Inference of the worldwide invasion routes of the pinewood nematode, *Bursaphelenchus*

***xylophilus*, using ABC analysis on microsatellite data**

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ABSTRACT

16   Population genetics have been greatly beneficial to improve knowledge about

biological invasions. Quantitative genetic methods, such as Approximate Bayesian

Computation (ABC), have brought this improvement up a notch and are now essential tools to

decipher the invasion routes of any invasive species. In this paper, we performed ABC

analyses to shed light on the pinewood nematode (PWN) worldwide invasion routes and

identify the source of the European populations. Originating from North America, this

microscopic worm has been invading Asia since 1905 and Europe since 1999, causing

tremendous damage on pine forests. Using microsatellite data, we demonstrated the existence

of multiple independent events of introduction in Japan (one from the USA and one with an

unknown origin) and China (one from the USA and one from Japan). We also found an

American origin of the Portuguese samples. Although, we observed some discrepancies

between traditional genetic methods and the ABC method, which are worth investigating and

are discussed here, the ABC approach definitely helped clarify the worldwide history of the

PWN invasion.

Keywords: population genetics, genetic diversity, differentiation, spread, Approximate

Bayesian Computation, PWN.

INTRODUCTION

Biological invasions are recognized as one of the main threats to biodiversity (Walker

& Steffen, 1997) and are fully integrated in global environmental changes induced by humans

(Ricciardi, 2007; Sala et al., 2000; Vitousek, Dantonio, Loope, & Westbrooks, 1996;

Wilcove, Rothstein, Dubow, Phillips, & Losos, 1998). Owing to their mainly irreversible

aspect (Mooney & Cleland, 2001; but see Simberloff, 2009, for examples of successful

eradications) and as their number is growing, studying them constitutes huge challenges at

ecological, economical, societal and scientific scales (Mack et al., 2000).

Understanding biological invasions necessitates the understanding of the history of the

invasive process, including the routes of invasion. The routes of invasion are defined as the

pathways followed by organisms (individuals, seeds...) between their source populations and

the invasive populations they formed in a new area. Their identification is a crucial step,

whose accuracy is improved by the use of genetic data and analyses (Estoup & Guillemaud,

2010).

Lots of studies have shown the usefulness of genetic data and analyses to decipher the

invasion routes (Boucher et al., 2013; Ciosi et al., 2008; Facon et al., 2003; Fontaine,

Gladieux, Hood, & Giraud, 2013; Kelager, Pedersen, & Bruun, 2013; Papura et al., 2012;

Perdereau et al., 2013; Rollins, Woolnough, Wilton, Sinclair, & Sherwin, 2009; Wan, Liu, &

Zhang, 2012). Recently, quantitative methods were proposed, among which the Approximate

Bayesian Computation (ABC) method (Beaumont, Zhang, & Balding, 2002). ABC relies on

models and presents several advantages compared to traditional methods, including (i) the

statistical comparison of different invasion scenarios, (ii) the integration of historical and

biological knowledge and (iii) the evaluation of the confidence in the scenario choice. In

brief, various models supposed to explain the data are compared based on the similarity

between summary statistics simulated from these models and from prior distributions of

historical and genetic parameters on one hand and summary statistics computed from the

actual samples on the other hand. This tool has already provided a great step forward in

several invasion studies (Ascunce et al., 2011; Barres et al., 2012; Boissin et al., 2012;

Boubou et al., 2012; Guillemaud et al., 2015; Lombaert et al., 2010a; Miller et al., 2005;

Pascual et al., 2007; Rius, Turon, Ordonez, & Pascual, 2012).

In this study, we were interested in the invasion of the pinewood nematode (PWN),

Bursaphelenchus xylophilus (Steiner & Buhrer, 1934; Nickle,   1970; Nematoda:

Aphelenchoididae). The PWN is a microscopic worm, responsible of the pine wilt disease

(Mamiya, 1972, 1976, 1983), which annually kills millions of pine trees worldwide (Mamiya,

1988; Soliman et al., 2012; Suzuki, 2002; Vicente, Espada, Vieira, & Mota, 2011). The PWN

is native to North America, namely Canada and the USA (Dropkin et al., 1981; Kiritani &

Morimoto, 2004). The first invasive outbreak of the PWN was observed in Asia, in Japan in

1905. The species was then observed in China in 1982, in Taiwan before 1985 and South

Korea in 1988 (Futai, 2013; Mamiya, 1988; Moon, Cheon, & Lee, 2007). In 1999, the PWN

was observed in Europe, first continental Portugal, and then Madeira Island and Spain

(Abelleira, Picoaga, Mansilla, & Aguin, 2011; Fonseca et al., 2012; Mota et al., 1999). To

date, the origin of PWN European populations is still unclear. Some authors have proposed an

Asian origin for European invasive populations of PWN (Figueiredo et al., 2013; Valadas,

Barbosa, Espada, Oliveira, & Mota, 2012). However, traditional genetic methods based on

genetic distances and clustering analyses recently failed to distinguish between an American

and an Asian origin for European outbreaks (Mallez et al., 2015). Analyses based on F_{ST}

values (Weir & Cockerham, 1984) and mean individual assignment likelihoods (Paetkau,

Slade, Burden, & Estoup, 2004) suggested an American origin for all Portuguese samples,

while analyses based on Cavalli-Sforza and Edwards' distances (Cavalli-Sforza & Edwards,

1967), Neighbor-Joining tree (Saitou & Nei, 1987) and Bayesian clustering (Pritchard,

Stephens, & Donnelly, 2000) suggested a Japanese origin for these samples. The aim of the

present study was to use ABC to clarify the invasion history of the PWN by performing

quantitative statistical comparison of various competing invasion scenarios, through the

calculation of their posterior probabilities (Miller et al., 2005; Pascual et al., 2007).

MATERIAL & METHODS

Sampling and genotyping

In this study, 48 site samples, representing a total of 1080 individuals were analyzed:

28 site samples from the USA (554 individuals) representing the native area, and 7 site

samples from Japan (210 individuals), 9 from Portugal/Madeira (169 individuals) and 4 from

China (147 individuals) representing the invaded areas. The features of the samples are listed

in Table 1. All the individuals were extracted from wood samples collected directly in the

field, using a sieve or the Baermann method (Viglierchio & Schmitt, 1983). The historical

knowledge about the invasion of the PWN and its spread within each invaded area (Abelleira

et al., 2011; Fonseca et al., 2012; Futai, 2013; Mamiya, 1988; Moon et al., 2007; Mota et al.,

1999) allowed us to associate a date of first observation to the different invasive outbreaks

(Table 1).

The DNA extraction of each single individual was carried out by thermal shock

(Castagnone, Abad, & Castagnone-Sereno, 2005) and 16 microsatellite loci were then

amplified as described by Mallez et al. (2015; 2013). The locus M26 was excluded due to a

lack of amplification for one Chinese site sample. The analyses presented here were thus

conducted with 15 microsatellite loci.

Standard genetic analyses

For each site sample, we computed the mean number of alleles (Mean *Na*) and the

observed (*Ho*) and expected (*He*) heterozygosities per site sample with Genetix version 4.05

(Belkhir, Borsa, Chikhi, Raufaste, & Bonhomme, 1996-2004). We tested each site sample for

a deviation to Hardy-Weinberg equilibrium (HWE) with Genepop version 4.1.3 (Rousset,

2008) and we quantified the inferred deviations to HWE by computing the F_{IS} estimate of

Weir and Cockerham (1984) with Fstat version 2.9.3.2 (Goudet, 2002). Linkage disequilibria

(LD) between all pairs of loci were also tested with Genepop (Rousset, 2008). The

significance level of multiple tests of HWE and LD were adjusted by applying the false

discovery rate (Benjamini & Hochberg, 1995) and sequential Bonferroni (Sokal & Rohlf,

1995) corrections, respectively.

Evolutionary relationships between the different site samples

We used traditional genetic methods to (i) determine the level of genetic structure

existing in each area under study, (ii) select the site samples (especially in the native area) to

include in the ABC analyses (see below) and (iii) guide the choice of the invasive scenarios to

compare in these latter analyses.

First, we evaluated the level of genetic differentiation between all site sample pairs

with Genepop (Rousset, 2008). Non independent multiple testing implied to adjust the

significance level with the sequential Bonferroni method (Sokal & Rohlf, 1995). F_{ST} (Weir &

Cockerham, 1984), corrected for null alleles (observed for some of the site samples and loci,

see Mallez et al., 2015) and D_{EST} of Jost (2008) between pairs of site samples were computed

with FreeNA (Chapuis & Estoup, 2007) and SMOGD (Crawford, 2010), respectively.

Second, we performed Bayesian clustering analyses within each geographical area

with Structure version 2.3 (Pritchard, Stephens, & Donnelly, 2000). These analyses were

carried out only within the USA and China given that the other areas (i. e., Japan and

Portugal) had been analyzed previously (Mallez et al., 2015). The number of clusters, K,

varied from 1 to the total number of site samples per area, i.e., 18 for the USA and 4 for

China. Twenty independent runs per K were carried out with 10^6 MCMC iterations each,

following a burn-in period of 2×10^5 iterations. For each Structure analysis, the Clumpak

server (Kopelman, Mayzel, Jakobsson, Rosenberg, & Mayrose, 2015), which combines the

Clumpp (Jakobsson & Rosenberg, 2007) and Distruct softwares (Rosenberg, 2004), was used

to postprocess the Structure outputs, i.e., (i) to determine the highest level of genetic structure

using the ΔK of Evanno (2005) and (ii) to identify the most frequent clustering pattern for

each value of K (across the 20 independent runs) and to display the corresponding bar plots.

Third, the Structure software was also run with all site samples from all areas as

previously described with K varying from 1 to 10 (which appeared to be high enough to

capture the genetic pattern existing between the native and the invaded areas, see Results).

For this Structure analysis, the Clumpak server (Kopelman et al., 2015) was also used to

postprocess the Structure outputs as presented above. Examination of Structure results

allowed us to propose putative historical relationships existing between the various site

samples.

Finally, a neighbor-joining tree based on Cavalli-Sforza & Edwards' distances

(Cavalli-Sforza & Edwards, 1967), corrected for null alleles with FreeNA (Chapuis & Estoup,

2007) was plotted and the mean individual assignment likelihoods (denoted Li→s, Paetkau,

Slade, Burden, & Estoup, 2004) of each invasive site sample to each putative native site

sample were computed with GeneClass2 version 2.0 (Piry et al., 2004).

ABC analyses

All ABC analyses were performed with DIYABC version 2.1.0 (Cornuet et al., 2014).

To define the scenarios to compare, we took into account the historical knowledge and results

from this study and previous analyses (Mallez et al., 2015). Particular site samples

representing genetic units inferred from Structure analyses were used for the ABC analyses.

As proposed by Lombaert et al. (2014), we performed a step-by-step ABC analysis: we first

tried to clarify the Asian invasion history and then integrated Europe into the scenarios. Thus,

we first focused on establishing the number of introduction events in Japan and in China,

independently (Figures S1 and S2 for a representation of the competing scenarios in each

analysis). Then, we studied the relationships between Japan and China (Figures S3 and S4 for

a representation of the competing scenarios in each analysis). Finally, we integrated the

Portuguese site samples into the inferred Asian invasion scenario to determine their origin

among the three possible sources: USA, Japan and China (Figure 1 for a representation of the

competing scenarios in this analysis).

For each ABC analysis, we simulated one million datasets per competing scenario.

The following summary statistics (SuSts) were used: the mean number of alleles, the mean

expected heterozygosity and the mean allelic size variance per population and pairs of

populations, the ratio of the number of alleles on the allelic size range (Garza & Williamson,

2001), the mean individual assignment likelihoods of population i to population j (Pascual et

al., 2007) and the shared allele distances between populations (Chakraborty & Jin, 1993).

Posterior probabilities were estimated using a polychotomous logistic regression on the 1% of

simulated datasets closest to the observed dataset (Cornuet, Ravigne, & Estoup, 2010; J. M.

Cornuet et al., 2008). The LDA option, allowing to perform a discriminant analysis on the

SuSts before the regression, was used (Cornuet et al., 2014; Estoup et al., 2012) to gain in

computation time.

To evaluate the robustness of our inferences, we performed the ABC analyses with

two different sets of prior distributions (referred to as prior set 1 and prior set 2 in the

Results). The generation time of the PWN *in natura* is not well-known due to a lack of direct

observations of the PWN inside the tree. Consequently, we tested the effect of choosing

various generation times (15, 30 and 45 generations per year, Table S1). Thirty generations

per year being the "reference", considering that the PWN reproduces from June to September

and estimating an average summer temperature of 25°C where it is currently distributed. We

also tested the effect of choosing different samples from the native area (NE5, NE9, NE22),

the closest genetically to NE2, the sample selected based on traditional genetic analyses (see

below). Then, we verified the non-overlapping of the 95 % confidence intervals of the

posterior probabilities of scenarios within each ABC analysis. We calculated the posterior

error rate by performing ABC analyses on 1000 simulated pseudo-observed datasets. These

datasets were simulated by drawing with replacement the scenario identity and the parameter

values among the 500 sets of scenarios and parameters values that generated simulated

datasets closest to the observed dataset. The posterior error rate corresponds to the proportion

of ABC analyses that wrongly identified scenarios and thus informs us on our ability to select

the true scenario. Finally, we checked the adequacy between the observed dataset and the

selected scenario, i.e., the capacity of this scenario to generate datasets similar to the observed

one. Basically, 1000 new datasets were simulated by drawing the parameter values from the

posterior distributions of the selected scenario. Then, each observed SuSt was confronted to

the distribution formed by the SuSts obtained from the new simulated datasets, which gave a

rejection probability for each observed SuSt. A large number of probabilities being obtained,

we adjusted the significance threshold with the false discovery rate method (Benjamini &

Hochberg, 1995). For this purpose, two new SuSts were added to the ones used in ABC

analyses: the F_{ST} (Weir & Cockerham, 1984) and the $d\mu^2$ distance (Goldstein, Linares,

Cavallisforza, & Feldman, 1995) between pairs of populations.

RESULTS

Standard genetic analyses

The genetic features of the site samples are shown in Table 1. The American site

samples displayed a genetic diversity low to moderate with mean Na from 1.27 to 3.87 and

He from 0.08 to 0.44. More than half of the site samples were not at HWE and significant LD

were observed for 13 tests out of the 1,084 realized, after correction. This may be due to the

presence of null alleles, as already observed by Mallez et al. (2015).

All site samples from the invaded areas exhibited very low levels of genetic diversity

owing to the presence of numerous monomorphic loci: 9 for Japan, 10 for China and 14 for

Portugal/Madeira out of the 15 loci under study. In Japan, three site samples (Jap212, Jap308

and Jap120) appeared completely "identical", with all the individuals sharing the same multi-

locus genotype. In Portugal/Madeira, 162 individuals out of the 169 analyzed had the same

multi-locus genotype and only one site sample, 128S, was different from the others. In China,

all loci were monomorphic within site samples and three site samples out of the four studied

presented the same multi-locus genotype for all their individuals.

Relationships between site samples within geographical areas

The analysis of the American site samples led to the general observation of a strong

genetic structure among PWN populations in USA, in good agreement with data from a

previous sampling in the same area (Mallez et al., 2015). Indeed, almost all the American site

samples appeared significantly differentiated after correction for multiple tests (Fisher method

of probability combination on loci, $p < 10^{-2}$). Only one test, between NC2 and GA3, was not

significant (Fisher method, $p = 0.11$). F_{ST}, corrected for null alleles, reached high values and

ranged from 0.036 to 0.76 (Table S2) and D_{EST} ranged from 0.0041 to 0.329 (Table S2). This

strong genetic differentiation pattern was supported by the Structure analysis, which revealed

biological relevant clustering of individuals within site samples for high values of K, although

the ΔK inferred $K = 2$ as the main structure (Figure S5). Different clustering patterns were

identified with Clumpp, but they exhibited similar results, i.e., a strong genetic structure

(Figure S5). In Japan, all site samples were also significantly differentiated, except the three

"identical" ones (Fisher method, $p < 10^{-5}$) with F_{ST} corrected for null alleles ranging from

0.627 to 0.995 and D_{EST} ranging from 0.0042 to 0.0938 (Table S2). In China, all site samples

were "identical" except Guangdong that appeared significantly differentiated from the others

(Fisher method, $p < 10^{-5}$) with a high mean corrected F_{ST} equal to 0.997 and a lower mean

D_{EST} equal to 0.106 (Table S2). This result was confirmed by the Bayesian clustering analysis,

which inferred a relevant clustering of individuals for $K = 2$, Guangdong being pulled apart

from the others Chinese site samples (Figure S6). In Portugal/Madeira, all site samples were

"identical" except the site sample 128S, that was significantly different from the others

(Fisher method, $p < 10^{-5}$) with a mean corrected F_{ST} equal to 0.341 and a mean D_{EST} equal to

0.0005 (Table S2). In view of the presence of "identical" site samples within Japan, China and

Portugal/Madeira, only one of these site samples was conserved in each invaded area for

simplification. The conserved site samples were the following: Jap212, Laoshan, TR1 and

Mad23PC for Japan, China, Portugal and Madeira, respectively.

Relationships between site samples among geographical areas

All the site samples from different invaded areas appeared significantly differentiated

between them and from the American site samples (Fisher method, $p < 10^{-5}$). NE2 was the

American site sample with the lowest differentiation from site samples from invaded areas

based on the D_{EST} (mean value = 0.034). Another American site sample, NE9, was selected

based on the corrected F_{ST} (mean value = 0.469), NE2 being the second lowest one with this

measure (mean value = 0.541). The neighbor-joining tree confirmed that the site sample NE2

is the closest site sample to all site samples from the invaded areas and that the Portuguese

site samples are closer to a Japanese site sample than a Chinese or an American site samples

(Figure 2). For the Bayesian clustering analysis, the ΔK inferred K=2 as the main structure

(Figure S7A), in which all the invasive site samples pulled apart from the native American

site samples (Figure 3). From K=3, different clustering patterns were observed (Figures 3 and

S7B), in congruence with the sub-structure existing within each of the invaded and native

areas. However, from K=8, the clustering pattern obtained within the invasive area appeared

consistent between runs and identified four groups: one group with Mad23PC, 128S, TR1,

Jap212 and Laoshan; another one with Kasumig2, Kasumig5 and Kosa; and Kasumig3 and

Guangdong, that clustered separately from one another (Figure 3). Finally, almost all the site

samples from the invaded areas were assigned to NE2 with the highest mean individual

assignment likelihood (Figure S8). Only Guangdong and Laoshan were assigned to different

samples, NE5 and Kosa, respectively (Figure S8).

Summarizing and considering the results presented above, the site sample NE2 was

the closest site sample of the native area to all invasive site samples but Guangdong. Thus, it

was used to represent the native area in the competing scenarios of the ABC analyses. In view

of the incomplete sampling of the native area (no sample from Canada was obtained) and the

strong genetic structuration observed in the USA, it is unlikely that NE2 is a sample of the

true source population. An unsampled "ghost" population diverging from NE2 was thus

included in all scenarios (see Figures 1, S1, S2, S3, S4 and Table S1). We also considered the

following relevant genetic units to characterize the invaded areas in the following ABC

analyses, such as:

-    Japanese genetic unit 1, hereafter called Japan1, represented by Kosa, the closest sample

to the introduction point in Japan and considered coming from the native area (Japan

being the first known outbreak area of the PWN).

-    Japanese genetic unit 2, hereafter called Japan2, represented by Kasumig3.

-    Chinese genetic unit 1, hereafter called China1, represented by Laoshan.

-    Chinese genetic unit 2, hereafter called China2, represented by Guangdong.

-    Portuguese genetic unit, hereafter called Portugal, represented by TR1.

ABC analyses

In Japan, the results clearly indicated that a multiple introduction scenario is more

likely than a single introduction scenario with posterior probabilities over 0.95 for prior sets 1

and 2 and non-overlapping confidence intervals, Japan2 being thus independent from Japan1

(Table 2). A similar result was observed in China. China1 was independent from China2 with

posterior probabilities of 0.94 and 0.70 for prior sets 1 and 2 respectively and non-overlapping

confidence intervals, suggesting two introduction events in China (Table 2). The ABC

analyses carried out between site samples from Japan and China revealed that: (i) Japan1 was

probably the source of China1 and (ii) China2 is independent from Japan1 (high posterior

probabilities, results robust to prior set changes, Table 2). The analysis between China2 and

Japan2 gave inconsistent results, relative to the prior set used (Table 2). With the prior set 1,

the scenario describing an admixture event was selected but the posterior probability was low,

and a high posterior error rate was observed (Table 2). On the contrary, with the prior set 2,

the scenario describing two independent events of introduction was selected with once again a

low posterior probability and a high posterior error rate value (Table 2). Consequently, Japan2

was not included in the Asian scenario.

The analyses with all invaded areas studied here showed that an American origin for

the Portuguese populations appeared more likely than an Asian origin, this scenario obtaining

the highest posterior probability and this result being robust to prior set changes and

generation time changes (Table 3). The use of the 3 samples (NE5, NE9, NE22) from the

native area genetically closest to NE2, instead of NE2, gave mixed results and three analyses

(out of six) pointed towards an American origin for the Portuguese populations (Table S3).

Overall, there was a negative correlation between the posterior probability of scenario 1 and

the genetic distance between the sample used and NE2 [correlation coefficient of -0.46 (prior

set 1) and -0.64 (prior set 2) for F_{ST}, and of -0.76 (prior set 1) and -0.79 (prior set 2) for D_{EST}].

We performed an independent ABC analysis with Japan2 and Portugal. In this latter analysis,

we selected an American origin for Portugal based on the posterior probability (Table S4). We

finally compared the final selected scenario (Scenario 1, Figure 1) to an "unique invasive

bridgehead" scenario (Lombaert et al., 2010b), in which an unsampled "ghost" invasive

population, originating from the native area, was the source for Japan1 (being then the source

for China1), Japan2, China2 and Portugal. The scenario selected in the global analysis

(Scenario 1, Figure 1) obtained the highest posterior probability (p > 0.85) whatever the prior

set used and thus appeared more likely than the "invasive bridgehead" scenario (Table S5).

338   The posterior error rates appeared quite low for all the analyses (except the Japan2 –

China2 analysis, Tables 2 and 3), demonstrating that the competing scenarios were actually

umiquedistinguishable. Finally, simulations from the selected scenarios and the posterior

distributions of the parameters were able to produce data close to the observed ones: none of

the observed SuSts was in the extreme 5% tails of the distribution of simulated SuSts after

correction (Table S6).

DISCUSSION

348   The main objective of this study was to infer the invasion routes of the PWN using the

ABC method. We demonstrated the existence of multiple independent introductions from the

native area to Japan, China and Europe. This inference has not been possible using traditional

population genetic methods performed in a previous study (Mallez et al., 2015). The ABC

method helped clarify the worldwide history of the PWN invasion even though the post-hoc

analyses tempered the conclusions. This highlights the importance of assessing the confidence

in the results, especially when dealing with atypical genetic data. The discrepancies observed

between traditional genetic methods and the ABC method are worth investigating and are

discussed here.

Invasion routes of the PWN and the origin of Portuguese populations

Despite the low genetic diversity generally present in the PWN populations, we were

able to reliably select scenarios (among the ones compared) with high posterior probabilities

and good quality analyses in most cases. The most probable worldwide invasion history of the

PWN identified in this study (Figure 4) includes multiple introductions in Asia and an

American origin of the European samples.

The analyses carried out with the Asian site samples allowed us to reveal the existence

of multiple independent events of introduction in China and Japan. Two events of

introduction were inferred in both China (one from the USA and one from Japan) and Japan

(one from the USA and one with an unknown origin). Especially, these results strengthen the

hypothesis of two events of introduction in China, the first one near Nanjing in 1982 and the

second one near Hong-Kong in 1988, already suggested in Robinet et al (2009; but see Cheng,

Cheng, Xu, & Xie, 2008). Concerning Europe, the main ABC analyses of the PWN

microsatellites data designated the USA as the most probable source of the Portuguese

populations and excluded Japan and China as possible sources, contrary to what is often

presented in the literature based on genetic distance trees (Figueiredo et al., 2013; Fonseca et

al., 2012; Metge & Burgermeister, 2008; Valadas et al., 2012; Valadas, Laranjo, Mota, &

Oliveira, 2013; Vieira, Burgermeister, Mota, Metge, & Silva, 2007). This result was quite

robust to changes in the parameters' priors and to the posterior model checking analyses, but

it depended on the samples considered in the native area, which may be considered a problem.

Interestingly, our results displayed a pattern similar to that obtained by Guillemaud et al.

(2010). These authors found a negative correlation between the genetic distance of the

samples considered to the most probable source and the probability of an independent

scenario when this scenario was true. In other words, when two invasive populations originate

from two independent introductions from a native population, the more distant is the native

sample from the true source, the more probable it is to erroneously point to a scenario with

two serial introductions. Therefore the "mixed" results obtained by changing the sample of

the native area are expected if the true scenario includes an American origin of the European

samples. Overall, the ABC method allowed us to clarify the history of the PWN invasion in

Europe with good statistical confidence.

This study highlights the central role of the native area in the worldwide history of

invasion of the PWN, adding up three as the minimum number of independent introductions

from the USA. Only one detectable event of introduction seems to have occurred in Portugal

since the first outbreak, i.e., 19 years ago. Considering the volume of global trade between

Portugal and the worldwide infested areas, this result shows that quarantine and inspection

efforts have been globally successful and efficient to date and confirms the importance of

prevention and pro-active actions to limit new introductions (Simberloff & Rejmanek, 2011).

At the European level, ecological and pedo-climatic conditions favorable to the PWN are

encountered in many countries (Robinet, Van Opstal, Baker, & Roques, 2011). However, the

invasion is currently limited to the Iberian Peninsula (plus Madeira Island). This reinforces

the idea that the quarantine regulation measures have been efficient to date to protect Europe

1) from the entry of new invasive PWN populations from abroad, and 2) from a large

dissemination of the PWN from the original Portuguese outbreak.

Different methods with different results

In the case of the PWN invasion, the results obtained from traditional genetic

approaches pointed out different putative origins for the European populations of PWN

(Mallez et al., 2015): American based on F_{ST} values (Weir & Cockerham, 1984) and mean

individual assignment likelihoods (Paetkau et al., 2004), and Japanese based on Cavalli-

Sforza and Edwards' distances (Cavalli-Sforza & Edwards, 1967), Neighbor-Joining tree

(Saitou & Nei, 1987) and Bayesian clustering (Pritchard et al., 2000). The ABC analyses,

quantitative and probability-based, allowed us to improve greatly the history of the PWN

invasion. Our study highlights discrepancies between the ABC methodology and some of the

traditional approaches, namely the Bayesian clustering of individuals (Pritchard et al., 2000)

and distance genetic trees (Saitou & Nei, 1987). To our knowledge, such discrepancies have

not been observed yet, although a recent simulation study showed that classic clustering

methods such as Structure analyses (Pritchard et al., 2000) may be erroneously interpreted in

the context of invasion route inference (Lombaert, Guillemaud, & Deleury, 2018). The origin

of the discrepancies we observed is not completely clear to us, but hypotheses may be

proposed.

They may have arisen from the very low level of genetic diversity observed in the

PWN populations, which might disturb the analyses of the traditional genetic methods and

thus their outcomes. In an invasion context, a low genetic diversity is usually the translation

of strong founder effects, which implies a strong genetic drift. Genetic drift, in turn, is a

stochastic process whose consequences are random. Employing methods such as ABC that

take these stochastic effects into account may thus be more reliable than other methods, such

as Bayesian clustering or distance-based methods that do not (Guillemaud et al., 2010). In

these conditions of low genetic diversity, we can expect traditional methods to lead to variable

results when applied to various realizations of the same historical scenario. We may expect

the low level of genetic diversity to produce a genetic structure pattern and/or relationship

between site samples that we misinterpret (as in Lombaert et al., 2018). For instance, if the

native area is weakly diversified so that it exhibits a few very frequent alleles, it is probable

that two independent introductions from this native area (native → invasive 1 and native →

invasive 2) lead to samples closer to each other than to their native area. When analyzed, such

results would suggest serial introductions (native → invasive 1 → invasive 2, because

invasive1 and 2 are closer to each other than to native) instead of the true independent

introductions scenario, because the interpretation is binary and not based on probability

computation. Using a model-based stochastic method, such as ABC, a genetic structure with

invading samples closer to each other than to the native area may lead to the choice of

independent introduction scenario because it is probable to get such structure with this

scenario when the diversity is low.

The observed discrepancies may thus just be the consequence of the absence of

statistical support associated with traditional genetic methods permitting to quantitatively base

the choice of evolutionary scenarios (criticism presented in A. Estoup & Guillemaud, 2010).

Carrying out simulations may help to disentangle the effect of the genetic diversity in

the native area on the outcomes of clustering methods and ABC analyses, a work partly

performed by Lombaert et al. (2018) who did not use ABC analyses. More precisely,

analyzing dataset, simulated under chosen scenarios and known levels of genetic diversity,

may allow to assess the behavior of these analyses in relation to the level of genetic diversity.

These simulations would thus be useful to (i) determine the real impact of the genetic

diversity on the outcomes and (ii) verify that the interpretations made from these methods in

the biological invasion framework are appropriate.

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REFERENCES

Abelleira, A., Picoaga, A., Mansilla, J. P., & Aguin, O. (2011). Detection of *Bursaphelenchus*
xylophilus, causal agent of pine wilt disease on *Pinus pinaster* in Northwestern Spain.
Plant Disease, 95(6), 776.
Ascunce, M. S., Yang, C. C., Oakey, J., Calcaterra, L., Wu, W. J., Shih, C. J., Shoemaker, D.
(2011). Global Invasion History of the Fire Ant Solenopsis invicta. *Science,*
331(6020), 1066-1068. doi:10.1126/science.1198734
Barres, B., Carlier, J., Seguin, M., Fenouillet, C., Cilas, C., & Ravigne, V. (2012).
Understanding the recent colonization history of a plant pathogenic fungus using
population genetic tools and Approximate Bayesian Computation. *Heredity, 109*(5),
269-279. doi:10.1038/hdy.2012.37
Beaumont, M. A., Zhang, W. Y., & Balding, D. J. (2002). Approximate Bayesian
computation in population genetics. *Genetics, 162*(4), 2025-2035.
Belkhir, K., Borsa, P., Chikhi, L., Raufaste, N., & Bonhomme, F. (1996-2004). GENETIX
4.05, logiciel sous Windows TM pour la génétique des populations. Laboratoire
Génome, Populations, Interactions, CNRS UMR 5000, Université de Montpellier
Montpellier (France).
Benjamini, Y., & Hochberg, Y. (1995). Controlling the false discovery rate - a practical and
powerful approach to multiple testing. *Journal of the Royal Statistical Society Series*
B-Methodological, 57(1), 289-300.
Boissin, E., Hurley, B., Wingfield, M. J., Vasaitis, R., Stenlid, J., Davis, C., . . . Slippers, B.
(2012). Retracing the routes of introduction of invasive species: the case of the Sirex
noctilio woodwasp. *Molecular Ecology, 21*(23), 5728-5744. doi:10.1111/mec.12065
Boubou, A., Migeon, A., Roderick, G. K., Auger, P., Cornuet, J. M., Magalhaes, S., &
Navajas, M. (2012). Test of Colonisation Scenarios Reveals Complex Invasion
History of the Red Tomato Spider Mite Tetranychus evansi. *Plos One, 7*(4).
doi:10.1371/journal.pone.0035601
Boucher, A. C., Mimee, B., Montarry, J., Bardou-Valette, S., Belair, G., Moffett, P., &
Grenier, E. (2013). Genetic diversity of the golden potato cyst nematode Globodera
rostochiensis and determination of the origin of populations in Quebec, Canada.
Molecular   Phylogenetics   and   Evolution,   69(1),   75-82.
doi:10.1016/j.ympev.2013.05.020
Castagnone, C., Abad, P., & Castagnone-Sereno, P. (2005). Satellite DNA-based species-
specific   identification   of   single   individuals   of   the   pinewood   nematode
Bursaphelenchus xylophilus (Nematoda : Aphelenchoididae). *European Journal of*
Plant Pathology, 112(2), 191-193. doi:10.1007/s10658-004-0580-2
Cavalli-Sforza, L. L., & Edwards, A. W. F. (1967). Phylogenetic analysis. Models and
estimation procedures. *The American Journal of Humans Genetics, 19*, 233 - 257.
Chakraborty, R., & Jin, L. (1993). A unified approach to study hypervariable polymorphisms:
statistical considerations of determining relatedness and population distances. *EXS, 67*,
153-175.
Chapuis, M.-P., & Estoup, A. (2007). Microsatellite null alleles and estimation of population
differentiation.   *Molecular   Biology   and   Evolution,   24*(3),   621-631.
doi:10.1093/molbev/msl191
Cheng, X. Y., Cheng, F. X., Xu, R. M., & Xie, B. Y. (2008). Genetic variation in the invasive
process of *Bursaphelenchus xylophilus* (Aphelenchida: Aphelenchoididae) and its
possible   spread   routes   in   China.   *Heredity,   100*(4),   356-365.
doi:10.1038/sj.hdy.6801082

Ciosi, M., Miller, N. J., Kim, K. S., Giordano, R., Estoup, A., & Guillemaud, T. (2008).
Invasion of Europe by the western corn rootworm, Diabrotica virgifera virgifera:
multiple transatlantic introductions with various reductions of genetic diversity.
Molecular Ecology, 17(16), 3614-3627.
Cornuet, J.-M., Santos, F., Beaumont, M. A., Robert, C. P., Marin, J.-M., Balding, D. J., . . .
Estoup, A. (2008). Inferring population history with DIY ABC: a user-friendly
approach to approximate Bayesian computation. *Bioinformatics, 24*(23), 2713-2719.
doi:10.1093/bioinformatics/btn514
Cornuet, J. M., Pudlo, P., Veyssier, J., Dehne-Garcia, A., Gautier, M., Leblois, R., . . . Estoup,
A. (2014). DIYABC v2.0: a software to make approximate Bayesian computation
inferences about population history using single nucleotide polymorphism, DNA
sequence and microsatellite data. *Bioinformatics, 30*(8), 1187-1189.
doi:10.1093/bioinformatics/btt763
Cornuet, J. M., Ravigne, V., & Estoup, A. (2010). Inference on population history and model
checking using DNA sequence and microsatellite data with the software DIYABC
(v1.0). *Bmc Bioinformatics, 11*. doi:40110.1186/1471-2105-11-401
Crawford, N. G. (2010). SMOGD: software for the measurement of genetic diversity.
Molecular Ecology Resources, 10(3), 556-557. doi:10.1111/j.1755-0998.2009.02801.x
Dropkin, V. H., Foudin, A., Kondo, E., Linit, M. J., Smith, M., & Robbins, K. (1981).
Pinewood nematode : a threat to US forests? *Plant Disease, 65*(12), 1022 - 1027.
Estoup, A., & Guillemaud, T. (2010). Reconstructing routes of invasion using genetic data:
why, how and so what? *Molecular Ecology, 19*(19), 4113-4130. doi:10.1111/j.1365-
294X.2010.04773.x
Estoup, A., Lombaert, E., Marin, J.-M., Guillemaud, T., Pudlo, P., Robert, C. P., & Cornuet,
J.-M. (2012). Estimation of demo-genetic model probabilities with Approximate
Bayesian Computation using linear discriminant analysis on summary statistics.
Molecular Ecology Resources, 12(5), 846-855. doi:10.1111/j.1755-0998.2012.03153.x
Evanno, G., Regnaut, S., & Goudet, J. (2005). - Detecting the number of clusters of
individuals using the software structure: a simulation study. - *14*(- 8), - 2620.
Facon, B., Pointier, J. P., Glaubrecht, M., Poux, C., Jarne, P., & David, P. (2003). A
molecular phylogeography approach to biological invasions of the New World by
parthenogenetic Thiarid snails. *Molecular Ecology, 12*(11), 3027-3039.
doi:10.1046/j.1365-294X.2003.01972.x
Figueiredo, J., Simões, M. J., Gomes, P., Barroso, C., Pinho, D., Conceição, L., . . . Egas, C.
(2013). Assessment of the Geographic Origins of Pinewood Nematode Isolates via
Single Nucleotide Polymorphism in Effector Genes. *Plos One, 8*(12), e83542.
doi:10.1371/journal.pone.0083542
Fonseca, L., Cardoso, J. M. S., Lopes, A., Pestana, M., Abreu, F., Nunes, N., . . . Abrantes, I.
(2012). The pinewood nematode, *Bursaphelenchus xylophilus*, in Madeira Island.
Helminthologia, 49(2), 96 - 103.
Fontaine, M. C., Gladieux, P., Hood, M. E., & Giraud, T. (2013). History of the invasion of
the anther smut pathogen on Silene latifolia in North America. *New Phytologist,*
198(3), 946-956. doi:10.1111/nph.12177
Futai, K. (2013). Pinewood nematode, Bursaphelenchus xylophilus. *Annual Review of*
Phytopathology, 51, 61-83.
Garza, J. C., & Williamson, E. G. (2001). Detection of reduction in population size using data
from microsatellite loci. *Molecular Ecology, 10*(2), 305-318. doi:10.1046/j.1365-
294X.2001.01190.x

Goldstein, D. B., Linares, A. R., Cavallisforza, L. L., & Feldman, M. W. (1995). An evaluation of genetic distances for use with microsatellite loci. *Genetics, 139*(1), 463-471.

Goudet, J. (2002). Fstat, a program to estimate and test gene diversities and fixation indices. Retrieved from http://www2.unil.ch/popgen/softwares/fstat.htm

Guillemaud, T., Beaumont, M. A., Ciosi, M., Cornuet, J.-M., & Estoup, A. (2010). Inferring introduction routes of invasive species using approximate Bayesian computation on microsatellite data. *Heredity, 104*(1), 88 - 99. doi:doi:10.1038/hdy.2009.92

Guillemaud, T., Blin, A., Le Goff, I., Desneux, N., Reyes, M., Tabone, E., . . . Lombaert, E. (2015). The tomato borer, Tuta absoluta, invading the Mediterranean Basin, originates from a single introduction from Central Chile. *Scientific Reports, 5*, 8371. doi:10.1038/srep08371

Jakobsson, M., & Rosenberg, N. A. (2007). CLUMPP: a cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. *Bioinformatics, 23*(14), 1801 - 1806.

Jost, L. (2008). G(ST) and its relatives do not measure differentiation. *Molecular Ecology, 17*(18), 4015-4026. doi:10.1111/j.1365-294X.2008.03887.x

Kelager, A., Pedersen, J., & Bruun, H. (2013). Multiple introductions and no loss of genetic diversity: invasion history of Japanese Rose, Rosa rugosa, in Europe. *Biological Invasions, 15*(5), 1125-1141. doi:10.1007/s10530-012-0356-0

Kiritani, K., & Morimoto, N. (2004). Invasive insect and nematode pests from North America *Global Environmental Research, 8*(1), 75 - 88.

Kopelman, N. M., Mayzel, J., Jakobsson, M., Rosenberg, N. A., & Mayrose, I. (2015). - Clumpak: a program for identifying clustering modes and packaging population structure inferences across K. - *15*(- 5), - 1191.

Lombaert, E., Guillemaud, T., Cornuet, J. M., Malausa, T., Facon, B., & Estoup, A. (2010a). Bridgehead Effect in the Worldwide Invasion of the Biocontrol Harlequin Ladybird. *Plos One, 5*(3). doi:e974310.1371/journal.pone.0009743

Lombaert, E., Guillemaud, T., Cornuet, J. M., Malausa, T., Facon, B., & Estoup, A. (2010b). Bridgehead Effect in the Worldwide Invasion of the Biocontrol Harlequin Ladybird. *Plos One, 5*(3), e9743. doi:10.1371/journal.pone.0009743

Lombaert, E., Guillemaud, T., & Deleury, E. (2018). Biases of STRUCTURE software when exploring introduction routes of invasive species. *Heredity, 120*(6), 485 - 499. doi:doi:10.1038/s41437-017-0042-1

Lombaert, E., Guillemaud, T., Lundgren, J., Koch, R., Facon, B., Grez, A., Estoup, A. (2014). Complementarity of statistical treatments to reconstruct worldwide routes of invasion: the case of the Asian ladybird *Harmonia axyridis*. *Molecular Ecology, 23*, 5979 - 5997.

Mack, R. N., Simberloff, D., Mark Lonsdale, W., Evans, H., Clout, M., & Bazzaz, F. A. (2000). Biotic invasions : Causes, Epidemiology, Global Consequences, and control. *Ecological Applications, 10*(3), 689-710.

Mallez, S., Castagnone, C., Espada, M., Vieira, P., Eisenback, J. D., Harrell, M., Guillemaud, T. (2015). Worldwide invasion routes of the pinewood nematode: What can we infer from population genetics analyses? *Biological Invasions, 17*(4), 1199 - 1213. doi:10.1007/s10530-014-0788-9

Mallez, S., Castagnone, C., Espada, M., Vieira, P., Eisenback, J. D., Mota, M., Castagnone-Sereno, P. (2013). First insights into the genetic diversity of the pinewood nematode in its native area using new polymorphic microsatellite loci. *Plos One, 8*(3), e59165. doi:10.1371/journal.pone.0059165

Mamiya, Y. (1972). Pine wood nematode *Bursaphelenchus lignicolus* Mamiya and Kiyohara, as a causal agent of pine wilting disease. *Review of Plant Protection Research, 5*, 46 - 60.

Mamiya, Y. (1976). Pine wilt disease caused by the pine wood nematode, *Bursaphelenchus lignicolus*, in Japan. *Japan Agricultural Research Quarterly, 10*(4), 206 - 211.

Mamiya, Y. (1983). Pathology of the pine wilt disease caused by *Bursaphelenchus xylophilus*. *Annual Review of Phytopathology, 21*, 201-220. doi:10.1146/annurev.py.21.090183.001221

Mamiya, Y. (1988). History of pine wilt disease in Japan. *Journal of Nematology, 20*(2), 219 - 226.

Metge, K., & Burgermeister, W. (2008). Analysis of *Bursaphelenchus xylophilus* (Nematoda: Parasitaphelenchidae) provenances using ISSR and RAPD fingerprints. In M. Mota & P. Vieira (Eds.), *Pine wilt disease: a worldwide threat to forest ecosystems* (pp. 175-186).

Miller, N., Estoup, A., Toepfer, S., Bourguet, D., Lapchin, L., Derridj, S., Guillemaud, T. (2005). Multiple transatlantic introductions of the western corn rootworm. *Science, 310*(5750), 992-992. doi:10.1126/science.1115871

Moon, Y. S., Cheon, H. M., & Lee, S. (2007). Occurence of pine wilt disease, caused by *Bursaphelenchus xylophilus*, from *Pinus koraiensis* in Korea. *Journal of Nematology, 39*(1), 87 - 88.

Mooney, H. A., & Cleland, E. E. (2001). The evolutionary impact of invasive species. *Proceedings of the National Academy of Sciences of the United States of America, 98*(10), 5446-5451. doi:10.1073/pnas.091093398

Mota, M., Braasch, H., Bravo, M. A., Penas, A. C., Burgermeister, W., Metge, K., & Sousa, E. (1999). First report of *Bursaphelenchus xylophilus* in Portugal and in Europe. *Nematology, 1*(7 - 8), 727 - 734.

Paetkau, D., Slade, R., Burden, M., & Estoup, A. (2004). Genetic assignment methods for the direct, real-time estimation of migration rate: a simulation-based exploration of accuracy and power. *Molecular Ecology, 13*(1), 55 - 65. doi:10.1046/j.1365-294X.2004.02008.x

Papura, D., Burban, C., van Helden, M., Giresse, X., Nusillard, B., Guillemaud, T., & Kerdelhué, C. (2012). Microsatellite and Mitochondrial Data Provide Evidence for a Single Major Introduction for the Neartic Leafhopper *Scaphoideus titanus* in Europe. *Plos One, 7*(5), e36882. doi:10.1371/journal.pone.0036882

Pascual, M., Chapuis, M. P., Mestres, F., Balanya, J., Huey, R. B., Gilchrist, G. W., Estoup, A. (2007). Introduction history of Drosophila subobscura in the New World: a microsatellite-based survey using ABC methods. *Molecular Ecology, 16*, 3069 - 3083.

Perdereau, E., Bagnères, A. G., Bankhead-Dronnet, S., Dupont, S., Zimmermann, M., Vargo, E. L., & Dedeine, F. (2013). Global genetic analysis reveals the putative native source of the invasive termite, *Reticulitermes flavipes*, in France. *Molecular Ecology, 22*(4), 1105-1119. doi:10.1111/mec.12140

Piry, S., Alapetite, A., Cornuet, J.-M., Paetkau, D., Baudouin, L., & Estoup, A. (2004). GENECLASS2: A Software for Genetic Assignment and First-Generation Migrant Detection. *Journal of Heredity, 95*(6), 536-539. doi:10.1093/jhered/esh074

Pritchard, J. K., Stephens, M., & Donnelly, P. (2000). Inference of population structure using multilocus genotype data. *Genetics, 155*(2), 945-959.

Ricciardi, A. (2007). Are modern biological invasions an unprecedented form of global change? *Conservation Biology, 21*(2), 329-336. doi:10.1111/j.1523-1739.2006.00615.x

654   Rius, M., Turon, X., Ordonez, V., & Pascual, M. (2012). Tracking Invasion Histories in the
655   Sea: Facing Complex Scenarios Using Multilocus Data. *Plos One, 7*(4).
656   doi:10.1371/journal.pone.0035815

657   Robinet, C., Roques, A., Pan, H. Y., Fang, G. F., Ye, J. R., Zhang, Y. Z., & Sun, J. H. (2009).
658   Role of human-mediated dispersal in the spread of the pinewood nematode in China.
659   *Plos One, 4*(2). doi:e464610.1371/journal.pone.0004646

660   Robinet, C., Van Opstal, N., Baker, R., & Roques, A. (2011). Applying a spread model to
661   identify the entry points from which the pine wood nematode, the vector of pine wilt
662   disease, would spread most rapidly across Europe. *Biological Invasions, 13*(12), 2981-
663   2995. doi:10.1007/s10530-011-9983-0

664   Rollins, L. A., Woolnough, A. P., Wilton, A. N., Sinclair, R. O. N., & Sherwin, W. B. (2009).
665   Invasive species can't cover their tracks: using microsatellites to assist management of
666   starling (Sturnus vulgaris) populations in Western Australia. *Molecular Ecology,*
667   *18*(8), 1560-1573. doi:10.1111/j.1365-294X.2009.04132.x

668   Rosenberg, N. A. (2004). DISTRUCT: a program for the graphical display of population
669   structure. *Molecular Ecology Notes, 4*(1), 137-138. doi:10.1046/j.1471-
670   8286.2003.00566.x

671   Rousset, F. (2008). GENEPOP ' 007: a complete re-implementation of the GENEPOP
672   software for Windows and Linux. *Molecular Ecology Resources, 8*(1), 103-106.
673   doi:10.1111/j.1471-8286.2007.01931.x

674   Saitou, N., & Nei, M. (1987). The neighbor-joining method: a new method for reconstructing
675   phylogenetic trees. *Molecular Biology and Evolution, 4*(4), 406-425.

676   Sala, O. E., Chapin, F. S., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., . . . Wall, D.
677   H. (2000). Biodiversity - Global biodiversity scenarios for the year 2100. *Science,*
678   *287*(5459), 1770-1774. doi:10.1126/science.287.5459.1770

679   Simberloff, D. (2009). We can eliminate invasions or live with them. Successful management
680   projects. *Biological Invasions, 11*(1), 149-157. doi:10.1007/s10530-008-9317-z

681   Simberloff, D., & Rejmanek, M. (2011). *Encyclopedia of Biological Invasions*. In U. o. C.
682   Press (Ed.).

683   Sokal, R. R., & Rohlf, F. J. (1995). *Biometry. The principles and practice of statistics in*
684   *biological research.* (N. Y. W. H. Freeman and Compagny Ed. Third Edition ed.).

685   Soliman, T., Mourits, M. C. M., van der Werf, W., Hengeveld, G. M., Robinet, C., &
686   Lansink, A. (2012). Framework for modelling economic impacts of invasive species,
687   applied to pine wood nematode in Europe. *Plos One, 7*(9).
688   doi:10.1371/journal.pone.0045505

689   Steiner, G., & Buhrer, E. M. (1934). *Aphelenchoides xylophilus*, N. SP. A nematode
690   associated with blue-stain and other fungi in timber. *Journal of agricultural research,*
691   *48*(10), 949 - 951.

692   Suzuki, K. (2002). Pine wilt disease - a threat to pine forest in Europe. *Dendrobiology, 48*, 71
693   - 74.

694   Valadas, V., Barbosa, P., Espada, M., Oliveira, S., & Mota, M. (2012). The pinewood
695   nematode, *Bursaphelenchus xylophilus,* in Portugal : possible introductions and spread
696   routes of a serious biological invasion revealed by molecular methods. *Nematology,*
697   *14*(8), 899 - 911.

698   Valadas, V., Laranjo, M., Mota, M., & Oliveira, S. (2013). Molecular characterization of
699   Portuguese populations of the pinewood nematode Bursaphelenchus xylophilus using
700   cytochrome b and cellulase genes. *Journal of Helminthology*, *87*(4), 457 - 466.

701   Vicente, C., Espada, M., Vieira, P., & Mota, M. (2011). Pine Wilt Disease : a threat to
702   European forestry. *European Journal of Plant Pathology, 133*, 89 - 99.

Vieira, P., Burgermeister, W., Mota, M., Metge, K., & Silva, G. (2007). Lack of genetic
variation of *Bursaphelenchus xylophilus* in Portugal revealed by RAPD-PCR analyses.
Journal of Nematology, 39(2), 118-126.
Viglierchio, D. R., & Schmitt, R. V. (1983). On the methodology of nematode extraction from
field samples : Baermann funnel modifications. *Journal of Nematology, 15*(3), 438 -
444.
Vitousek, P. M., Dantonio, C. M., Loope, L. L., & Westbrooks, R. (1996). Biological
invasions as global environmental change. *American Scientist, 84*(5), 468-478.
Walker, B. H., & Steffen, W. (1997). An overview of the implications of global change for
natural and managed terrestrial ecosystems. *Conservation Ecology [online]*. Retrieved
from http://www.consecol.org/vol1/iss2/art2/
Wan, X. W., Liu, Y. H., & Zhang, B. (2012). Invasion History of the Oriental Fruit Fly,
Bactrocera dorsalis, in the Pacific-Asia Region: Two Main Invasion Routes. *Plos One,
7*(5). doi:10.1371/journal.pone.0036176
Weir, B. S., & Cockerham, C. C. (1984). Estimating F-Statistics for the analysis of population
structure. *Evolution, 38*(6), 1358-1370.
Wilcove, D. S., Rothstein, D., Dubow, J., Phillips, A., & Losos, E. (1998). Quantifying threats
to imperiled species in the United States. *BioScience, 48*(8), 607-615.
doi:10.2307/1313420
722

Table 1: Characteristics and population genetics summary statistics of the 48 samples of the pinewood nematode used in this study.

Code	Nb. Individuals	Location	First obs. (years)	Mean Na	He	Ho	F_{IS}
MO1	31	USA - Missouri - Columbia	-	3.33	0.35	0.29	0.19*
MO2	23	USA - Missouri - Columbia	-	3.53	0.38	0.28	0.27*
NE1	16	USA - Nebraska - Davey	-	2.07	0.21	0.19	0.09
NE2	15	USA - Nebraska - Davey	-	1.87	0.22	0.17	0.24
NE5	14	USA - Nebraska - Pawnee Lake	-	2.00	0.22	0.19	0.14
NE6	21	USA - Nebraska - Pawnee Lake	-	1.40	0.20	0.10	0.48*
NE9	29	USA - Nebraska - Pawnee Lake	-	2.73	0.33	0.25	0.26*
NE10	26	USA - Nebraska - Pawnee Lake	-	1.80	0.20	0.16	0.19*
NE12	28	USA - Nebraska - Conestoga Lake	-	1.53	0.16	0.11	0.34*
NE13b	19	USA - Nebraska - Pioneers Park	-	2.40	0.30	0.21	0.31*
NE14	28	USA - Nebraska - Pioneers Park	-	1.73	0.25	0.17	0.34*
NE15	23	USA - Nebraska - Pioneers Park	-	2.27	0.23	0.14	0.39*
NE19	16	USA - Nebraska - UNL East Campus	-	1.87	0.25	0.21	0.15
NE22	17	USA - Nebraska - Lincoln	-	2.07	0.30	0.22	0.28*
NE23	25	USA - Nebraska - Lincoln	-	1.27	0.08	0.07	0.09
NE24	19	USA - Nebraska - Lincoln	-	1.33	0.08	0.07	0.15
VI9	22	USA - Virginia - Midlothian	-	1.53	0.15	0.16	0.14
VI10	19	USA - Virginia - Virginia Beach	-	2.80	0.32	0.26	0.19
MA1	19	USA - Massachusetts - Worchester	-	1.53	0.17	0.15	0.00
NC1	12	USA - North Carolina	-	2.93	0.41	0.38	0.06
NC2	23	USA - North Carolina	-	3.27	0.38	0.30	0.22*
NY1	10	USA - New York	-	2.40	0.37	0.20	0.47*
GA1	7	USA - Georgia	-	2.00	0.27	0.23	0.16
GA2	20	USA - Georgia	-	2.67	0.36	0.31	0.13*
GA3	8	USA - Georgia	-	2.80	0.39	0.28	0.31
GA4	15	USA - Georgia	-	2.80	0.42	0.28	0.34*
KS1	17	USA - Kansas	-	3.87	0.44	0.36	0.18*
Boylston	32	USA - Mississippi - Boylston	-	2.20	0.34	0.24	0.29*
Jap120	23	Japan - Iwate - Shiwa	1978	1	-	-	-
Jap212	27	Japan - Iwate - Shiwa	1978	1	-	-	-
Jap308	25	Japan - Iwate - Shiwa	1978	1	-	-	-
Kasumig2	36	Japan - Ibaraki - Kasumigaura	1971	1	-	-	-
Kasumig3	29	Japan - Ibaraki - Kasumigaura	1971	1.20	0.07	0.07	-0.02
Kasumig5	35	Japan - Ibaraki - Kasumigaura	1971	1	-	-	-
Kosa	35	Japan - Kumamoto - Kosa	1905	1.27	0.06	0.06	0.03
GuangDong	41	Chine - Guangdong	1988	1	-	-	-
GuangXi	32	Chine - GuangXi	-	1	-	-	-
Laoshan	39	Chine - Shandong - Laoshan	1982	1	-	-	-
ShanDong	35	Chine - Shandong	1982	1	-	-	-

Mad23PC	12	Madeira Island - Porto da Cruz	2008	1	-	-	-
Mad24C	7	Madeira Island - Calheta	2008	1	-	-	-
128S	17	Portugal - Setubal - Grândola	1999	1.07	0.03	0.01	0.62*
TR1	30	Portugal - Setubal - Troia	1999	1	-	-	-
TR2	27	Portugal - Setubal - Troia	1999	1	-	-	-
AM2	21	Portugal - Setubal - Aguas de Moura	1999	1	-	-	-
Comporta	28	Portugal - Setubal - Comporta	1999	1	-	-	-
E182	13	Portugal - Coimbra - Penela	2008	1	-	-	-
E1069	14	Portugal - Viseu - Castro Daire	2008	1	-	-	-

Note: Mean *Na*, mean number of alleles per sample; *He*, expected heterozygosity and *Ho*, observed heterozygosity. F_{IS} was calculated as described by Weir & Cockerham (1984). "*" indicates that the result of the HWE test was significant at the 5% level after correction for multiple comparison. "-" indicates that *He*, *Ho* and F_{IS} were not computed for monomorphic samples on one hand and that the date of first observation was unknown or not applicable on the other hand.

Table 2: Results of the ABC analyses carried out to infer the history of the pinewood nematode invasion in Asia. Posterior probabilities and their 95% confidence intervals (CI) were obtained using a polychotomous logistic regression (J. M. Cornuet et al., 2010; J. M. Cornuet et al., 2008) and the LDA option (J. M. Cornuet et al., 2014; Arnaud Estoup et al., 2012). The selected scenario in each analysis is shown in bold.

Analysis and scenario	Posterior probability [95% CI]	PER
Analysis - Japan		
Prior set 1 - 30 generations/year		
Native → Japan1 ; Native → Japan2	**0.965 [0.962 ; 0.968]**	0.117
Native → Japan1 → Japan2	0.035 [0.032 ; 0.038]	
Prior set 2 - 30 generations/year		
Native → Japan1 ; Native → Japan2	**0.985 [0.983 ; 0.987]**	0.106
Native → Japan1 → Japan2	0.015 [0.013 ; 0.017]	
Analysis - China		
Prior set 1 - 30 generations/year		
Native → China1 ; Native → China2	**0.932 [0.928 ; 0.935]**	
Native → China1 → China2	0.018 [0.017 ; 0.020]	0.163
Native → China2 → China1	0.050 [0.048 ; 0.053]	
Prior set 2 - 30 generations/year		
Native → China1 ; Native → China2	**0.696 [0.686 ; 0.706]**	
Native → China1 → China2	0.088 [0.082 ; 0.093]	0.277
Native → China2 → China1	0.216 [0.207 ; 0.225]	
Analysis - China/Japan		
Prior set 1 - 30 generations/year		
Native → Japan1 ; Native → China1	0.256 [0.246 ; 0.265]	0.100
Native → Japan1 → China1	**0.744 [0.735 ; 0.754]**	
Native → Japan1 ; Native → China2	**0.827 [0.818 ; 0.836]**	0.092
Native → Japan1 → China2	0.173 [0.164 ; 0.182]	
Native → Japan2 ; Native → China2	0.342 [0.334 ; 0.350]	
Native → Japan2 → China2	0.170 [0.164 ; 0.176]	
Native → China2 → Japan2	0.079 [0.074 ; 0.086]	0.469
Native → China2 ; Native+China2 → Japan2	**0.409 [0.402 ; 0.416]**	
Prior set 2 - 30 generations/year		
Native → Japan1 ; Native → China1	0.112 [0.105 ; 0.119]	0.105
Native → Japan1 → China1	**0.888 [0.881 ; 0.895]**	
Native → Japan1 ; Native → China2	**0.794 [0.784 ; 0.805]**	0.089
Native → Japan1 → China2	0.206 [0.195 ; 0.216]	

Native → Japan2 ; Native → China2	**0.380 [0.372 ; 0.388]**	
Native → Japan2 → China2	0.222 [0.216 ; 0.229]	0.513
Native → China2 → Japan2	0.104 [0.099 ; 0.110]	
Native → China2 ; Native+China2 → Japan2	0.294 [0.287 ; 0.301]	

Note: PER stands for Posterior Error Rate. Native stands for the native area, represented by the sample NE2. Japan1 stands for the Japanese genetic unit 1, represented by the sample Kosa. Japan2 stands for the Japanese genetic unit 2, represented by the sample Kasumig3. China1 stands for the Chinese genetic unit 1, represented by the sample Laoshan. China 2 stands for the Chinese genetic unit 2, represented by the sample Guangdong. See Results for details.

Table 3: Results of the ABC analyses carried out to determine the origin of the pinewood nematode invasion in Portugal. Posterior probabilities and their 95% confidence intervals (CI) were obtained using a polychotomous logistic regression (J.-M. Cornuet, Ravigné, & Estoup, 2010; J.-M. Cornuet et al., 2008) and the LDA option (J. M. Cornuet et al., 2014; Arnaud Estoup et al., 2012). The selected scenario in each analysis is shown in bold.

Analysis and scenario	Posterior probability [95% CI]	PER
Prior set 1		
30 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.529 [0.509 ; 0.550]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.013 [0.000 ; 0.040]	
Native → Japan1 → China1 and Portugal ; Native → China2	0.225 [0.203 ; 0.246]	0.231
Native → Japan1 → China1 → Portugal ; Native → China2	0.233 [0.200 ; 0.266]	
15 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.574 [0.546 ; 0.603]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.0210 [0.000 ; 0.067]	
Native → Japan1 → China1 and Portugal ; Native → China2	0.209 [0.174 ; 0.244]	0.235
Native → Japan1 → China1 → Portugal ; Native → China2	0.196 [0.143 ; 0.250]	
45 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.497 [0.479 ; 0.514]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.021 [0.000 ; 0.044]	
Native → Japan1 → China1 and Portugal ; Native → China2	0.258 [0.241 ; 0.275]	0.248
Native → Japan1 → China1 → Portugal ; Native → China2	0.224 [0.197 ; 0.252]	
Prior set 2		
30 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.744 [0.716 ; 0.772]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.039 [0.000 ; 0.118]	
Native → Japan1 → China1 and Portugal ; Native → China2	0.049 [0.000 ; 0.131]	0.158
Native → Japan1 → China1 → Portugal ; Native → China2	0.168 [0.079 ; 0.257]	
15 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.605 [0.568 ; 0.642]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.077 [0.023 ; 0.130]	
Native → Japan1 → China1 and Portugal ; Native → China2	0.123 [0.068 ; 0.178]	0.134
Native → Japan1 → China1 → Portugal ; Native → China2	0.195 [0.123 ; 0.267]	

45 generations/year		
Native → Japan1 → China1 ; Native → China2 ; Native → Portugal	**0.598 [0.569 ; 0.627]**	
Native → Japan1 → China1 ; Native → China2 → Portugal	0.0372 [0.000 ; 0.078]	0.182
Native → Japan1 → China1 and Portugal ; Native → China2	0.082 [0.042 ; 0.124]	
Native → Japan1 → China1 → Portugal ; Native → China2	0.282 [0.234 ; 0.330]	

Note: PER stands for Posterior Error Rate. Native stands for the native area, represented by the sample NE2. Japan1 stands for the Japanese genetic unit 1, represented by the sample Kosa. Japan2 stands for the Japanese genetic unit 2, represented by the sample Kasumig3. China1 stands for the Chinese genetic unit 1, represented by the sample Laoshan. China 2 stands for the Chinese genetic unit 2, represented by the sample Guangdong. Portugal stands for the Portuguese genetic unit, represented by the sample TR1. See Results for details.

 FIGURE LEGENDS

Figure 1: Competing scenarios in the ABC analyses performed to investigate the origin of the pinewood nematode invasion in Portugal. Native stands for the native area, represented by the sample NE2. Japan1 stands for the Japanese genetic unit 1, represented by the sample Kosa. Japan2 stands for the Japanese genetic unit 2, represented by the sample Kasumig3. China1 stands for the Chinese genetic unit 1, represented by the sample Laoshan. China2 stands for the Chinese genetic unit 2, represented by the sample Guangdong. Portugal stands for the Portuguese genetic unit, represented by the sample TR1. The different colors represent different effective population sizes. See Table S1 and Results for details.

Figure 2: Neighbor-Joining tree built with all the pinewood nematode samples used in this study. Tree based on Cavalli-Sforza and Edwards distances (Cavalli-Sforza and Edwards 1967), corrected for null alleles (Chapuis and Estoup, 2007). The American samples were highlighted in orange, the Japanese samples in green, the Chinese samples in purple and the Portuguese samples in blue.

Figure 3: Genetic structure of the pinewood nematode samples used in this study. Bar plots of the coefficients of co-ancestry obtained in STRUCTURE analyses with several values of K. Each bar corresponds to one individual nematode and each cluster is represented with a particular color. The major clustering patterns are shown (see Figure S7B for the minor ones, when applicable).

Figure 4: Worldwide invasion routes of the pinewood nematode as inferred from this study. The year of first infestation for every invaded area is indicated in parenthesis.

801 **Figure 1**

803   **Figure 2**

Figure 3

807 **Figure 4**