

1 **Title Page**

2

3 **Title:** Species limits in butterflies (Lepidoptera: Nymphalidae): Reconciling classical taxonomy  
4 with the multispecies coalescent

5

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20

21 *Abstract.*—Species delimitation is at the core of biological sciences. During the last decade,  
22 molecular-based approaches have advanced the field by providing additional sources of evidence  
23 to classical, morphology-based taxonomy. However, taxonomy has not yet fully embraced  
24 molecular species delimitation beyond threshold-based, single-gene approaches, and taxonomic  
25 knowledge is not commonly integrated to multi-locus species delimitation models. Here we aim  
26 to bridge empirical data (taxonomic and genetic) with the latest coalescent-based species  
27 delimitation approaches. We use the multispecies coalescent model as implemented in two  
28 recently developed Bayesian methods (DISSECT/STACEY and BP&P) to infer species  
29 hypotheses. In both cases, we account for phylogenetic uncertainty (by not using any guide tree)  
30 and taxonomic uncertainty (by measuring the impact of using or not a priori taxonomic  
31 assignment to specimens). We focus on an entire Neotropical tribe of butterflies, the Haeterini  
32 (Nymphalidae: Satyrinae). We contrast divergent taxonomic opinion—splitting, lumping and  
33 misclassifying species—in the light of different phenotypic classifications proposed to date. Our  
34 results provide a solid background for the recognition of 22 species. The synergistic approach  
35 presented here overcomes limitations in both traditional taxonomy (e.g., by recognizing cryptic  
36 species) and molecular-based methods (e.g., by recognizing structured populations, and not raise  
37 them to species). Our framework provides a step forward towards standardization and increasing  
38 reproducibility of species delimitations.

39

40 **KEYWORDS:** Bayes factor; clearwing Satyrinae; Haeterini; Neotropics; speciation; systematics

41

42 INTRODUCTION

43 Traditionally, taxonomic delimitations have relied on diagnostic phenotypic characters to  
44 classify distinct populations into species and subspecies (hereafter, the ‘traditional taxonomic  
45 approach’). More recently, coalescent-based methods that quantify reproductive isolation using  
46 genetic data have been proposed as a means to calculate the probability of speciation (hereafter,  
47 the ‘coalescent approach’; e.g., Knowles and Carstens 2007; Yang and Rannala 2010; Fujita et  
48 al. 2012). Both approaches do not necessarily agree in their species hypotheses because their  
49 scopes are centered on different sections of the speciation continuum; while traditional taxonomy  
50 depends on the evolution of informative and consistent morphological characters, the coalescent  
51 approach is guided by any gene flow reduction that could be associated with the onset of  
52 reproductive isolation. As a consequence, phenotype-based delimitation may not identify cryptic  
53 and incipient species (i.e., genetically divergent lineages embarked in the process of speciation;  
54 Rosindell et al. 2010) whereas coalescent-based delimitation may simply reveal population  
55 structure (i.e., subpopulations with a long non-breeding history) (Sukumaran and Knowles  
56 2017). Despite the paramount importance of delimiting species for multiple disciplines and  
57 practices in science and society (e.g., ecology, evolution, conservation biology, among others), it  
58 still remains unclear how to reconcile conflicts between traditional taxonomy and the coalescent  
59 approach while taking into account their respective benefits and limitations.

60

61 Taxonomists have not fully embraced the recent developments in molecular species delimitation  
62 beyond threshold-based, single-gene approaches (e.g., DNA barcoding based on COI; Hebert et  
63 al. 2003) (Pons et al. 2006; Puillandre et al. 2012). End-users of the coalescent approach, on the  
64 other hand, do not usually incorporate taxonomic knowledge to inform their models (e.g., Leaché

65 and Fujita 2010; Olave et al. 2014; but see Aydin et al. 2014; Jones et al. 2015). In a Bayesian  
66 framework, taxonomic information could be explicitly acknowledged in the form of prior  
67 distributions, and thus alternative species hypotheses can be statistically weighed. Indeed, the  
68 latest Bayesian implementations, which include the multispecies coalescent model (MSC;  
69 Degnan and Rosenberg 2009), can accommodate parameters that control the number of species  
70 in a dataset, their divergence times and ancestral population sizes, all in a single probabilistic  
71 framework. These properties among others, such as the recognition of genealogical incongruence  
72 and incomplete lineage sorting, arguably make MSC-based methods more biologically realistic  
73 than threshold-based molecular species delimitation (Knowles and Carstens 2007; Fujita et al.  
74 2012). However, it remains unexplored how divergent taxonomic opinions affect species  
75 delimitations, when these opinions (such as “splitters” vs. “lumpers”) are translated into prior  
76 distributions for molecular species delimitation analyses.

77  
78 Here we aim to reconcile genetic and taxonomic data using species delimitation models that  
79 estimate coalescence and species divergence in a fully Bayesian framework. We study the  
80 butterflies classified in the tribe Haeterini (Nymphalidae: Satyrinae), insects that exclusively  
81 inhabit tropical rainforests in southern Mexico, Central and South America. Transparent wings  
82 are the most obvious characteristic of this group, an attribute shared by four out of five genera  
83 within the tribe—*Pierella* being the exception by having full scale-cover (Fig. 1).

84  
85 Taxonomic work on this charismatic group involves numerous researchers over the last 150  
86 years (e.g., Herrich-Schäffer 1864; Weymer 1910; Miller 1968; Constantino 1995; Lamas 1997;  
87 Penz et al. 2014; Paluch et al. 2015; Willmott 2015; Zacca et al. 2016). Haeterini is a

88 monophyletic tribe (Wahlberg et al. 2009; Chazot et al. 2018), and current taxonomic  
89 understanding (Lamas 2004; Penz et al. 2014; Willmott 2015; Zacca et al. 2016) is that Haeterini  
90 consists of 29 described species and 39 subspecies. However, these estimates remain contentious  
91 because published taxonomic studies have mostly focused on smaller groups within Haeterini  
92 and have relied on different morphological character systems (wing coloration, genitalia shape or  
93 male androconial organs), and have not taken into account variation in DNA sequence data. The  
94 taxonomic opinions among these authors reflect the well-recognized subjectivity of species-level  
95 taxonomic work, even though Haeterini has a comprehensive taxonomic knowledge as compared  
96 to most other tropical insects. This offers the opportunity to quantify taxonomic opinion and  
97 definitions among authors, as a preliminary step for standardization.

98  
99 Specifically, we aim to test alternative species delimitation hypotheses and to evaluate the  
100 following two interconnected expectations:

- 101
- 102 1. Lineage delimitation using the multispecies coalescent model (MSC) will recover the  
103 taxonomic subspecies in Haeterini. Butterfly subspecies are commonly described based on  
104 parapatric or allopatric geographical distributions, and reduced gene flow could facilitate  
105 phenotypic divergence. If phenotype divergence resulted from genetic differentiation, then any  
106 reproductive isolation among subspecies will be recovered by the MSC.
  - 107  
108 2. Taxonomic knowledge modeled as prior distributions will group specimens into MSC Clusters  
109 (MSCC). However, regardless of prior distribution, the MSC will standardize species  
110 delimitation because divergent taxonomic opinion among authors, if any, will be evident in the

111 phylogeny of Haeterini. Therefore, the MSC will inform on whether different morphological  
112 characters (wing coloration, genitalia shape or male androconial organs) result in narrower or  
113 broader species definitions.

114

## 115 MATERIALS AND METHODS

### 116 *Taxon sampling and molecular dataset*

117 Haeterini butterflies were collected by the authors and collaborators throughout most of the  
118 geographical range of the tribe, including localities from northern Costa Rica to southeastern  
119 Brazil. Specimens were identified to the species and subspecies level following published  
120 taxonomic revisions (Constantino 1995; Lamas 1997; Penz et al. 2014; Paluch et al. 2015;  
121 Willmott 2015; Zacca et al. 2016) and comparing wing morphology to type specimen  
122 photographs at <https://www.butterfliesofamerica.com> (last accessed August 2018). We sampled  
123 all five Haeterini genera, including 18 of 29 currently valid species, and 18 of 39 subspecies  
124 (Lamas 2004; Penz et al. 2014; Paluch et al. 2015; Willmott 2015; Zacca et al. 2016). Note that  
125 several subspecies were not represented in our sample because they are only known from type  
126 collections (hampering their access for genetic studies) or have rarely been collected in recent  
127 years.

128

129 We followed standard lab protocols (Wahlberg and Wheat 2008) to sequence 6 gene fragments  
130 from 63 specimens: the mitochondrial locus COI (1,475 bp) and the nuclear loci CAD (850 bp),  
131 EF1 $\alpha$  (1,240 bp), GAPDH (691 bp), RpS5 (617 bp), and *wingless* (400 bp). Sanger sequencing  
132 was conducted by the company Macrogen (South Korea), and sequence quality control and DNA  
133 alignments were carried out using the program Geneious R7. We retrieved from GenBank the

134 DNA sequences of sixteen species classified in the subfamily Satyrinae in order to root the  
135 phylogeny of Haeterini. All DNA sequences were deposited in GenBank (Table 1) and the DNA  
136 alignments used in this study were archived in TreeBASE (ID: 23439). Detailed specimen  
137 voucher information can be found in Table 1 and voucher photographs can be found at  
138 <http://www.nymphalidae.net/db.php>.

139

#### 140 *Phylogenetic analyses*

141 We inferred phylogenies using single-gene datasets partitioned by codon position to rule out any  
142 tree topology conflict due to contamination. In addition, in order to test the monophyly of each  
143 genus, we inferred a phylogeny using the concatenated multi-locus dataset consisting of 6  
144 genetic markers that proved to be phylogenetically informative (Table 2), and 63 Haeterini  
145 specimens and 16 outgroup taxa. We used PartitionFinder v2.1.1 (Lanfear et al. 2017) to estimate  
146 the best-fit partitioning strategy for the concatenated dataset using 18 data blocks (each codon  
147 position separately for each gene region) and the following settings: branchlengths = linked  
148 (higher likelihood than the unlinked option), models = mrbayes, model\_selection = bic and  
149 search = greedy. All phylogenetic analyses using single-gene and concatenated datasets were  
150 carried out using MrBayes v3.2.3 (Ronquist et al. 2012) through the CIPRES portal (Miller et al.  
151 2010). We used the reversible-jump Markov chain Monte Carlo approach (rjMCMC) to allow  
152 moving across nucleotide substitution schemes (nst = mixed) with different rate variation across  
153 sites (+I and + $\Gamma$ ). Two independent analyses and four chains, one cold and three heated, were run  
154 for 10 million cycles and sampled every 1,000 cycles, discarding the first 25% sampled  
155 parameters as burn-in. We evaluated convergence using the average standard deviation of split  
156 frequencies ( $< 0.005$ ), potential scale reduction factor ( $\sim 1.000$ ), estimated sample sizes (ESS  $>$

157 200), and by inspection of stationary distribution of log-probabilities in both independent runs.  
158 We summarized the 7,500 sampled trees using the 50% majority-rule consensus method.

159

### 160 *Molecular species delimitation*

161 Once we had confidence on the monophyly of each genus and a notion of the phylogenetic  
162 relationships among the 63 specimens, we carried out molecular species delimitation under the  
163 MSC framework. Although the inference of phylogenetic relationships is not a pre-requisite for  
164 subsequent delimitation analyses, it becomes informative when large phylogenies need to be  
165 divided into smaller well-supported subclades, and analyzed separately due to computational  
166 limitation (see BP&P analyses below). We used a comprehensive taxon sampling of Haeterini, a  
167 multi-locus dataset, and two popular Bayesian implementations of the MSC, namely STACEY  
168 (Jones 2017) and BP&P (Yang 2015):

169

170 *STACEY*.—We used DISSECT (Jones et al. 2015) which is a taxonomic assignment-free  
171 Bayesian method for grouping individuals into multispecies coalescent clusters (MSCC). The  
172 method is implemented in the *STACEY* v1.2.4 package (Jones 2017) available in BEAST v2.4.7  
173 (Bouckaert et al. 2014). All gene markers sequenced for this study are likely unlinked in the  
174 genome of the butterflies and thus gene trees, substitution and clock models were all treated as  
175 unlinked in the analysis. We assigned to the mitochondrial COI locus a gene ploidy of 0.5 and to  
176 the remaining nuclear loci a gene ploidy of 2.0 (diploid). Uncorrelated relaxed-clock models  
177 were chosen for all loci, and we estimated nuclear clock rates relative to the COI mean clock rate  
178 fixed to 1.0. The relative clock mean priors were all log normal ( $M = 0$ ,  $S = 1$ ). We used the  
179 birth-death-collapse model following Jones et al. (2015) with GrowthRate prior as log normal ( $M$



180 = 5,  $S = 2$ ) and relative DeathRate as uniform in  $[0, 1]$ , while the popMean prior was set to log  
181 normal ( $M = -7$ ,  $S = 2$ ).

182

183 In STACEY, the discovery of MSCCs relies on two parameters that control node collapsing in  
184 the phylogeny, the collapseHeight ( $\epsilon$ ) and collapseWeight ( $\omega$ ). The parameter  $\epsilon$  distinguishes  
185 very shallow species divergences (node heights) and should be assigned a small value (Jones et  
186 al. 2015), thus, we set  $\epsilon$  to  $1e-4$ . The parameter  $\omega$  controls the number of MSCCs and can be  
187 used as a proxy for prior taxonomic knowledge. The  $1/X$  distribution of the prior for the number  
188 of MSCCs has a mean of  $1 + (n - 1) \times (1 - \omega)$ , where  $n$  is the number of individuals in the  
189 dataset. We set  $\omega$  to 0.73 or 0.59, corresponding to 18 described species and 26 taxa (the 18  
190 sampled subspecies elevated to species), respectively. In addition, we carried out a third analysis  
191 that did not take into account prior taxonomic information on the number of species by using a  
192 Beta distribution ( $\alpha = 2$ ,  $\beta = 2$ ) as a prior for the parameter  $\omega$ . The analyses were run four  
193 independent times for 200 million cycles each, with parameters sampled every 20,000 cycles.  
194 Sampled trees were combined after discarding the first 25% samples as burn-in and checking that  
195 ESS values were  $> 200$ . MSCCs, their posterior probabilities and pairwise similarity  
196 probabilities were obtained using SpeciesDelimitationAnalyser v1.8.0 (Jones et al. 2015)  
197 acknowledging  $\epsilon = 1e-4$ . We chose the clustering with the highest posterior probability (counts)  
198 as the working species hypothesis.

199

200 *BP&P*.—We used the multispecies coalescent model as implemented in BP&P v3.4 (Yang 2015)  
201 to jointly infer species trees and delimit MSCCs, without using any guide tree (Yang and  
202 Rannala 2014; Rannala and Yang 2017) nor taxonomic assignment to specimens (Yang and

203 Rannala 2010; Rannala and Yang 2013). The heredity scalar was set to 1 for all nuclear loci and  
204 to 0.25 for the mitochondrial locus. The prior for ancestral population sizes controlled by the  
205 parameter  $\theta$  was assigned the inverse gamma distribution (IG[ $\alpha$ ,  $\beta$ ]), and we evaluated three  
206 different scenarios: i) large ancestral population size (IG[3, 0.2]), ii) medium ancestral  
207 population size IG[3, 0.1], and iii) small ancestral population size IG[3, 0.02]. These three  
208 different scenarios are expected to impact the inferred number of MSCCs, so that a larger  
209 ancestral population prior would favor fewer species in the model. We separately analyzed two  
210 Haeterini subclades, namely i) *Pierella* and ii) the remaining genera, to overcome computational  
211 limitation in BP&P. We used the inverse gamma distribution for the divergence time of the root  
212 in the species trees, which is controlled by the parameter  $\tau_0$ . The prior distribution of  $\tau_0$  was  
213 diffuse ( $\alpha = 3$ ) and we specified  $\beta = 0.042$  for the *Pierella* species tree and  $\beta = 0.098$  for the  
214 remaining Haeterini genera. These values enforced a sequence divergence mean of 2.1% for  
215 *Pierella* and 4.9% for the remaining genera, which translate into absolute times of ~7 Ma for  
216 *Pierella*'s crown age and ~17 Ma for the remaining genera, assuming a butterfly mutation rate of  
217  $2.9 \times 10^{-9}$  (Keightley et al. 2015). The analyses were run two independent times using the  
218 rjMCMC algorithm 1 with gamma variable fine-tuning shape  $\alpha = 2$  and mean  $m = 1$ , each for  
219 500,000 cycles with sampling frequency of 50 and burn-in of the first 10,000 cycles. We chose  
220 the species delimitation model with the highest posterior probability as the working species  
221 hypothesis.

222

### 223 *Species hypothesis testing and divergence time estimation*

224 We compared the statistical support (model adequacy) for eight species hypotheses in a fully  
225 Bayesian framework:

- 226 i) Taxonomic species (18 lineages);
- 227 ii) Taxonomic subspecies elevated to species (26 lineages);
- 228 iii) STACEY's clusters under  $\omega = 0.73$  (accounting for number of taxonomic species; 22
- 229 lineages);
- 230 iv) STACEY's clusters under  $\omega = 0.59$  (accounting for number of subspecies elevated to species;
- 231 24 lineages);
- 232 v) STACEY's clusters under  $\omega = \text{Beta}(\alpha = 2, \beta = 2)$  (non-informative prior on the number of
- 233 MSCCs; 63 lineages);
- 234 vi) BP&P's clusters under  $\theta = \text{IG}[3, 0.2]$  (large ancestral population size prior; 21 lineages);
- 235 vii) BP&P's clusters under  $\theta = \text{IG}[3, 0.1]$  (medium ancestral population size prior; 28 lineages);
- 236 viii) BP&P's clusters under  $\theta = \text{IG}[3, 0.02]$  (small ancestral population size prior; 55 lineages).

237

238 In order to account for incomplete lineage sorting and to avoid any of the pitfalls of using

239 concatenated datasets (Edwards et al. 2016; Bravo et al. 2018), we inferred species tree topology

240 and divergence times using the Bayesian multispecies coalescent method implemented in

241 StarBEAST2 (Ogilvie et al. 2017). We used jModelTest v2.1.7 (Darriba et al. 2012) to evaluate

242 the substitution models available in starBEAST2, including or not rate variation among sites (+I

243 and + $\Gamma$ ). Nucleotide substitution models for each locus were chosen on the basis of the Bayesian

244 Information Criterion (BIC). We preliminarily evaluated the fit of two tree models, namely Yule

245 and birth-death, using 50 path-sampling steps under thermodynamic integration (Lartillot and

246 Philippe 2006), each running for 60 million cycles to ensure final ESS values  $> 200$ . The Yule

247 tree model had the highest marginal likelihood estimate in all cases and it was preferred over the

248 birth-death model. We therefore report here the path-sampling analyses based on the Yule tree

249 model for the eight species delimitation hypotheses. Other parameters, including gene ploidy,  
250 clock models and popMean prior, were set as in the STACEY analyses. The support for each of  
251 the eight species delimitation hypotheses was assessed via Bayes factors ( $lnBf$ ) (Kass and  
252 Raftery 1995) calculated from the posterior tree samples, and we considered  $lnBf = 2-10$  to  
253 represent positive but not conclusive support and  $lnBf > 10$  as decisive support for the species  
254 hypothesis with the highest marginal likelihood estimated through path sampling.

255  
256 We must rely on secondary calibrations to date the phylogeny of Haeterini because there are no  
257 described fossils assigned to the tribe. Based on a densely sampled, fossil-calibrated butterfly  
258 phylogeny (Chazot et al. 2018), we constrained the ages of six well-supported nodes that do not  
259 belong to Haeterini. We followed a conservative approach by using uniform priors encompassing  
260 the 95% highest posterior density (HPD) intervals from Chazot et al. (2018). The constrained  
261 nodes included the divergences of:

- 262 i) Brassolini and Morphini to 32–58 Ma;
- 263 ii) Melaniti and Dirini to 23–47 Ma;
- 264 iii) Lethina, Parargina and Mycalesina to 25–44 Ma;
- 265 iv) The crown age of Satyrini to 32–53 Ma;
- 266 v) The crown age of the Satyrini’s subclade encompassing the tribes Pronophilina, Euptychiina,  
267 Satyrina, Erebiina, and other closely related subtribes, to 25–43 Ma;
- 268 vi) The crown age of Satyrinae to 41–67 Ma.

269  
270 Time-calibrated species trees were inferred using BEAST v2.4.7 and the analyses were run four  
271 independent times for 200 million cycles each, with parameters sampled every 20,000 cycles.

272 Trees were summarized in TreeAnnotator (part of BEAST v2.4.7) using the maximum clade  
273 credibility method after discarding the first 25% samples as burn-in and merging the four  
274 independent runs in LogCombiner (part of BEAST v2.4.7). Convergence among runs was  
275 evaluated on the basis of ESS values  $> 200$ . For qualitative evaluation we present “cloudograms”  
276 (Figs 2 & 3), which are phylogenetic diagrams that reflect topological uncertainty of species  
277 trees. Cloudograms were recovered using DensiTree v2.2.5 (Bouckaert 2010) based on 500 trees  
278 from the posterior distribution.

279

## 280 RESULTS

### 281 *Data compilation and potential phylogenetic biases*

282 The single-gene and concatenated multi-locus tree topologies were congruent and showed no  
283 evident signature of cross-contamination (Fig. S1). The inferred inter-generic relationships were  
284 robust as indicated by high posterior probabilities (PP = 1.0 for the concatenated multi-locus  
285 tree) and the posterior MSC trees (Figs 2 & 3). All clearwing Haeterini form a monophyletic  
286 group sister to the Haeterini butterflies having full scale-cover on wings. Therefore, in terms of  
287 phylogenetic branching pattern, the genus *Pierella* diverged early in the evolution of Haeterini,  
288 followed by the monotypic genus *Dulcedo*, the genus *Pseudohaetera*, and the divergence  
289 between the genera *Haetera* and *Cithaerias*. Mixed node support for inter-specific relationships  
290 were recovered in single-gene and in the concatenated multi-locus datasets, with PP ranging  
291 from  $\sim 0.6$  to 1.0. The six loci chosen for this study have been previously utilized in butterfly  
292 species-level systematics, thus, we expected these loci to be phylogenetically informative (Table  
293 2). Instead, we recovered low node support among certain Haeterini species (Fig. S1), such as

294 those that rapidly diverged early in the evolution of *Pierella*, or in the recent radiation (< 2.5 Ma)  
295 of the genera *Pierella* and *Cithaerias*.

296

### 297 *Species delimitation*

298 Regardless of which prior was used to take into account taxonomic knowledge for the  
299 collapseWeight parameter ( $\omega$ ), STACEY converged in similar MSCCs suggesting that either 22  
300 (under parameter  $\omega = 0.73$ ) or 24 (under parameter  $\omega = 0.59$ ) lineages are the most adequate  
301 representation of species in our dataset. Indeed the pairwise similarity matrices generated for the  
302 two analyses are highly congruent, though the pairwise posterior probabilities seem to decrease  
303 in the STACEY analysis using  $\omega = 0.59$  (Figure S2). On average, the posterior probability that  
304 two or more specimens inferred by STACEY under parameter  $\omega = 0.73$  belong to a single MSCC  
305 ranged from 0.42 to 0.99, with a median of 0.85 and mode of 0.93. The lowest posterior  
306 probability (0.42) was for the MSCC encompassing *Pierella lena lena*. In contrast, the STACEY  
307 analysis that did not take into account any taxonomy information (parameter  $\omega$  with Beta  
308 distribution [ $\alpha = 2, \beta = 2$ ]) did not group any specimen in the dataset into MSCCs.

309

310 The MSCCs inferred by BP&P showed high sensitivity to prior distributions for the parameter  $\theta$ .  
311 The scenario with small ancestral population size, i.e.,  $\theta = \text{IG}[3, 0.02]$ , suggested that nearly  
312 every specimen in the dataset represent a single divergent lineage. This extreme scenario  
313 suggesting 55 MSCCs was considered in the hypothesis testing exercise using Bayes factors,  
314 even though it significantly departs from current taxonomic understanding of Haeterini. The  
315 most likely numbers of MSCCs recovered by the other two BP&P analyses were 21 (when  $\theta =$   
316  $\text{IG}[3, 0.2]$ ; prior mean of 0.1) or 28 (when  $\theta = \text{IG}[3, 0.1]$ ; prior mean of 0.05). The two

317 independent runs for each analysis converged in the same MSCCs, and the posterior probabilities  
318 of the delimited species for the analysis under  $\theta = \text{IG}[3, 0.2]$  ranged from 0.44 to 1.00, with a  
319 median of 0.96 and mode of 0.99. The two lowest posterior probabilities (0.44) were for the  
320 MSCCs encompassing *Pierella lena lena* and for *Pierella hyceta hyceta*. Overall, the large  
321 impact of the prior distribution of the parameter  $\theta$  in BP&P, resulting in 21, 28 or 55 MSCCs,  
322 suggests that more data, both genetic and taxon sampling, is needed to strongly inform the  
323 molecular species delimitation. However, the resulting most-probable number of MSCCs favored  
324 under the two scenarios for large and medium ancestral population sizes are congruent with the  
325 taxonomic knowledge of the group, as well as with the species delimitation exercises carried out  
326 in STACEY (Figure 2).

327  
328 The eight species delimitation hypotheses ranged from 18 to 63 species (or MSC clusters)  
329 (Figure 2). The species hypotheses recovered by STACEY using  $\omega = 0.73$  (22 MSCCs) and  
330 BP&P using  $\theta = \text{IG}[3, 0.2]$  (21 MSCCs) were strongly supported compared to the remaining six  
331 species delimitation hypotheses ( $\ln Bf \gg 10$ ). The STACEY analysis recovering 22 MSCCs is  
332 supported, but not conclusively ( $\ln Bf = 9.69$ ), over the BP&P analysis recovering 21 MSCCs  
333 (Table 3). The differences between these two species hypotheses are: 1) the delimitation of two  
334 lineages of *Pseudohaetera hypaesia*, one from central Peru (Chanchamayo Valley) and the other  
335 from southeastern Ecuador (Morona Santiago and Zamora Chinchipe Provinces); BP&P  
336 suggested a single MSCC for these two lineages whereas STACEY suggested two separate  
337 MSCCs, 2) split of the subspecies *Pierella helvina ocreata* and *P. helvina incanescens* into two  
338 separate species only by STACEY, and 3) split of *Pierella chalybaea* from Ecuador (Sucumbios)  
339 from remaining *P. chalybaea* only by BP&P. On the other hand, the BP&P analyses under the

340 medium ancestral population size ( $\theta = \text{IG}[3, 0.1]$ ) recovered most of the described subspecies in  
341 Haeterini, thus, we used this delimitation as a proxy for recognizing taxonomic subspecies in the  
342 framework of the MSC despite it not being supported by the Bayes factor model comparison  
343 (Figure 2).

344

### 345 *Absolute divergence times*

346 We show in Figure 3 two time-calibrated species trees, one approximating species in Haeterini  
347 (i.e., STACEY's clusters under  $\omega = 0.73$ ; which had the highest marginal likelihood estimate  
348 among the eight species delimitation hypotheses) and the second approximating subspecies in  
349 Haeterini (BP&P's clusters under  $\theta = \text{IG}[3, 0.1]$ ; which recovered most of the described  
350 taxonomic subspecies). The remaining six species trees can be found in the Supplementary  
351 Information (Fig S3). Species tree topologies and branch lengths in absolute time remained  
352 virtually the same regardless of species delimitation hypothesis. The median crown age of extant  
353 Haeterini was estimated at 27 Ma (95% HPD: 22 to 33 Ma). The rapid early radiation of *Pierella*  
354 happened in the late Miocene, at about 5 to 7 Ma (95% HPD: 3 to 9 Ma), whereas the recent  
355 radiation of *Pierella* and *Cithaerias* occurred in the Pleistocene, at about 1 to 2 Ma (95% HPD:  
356 0.5 to 2.5 Ma). We estimate that these two diversification events gave rise to about 70% of extant  
357 Haeterini species.

358

### 359 DISCUSSION

360 Regardless of the species concept advocated by different researchers, traditional taxonomy and  
361 coalescent-based approaches can act in synergy to infer statistically robust species hypotheses.  
362 Importantly, the multispecies coalescent (MSC) model and the Bayesian framework to delimit



363 MSC clusters (MSCCs) allow the quantification and testing of species boundaries informed by  
364 taxonomic knowledge. Furthermore, divergent taxonomic opinion among authors working on  
365 particular species groups has been statistically weighed by the approach followed in this study.  
366 This is important as a first step towards reliable standardization of taxonomy and higher-level  
367 systematics based on models that take into account the process of speciation (e.g., incomplete  
368 lineage sorting) and not just arbitrary genetic distances as in threshold-based approaches (e.g.,  
369 based on COI barcoding). More realistic models of speciation that include, for example, inter-  
370 specific gene flow might be more accurate at estimating species histories (Müller et al. 2018).  
371 For example, in Europe alone, around 16% of butterfly species are known to hybridize  
372 (Descimon and Mallet 2009), which is in line with the idea of speciation as a continuum (De  
373 Queiroz 2007) and that inter-specific gene flow in animals might not be uncommon (Mallet et al.  
374 2007). Another source of information for taxonomic conclusions is the development of new  
375 methods that jointly model phenotypic traits and genetic data (Solís-Lemus et al. 2015). Taken  
376 together, these methodological advances and the generation of biological data from various  
377 sources suggest that in the near future coalescent-based approaches based on multi-locus data  
378 may be able to overcome many current shortcomings in delimiting species.

379

### 380 *Limitations and strengths*

381 The dataset and approach that we have followed here are not exempt of limitations. First, the  
382 amount of missing information, including taxonomic sampling and gapped molecular dataset,  
383 may have reduced the power of our analyses. However, it has been noted previously that the  
384 MSC might still be accurate with sampling schemes including fewer than five individuals per  
385 lineage, as long as multi-locus datasets (in our case 6 unlinked loci) are utilized (Zhang et al.

2011). Although the impact of missing data in the MSC framework needs to be further  
examined, a recent simulation study suggested that coalescent-based species tree inference might  
be highly accurate even with severely gapped multi-locus datasets (Nute et al. 2018). Second,  
inter-specific gene flow may impact current implementations of the MSC by obscuring lineages  
divergence (Luo et al. 2018). Nevertheless, unless there is a high level of gene flow and  
hybridization, the MSC-based delimitation and species tree inference should be robust (Eckert  
and Carstens 2008; Zhang et al. 2011). However, we note that this issue needs to be further  
studied in the light of more data and using recently-developed approaches, such as the isolation-  
with-migration model (Müller et al. 2018). Third, the impact of priors in MSC-based species  
delimitation might be high when the molecular dataset does not hold sufficient signal to  
converge on the same MSCCs (Leaché and Fujita 2010; Jones et al. 2015). Our results showed  
that the choice of priors heavily influenced the number of MSCCs estimated by MSC-based  
methods. However, our priors rely on existing taxonomic knowledge and thus we have reduced  
the clustering space of 63 specimens based on actual evidence coming from previous  
morphology-based studies. The number of most likely MSCCs inferred here (from 21 to 28  
clusters) remains highly congruent with the morphological diversity encountered in this group of  
butterflies.

The strengths of our study rely on three key aspects. First, we have not used any a priori  
taxonomic assignment of individuals to species, nor any guide species tree to delimit species.  
This allowed us to include both tree topology and taxonomic assignment uncertainties explicitly  
into the models, avoiding potential biases that may have precluded accurate estimation of  
speciation probabilities (Leaché and Fujita 2010). However, we note that the posterior

409 probabilities of certain species delimitations, such as *Pierella lena lena*, remain low, which may  
410 be explained by non-sufficient signal of our molecular dataset. We expect that adding more  
411 genetic data might increase the posterior probabilities while holding the same delimitations,  
412 given that different approaches converged in the same species hypotheses regardless of the low  
413 posterior probabilities (e.g., *Pierella lena lena* was recovered as most probable by both STACEY  
414 and BP&P, Fig. 2). Second, taxonomic knowledge has been formally taken into account in our  
415 probabilistic scenarios. Previous studies have put forward the statistical comparison (e.g., via  
416 Bayes factors) of alternate species delimitation models, but these mostly evaluated different  
417 individual reassignments based only on node collapsing of sister lineages in a phylogeny  
418 (Grummer et al. 2014; Yu et al. 2017). Our pipeline, on the contrary, has aided the exploration of  
419 alternate MSCCs that are congruent as well with other sources of information in butterfly  
420 taxonomy, such as morphology and geography. Third, the MSCCs found in this study are not  
421 nested, meaning that specimens could be re-assigned to any combination of clusterings. The use  
422 of Bayes factors as a selection tool is appropriate because of its flexibility in testing non-nested  
423 models (Leaché et al. 2014). We used path sampling under thermodynamic integration which has  
424 been shown to be highly accurate in testing non-nested species delimitation models (Grummer et  
425 al. 2014).

426

#### 427 *Performance of methods*

428 Our approach focused on using the multispecies coalescent model, informed by taxonomic  
429 knowledge, to assign individuals to MSCCs based on different priors and methods implemented  
430 in STACEY and BP&P. This is arguably a less arbitrary approach to reduce the space of all  
431 possible clusterings for model testing using Bayes factors, compared to other approaches based

432 solely in taxonomic expertise (Leaché et al. 2014) or in multi-locus networks (Grummer et al.  
433 2014) and population assignments (Olave et al. 2014). Furthermore, the approach outlined here is  
434 flexible, simple and fast because it avoids preliminary estimation of guide trees and *a priori*  
435 taxonomic assignments of individuals. Taxonomic knowledge modeled as prior distribution has  
436 the potential to ameliorate potential biases (e.g., population structure raised to species), which  
437 have been observed when relying solely on genetic information and guide trees (Olave et al.  
438 2014; Sukumaran and Knowles 2017). We showed that the number of inferred MSCCs heavily  
439 depends on priors, and indeed, under the scenario of medium ancestral population sizes ( $\theta =$   
440  $IG[3, 0.1]$ ) BP&P recovered traditionally viewed infra-specific diversity, and perhaps structured  
441 populations. Note that subspecies in butterfly taxonomy are seen as populations with limited  
442 gene flow due to allopatric distributions (Lamas 2008). Overall, we highlight the importance of  
443 taxonomically-informed molecular species delimitation and the use of Bayes factor model  
444 comparison.

445  
446 Both MSC-based methods used in this study, STACEY and BP&P, were fast and simple in their  
447 implementation thanks to available guidelines and manuals (e.g., Jones et al. 2015; Yang 2015).  
448 The most supported BP&P's delimitation followed a conservative prior distribution for the size  
449 of ancestral population size, that is  $\theta = IG[3, 0.2]$  with a prior mean = 0.1, which was pointed out  
450 as the most appropriate prior mean (Leaché and Fujita 2010). This species delimitation model is  
451 highly similar to the most supported STACEY model, except for three species hypotheses. In  
452 line with morphological differences, STACEY favored the split of *Pierella helvina incanescens*  
453 and *P. helvina ocreata*, two subspecies that were originally considered separate species but  
454 synonymized due to allopatric geographical distribution by Constantino (1995). On the other

455 hand, BP&P favored the split of one *Pierella chalybaea* specimen (Ecuador, Sucumbíos) from  
456 other conspecific individuals, including others from Ecuador. STACEY did not favor such a  
457 split, which again is in line with the absence of any clear morphological difference in *P.*  
458 *chalybaea* from western Amazonia (Zacca et al. 2016). Finally, STACEY delimited three  
459 separate species within *Pseudohatera hypaesia* while BP&P delimited only two. However, there  
460 are no clear wing coloration differences among populations of *P. hypaesia* throughout its range  
461 from Colombia to Bolivia (Gerardo Lamas, *pers. comm.*). The supported STACEY's  
462 delimitation here thus points out that large genetic divergences exist in the otherwise similar-  
463 looking *P. hypaesia*, warranting a taxonomic revision of these montane butterflies from the  
464 tropical Andes.

465

#### 466 *Taxonomic implications*

467 The probabilistic framework applied in this study allows the statistical test of alternative species  
468 hypotheses in a taxonomic group that has likely evolved for nearly 27 million years. The most  
469 likely scenario among those tested here suggests that at least four divergent lineages should be  
470 elevated to species by current taxonomic standards. It is clear that a more comprehensive  
471 sampling and datasets, including morphological and molecular characters, are needed to robustly  
472 delineate species boundaries in Haeterini. Nonetheless, the four divergent lineages found in this  
473 study corroborate morphological differences that have been previously acknowledged in these  
474 groups but that taxonomists considered conspecific variations (subspecies) mainly because of  
475 allopatric distributions, a historical practice in butterfly taxonomy (Lamas 2008). In almost all  
476 MSC-based analyses, the following taxa were considered independent evolutionary lineages: i)  
477 *Pierella hyceta hyceta* and *Pierella hyceta ceryce*, ii) *Pierella lena lena* and *Pierella lena*

478 *brasiliensis*, iii) *Pierella helvina incanescens* and *Pierella helvina ocreata*, iv) two divergent  
479 sympatric populations in central Peru (Chanchamayo Valley) of *Pseudohaetera hypaesia*.  
480 According to our results, these eight lineages are likely full species (a taxonomic revision in  
481 preparation is addressing this).  
482  
483 Zacca et al. (2016) split *Pierella lamia* into seven species mainly based on allopatric  
484 geographical distributions, and also on genitalia and androconial patches on male wings (despite  
485 their similar wing coloration). We did not find support for the species status of *Pierella*  
486 *keithbrowni* from southeastern Brazil proposed by these authors. In the case of *P. keithbrowni*,  
487 the main diagnostic characters were the androconial patch shape, and ductus bursae in female  
488 genitalia longer than in *P. nice*, albeit male genitalia in *Pierella lamia* complex are not  
489 differentiated. In all MSC-based analyses, *P. keithbrowni* is not genetically different from other  
490 populations in central Brazil and Ecuador. Our results therefore suggest that differences in the  
491 aforementioned morphological characters may represent variation within a single evolutionary  
492 lineage, and thus their usage needs to be complemented with other lines of evidence. Note that  
493 androconial patch morphology has been widely used to diagnose butterfly species boundaries in  
494 other satyrine groups (e.g., Núñez Aguila et al. 2013; Penz et al. 2017), and our results suggest  
495 that the reliability of this character system in different lineages should be further investigated. As  
496 noted previously, granting species status to butterfly populations primarily based on geographical  
497 distribution (allopatric populations) might be unjustified if other evidence fails to recognize clear  
498 divergence (Descimon and Mallet 2009). Therefore, we suggest that *P. keithbrowni* should be  
499 synonymized with *P. chalybaea* (work in preparation will address this).

500

501 CONCLUSIONS

502 Here we show that for the butterfly tribe Haeterini, the multispecies coalescent model generally  
503 recognizes traditionally viewed butterfly subspecies and species, with some exceptions linked to  
504 the use of different morphological character systems. By using a probabilistic framework, we  
505 have shown that divergent taxonomic opinion (concepts) were used by different authors,  
506 including butterfly species that were over-split (*lamia* complex), lumped (at least 6 subspecies  
507 raised to species here), or misclassified (e.g. *Pierella lena browni* was previously synonymized  
508 with *P. lena brasiliensis* (Lamas 2004), despite *browni* being evolutionarily more closely related  
509 to *P. lena lena* than to *brasiliensis*). Androconial patch morphology is commonly used as  
510 informative character to diagnose species, but we showed that at least in *Pierella* it alone may  
511 not be well-suited to distinguish infra- and inter-specific diversity. Furthermore, taxonomic  
512 knowledge informed as priors in MSC-based species delimitation using genetic data is a robust  
513 approach to reduce the clustering space in model testing.

514  
515 The low node support recovered among certain Haeterini species may be attributed to incomplete  
516 lineage sorting due, for example, to ancient rapid radiation as in the crown node of *Pierella* or to  
517 recent speciation as in *Pierella* and *Cithaerias*. Haeterini butterflies evolved for nearly 27 million  
518 years but most extant species (*ca.* 80%) likely diverged rather recently—within the past 2 million  
519 years. Future macroevolutionary studies using the revised species boundaries of Haeterini might  
520 address this puzzling diversification history, whether it was the result of high tropical species  
521 turnover over millions of years (with constant and high extinction and speciation rates, in line  
522 with Antonelli et al. 2015) or a Pleistocene major burst in diversification (in line with the  
523 Quaternary diversification model, as characterized via simulated data; Matos-Maraví 2016).

524

525 SUPPLEMENTARY MATERIAL

526 Figure S1: Consensus trees based on single-gene and concatenated loci datasets (pages 2–9).

527 Figure S2: Pairwise similarity matrices based on delimitation analyses with STACEY (pages 10–  
528 14).

529 Figure S3: Time-calibrated Maximum Clade Credibility species trees of the eight species  
530 delimitation hypotheses (pages 15–24).

531 DNA alignments and phylogenies of the single-gene and concatenated datasets were archived in  
532 TreeBASE (ID: 23439). Accession URL:

533 <http://purl.org/phylo/treebase/phylovs/study/TB2:S23439>

534

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549

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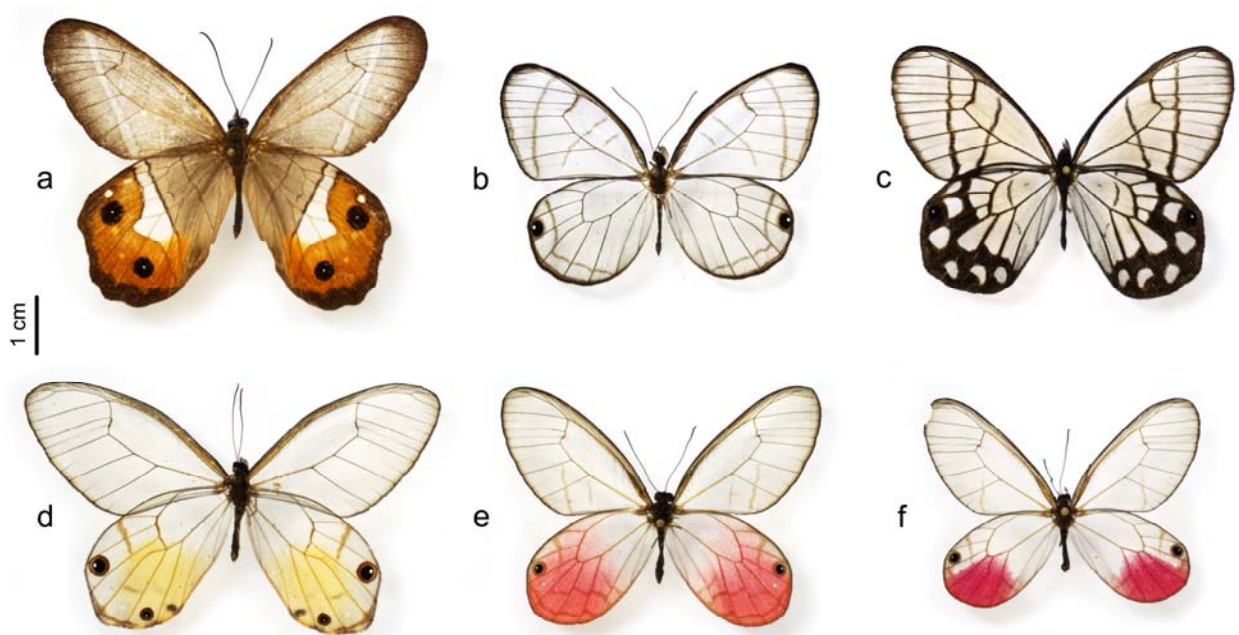
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- 702



703 FIGURES AND TABLES

704 Figure 1: Representatives of the tribe Haeterini. a) *Pierella nereis* (Brazil, Minas Gerais, Santa  
705 Barbara; Milwaukee Public Museum), b) *Dulcedo polita* (Costa Rica, Sarapiquí, Tirimbina  
706 Biological Station; Phil DeVries Collection), c) *Pseudohaetera mimica* (Peru, Junin, Satipo;  
707 Natural History Museum of Los Angeles County), d) *Haetera piera* (Ecuador, Napo, Garza  
708 Cocha; Phil DeVries Collection), e) *Cithaerias cliftoni* (Ecuador, Oriente; Natural History  
709 Museum of Los Angeles County), f) *Cithaerias aurora tambopata* (Peru, Madre de Dios,  
710 Pakitza, Manu National Reserve; Smithsonian Institution).

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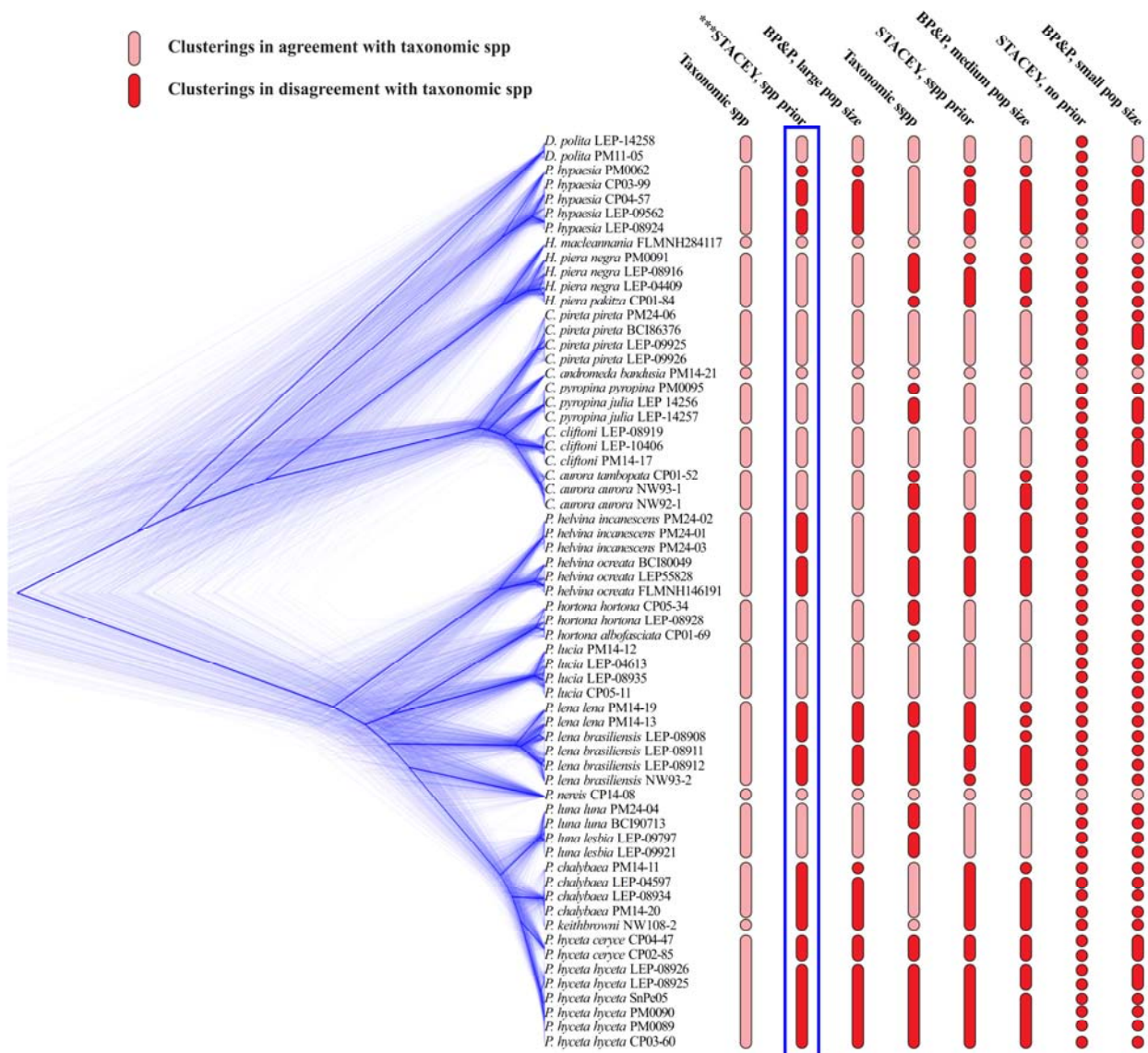


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714

715 Figure 2: Evaluated species delimitation hypotheses using Bayes factors. The eight scenarios  
716 were: Taxonomic species (*spp*, 18 lineages) or subspecies raised to species (*sspp*, 26 lineages),  
717 STACEY's delimited species under prior accounting for number of taxonomic species (*spp*  
718 *prior*, 22 lineages) or number of subspecies raised to species (*sspp prior*, 24 lineages), as well as  
719 with prior not informed by taxonomy (*no prior*, 63 lineages), and BP&P's delimited species  
720 under prior for ancestral population size as large (21 lineages), medium (28 lineages), or small  
721 (53 lineages). The "cloudogram", which is a diagram representing phylogenetic uncertainty of  
722 the 63 Haeterini specimens, was generated based on 500 posterior trees from STACEY analysis  
723 (thicker blue line represents the consensus phylogeny). The delimitation model STACEY under  
724 prior accounting for taxonomic species (outlined by a surrounding box) received significant  
725 support based on Bayes factors over all other models, and thus is the classification that we  
726 propose here.



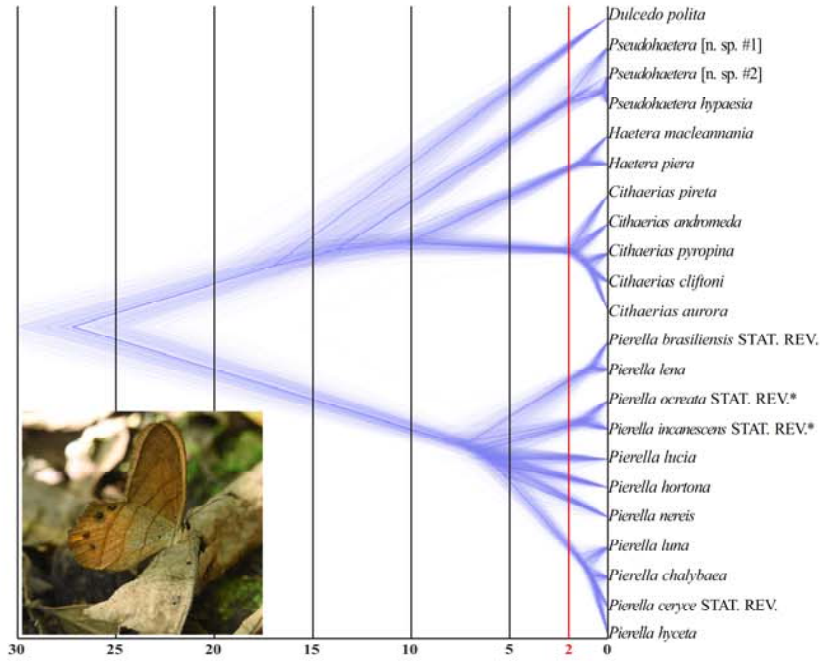
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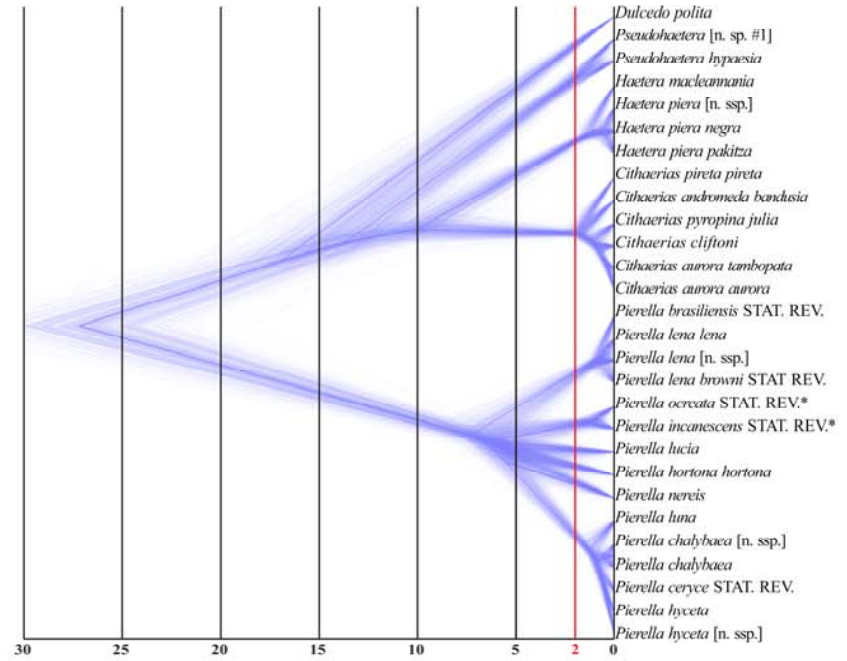
729

730 Figure 3: Time-calibrated phylogenetic hypotheses of models that best approximate species and  
731 subspecies in Haeterini. A) “Cloudogram” of the best-fit species delimitation model based on  
732 Bayes factors, STACEY under prior accounting for taxonomic species. B) “Cloudogram” of the  
733 delimitation model that best approximated described subspecies, BP&P under prior for medium  
734 ancestral population size. Time axis in both panels is scaled to million years. \*The species status  
735 of *Pierella helvina ocreata* and *P. helvina incanescens* may change with the inclusion of *P.*  
736 *helvina helvina*, but it is likely that *P. helvina ocreata* and *P. helvina incanescens* are separate  
737 species. Inset butterfly: *Pierella hyceta hyceta*; Peru, Pasco, Cañón de Huancabamba, 1200 masl,  
738 29.vii.2017. Photo: Markéta Aubrechtová.

**A: Haeterini species tree, STACEY, spp prior**



**B: Haeterini (sub)species tree, BP&P, medium pop size**



739

740 Table 1: Voucher locality information and associated genetic data. All Haeterini specimens were  
741 identified to the species and subspecies level based on the most recent taxonomic revisions  
742 (Constantino 1995; Lamas 1997; Penz et al. 2014; Paluch et al. 2015; Willmott 2015; Zacca et al.  
743 2016). GenBank accession numbers for each of the sequenced locus are presented for every  
744 specimen.

Code	Genus	Species	Subspecies	Country	Locality	CAD	COI	EF1a	GAPDH	RpS5	wingless
EW1-1	Pararge	aegeria		FRANCE	Carcassonne, 2 km S Trassonel	EU141293	DQ176379	DQ338913	EU141476	EU141372	DQ338620
NW121-17	Lethe	minerva		INDONESIA	Bali	EU141309	DQ338768	DQ338909	EU141492	EU141387	DQ338616
EW10-5	Bicyclus	anyana		ZIMBABWE	Harare	EU141295	AY218238	AY218258	EU141478	EU141374	AY218276
CP06-89	Oressinoma	sorata		PERU	Pasco, Oxapampa	MH802140	GQ357209	GQ357278	GQ357440	GQ357570	GQ357342
EW25-17	Orsotriaena	medus		BANGLADESH	Sylhet Div., Lowacherra Forest	-	DQ338766	DQ338906	EU528405	EU528453	DQ338633
NW108-3	Taygetis	virgilia		BRAZIL	SP, Campinas, Santa Genebra	EU141305	DQ338812	DQ338958	EU141487	EU141383	DQ338683
PM07-05	Calisto	smintheus		CUBA	Alrededores de La Platica	JN881807	JN881904	JN881778	JN881827	JN881845	JN881870
CP03-63	Manerebia	cyclopina		PERU	Junín, Quebrada Siete Jeringas	-	DQ338785	DQ338928	EU528397	EU528443	GQ864477
NW162-21	Satyrus	actaea		FRANCE	Aude, Villegly	GQ864709	GQ864807	GQ864901	GQ865030	GQ865494	GQ864495
EW24-7	Erebia	oeme		FRANCE	Languedoc, Ariège 09, Ustou	EU141296	DQ338780	DQ338923	EU141479	EU141375	DQ338640
CP13-01	Aeropetes	tulbaghia		S. AFRICA	Mpumalanga Verloren Valei	-	DQ338579	DQ338907	EU528381	EU528419	DQ338634
NW66-6	Melanitis	leda		AUSTRALIA	Queensland, Cairns	EU141330	AY090207	AY090173	EU141508	EU141408	AY090140
NW122-21	Brassolis	sophorae		BRAZIL	SP, Campinas	-	EU528314	EU528291	GQ357384	EU528425	EU528270
NW66-5	Morpho	helenor		-	London Pupae Supplies	EU141329	AY090210	AY090176	EU141507	EU141407	AY090143
NW121-20	Elymnias	casiphone		INDONESIA	Bali	EU141310	DQ338760	DQ338900	-	EU141388	DQ338627
NW101-2	Zeuxidia	dorhni		INDONESIA	Java, Bandung	-	DQ338752	DQ338892	EU528417	EU528471	DQ338609
PM14-21	Cithaerias	andromeda	bandusia	BRAZIL	MT, Alta Floresta	KJ145993	KJ145988	KJ145996	KJ146001	KJ146004	KJ146008
NW92-1	Cithaerias	aurora	aurora	PERU	Loreto, Río Paiwa	-	MH802187	-	MH802251	MH802286	MH802332
NW93-1	Cithaerias	aurora	aurora	PERU	Loreto, Río Paiwa	-	DQ338756	DQ338896	MH802252	MH802287	DQ338613
CP01-52	Cithaerias	aurora	tambopata	PERU	Tambopata Research Center	MH802134	MH802157	MH802205	MH802239	MH802258	MH802304
LEP-08919	Cithaerias	cliftoni		ECUADOR	Pastaza, Kapawi lodge	MH802148	MH802174	MH802218	MH802246	MH802276	MH802320
LEP-10406	Cithaerias	cliftoni		ECUADOR	Sucumbios, Cuyabeno lodge	KM013123	KM012939	KM012990	KM013275	KM013174	KM013055
PM14-17	Cithaerias	cliftoni		ECUADOR	Sucumbios, Garza Cocha	-	MH802196	-	-	MH802295	MH802341
PM24-06	Cithaerias	pireta	pireta	COSTA RICA	Sarapiquí, Agrícola Sofia, nr. Tirimbina	-	-	MH802238	-	MH802299	MH802345
LEP-09925	Cithaerias	pireta	pireta	ECUADOR	Carchi, Finca San Francisco	MH802154	MH802182	MH802225	MH802249	MH802282	MH802327
LEP-09926	Cithaerias	pireta	pireta	ECUADOR	Pichincha, km. 20 Pacto-Guayabillas Rd.	KM013135	KM012940	KM013007	KM013282	KM013181	KM013069
BCI86376	Cithaerias	pireta	pireta	PANAMA	BCI	-	HM406591	MH802203	-	MH802256	MH802302
LEP-14256	Cithaerias	pyropina	julia	ECUADOR	Morona Santiago	-	MH802183	MH802226	-	MH802283	MH802328
LEP-14257	Cithaerias	pyropina	julia	ECUADOR	Morona Santiago	-	MH802184	MH802227	-	MH802284	MH802329
PM0095	Cithaerias	pyropina	pyropina	PERU	Pasco, Cañón de Huancabamba	-	MH802192	MH802232	-	MH802291	MH802337

PM11-05	Dulcedo	polita		COSTA RICA	Heredia, Tirimbina	-	KJ145990	KJ145998	KJ146002	KJ146006	KJ146010
LEP-14258	Dulcedo	polita		ECUADOR	Esmeraldas, Río Chuchubi	-	MH802185	MH802228	MH802250	-	MH802330
FLMNH284117	Haetera	macleanmania		ECUADOR	Esmeraldas, Finca Cypris	-	MH802165	MH802209	MH802242	MH802267	MH802311
LEP-04409	Haetera	piera	negra	ECUADOR	Morona Santiago, nr. Yaupi	MH802142	MH802167	MH802211	MH802244	MH802269	MH802313
LEP-08916	Haetera	piera	negra	ECUADOR	Pastaza, Kapawi lodge	MH802147	MH802173	MH802217	MH802245	MH802275	MH802319
PM0091	Haetera	piera	negra	PERU	Pasco, Cañón de Huancabamba	-	MH802191	MH802231	MH802253	MH802290	MH802336
CP01-84	Haetera	piera	pakitza	PERU	Tambopata Research Center	EU141292	DQ018959	DQ018926	EU141475	EU141371	DQ018897
PM14-20	Pierella	chalybaea		BRAZIL	MT, Alta Floresta	KM013148	KM012978	KM013025	-	KM013197	KM013087
LEP-04597	Pierella	chalybaea		ECUADOR	Orellana, Napo Wildlife Center	-	MH802168	MH802212	-	MH802270	MH802314
LEP-08934	Pierella	chalybaea		ECUADOR	Pastaza, Kapawi lodge	MH802151	MH802178	MH802221	-	MH802279	MH802323
PM14-11	Pierella	chalybaea		ECUADOR	Sucumbios, Garza Cocha	MH802155	MH802193	MH802233	-	MH802292	MH802338
PM24-01	Pierella	helvina	incanescens	COSTA RICA	Sarapiquí, Finca Starke	-	MH802197	-	-	-	-
PM24-02	Pierella	helvina	incanescens	COSTA RICA	Heredia, Tirimbina	-	MH802198	MH802235	-	MH802296	MH802342
PM24-03	Pierella	helvina	incanescens	COSTA RICA	Heredia, Tirimbina	-	MH802199	MH802236	MH802254	MH802297	MH802343
FLMNH146191	Pierella	helvina	ocreata	ECUADOR	Esmeraldas, Río Chuchubi	-	MH802164	MH802208	-	-	MH802310
LEP55828	Pierella	helvina	ocreata	ECUADOR	Esmeraldas, San Francisco Ridge	-	MH802166	MH802210	MH802243	MH802268	MH802312
BCI80049	Pierella	helvina	ocreata	PANAMA	BCI	-	MH802156	MH802202	-	MH802255	MH802301
CP01-69	Pierella	hortona	albofasciata	PERU	Tambopata Research Center	MH802135	-	MH802206	-	MH802259	-
LEP-08928	Pierella	hortona	hortona	ECUADOR	Pastaza, Kapawi lodge	MH802150	MH802177	MH802220	-	MH802278	MH802322
CP05-34	Pierella	hortona	hortona	PERU	Paraíso, Río Momón	MH802139	-	-	-	MH802265	MH802308
CP02-85	Pierella	hyceta	ceryce	PERU	Junín, Aldea	MH802136	MH802158	-	-	MH802260	MH802305
CP04-47	Pierella	hyceta	ceryce	PERU	Junín, Río Colorado, Quebrada Perla	MH802138	MH802160	-	-	MH802263	MH802307
LEP-08925	Pierella	hyceta	hyceta	ECUADOR	Zamora-Chinchipe, San Roque, Ridge E	KJ145994	KJ145989	KJ145997	-	KJ146005	KJ146009
LEP-08926	Pierella	hyceta	hyceta	ECUADOR		-	MH802176	-	-	-	-
CP03-60	Pierella	hyceta	hyceta	PERU	Yanachaga-Chemillén	MH802137	MH802159	-	-	MH802261	MH802306
PM0089	Pierella	hyceta	hyceta	PERU	Pasco, Cañón de Huancabamba	-	MH802189	MH802230	-	MH802289	MH802334
PM0090	Pierella	hyceta	hyceta	PERU	Pasco, Cañón de Huancabamba	-	MH802190	-	-	-	MH802335
SnPe05	Pierella	hyceta	hyceta	PERU	Cuzco, San Pedro	-	MH802201	-	-	MH802300	MH802346
NW108-2	Pierella	keithbrowni		BRAZIL	SP, Ilha do Cardoso, Cananéia	-	MH802186	MH802229	-	MH802285	MH802331
LEP-08908	Pierella	lena	brasiliensis (browni)	ECUADOR	Pastaza, Kapawi lodge	MH802144	MH802170	MH802214	-	MH802272	MH802316
LEP-08911	Pierella	lena	brasiliensis	ECUADOR	Morona-Santiago, Wachirpas airfield	MH802145	MH802171	MH802215	-	MH802273	MH802317
LEP-08912	Pierella	lena	brasiliensis	ECUADOR	Pastaza, Kapawi lodge	MH802146	MH802172	MH802216	-	MH802274	MH802318



NW93-2	Pierella	lena	brasiliensis	PERU	Loreto, Río Paiwa	-	DQ338757	DQ338897	-	MH802288	DQ338614
PM14-19	Pierella	lena	lena	BRAZIL	MT, Alta Floresta	KM013139	KM012979	KM013011	-	KM013185	KM013073
PM14-13	Pierella	lena	lena	ECUADOR	Sucumbios, Garza Cocha	-	MH802195	-	-	MH802294	MH802340
LEP-04613	Pierella	lucia		ECUADOR	Orellana, Napo Wildlife Center	MH802143	MH802169	MH802213	-	MH802271	MH802315
LEP-08935	Pierella	lucia		ECUADOR	Pastaza, Kapawi lodge	MH802152	MH802179	MH802222	-	MH802280	MH802324
PM14-12	Pierella	lucia		ECUADOR	Sucumbios, Garza Cocha	-	MH802194	MH802234	-	MH802293	MH802339
CP05-11	Pierella	lucia		PERU	Cordillera del Cóndor	-	MH802162	-	-	-	-
LEP-09797	Pierella	luna	lesbia	ECUADOR	Manabi, above Camarones, Pedernales-Jama Rd.	MH802153	MH802181	MH802224	-	MH802281	MH802326
LEP-09921	Pierella	luna	lesbia	ECUADOR	Esmeraldas, Tundaloma lodge	KM013149	KM012980	KM013026	-	KM013198	KM013088
PM24-04	Pierella	luna	luna	COSTA RICA	Heredia, Tirimbina	-	MH802200	MH802237	-	MH802298	MH802344
BCI90713	Pierella	luna	luna	PANAMA	Argos Plantations	-	HM406585	MH802204	-	MH802257	MH802303
CP14-08	Pierella	nereis		BRAZIL	SP, São Luiz do Paraitingo	MH802141	MH802163	MH802207	-	MH802266	MH802309
LEP-08924	Pseudohaetera	hypaesia		ECUADOR	Zamora-Chinchiipe, Quebrada Guayzimi (close to La Merced)	MH802149	MH802175	MH802219	MH802247	MH802277	MH802321
LEP-09562	Pseudohaetera	hypaesia		ECUADOR	Morona Santiago, El Boliche	-	MH802180	MH802223	MH802248	-	MH802325
CP03-99	Pseudohaetera	hypaesia		PERU	Junín, Quebrada Siete Jeringas	-	DQ338758	DQ338898	MH802240	MH802262	DQ338625
CP04-57	Pseudohaetera	hypaesia		PERU	Junín, Mina Pichita	-	MH802161	-	MH802241	MH802264	-
PM0062	Pseudohaetera	hypaesia		PERU	Junín, San Ramón, Nueva Italia	-	MH802188	-	-	-	MH802333

745 Table 2: Characteristics of the molecular dataset used in this study, including gene and specimen  
746 coverage, GC content and the number of variable sites.

747

<b>Genes</b>	<b>Specimens</b>	<b>Length (bp)</b>	<b>Variable sites</b>	<b>Missing data (%)</b>	<b>GC content (%)</b>
CAD	29 (46%)	850	157	25.8	33.5
COI	60 (95%)	1475	474	26.5	29.5
EF1 $\alpha$	49 (78%)	1240	230	17.5	48.6
GAPDH	21 (33%)	691	129	2.6	45.4
RpS5	55 (87%)	617	124	1.9	45.2
<i>wingless</i>	58 (92%)	412	96	9.3	58.6
TOTAL	63	5285	1210	39.2	41.4

748

749

750 Table 3: Marginal-likelihood calculations using path sampling and Bayes factor model testing.  
751 Eight competing, non-nested species delimitation models were compared, and the STACEY  
752 analysis under prior accounting for the number of taxonomic species (*spp prior*) had the highest  
753 marginal likelihood estimate. Bayes factors ( $\ln BF$ ) = 2–10 were considered to represent positive  
754 support, while  $\ln Bf > 10$  were considered as decisive support.

755

Delimitation model	Marginal L (Path Sampling)	Delimited species	$\ln BF$ STACEY, <i>spp prior</i>
STACEY, <i>spp prior</i>	-43522.54	22	0
BP&P, large pop size	-43532.23	21	9.69
Taxonomic <i>spp</i>	-43537.35	18	14.81
STACEY, <i>sspp prior</i>	-43674.24	24	151.70
BP&P, medium pop size	-43689.95	28	167.41
Taxonomic <i>sspp</i>	-43701.81	26	179.27
STACEY, no prior	-43719.44	63	196.90
BP&P, small pop size	-43737.40	55	214.86

756