

1 **Gut microbiota and phytoestrogen-associated infertility in southern white rhinoceros**

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28

29 **Abstract**

30 With recent poaching of southern white rhinoceros (*Ceratotherium simum simum*; SWR)
31 reaching record levels, the need for a robust assurance population is urgent. However, the global
32 captive SWR population is not currently self-sustaining due to the reproductive failure of
33 captive-born females. Dietary phytoestrogens have been proposed to play a role in this
34 phenomenon, and recent work has demonstrated a negative relationship between diet
35 estrogenicity and fertility of captive-born female SWR. To further examine this relationship, we
36 compared gut microbial communities, fecal phytoestrogens, and fertility of SWR to another
37 rhinoceros species—the greater one-horned rhinoceros (*Rhinoceros unicornis*; GOHR), which
38 consumes a similar diet but exhibits high levels of fertility in captivity. Using 16S rRNA
39 amplicon sequencing and mass spectrometry, we identified a species-specific fecal microbiota
40 and three dominant fecal phytoestrogen profiles. These profiles exhibited varying levels of
41 estrogenicity when tested in an *in vitro* estrogen receptor activation assay for both rhinoceros
42 species, with profiles dominated by the microbial metabolite, equol, stimulating the highest
43 levels of receptor activation. Finally, we found that SWR fertility varies significantly with
44 respect to phytoestrogen profile, but also with the abundance of several bacterial taxa and
45 microbially-derived phytoestrogen metabolites. Taken together, these data suggest that in
46 addition to species differences in estrogen receptor sensitivity to phytoestrogens, reproductive
47 outcomes may be driven by gut microbiota's transformation of dietary phytoestrogens in captive
48 SWR females.

49

50 **Background**

51 The southern white rhinoceros (SWR; *Ceratotherium simum simum*) has returned from
52 the brink of extinction through extensive *in situ* and *ex situ* conservation efforts, with wild
53 populations increasing from approximately 100 to 20,000 over the last century (1). However,
54 wild SWR now face an uncertain future due to the recent dramatic increase in poaching (2). An
55 additional challenge facing the species is the reproductive failure of the once robust *ex situ*
56 assurance populations (3,4). Together, poaching, long gestational length (~16 months) and inter-
57 calving interval (~ 2.5 years) (5), and captive infertility (3,4) have rendered both wild and
58 captive populations no longer self-sustaining. Without any change in poaching rates, wild SWR
59 populations will likely face the threat of extinction within the next two decades (6).

60 Previous work has implicated captive diets in the reproductive failure of captive SWR
61 (4,7). In the wild, SWR are pure grazers, consuming up to ~40 kg/day of various grasses (8,9). In
62 contrast, diets in managed settings typically contain phytoestrogen-rich legume hays and soy-
63 and alfalfa-based concentrated feeds (4). A survey of nine SWR-breeding institutions
64 demonstrated that diet estrogenicity was strongly associated with the amount of soy and/or
65 alfalfa-based pellets fed. Moreover, female SWR born at institutions feeding highly estrogenic
66 diets exhibit lower fertility than female SWR born at institutions feeding low phytoestrogen diets
67 (4).

68 Due to their structural similarity to endogenous estrogens, phytoestrogens may interact
69 with estrogen receptors (ERs) and disrupt normal endocrine function, reproduction, and
70 development (10-13). Previously, we showed that SWR ERs exhibit higher maximal activation
71 by phytoestrogens than ERs of the greater one-horned rhinoceros (*Rhinoceros unicornis*; GOHR)
72 (5). Both species consume similar high-phytoestrogen diets in captivity, but GOHR do not

73 exhibit the decrease in fertility observed in SWRs. These data suggest that at the receptor level,
74 SWR are particularly vulnerable to the deleterious effects of phytoestrogen exposure. Whether
75 SWR possess additional species-specific characteristics that predispose them to phytoestrogen
76 sensitivity remains unclear.

77 Due to the limitations of collecting biological samples from a threatened megafaunal
78 species, little is understood about the specific physiological consequences of SWR consuming
79 estrogenic diets. Altered endocrine and reproductive function by phytoestrogen exposure has
80 been described in humans, rodents, and livestock species (11,13,14). Many of these effects,
81 including reproductive tract pathologies, erratic or absent luteal activity, and reduced fertility,
82 parallel findings in captive female SWR (15-17). However, the potential role of phytoestrogens
83 in the onset of these pathologies has not been investigated. In other species, the physiological
84 outcomes of phytoestrogen exposure are profoundly affected by transformation of parent
85 compounds following consumption. For example, in ewes, reproductive pathologies and
86 infertility develop following consumption of diets high in the isoflavone daidzein, but it is equol,
87 a daidzein metabolite, that is thought to be the driver of this effect (10). Equol production relies
88 exclusively on microbial transformation, and several other phytoestrogens are metabolized by
89 members of the gastrointestinal tract microbiota to produce metabolites that vary in estrogenicity
90 (18-21). Coumestrol, a compound from another class of phytoestrogens, the coumestans, also has
91 been associated with sheep infertility (9), but to date, the microbial metabolism of coumestans
92 has not been explored. Whether gut microbiota may play a similar role in SWR responses to
93 dietary phytoestrogens is unclear.

94 The relationship between animals and their associated microbes is important, as
95 microbiota are essential for many biological processes within their hosts (22). However, an

96 understanding of how interactions between phytoestrogens and resident gut microbiota may
97 affect fertility is lacking for any vertebrate species. Given what is known about bioactivation of
98 phytoestrogens by gut microbiota in other mammalian species (23) and the strong link between
99 dietary phytoestrogens and reproductive failures in rhinoceros (4), an investigation into
100 phytoestrogen metabolism by rhinoceros gut microbiota is warranted. To examine these
101 interactions, we characterized SWR and GOHR fecal microbiota as a proxy for gut microbiota.
102 In addition, we compared fecal phytoestrogen composition and metabolite profile estrogenicity,
103 using mass spectrometry and ER activation assays, respectively, between the two species. By
104 sampling separately housed, but similarly managed SWR and GOHR females from the same
105 institution, we sought to reduce variation by eliminating known drivers of gut microbiota
106 composition, such as diet and geographic location (24-26), to better identify species differences.
107 Finally, we used historical breeding records to examine the relationships between specific
108 microbial taxa, phytoestrogen metabolites, and SWR reproductive success. With these data, we
109 shed light on the role microbiota may play in captive SWR infertility with the aim to develop
110 techniques to support and increase this species' assurance population.

111

112 **Results**

113

114 ***Composition of fecal microbiota, but not phytoestrogens, differ by species***

115

116 Sequencing of 16S rRNA fecal samples (SWR: $n = 42$; GOHR: $n = 16$; Table S1) from
117 eight individual rhinoceros (SWR: $n = 6$; GOHR: $n = 2$; Table S1) revealed that GOHR samples
118 had significantly higher inter-sample diversity compared to SWR despite SWRs having a higher
119 number of unique, low (>1 %) relative abundance operational taxonomic units (OTUs) overall
120 (Table S2). Significant differences in fecal community structure and composition between rhino

121 species were also observed at the phylum, family, and OTU level using permutational analysis of
122 variance (PERMANOVA) and accounting for relative abundances using weighted UniFrac (all P
123 < 0.001). A difference in microbial communities was also observed by nonmetric
124 multidimensional scaling (nMDS; Fig. 1, inset). Members of four phyla were found to
125 significantly contribute to variation (Fig. 1), with the relative abundance of the Bacteroidetes
126 (SWR: 55 ± 1.1 %; GOHR: 30 ± 1.8 %) and the Firmicutes (SWR: 33 ± 1.2 %; GOHR: 55 ± 2.2
127 %) found to significantly differ with respect to rhino species (Welch's t -test, both $P < 0.001$; Fig.
128 1). Several members of these phyla were also found to be significantly different at both the
129 family and OTU level, with six families and eleven OTUs contributing to these significant
130 differences (Fig. 1).

131 The observed differences in microbial community are likely related to the different
132 foraging strategies exhibited by the two species. All individuals in this study live in large
133 exhibits where they are provided diet of soy and alfalfa based pellets supplemented with either
134 grasses and browse. SWR, which in the wild are grazers, consume additional hay and fresh
135 grasses (8,9). In contrast, GOHR, a predominantly browsing species, consume a more varied diet
136 that includes fruits and leaves (27). This difference in foraging may be driving species
137 differences in gut microbiota, as observed in other closely related species (28). Nevertheless,
138 both species are herbivorous and their gut microbiota are similar in that the dominant
139 microorganisms present in both species are those capable of fiber degradation and therefore
140 fulfill similar functional niches (29).

141 Despite species differences in microbial communities, neither overall structure nor
142 composition of detected phytoestrogen analytes varied significantly between SWR and GOHR
143 (Fig. 2A, PERMANOVA, $P > 0.05$). However, species differences were observed at the

144 individual analyte level. Concentrations of equol (EQ), enterolactone (EL), methoxycoumestrol
145 (MOC), and coumestrol (CO) were significantly higher in the GOHR (Fig. 2B-D, Table S3).
146 Several phytoestrogens were detected exclusively in the diet, the isoflavones, formononetin (FM)
147 and genistein (GN) (Fig. 2B), whereas microbially-derived metabolites EQ, 4'-ethylphenol
148 (PEP), EL, and enterodiol (ED) were detected only in feces (Fig. 2BC). Two other
149 phytoestrogens, biochanin-A and *o*-demethylangolesin, were not detected in any sample type
150 (both, < 65 ppb). In general, there was an overall trend for excreted quantities of phytoestrogens
151 and metabolites to be higher in GOHR compared to SWR (Fig. 2B-D).

152 The relative abundances of specific OTUs provide some insight into the observed
153 phytoestrogen and metabolite concentrations described above. Overall, 77 OTUs were found to
154 significantly correlate with phytoestrogen concentration (Fig. 3A), which were overall
155 significantly more abundant in SWR compared to GOHR (Welch's *t*-test, $P < 0.0001$; Fig. 3B).
156 Interestingly, the 27 OTUs negatively correlated with phytoestrogen metabolite concentration are
157 nearly four times more abundant in SWR, while the 41 positively correlated OTUs are
158 approximately 5-fold less abundant in SWR (both, Welch's *t*-test, $P < 0.0001$; Fig. 3CD). Taken
159 together, these findings provide a plausible explanation for why there was an overall trend of
160 lower concentrations of individual phytoestrogen analytes in SWR samples.

161 One possible explanation for similarity in phytoestrogen composition between SWR and
162 GOHR is that the OTUs positively associated with phytoestrogen concentrations belong to taxa
163 of known fiber degraders. Many of these taxa, like members of the Bacteroidetes (Rikenellaceae)
164 in SWR, and the Firmicutes (Ruminococcaceae and Lachnospiraceae) in GOHR accomplish this
165 via β -glucosidase activity, as this enzyme also catalyzes early steps of phytoestrogen

166 transformation (23). Thus, the lack of species differences in phytoestrogen composition may be
167 driven by the overall functional similarity of the two species' gut microbial communities.

168

169 ***Three distinct phytoestrogen profiles examined***

170 With no clear species difference in metabolite composition, hierarchical clustering was
171 used to group similar fecal samples from both species of rhinoceros according to their
172 phytoestrogen composition. This approach identified three distinct phytoestrogen profiles
173 representing the most commonly observed fecal metabolite profiles in individual samples from
174 both SWR and GOHR (Fig. 4A-C, Fig. S1, Fig. S1D). For the two most similar profiles, the
175 moderately estrogenic EQ was the dominant metabolite produced, followed by the weakly
176 estrogenic EL (Profile B & C; Fig. 4BC, Table S3, Fig. S1). However, total phytoestrogen
177 concentrations in Profile C were approximately twice the total concentration of phytoestrogens
178 detected Profile B ($8,884 \pm 970$ ppb and $4,254 \pm 315$ ppb, respectively; Fig. 4BC, Fig. S1). A
179 third, less similar profile was also identified, in which the dominant metabolite was EL (Profile
180 A; Fig. 4A, Table S3). The total concentration of phytoestrogens in this profile was significantly
181 lower ($1,510 \pm 229$ ppb) (Fig. 4A, Fig. S1). Despite no visual difference in the overall
182 communities using nMDS (Fig. S1E), several bacterial taxa were found to differ significantly
183 with respect to phytoestrogen profiles (OTU 03, YRC22; OTU 07, OTU 27 Ruminococcaceae).
184 However, no individual OTU contributed to variation $> 8.5\%$, indicating that a group of
185 microbiota, not individual OTUs, may be important in driving differences between
186 phytoestrogen profiles.

187 To quantify the relative estrogenicity of phytoestrogen profiles (Profiles A, B, C), each
188 observed mixture was formulated *in vitro* and tested in estrogen receptor (ER) activation assays

189 using ER α or ER β from SWR and GOHR (Fig. 4D-F, Fig. S1F-G), as described previously (5).
190 All three phytoestrogen profiles activated SWR and GOHR ERs (Fig. 4D-F, Fig. S1F-G), with
191 Profile C, the most potent agonist for both SWR ERs reaching maximal activation relative to
192 17 β -estradiol (E₂) (Fig. 4F, Fig. Fig. S1F-G, Table S4). Similarly, Profile B stimulated maximal
193 activation of SWR ER α , and near maximal activation of SWR ER β relative to E₂, despite having
194 less than half the total concentration of analytes of Profile C (Fig. 4E, Fig. Fig. S1F-G, Table
195 S4). We attribute this high activation primarily to EQ, which is a dominant metabolite in both
196 profiles and a known potent agonist to rhino ERs (5). However, it is interesting that activation of
197 SWR ER α by Profiles B & C was significantly greater than that of GOHR ER α (Table S4), as
198 previous work has shown GOHR ER α to be more sensitive to EQ than its SWR homologue (5).
199 In contrast, the least potent Profile A stimulated significantly greater activation of GOHR ER β
200 relative to SWR ER β (Fig. 4D, Fig. Fig. S1F-G, Table S4). This is also noteworthy, as no
201 phytoestrogen tested in previous studies has ever been shown to be a more potent agonist of
202 GOHR ER β than SWR β . What is driving these differences is unclear, as the dominant
203 metabolites in Profile A, EL and PEP, do not appreciably bind or activate ER β s from either
204 species, and the known agonists present in Profile A (DZ, EQ and CO) are more potent activators
205 of SWR ER β s than those from GOHR (5). Nevertheless, this observation highlights the
206 importance of evaluating the effects of mixtures of suspected endocrine disrupting chemicals on
207 receptors, in addition to individual chemicals, as this method better mimics *in vivo* conditions.

208

209 ***Interactions with SWR fertility explored***

210 To assess fertility of our SWR population, the number of pregnancies achieved and/or
211 calves born were determined for both the period of sample collection as well as for the lifetime

212 of each of the SWR included in this study (Table S1). Pregnancies achieved (Pregnancy_{study}, PS;
213 Pregnancy_{life}, PL) were confirmed via elevations in fecal progestagen levels and were included in
214 the analysis since rhino gestation length (~16 months) exceeded the duration of sample
215 collection (4 months). Fertility (Calf_{study}, CS; Calf_{life}, CL) represents calves born per reproductive
216 year using calculations described previously (4). When comparing phytoestrogen profiles using
217 CS, we did not find any significant difference in mean fertility (Fig. 4G-I; Table S5). Using the
218 PS calculation, however, we showed that individuals exhibiting Profile A had the lowest mean
219 pregnancy rate, and those producing Profile C had the highest (Fig. 4G & I). For lifetime
220 measures, we found a similar relationship, with PL and CL for Profile C producers being
221 significantly greater than Profile A producers (Fig. 4G & I). Although not significantly different,
222 SWR producing Profile B profiles tended to have higher mean fertility than individuals
223 belonging to Profile A across all measures (Fig. 4GH).

224 Mean reproductive success, in terms of pregnancies achieved and calves born, was
225 highest in individuals with the greatest concentrations of fecal metabolites (Fig. 4I; Profile C).
226 For some of the metabolites produced, these findings parallel observations by others. For
227 example, all measures of SWR fertility were positively correlated with production of EL (Table
228 S5). This finding is consistent with studies in humans that have demonstrated a link between
229 high levels of EL and increased reproductive success (31). Our previous work shows EL does not
230 appreciably bind or activate SWR ERs and therefore possesses little endocrine disrupting
231 potential as a xenoestrogen (7). However, the positive relationship between EQ and calf-based
232 fertility measures is unexpected (Table S5). *In vitro*, EQ is a relatively potent agonist of both
233 SWR ER α and ER β (7), and in other vertebrate species EQ is cleared from the circulation less
234 quickly than other isoflavones, increasing its bioavailability (32). This suggests that high levels

235 of EQ production should negatively affect SWR fertility, as is well documented in other grazing
236 species (11,12).

237 Also unexpected was the finding that individual SWR producing the most estrogenic
238 profiles (Profile B & C) exhibited the highest fertility (Fig. 4HI), while SWR fertility was lowest
239 in individuals producing profiles with the lowest overall estrogenicity (Profile A). These
240 observations lead to several new questions. Do SWR belonging to Profile A produce novel
241 phytoestrogen metabolites that are more estrogenic? We observe high levels of certain
242 compounds, such as MOC and CO in feeds, but low levels are detected in feces. CO is a potent
243 SWR ER agonist (7) and has been associated with infertility in sheep (9), but little is known
244 about possible microbial metabolites and their relative estrogenicity. It is possible that these
245 coumestans are converted into a novel metabolite that could be highly estrogenic to SWR.
246 Another possible explanation for the positive association between profile estrogenicity and
247 fertility is that the varying degrees of fecal profile estrogenicity result from differences in
248 phytoestrogen absorption or excretion between individuals. Specifically, it could be hypothesized
249 that elevated excretion of phytoestrogens and metabolites would reduce circulating levels, thus
250 limiting the potential for these chemicals to cause reproductive harm. This is supported not only
251 by our findings in individual animals, but also by our species-level observations where the more
252 fertile GOHR generally excrete higher levels of phytoestrogens than the less fertile SWR. This
253 does not appear to be case in in sheep and cattle, where concentrations of phytoestrogens and
254 metabolites in excreta (i.e., urine) generally correlate to plasma levels (33-35). However, detailed
255 studies examining the generation and clearance of phytoestrogen metabolites, and their
256 subsequent endocrine disrupting effects on target tissues, are lacking even for relatively well-
257 studied species. Addressing such relationships in SWR will be challenging, if not impossible.

258 Nevertheless, the findings presented here do provide the opportunity to apply potentially
259 innovative approaches, like using non-targeted mass spectrometry to identify novel metabolites,
260 or using fecal EL or EQ concentrations to identify individual SWR with high reproductive
261 potential.

262 Few studies have examined the interaction between mammalian fertility and gut
263 microbiota, and defining this link is difficult. Here, we found the abundance of six OTUs to
264 correlate to fertility measures (Table S6). Of the four OTUs that negatively correlate to fertility,
265 three also have significant negative correlations to EL and EQ. However, OTUs positively
266 correlated to fertility did not display significant correlations to any phytoestrogen measured in
267 our study (Table S6), and it is unknown what role these microbiota may play. Using correlations
268 between microbial abundance and phytoestrogen metabolites to determine microbial activity is
269 biased, as compositional data, like presented here, do not directly correlate to microbial activity
270 (36). Therefore, it is possible that less abundant taxa may significantly contribute to the
271 transformation of phytoestrogens. For example, members of the Coriobacteriaceae, which have
272 been shown to convert DZ to EQ (30), the Eubacteriaceae, which are capable of dehydroxylation
273 of lignans to produce ED and EL (37), and the *Blautia* spp., which have displayed in both lignan
274 and isoflavone metabolism (20, 30) are found in samples collected from both SWR and GOHR
275 in low abundances (all, < 1.0 %). However, further work is needed to determine their
276 contributions to phytoestrogen metabolism within the rhinoceros, including *in vitro* culture
277 experiments to measure their microbial activity.

278

279

280

281 **Discussion**

282 Working with threatened species, such as the two species studied here, presents its own unique
283 set of challenges. Despite these challenges, however, our combining of parallel sequencing, mass
284 spectrometry, and estrogen receptor activation assays, provides insight into the host-microbe
285 relationship with fertility that, to our knowledge, is novel for any vertebrate species. Such an
286 approach is needed to understand and apply novel application of techniques within non-
287 traditional systems.

288 Our work sheds light on how microbiota may drive reproductive outcomes in SWR, but
289 they are not the only species that may benefit from the work presented here. Among its broader
290 application to other vertebrates, our findings may be critical for the management of SWR's
291 closest relative, the northern white rhinoceros (*Ceratotherium simum cottoni*, NWR), a
292 subspecies with only two living members (38). Like SWR, NWR experience low fertility and a
293 prevalence of reproductive pathologies in managed settings (17). As a closely related, grazing
294 subspecies, it is likely sensitive to phytoestrogens as well. Currently, several rescue attempts are
295 underway to prevent NWR extinction (39,40). Should these attempts to save the NWR be
296 successful and with SWR facing a similar uncertain fate, any novel approaches to promote high
297 fertility, such as managing microbial phytoestrogen transformation by altering microbiota
298 through diet modifications and other therapeutic approaches, will be needed. With the
299 information presented here, we plan to direct future work aimed at developing strategies to
300 improve captive SWR reproduction, with the ultimate goal of alleviating their threat of
301 extinction.

302

303

304

305 **Methods**

306 **Study animals.** Female greater one-horned rhinoceros ($n = 2$) and southern white rhinoceros (n
307 = 6) used in this study were housed at the San Diego Zoo Safari Park, Escondido, CA, USA in
308 two separate 24 ha mixed species exhibits (Table S1). All procedures were approved by San
309 Diego Zoo Global's Institutional Animal Care and Use Committee (#15-013).

310

311 **Sample collection.** Fresh fecal samples (SWR, $n = 42$; GOHR, $n = 16$, Table S1) were collected
312 weekly beginning September 3, 2015 through January 1, 2016, alternating weeks between SWR
313 and GOHR. Samples were collected from animals at the same time of day using binoculars to
314 identify individuals based on their unique horn structure. Following defecation, collection
315 occurred between one to twenty minutes, and samples were transported on dry ice and stored at -
316 80 °C prior to processing.

317

318 **DNA extraction.** Total genomic DNA from fecal samples and negative control was extracted via
319 mechanical disruption and hot/cold phenol extraction following Stevenson *et al.*'s protocol
320 (2007) with the exception that 25:24:1 phenol:chloroform:isoamyl alcohol was used in place of
321 phenol:chloroform at all steps. DNA was quantified using a Qubit Fluorometer (Invitrogen,
322 Carlsbad, CA, USA) and stored at -20 °C following extraction.

323

324 **Library Preparation & Sequencing.** Sequencing library preparation was carried out following
325 manufacturer's recommendations (Illumina, 2013) with some modifications. In brief, amplicon
326 PCR targeted the V4 region of the 16S rRNA gene using a forward (V4f:

327 TATGGTAATTGTGTGCCAGCMGCCGCGGTAA) and reverse (V4r: AGTCAGTCAGC
328 CGGACTACHVGGGTWTCTAAT) primers in a 25- μ L reaction with 1X KAPA HiFi Hot Start
329 Ready Mix (Kapa Biosystems), 0.2 mM each primer, and 1.0 - 5.0 ng DNA (31). Amplification
330 conditions were as follows: 95 °C for 2 min, 25 cycles of 95 °C for 20 s, 55 °C for 15 s, 72 °C
331 for 30 s, and a final 10 min extension at 72 °C. PCR products were purified via gel extraction
332 (Zymo Gel DNA Recovery Kit; Zymo, Irvine, CA) using a 1.0 % low melt agarose gel (National
333 Diagnostics, Atlanta, GA) and quantified with a Qubit Fluorometer (Invitrogen). With the
334 negative control producing no band, the expected area was excised. All samples were combined
335 to yield an equimolar 4 nM pool. Following manufacturer's protocol, sequencing was conducted
336 on an Illumina MiSeq using reagent kit V2 (2 x 250 bp cycles), as described previously
337 (Illumina, 2013). All sequences were deposited into the National Center for Biotechnological
338 Information's Short Read Archive under Accession Number SRP136468.

339
340 **16S rRNA sequence analyses.** Sequence analysis was carried out using mothur v.1.39.5 (43)
341 following the MiSeq SOP (42). In brief, contigs were formed from 16S rRNA reads, and poor
342 quality sequences were removed. Sequences were trimmed and filtered based on quality
343 (maxambig = 0, minlength = 250, maxlength = 500). Unique sequences were aligned against the
344 SILVA 16S rRNA gene alignment database (44) and classified with a bootstrap value cutoff of
345 80, and operational taxonomic units (OTUs) found with < 2 sequences in the total dataset were
346 removed. Chimeras (chimera.uchime) and sequences identified as members of Eukaryota,
347 Archaea, Cyanobacteria lineages, and mitochondria were also removed. Sequences were
348 clustered into OTUs at a 97 % similarity cutoff using OptiClust (OTU table, Table S7). Negative

349 control yielded 273 sequences, comprised of low-level cross-sample contaminants; therefore,
350 OTUs were not removed from dataset.

351 Sequence coverage was assessed in mothur by rarefaction curves (Fig. S2) and Good's
352 coverage (45). Samples were then iteratively subsampled 10 times to 6,825 sequences per
353 sample, and OTU abundances were calculated as whole number means across iterations.
354 Additionally, richness and diversity were calculated for each sample. All other calculations were
355 carried out in R using both *vegan* and *phyloseq* packages (46,47). The similarity indices Bray-
356 Curtis (48), Jaccard (49), and weighted UniFrac (50) were used to assess differences in bacterial
357 community, and these differences were visualized by nonmetric multidimensional scaling plots
358 (nMDS, iters=10,000) (51). Permutational analysis of multivariate dispersions (PERMDISP2)
359 was used to test for heterogeneity of community structure and composition between rhino
360 species, and with unequal variances observed, data were down-sampled to create even sample
361 sizes using the *caret* package (52) prior to permutational analysis of variance (PERMANOVA,
362 *vegan::adonis*, SWR, $n = 16$; GOHR, $n = 16$) to determine species differences. Similarity
363 percentages (SIMPER, *vegan*) analyses then determined the contributions from each taxonomic
364 group to PERMANOVA reported differences. Species-related differences in individual OTUs
365 were examined by Welch's *t*-test (two-sided, SWR, $n = 16$; GOHR, $n = 16$). All data are
366 expressed as the mean \pm SE and considered significant if $P < 0.05$ unless otherwise stated.

367
368 **Phytoestrogen extraction and quantification.** Samples collected were batched into groups of
369 ten and accompanied by quality control samples. Phytoestrogens were extracted from fecal
370 samples by a two-phase extraction as described previously by Palme *et al.* (2013) with few
371 modifications. In the first phase, fecal samples were diluted ten-fold using 80 % methanol in

372 water (Fisher Scientific), homogenized for 20 min using a Geno/Grinder[®] at 1,000 rpm,
373 centrifuged for 10 min at 4,000 x g, and the supernatant was recovered. In the second-phase, 1.0
374 mL of methanol extract was added to 4.0 mL diethyl ether (Fisher Scientific), 0.5 mL of 5.0%
375 NaH₂CO₃ (Sigma), and 4.0 mL of water (54), inverted four times, and centrifuged for 10 min at
376 4000 x g. The ether phase was removed, evaporated at 45 °C by a nitrogen flow of 0.4 psi and re-
377 suspended in methanol. Extracts were further filtered (0.22 µm) and analyzed by liquid
378 chromatography-coupled tandem mass spectrometry (LC/MS/MS) for all analytes with the
379 exception of 4'-ethylphenol (PEP) which was analyzed by gas chromatography-mass selective
380 detector (GC/MSD). Quality control samples included a blank matrix sample (grass) that was
381 absent of phytoestrogens to assess contamination during the extraction and a matrix spiked
382 sample which was fortified with a known concentration of phytoestrogens. The spiked matrix
383 sample was used to determine the efficiency of the extraction for every batch; recoveries ranged
384 between 50-150 %.

385
386 **LC/MS/MS method.** Analysis was performed using an Agilent 1260 liquid chromatograph
387 coupled to an Agilent 6430 Triple Mass Spectrometer. Chromatographic separation was
388 performed using an Agilent Zorbax Eclipse Plus (2.1 x 50 mm id, 1.8 µm) Rapid Resolution
389 column maintained at 40 °C. The mobile phases consisted of 5 mM ammonium formate and 0.1
390 % formic acid in water for the aqueous phase (A), with 5 mM ammonium formate and 0.1 %
391 formic acid in methanol as the organic phase (B). The flow rate was held at 0.4 mL/min and
392 gradient program was as follows: 0-0.5 min 10 % B, 0.5-3.0 min increasing to 90 % B. The
393 ionization of phytoestrogens was performed using electrospray ionization (ESI) in positive mode
394 with an auxiliary gas (N₂), source temperature of 300 °C, and a gas flow rate of 12 L/min, with

395 the exception of enterodiol which was run in negative mode. Optimized MRM conditions are
396 listed in Table S8.

397 **GC/MSD method.** The analysis of PEP (Indofine, CAS: 123-07-09) was performed on an
398 Agilent 7890B gas chromatograph (GC) coupled to an Agilent 5977A mass selective detector
399 (MSD). The GC inlet temperature was set to 280 °C run in pulsed splitless mode with an
400 injection volume of 1 µL. The GC oven temperature was set to 80 °C and increased to 200 °C
401 between 1 and 13 minutes at a rate of 10 °C/min. The oven temperature was then increased to
402 300 °C between 13 and 22 minutes at a rate of 25 °C/min for a total run time of 22 minutes. Ultra
403 high purity helium (carrier gas) was used at a constant flow rate of 1.5 mL/min with an Agilent
404 DB-5MS UI (30 m x 0.250 mm) 0.25 µm analytical column. PEP was analyzed using an electron
405 ionization (EI) source with a source temperature of 230 °C. Selected ion monitoring (SIM) mode
406 was to monitor 77, 107, and 122 (m/z) ions with a gain factor of 10 and a scan speed of 1,562
407 (u/s).

408
409 **Phytoestrogen analyses.** Similar methods to those used for 16S rRNA analyses were used in
410 determining species differences in phytoestrogen analyte composition. Differences were
411 visualized following nMDS of Bray-Curtis and Jaccard similarity indices, and following
412 normalization, normality testing, and down-sampling, PERMANOVA was used to determine if
413 species differences were observed (SWR, $n = 16$; GOHR, $n = 16$). Welch's *t*-test was again used
414 to measure significant difference between rhino species for individual analytes (SWR, $n = 16$;
415 GOHR, $n = 16$). Since we did not observe a species-related difference using PERMANOVA,
416 and no apparent clustering was observed with nMDS, hierarchical clustering (Bray-Curtis) was
417 used to group phytoestrogen data into three profiles (A, B, C) based on their compositional

418 similarity for further analyses. SIMPER analysis was used to determine contributions of each
419 analyte to differences observed, and significant differences between groups were tested using
420 analysis of variance (ANOVA; Profiles A, $n = 23$, B, $n = 26$, C, $n = 9$) with FDR correction.

421
422 **Receptor Activation.** The ability of phytoestrogens and metabolites to activate SWR ERs was
423 assessed using a SWR and GOHR estrogen receptor (ER) activation assay described previously
424 by Tubbs *et al.* (2012; 2016) with minor modification. For each species, ER α or ER β sub-cloned
425 into pcDNA3.1(+) expression plasmid (Invitrogen) was co-transfected into human embryonic
426 kidney (HEK293) cells along with pCMX- β -galactosidase (β -gal), pGL2-3xERE luciferase
427 reporter plasmid. After 24 hr cells were treated with phytoestrogens or metabolites, alone or in
428 combination, and incubated for an additional 24 hr. For single test compounds, cells were treated
429 with 100 pM-10 μ M of each compound or vehicle (DMSO) alone. To assess the estrogenicity of
430 phytoestrogen/metabolite profiles produced by SWR and GOHR microbial communities, cells
431 were treated with serial dilutions of mixtures created *in vitro* to reflect those generated *in vivo*.
432 Within each assay, a series of cells was treated with the endogenous estrogen, 17-estradiol (E₂;
433 0.001-100 nM), to determine maximal E₂ activation. Following incubation, cells were lysed and
434 luciferase and β -Gal activity was measured as described previously (Tubbs *et al.*, 2012). All data
435 are presented as mean \pm SE fold activation over vehicle treatment for each metabolite or mixture
436 relative to the maximal activation E₂. Differences in mean activation for both ERs were
437 determined by ANOVA with FDR correction (each treatment, $n = 9$ (inter-assay, $n = 3$, intra-
438 assay, $n = 3$)), and significant interactions were observed between rhinoceros species and
439 phytoestrogen profile. Data were sliced according the main effects, and differences within factor
440 were observed.

441

442 **Fertility.** Four calculations of SWR fertility were conducted for this study. Similar to Tubbs *et*
443 *al.*, (2016), two calculations are based on the number of offspring produced by the female per
444 reproductive year through the completion of the study (Calf_{study}; CS) and current levels (Calf_{life};
445 CL). With pregnancies also considered a success, we conducted two additional calculations
446 based on the number of pregnancies per reproductive year ending with the completion of the
447 study (Pregnancy_{study}; PS) through current (Pregnancy_{life}; PL). GOHR samples were removed
448 from the dataset so that data were not skewed by species differences.

449

450 **Correlations.** Using the *microbiome* package in R (55), we examined significant correlations
451 between OTUs (≥ 1.0 % relative abundance) and phytoestrogen analytes, OTUs and fertility
452 measures, and those between phytoestrogen analytes and fertility measures. These correlations
453 were carried out using the Spearman correlation method with multiple testing correction by FDR.
454 OTUs found to correlate to fertility were further examined using a linear model (lm, *stats*
455 package) (56).

456

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467

468 **Author contributions**

469

470 CW and CT designed the experiment, and with AY and AM, CW conducted experimentation.

471 Data analysis was carried out by CW. The manuscript was written by CW, CT, AM, and AY,

472 with editorial assistance by BD.

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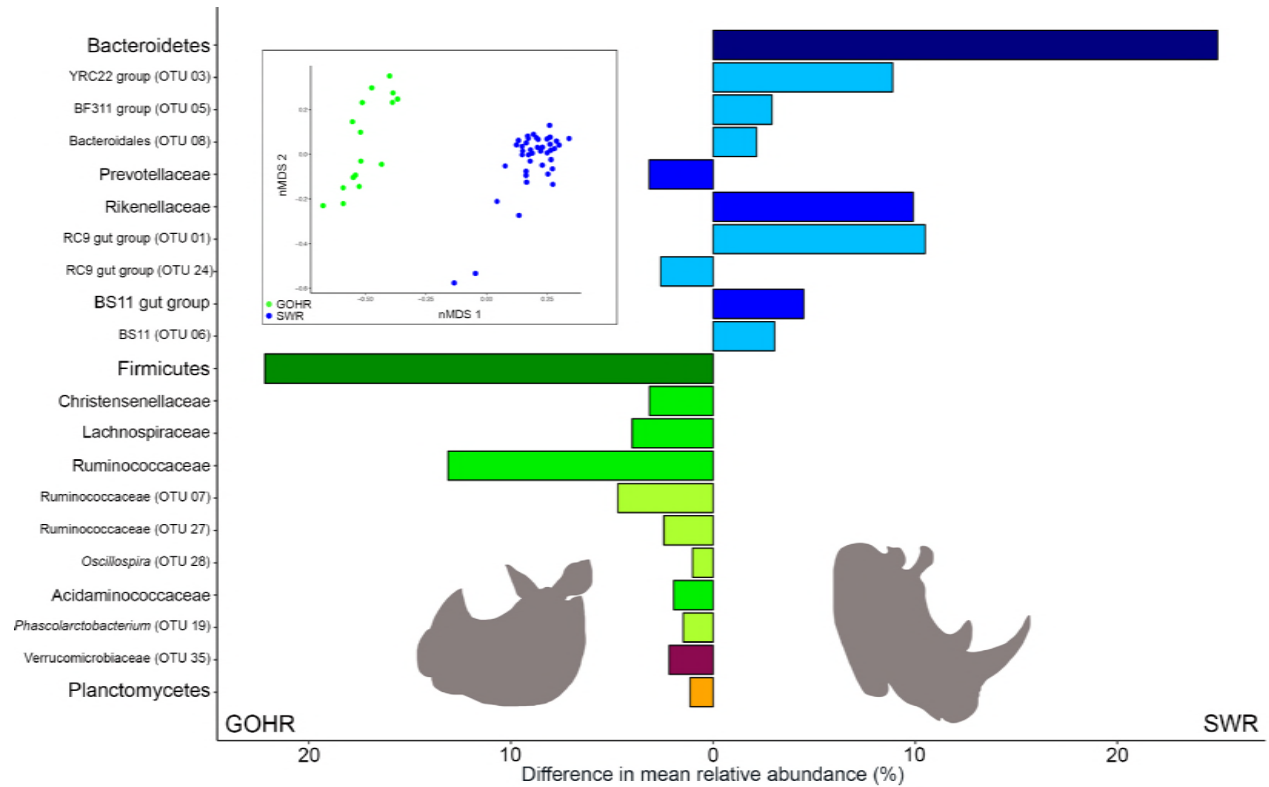
474 **Competing interests**

475 The authors declare that they have no competing interests.

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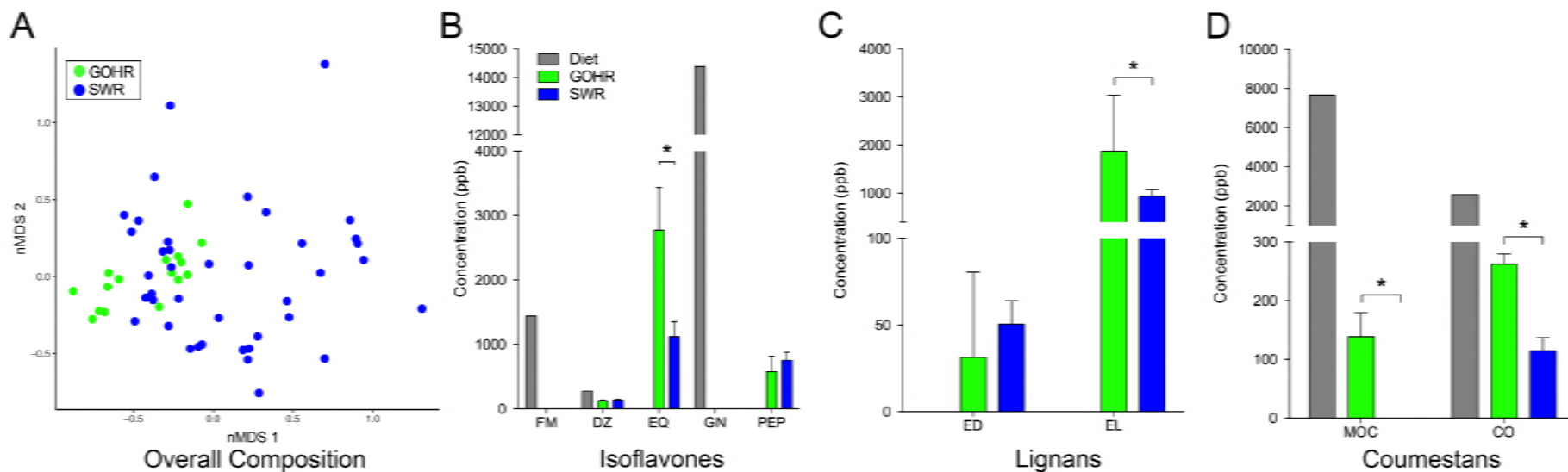
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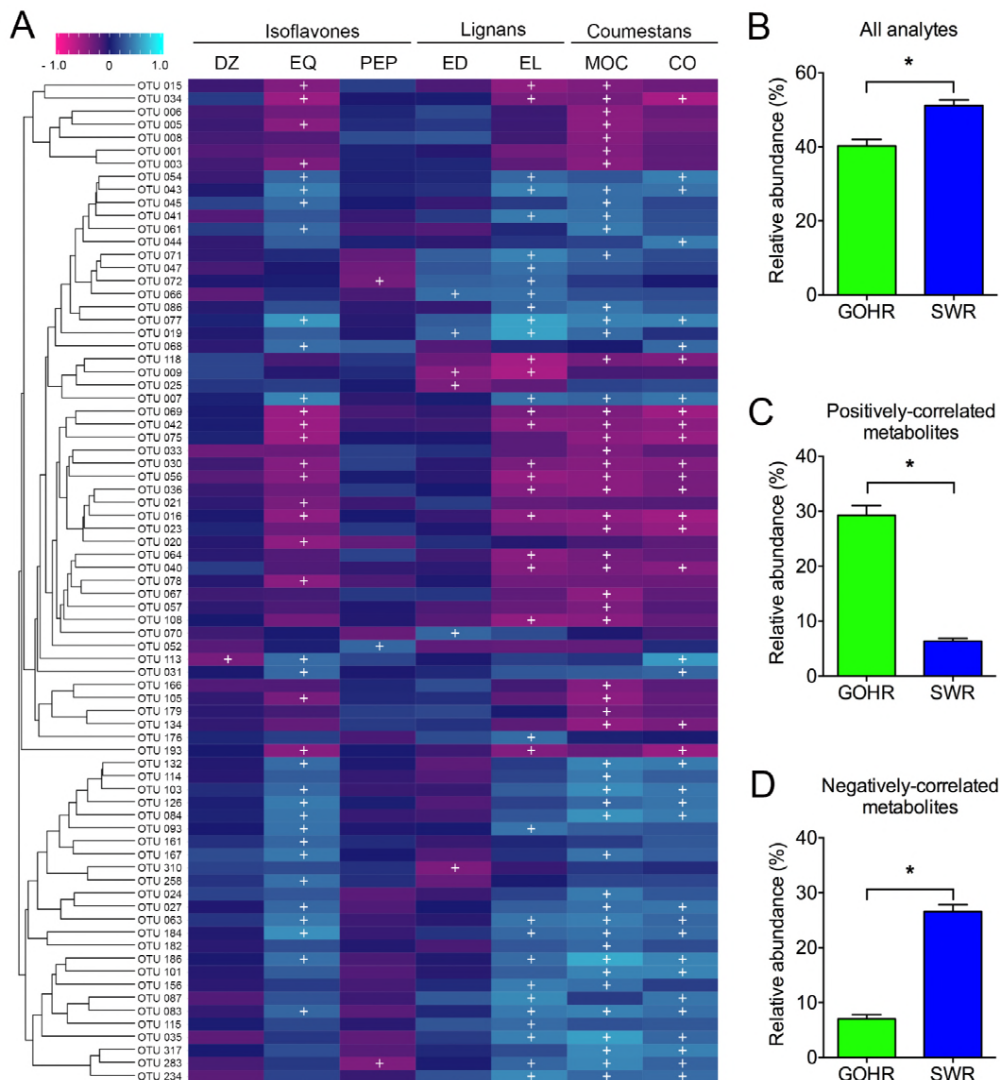


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Figure 1. Differences in fecal microbiota between southern white rhinoceros (SWR) and greater one-horned rhinoceros (GOHR). Nonmetric multidimensional scaling (nMDS, inset) analysis displaying differences in microbiota observed by 16S rRNA amplicon sequencing based on Bray-Curtis distances (PERMANOVA, $P < 0.001$; stress: 0.13). Differences in mean relative abundance of bacterial taxa found to significantly contribute to variation between rhinoceros species (SIMPER ≥ 2.0 %; Welch's t -test, $P < 0.05$) are organized by color, with all members of a particular phylum sharing a similar color, with intensity decreasing from phylum to family to OTU level

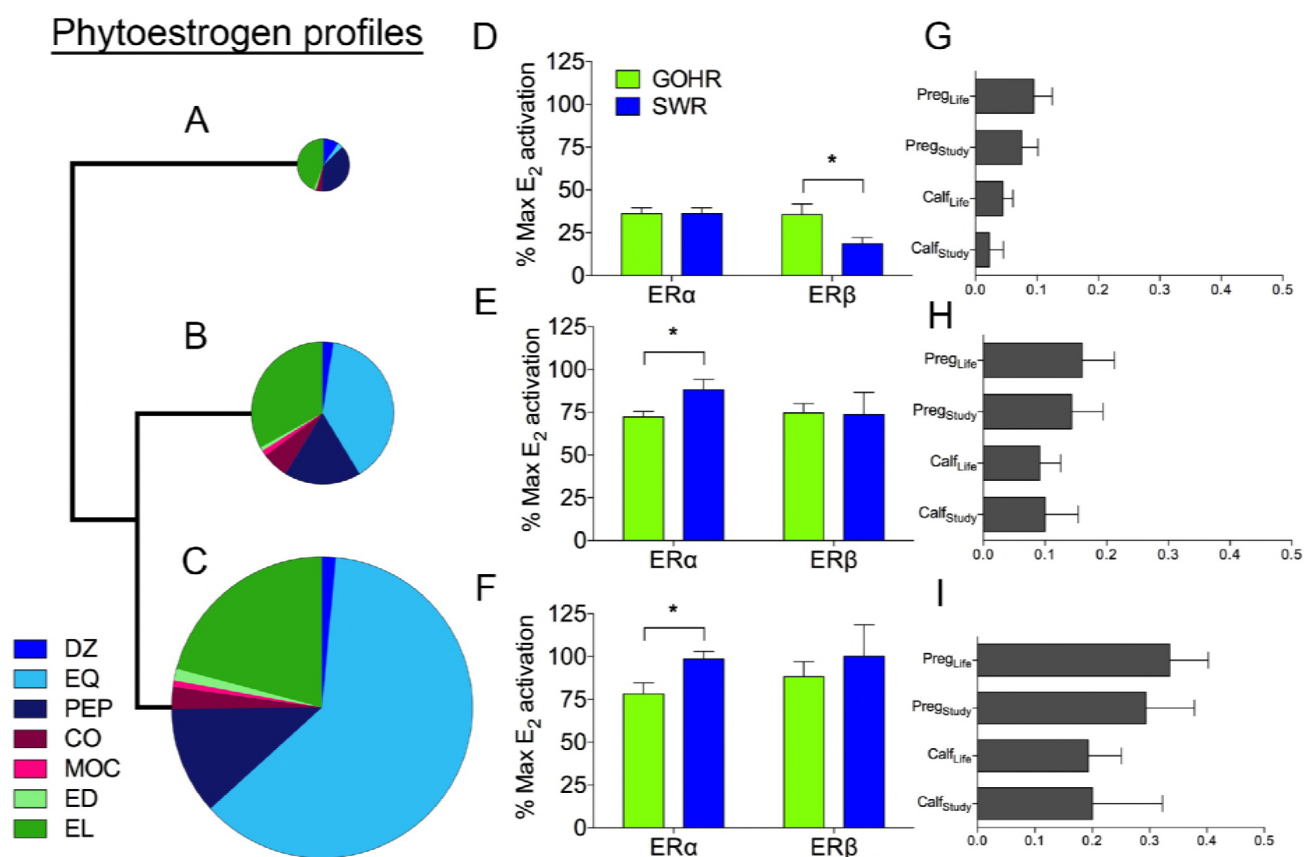


490
 491 **Figure 2.** Comparison of fecal phytoestrogen composition between southern white rhinoceros (SWR) and greater one-horned
 492 rhinoceros (GOHR) A) Nonmetric multidimensional scaling (nMDS) analysis displaying overall composition of fecal phytoestrogens
 493 detected by mass spectrometry based on Bray-Curtis distances (PERMANOVA, $P > 0.05$; stress: 0.13). Mean \pm SE analyte
 494 concentrations in parts per billion (ppb) of B) isoflavones, C) lignans, and D) coumestans for both SWR and GOHR and their diet.
 495 *Significantly different concentrations of fecal analytes (Welch's t -test, $P < 0.05$). FM: formononetin, DZ: daidzein, EQ: equol, GN:
 496 genistein, PEP: 4'ethylphenol, ED: enterodiol, EL: enterolactone, MOC: methoxycoumestrol, CO: coumestrol.



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Figure 3. Relative abundance of OTUs and phytoestrogen concentrations significantly correlate. A) Heatmap depicting significant correlations between phytoestrogen analytes and microbiota (≥ 1.0 % relative abundance) using Spearman correlation method with FDR correction (+Significance at $P < 0.05$). Dendrogram displays OTUs that commonly co-occur by hierarchical clustering (Bray-Curtis). Species differences in mean \pm SE relative abundance of observed OTUs correlating to B) phytoestrogen analytes, C) positively correlated metabolites, and for D) negatively correlated metabolites. * Welch's *t*-test, $P < 0.05$. DZ: daidzein, EQ: equol, PEP: 4'ethylphenol, ED: enterodiol, EL: enterolactone, MOC: methoxycoumestrol, CO: coumestrol, GOHR: greater one-horned rhinoceros, SWR: southern white rhinoceros.



509

510 **Figure 4.** Relative estrogenicity and fertility of phytoestrogen profiles identified by hierarchical
 511 clustering. Phytoestrogen composition, as depicted by hierarchical clustering, with each profile's
 512 size relative to total concentration detected by mass spectrometry, A) Profile A, B) Profile B,
 513 and C) Profile C. The mean \pm SE activation of ER α and ER β of both southern white rhinoceros
 514 (SWR) and greater one-horned rhinoceros (GOHR), relative to maximal activation by 17 β -E₂ by
 515 respective phytoestrogen profiles D) Profile A, E) Profile B, and F) Profile C, when tested at
 516 concentrations found *in vivo*. Differences in mean \pm SE fertility measurements with respect to
 517 phytoestrogen profiles, G) Profile A, H) Profile B, and I) Profile C, *Significantly different
 518 activation (ANOVA, $P < 0.05$). DZ: daidzein, EQ: equol, PEP: 4'ethylphenol, ED: enterodiol,
 519 EL: enterolactone, MOC: methoxycoumestrol, CO: coumestrol.

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