	58 (92%)	412	96	9.3	58.6
TOTAL	63	5285	1210	39.2	41.4

794 Table 3: Marginal-likelihood calculations using path sampling and Bayes factor model testing.

795 Eight competing, non-nested species delimitation models were compared, and the STACEY

796 analysis under prior accounting for the number of taxonomic species (*spp prior*) had the highest

797 marginal likelihood estimate. Bayes factors (*ln*BF) = 2–10 were considered to represent positive

798 support, while *ln*Bf > 10 were considered as decisive support.

Delimitation model	Marginal L (Path Sampling)	Delimited species	*ln*BF STACEY, spp prior
STACEY, spp prior	-43522.54	22	0
BP&P, large pop size	-43532.23	21	9.69
Taxonomic spp	-43537.35	18	14.81
STACEY, sspp prior	-43674.24	24	151.70
BP&P, medium pop size	-43689.95	28	167.41
Taxonomic sspp	-43701.81	26	179.27
STACEY, no prior	-43719.44	63	196.90
BP&P, small pop size	-43737.40	55	214.86