1   **Changes in the Active, Dead, and Dormant Microbial Community Structure Across**
2   **a Pleistocene Permafrost Chronosequence**
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4   Running Title:
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6   Alexander Burkert[1], Thomas A. Douglas[2], Mark P. Waldrop[3], and Rachel Mackelprang[1]
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8   [1] Department of Biology, California State University, Northridge, CA 91330, USA
9
10   [2] U.S. Army Cold Regions Research and Engineering Laboratory, Fort Wainwright,
11   Alaska 99703, US
12
13   [3] Geology, Minerals, Energy, and Geophysics Science Center, U.S. Geological Survey,
14   Menlo Park, CA, 94025, USA
15
16   *Corresponding author information:
17   Address: 18111 Nordhoff St. Northridge, CA 91330
18   Email: rachel.mackelprang@csun.edu
19   Phone: 818-677-4589
20   Fax: 818-677-2034

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The authors declare no conflict of interest

ABSTRACT

Permafrost hosts a community of microorganisms that survive and reproduce for

millennia despite extreme environmental conditions such as water stress, subzero

temperatures, high salinity, and low nutrient availability. Many studies focused on

permafrost microbial community composition use DNA-based methods such as

metagenomic and 16S rRNA gene sequencing. However, these methods do not

distinguish between active, dead, and dormant cells. This is of particular concern in

ancient permafrost where constant subzero temperatures preserve DNA from dead

organisms and dormancy may be a common survival strategy. To circumvent this

we applied: (i) live/dead differential staining coupled with microscopy, (ii)

endospore enrichment, and (iii) selective depletion of DNA from dead cells to

permafrost microbial communities across a Pleistocene permafrost chronosequence

(19K, 27K, and 33K). Cell counts and analysis of 16S rRNA gene amplicons from

live, dead, and dormant cells revealed how communities differ between these pools

and how they change over geologic time. We found clear evidence that cells capable

of forming endospores are not necessarily dormant and that the propensity to form

endospores differed among taxa. Specifically, Bacilli are more likely to form

endospores in response to long-term stressors associated with permafrost

environmental conditions than members of Clostridia, which are more likely to

persist as vegetative cells over geologic timescales. We also found that exogenous

DNA preserved within permafrost does not bias DNA sequencing results since its

removal did not significantly alter the microbial community composition. These

results extend the findings of a previous study that showed permafrost age and ice

content largely control microbial community diversity and cell abundances.

IMPORTANCE

The study of permafrost transcends the study of climate change and

exobiology. Permafrost soils store more than half earth's soil carbon despite

covering ~15% of the land area (Tarnocai et al 2009). This permafrost carbon is

rapidly degraded following thaw (Tarnocai C et al 2009, Schuur et al 2015).

Understanding microbial communities in permafrost will contribute to the

knowledge base necessary to understand the rates and forms of permafrost C and N

cycling post thaw. Permafrost is also an analog for frozen extraterrestrial

environments and evidence of viable organisms in ancient permafrost is of interest

to those searching for potential life on distant worlds. If we can identify strategies

microbial communities utilize to survive permafrost we can focus efforts searching

for evidence of life on cryogenic cosmic bodies. Our work is significant because it

contributes to an understanding of how microbial life adapts and survives in the

extreme environmental conditions in permafrost terrains across geologic timescales.

INTRODUCTION

Permafrost contains active microbial communities that are moderately diverse (1-

2). DNA-based methods, such as metagenomics and 16S rRNA gene sequencing, are

commonly used to interrogate these communities with the underlying assumptions that

the data represent intact viable cells or that nonviable cells do not strongly affect

conclusions drawn from whole community DNA. However, DNA from dead cells may

drastically alter estimates of diversity and abundance. Furthermore, because many

communities host dormant cells, DNA-based approaches may not represent active

members. In temperate soils, up to 40% of DNA is from dead or compromised cells (3).

In permafrost, the amount of 'relic' DNA may be even higher because frozen conditions

preserve DNA from dead cells. In non-permafrost environments, multi-omic approaches

provide functional information from RNA and protein, largely overcoming the problems

of dormancy and relic DNA. In permafrost, this strategy has only been successfully

applied in young near-surface permafrost due to low biomass and activity in older

samples (4-5). Therefore, in older deeper permafrost, other methods must be used to

differentiate between live, dead, and dormant cells.

Alternative methods to multi-omics investigations, include microscopy, stable

isotope probing, and physiological measurements with microbial isolates, have been used

on permafrost samples to demonstrate that an active community exists. Electron-

microscope examinations have shown evidence of apparently intact (no visual damage to

cell envelopes), compromised (cell envelope ruptures), and dormant (endospores and

cells with thick capsules) cells in permafrost (6-8). Using live/dead differential staining

coupled with fluorescence microscopy, Hansen et al (2007) estimated that 26% of cells

from a permafrost microbial community on Svalbard were viable (2). Stable isotope

probing revealed that permafrost microorganisms can build biomass and replicate their

genomes at subzero temperatures as demonstrated through the incorporation of ^{14}C-

labeled acetate into lipids (9) and DNA (10). Similarly, studies involving permafrost

microbial isolates have discovered microorganisms capable of reproduction at -15°C and

metabolism down to -25°C (11-12).

Though these studies show permafrost microbes exist in active states, dormancy is

still a viable strategy for many taxa. Microorganisms enter dormancy in variety of ways,

though the hardiest and most persistent is the endospore formed by some Gram-positive

taxa in response to nutrient limitation, temperature extremes, or other stressors (13).

However, endospore formation does not appear to be a universal survival strategy in

permafrost because the relative abundance of endospore-forming taxa varies substantially

across the Arctic and sub-Arctic, ranging from vanishingly rare to almost 80% (14-16)

The abundance may be related to soil physicochemical properties including depth (17-

18), ice content, and permafrost age (16) (19). Furthermore, endospore-forming taxa in

permafrost are not necessarily dormant. Hultman et al (2015) used RNA to DNA ratios

to show that Firmicutes in young Holocene permafrost are more active than expected

based on DNA abundance alone, demonstrating that dormancy cannot be inferred based

solely on 16S rRNA gene amplicon sequencing (4).

While endospore formation may contribute to long-term survivorship in

permafrost it is unclear whether this strategy is optimal across geologic timescales.

Despite resistance to extreme conditions, DNA within an endospore can still accumulate

damage (20). Typically, DNA damage is repaired upon germination by DNA repair

machinery (13). However, damage accumulated over geologic timescales may be beyond

the ability of repair enzymes to remedy. Willerslev et al (2004) amplified 16S rRNA

genes from globally distributed permafrost soils ranging from 0 – 600 kyr old and found

non-endospore-forming Actinobacteria were more highly represented than the endospore-

forming Firmicutes in samples of increasing age (>100 kyr) (19). Johnson et al (2007)

used a uracil-N-glycosylase treatment to break down damaged DNA extracted from

ancient permafrost and found Actinobacteria rather than endospore-forming Firmicutes

were more highly represented in the oldest samples (>600 kyr) (21). This suggests

metabolic activity and active DNA repair may be a better survival strategy than

dormancy in increasingly ancient permafrost. Therefore, endospore formation may not

be an optimal survival strategy in permafrost for timescales beyond the Late Pleistocene.

Here we address the unresolved question of whether indicators of life (DNA,

spores) in increasingly ancient permafrost are from live viable microbial communities or,

instead, originate from only dead and dormant cells. We hypothesized that dormancy

would increase with age but endospore formation wouldn't be the sole mechanism for

survival. Viable non-endospore forming taxa should also be present and some endospore-

forming cells should exist in a non-dormant state. We also hypothesized that cell

abundance would decrease with age, but cell abundance could also be affected by the

biophysical conditions of the site when it was frozen into permafrost. Further, we asked

whether preserved DNA from dead cells is so compositionally dissimilar from DNA from

live cells that it alters inferences about microbial community structure in permafrost

made with total soil DNA.

MATERIALS & METHODS

Permafrost Sample Collection. We collected frozen permafrost samples from the

United States Cold Regions Research and Engineering Laboratory (CRREL) Permafrost

Tunnel research facility located 16 km north of Fairbanks, Alaska (64.951°N, -

147.621°W) (Figure S1A). The facility includes 300 m of tunnels excavated into the

permafrost. The tunnel where samples were collected extends 110 m horizontally at a

depth of ~15 m into a hillside exposing a chronosequence of late Pleistocene permafrost

(Figure S1B) (22-23). The temperature of the tunnel is maintained by refrigeration at -3

^{o}C. It contains massive ice wedges (24-25) surrounded by high organic content ice

cemented windblown silt (26). In April 2016, we collected ice cemented silt from three

locations inside the tunnel representing three age categories: 19K (approximately 10 m

from the portal), 27K (54 m), and 33K (88 m) as determined previously by radiocarbon

dating (16). After removing the sublimated surface layer (~5 cm) from the walls of the

tunnel, we collected five replicate cores per age category using a 7.5 x 5 cm key hole saw

attached to a power drill as described previously (16). Cores were shipped back to

California State University, Northridge (CSUN) on dry ice and stored at -20 °C.

Permafrost Subsampling. For subsampling, we placed cores on autoclaved foil at room

temperature for 10 minutes to allow the outer ~1 cm to thaw and soften. Surface

contamination was removed by scraping the outer layer with an autoclaved knife to

expose the uncontaminated frozen interior. We sub-sectioned the remaining

uncontaminated material using a fresh knife into sterile 50 ml falcon tubes and high-

density polyethylene bags in preparation for downstream treatment.

Soil Chemistry. Ice content was measured via gravimetric moisture analysis. To

determine pH, soil was diluted 1:1 in a $CaCl_2$ solution and measured using a Hannah

benchtop meter with attached probe (Hanna Instruments, Woonsocket, RI). Percent total

carbon, organic carbon, total nitrogen, and carbon:nitrogen were measured via dry

combustion and direct measurement of total nutrients using an Elementar analyzer

(Elementar, Langenselbold, Germany). Dissolved organic carbon (DOC) was measured

using diluted meltwater on a Shimadzu total organic carbon (TOC) analyzer (Shimadzu

Corporation, Kyoto, Japan). Electrical conductivity (EC) was measured using a digital

benchtop meter with a potentiometric probe submerged in a diluted soil solution (Hanna

Instruments, Woonsocket, RI).

Cell Separation for Enumeration via Microscopy. For cell enumeration, cells were

separated from the permafrost soil matrix using Nycodenz density cushion centrifugation

as described previously (27-31). To separate cells from soil debris we disrupted 1.5 g of

soil in a mild detergent consisting of 2 ml of 0.05% Tween 80 and 50 mM tetrasodium

pyrophosphate buffer (TTSP) (28) (32) and sonicated for 1 minute at 20V using a

QSonica ultrasonicator (QSonica, Newtown, CT) with a 0.3 cm probe. Sonicated samples

were centrifuged at 750 x g for 7 min at 4 °C to remove large particles and debris. We

extracted 600 µl of the supernatant and layered it over 600 µl of 1.3 g/L Nycodenz

(Accurate Chemical, Westbury, NY) solution in a 2 ml tube. The tubes were centrifuged

at 14,000 x g for 30 min at 4 °C. We transferred 600 µl of the upper and middle phase

containing bacterial cells into a sterile 2 ml tube and centrifuged at 10,000 x g for 15 min

at 4 °C. The supernatant was discarded, and the pellet was resuspended in 1 ml of 0.85%

NaCl solution.

Live/Dead Staining. The Live/Dead BacLight Bacterial Viability Kit (Invitrogen

Detection Technologies, Carlsbad, CA) was used to differentially stain live and dead

cells. We added 3 µl of a 1:1 mixture of 3.34 mM SYTO 9 and 20 mM propidium iodide

solution to 1 ml cell suspensions as per the manufacturer's protocols. Stained

suspensions were incubated at room temperature for 15 minutes in the dark to allow the

dyes to permeate cells and bind to DNA.

DAPI Staining. DAPI staining was performed to obtain total cell counts. After removal

of soil debris (as described above in Cell Separation for Enumeration via Microscopy),

we added 3 µl of 14.3 mM DAPI stock solution to each 1 ml cell suspension. Stained

suspensions were incubated in the dark at room temperature for 15 minutes.

Cell Enumeration. We diluted and vacuum filtered the stained suspensions onto a 25

mm diameter 0.2 µm pore size black polycarbonate membrane which we placed on a

slide with sterile forceps. Samples were observed at 100X magnification on a single

focal plane using a Zeiss Axio Imager M2 fluorescence microscope coupled to an

Apotome 2.0 System with appropriate filters for each stain (Zeiss, Oberkochen,

Germany). We counted fifteen fields of view for live/dead and DAPI stained cells for

each sample (33-34). The average number of cells per field of view was multiplied by the

area of the filter and the dilution factor then corrected for dry weight to calculate the

average number of cells per gram of dry weight.

Cell Separation for DNA Extraction (SHMP Method). While Nycodenz density

centrifugation is effective at removing soil debris, making it ideal for microscopic

visualization, it is biased against endospores and heavily attached cells (35). To address

this we used a second less-biased method to extract cells for DNA-based analyses (36).

We disrupted 5 g of sample in 25 ml of 1% sodium hexametaphosphate buffer (SHMP)

and sonicated for 1 minute at 20V using a QSonica ultrasonicator with a 0.6 cm probe.

The samples were left for 15 minutes to allow large particles and debris to settle before

transferring the supernatant to a clean 50 ml tube. We added 15 ml of 1% SHMP to the

pellet and sonicated again with a 0.6 cm probe for 1 minute at 20V. The mixture was

incubated for another 15 minutes to allow debris to settle and then we combined the new

supernatant with the supernatant from the previous step. To further remove large

particles and debris we centrifuged the combined supernatant at 20 x g for 2 minutes. We

then divided the supernatant equally into two 50 ml tubes as an experimental group

(which received either the propidium monoazide treatment or the lysozyme enzyme

treatment) and a control group (which received no treatment). These tubes were

centrifuged at 10,000 x g for 15 minutes to pellet biomass. The biomass pellet was either

stored at -20 °C to await DNA extraction (in the case of the control pellets) or

immediately used for downstream treatments.

Depletion of DNA from Dead Cells via Propidium Monoazide Treatment. To deplete

DNA from dead cells we treated cells extracted using the SHMP method with propidium

monoazide (PMAxx, Biotium Inc., Hayward, CA), which is a DNA-intercalating dye

similar to the nucleic acid dye propidium iodide (37). It is selectively permeable, passing

through the impaired membranes of dead cells, but it is unable to penetrate the

membranes of living cells. In the presence of intense bright light, the azide group enables

propidium monoazide to covalently cross link double stranded DNA, preventing its

amplification via PCR (38).

For the propidium monoazide treatment we resuspended the extracted cell pellets

in 500 μl of 0.85% NaCl solution and placed them in clear 1.5 ml microcentrifuge tubes.

We added 2.5 μl of 20 mM propidium monoazide solution to each microcentrifuge tube

resulting in a final concentration of 100 μM. We increased the concentration from the

commonly used 50 μM due to the presence of leftover soil debris following cell

extraction as recommended for environmental samples by Heise et al (2016) (39-40).

Tubes were incubated in the dark at room temperature for 10 minutes. After incubation

we placed the tubes on a sheet of foil in an ice bucket to prevent warming. A 500 W

halogen work lamp was placed 20 cm above the samples for 15 minutes.   Every five

minutes, we mixed the samples gently to ensure even light distribution. Following light

exposure, we centrifuged samples at 10,000 x g for 15 minutes and discarded the

supernatant. We stored these propidium monoazide treated pellets at -20 °C until use in

downstream DNA extractions.

Cell extraction from soil fails to remove all soil particles, which can subsequently

block light penetration and prevent propidium monoazide from cross-linking to DNA.

To verify that our propidium monoazide treatment is effective in the presence of the

small number of remaining particles we extracted cells from temperate control soils

collected from the California State University, Northridge (CSUN) campus, spiked the

sample with 3.6 x 10^8 isopropanol-killed *Escherichia coli* cells, and treated the mixture

with propidium monoazide.   We also performed the treatment without *E. coli* spike-ins

on the same temperate control samples. The amount of DNA removed was determined

by comparing the number of copies of the 16S rRNA gene in the spiked and non-spiked

samples before and after treatment (as determined by qPCR—see below for detailed

protocols). The number of 16S rRNA gene copies decreased by ~66% after treatment

(Student's t-test, t (8.64) = 3.6, p < 0.01) showing that even in the presence of soil

particles, treatment removed 66% of added exogenous DNA.

Endospore Enrichment via Lysozyme Enzyme Treatment. To separate endospores

from vegetative cells we used a lysozyme enzyme treatment involving three steps:

physical, enzymatic, and chemical cell lysis (hereafter referred to as lysozyme enzyme

treatment following the convention of Wunderlin et al. (2016)) (36). The first physical

treatment uses heat to lyse vegetative cells. Second, lysozyme dissolves the cell

membrane followed by a solution of sodium hydroxide (NaOH) and sodium dodecyl

sulfate (SDS) to further disrupt cellular membranes. Finally, a DNAse treatment is used

to degrade the DNA from ruptured cells.

We resuspended cell pellets extracted using the SHMP method with 900 µl of 1X

Tris – EDTA buffer (10 mM Tris and 1 mM EDTA; pH 8) and placed them into 2 ml

tubes. Tubes were placed in a heat block at 60 °C for 10 min with shaking at 80 rpm.

After incubation, we let the tubes cool for 15 min to 37 °C before adding 100 µl of

lysozyme solution (20 mg/ml in 1X TE Buffer) and incubating in a heat block at 37 °C

for 60 min with shaking at 80 rpm. After lysis 250 µl of 3N NaOH and 100 µl of 10 %

SDS was added to the sample, reaching a final volume of 1.35 ml, which we incubated at

room temperature for 60 min at 80 rpm. After the final incubation, we centrifuged the

solution at 10,000 x g for 15 minutes to pellet cell debris and discarded the supernatant.

We resuspended the pellet with 450 µl of water, 50 µl of 1X DNAse reaction buffer and 1

µl DNAse enzyme (New England Biolabs, Ipswich, MA) and left it for 15 min to remove

DNA from lysed and dead cells. Following the DNase treatment we centrifuged the tubes

for 15 min at 10,000 x g to pellet endospores and discarded the supernatant. The pellet

was then resuspended in 1 ml 0.85% NaCl solution to wash away residual DNAse

enzyme. We centrifuged the suspension at 10,000 x g for 15 min, discarded the

supernatant, and stored the lysozyme enzyme treated pellet at -20°C in a sterile

Eppendorf tube until it was used for downstream DNA extraction.

To confirm that lysozyme treatment depletes vegetative cells we extracted cells

from temperate control soils collected on the CSUN campus, spiked the samples with 1.4

x 10^7 live *E. coli* cells, and performed the lysozyme enzyme treatment. We used the

same treatment on the temperate control soils without *E. coli* spike-ins. By comparing

the number of copies of 16S rRNA genes before and after treatment in the spiked and

non-spiked samples we were able to measure the amount of DNA removed (as

determined by qPCR—see below for detailed protocols). Endospore enrichment

treatment resulted in significant (82%) removal of DNA from vegetative *E.coli* cells

(Student's t-test, t (8.02) = -9.68, p < 0.01), demonstrating the effectiveness of lysozyme

enzyme treatment in removing vegetative cells.

DNA Extraction. We performed DNA extractions using a modified bead-beating

protocol capable of lysing endospores, cysts, and cells with thickened walls, all of which

are known to exist in permafrost (8) (41-43). We resuspended propidium monoazide

treated, lysozyme enzyme treated, and control pellets in 775 µl of lysis buffer (0.75M

sucrose, 20 mM EDTA, 40 mM NaCl, 50 mM Tris) and transferred them to a MP Bio

Lysis Matrix E Tube. We added 100 µl of 20 mg/ml lysozyme and incubated the samples

at 37°C for 30 minutes. Following incubation, 100 µl 10% SDS was added and the

samples were homogenized in an MP Biomedicals FastPrep 24 homogenizer for 20

seconds at 4.0 m/s. We placed the samples in a heat block at 99 °C for 2 minutes and

allowed them to cool at room temperature for 5 minutes. We added 25 µl of 20 mg/ml

Proteinase K and incubated samples at 55 °C overnight. The next day we centrifuged the

tubes at 10,000 x g for 15 minutes and transferred the supernatant to a clean 2 ml

Eppendorf tube. We used a FastDNA Spin Kit for Soil (MP Biomedicals, Santa Ana, CA)

for DNA extractions from the lysed cells following the manufacturer's protocols but

omitting the initial lysis step. The DNA was cleaned using a PowerClean DNA Clean-

Up (Mo Bio, Carlsbad, CA) and quantified using a Qubit 2.0 fluorometer (Thermofisher

Scientific, Canoga Park, CA).

PCR Amplification and Sequencing of the 16S rRNA gene. The variable region four

(V4) of bacterial and archaeal 16S rRNA genes from the propidium monoazide treated,

lysozyme enzyme treated, and control groups was amplified as described by Caporaso et

al (2012) but with the addition of 2 µl of 20 mg/ml bovine serum albumin in each PCR

reaction. We used the golay barcoded primer set 515F/806R with an added degeneracy

to enhance amplification of archaeal sequences on the 515F primer and thermal cycling

steps recommended by the Earth Microbiome Project protocol version 4.13 (44). The

amplified PCR products from triplicate reactions for each sample were pooled at

approximately equal concentrations, as measured using a PicoGreen dsDNA

Quantification Assay Kit (Thermofisher Scientific, Canoga Park, CA). We quantified the

pooled 16S rRNA gene amplicons by qPCR using an Illumina Library Quantification Kit

(Kapa Biosystems, Wilmington, MA) on a CFX96 Real Time PCR Detection System

(Bio-Rad, Hercules, CA). 16S rRNA gene amplicons were sequenced with a 2 x 150 bp

v2 Reagent Kit on an Illumina MiSeq instrument.

We demultiplexed and quality filtered raw fastq data using the Quantitative

Insights Into Microbial Ecology (QIIME) software package version 1.9.1 (45).

Sequences were truncated at the first position with a quality score less than 3. Then

forward and reverse sequences were merged with a minimum merged sequence length of

200 bp, a minimum overlapping length of 20 bp, and a maximum of one difference in the

aligned sequence. All sequences that passed quality filtering were *de novo* clustered into

operational taxonomic units (OTUs) at 97% sequence identity using USEARCH (46).

We assigned taxonomy using the RDP classifier (47) with a confidence score of 0.5 (48-

49). For phylogenetic metrics of diversity, a phylogenetic tree was constructed using

FastTree (50) as implemented in QIIME. We rarefied samples to an equal depth (N =

5000 sequences/sample) for all subsequent analyses. One OTU was only abundant (>3%

relative abundance) in the blank samples and samples which had the lowest DNA yields.

This OTU, from the genus Burkholderia, was removed from all other samples because it

was likely a result of laboratory contamination (51). 16S rRNA gene sequence data were

uploaded to the NCBI's sequence read archive under accession number SRP158034.

Quantitative PCR of the 16S rRNA gene. qPCR of the V4 region of the 16S rRNA

gene was conducted on propidium monoazide treated and lysozyme treated temperate soil

DNA after receiving a spike in of *E. coli* to evaluate these methods. Amplification was

accomplished using the 515F/806R primer set. Triplicate qPCR reactions were done in

25 μl volumes (12.5 μl of GoTaq qPCR Master Mix (Promega, Madison, WI), 1.25 μl of

each primer (17) (44), 5 μl nuclease free water, and 5 μl of template DNA) in 96 well

plates on a CFX96 Real Time PCR Detection System (Bio-Rad, Hercules, CA). The

thermal cycler program was as follows: 95 °C for 2 min, 40 cycles of (95 °C 15 sec, 60

°C 60 sec), and a melt curve analysis (60-95 °C). Quantified full-length *E. coli* 16S rRNA

gene amplicons were used to make a standard curve. A negative control lacking template

DNA was performed along with each qPCR.

Statistical Analysis. We tested for differences in soil chemical characteristics between

age categories using a Kruskal Wallis and a Dunn's post hoc test for nonparametric data

using the 'PMCMR' package in R (52). P-values were corrected using the False

Discovery Rate (FDR). Significant differences in the ratio of live to dead cells, the

proportion of live cells, and the direct cell counts for each stain between age categories

was also tested using the same methods.

We used the Shannon Diversity Index and phylogenetic diversity metrics to calculate

alpha diversity using the 'alpha_diversity.py' script in QIIME (53). Differences in alpha

diversity between age categories and treatments were tested the same methods as for soil

chemistry. We computed beta diversity using the weighted UniFrac metric (54).

Differences between samples were visualized using principal coordinate analysis (PCoA)

using the 'phyloseq' package in R (55). We also used non-metric multidimensional

scaling coordinates of microbial community composition between samples for

PerMANOVA analysis as a function of soil chemistry data using the function 'adonis'

from the 'vegan' package in R (54)(56).

Differences in the relative abundance of specific taxa between treatments and age

categories were tested using a linear mixed effect model on rank transformed taxa

abundances and nested treatment as a factor within age using the 'nlme' package in R

(57). P-values were corrected using the FDR. The specific taxa indicated were tested for

various pairwise comparisons between treatments and untreated controls for each age

category using a Mann-Whitney-Wilcoxon test using the 'wilcoxon.test' function in R.

Differences in 16S rRNA gene copy number, as quantified using qPCR, between

propidium monoazide treated samples and untreated controls (including samples given a

spike in with dead *E. coli* cells) and the endospore enriched samples (given a spike in

with live *E. coli* cells) were tested with a t-test using the 't.test' function in R.

RESULTS

Soil Chemistry. Soil physicochemical properties including ice content, carbon, nitrogen,

pH, and conductivity varied significantly among age categories (Kruskal Wallis, $p < 0.01$

for all measurements, Table 1). These values were consistently higher in the intermediate

age category compared to the older and younger samples.

Cell Enumeration. We performed cell counts across the chronosequence using live/dead

and DAPI staining coupled with fluorescent microscopy. Average cell counts ranged

from 3.6 x 10^6 to 9.2 x 10^6 cells [gram dry weight $(gdw)^{-1}$] (live cells), 1.7 x 10^7 to 4.5 x

10^7 cells gdw^{-1} (dead cells), and 2.3 x 10^7 to 4.7 x 10^7 cells gdw^{-1} (total cell count). Live

(Kruskal Wallis, X^2 (2) = 46.25, $p < 0.001$), dead (Kruskal Wallis, X^2 (2) = 53.16, $p <$

0.001), and total (Kruskal Wallis, X^2 (2) = 53.58, $p < 0.001$) counts were significantly

different among categories and higher in the intermediate age samples compared with the

oldest (Dunn's test, $p < 0.01$) and youngest (Dunn's test, $p < 0.01$) samples (Figure 2A).

The proportion of live cells was significantly higher in intermediate aged samples (26%)

compared to the youngest samples (14%) (Kruskal Wallis, X^2 (2) = 9.84, $p < 0.01$,

Dunn's test, $p < 0.01$, Figure 2B). In the oldest samples, 25% of the cells were live

though this was not significantly different than the values observed for the youngest or

intermediate aged samples. The ratio of live cells to dead cells did not change

significantly across the chronosequence (Kruskal Wallis, X^2 (2) = 1.38, $p > 0.05$).

Endospore Enrichment via Lysozyme Enzyme Treatment. The community of cells

remaining after lysozyme enzyme treatment had reduced alpha-diversity compared with

non-treated controls as measured via phylogenetic richness (Kruskal Wallis, $p < 0.05$,

Figure 2D) and the Shannon Index (Kruskal Wallis, $p < 0.05$, Figure S2). The enzyme

treatment increased the relative abundance of three phyla—Firmicutes, Actinobacteria,

and Chlamydiae (Figure 3A). Endospore enrichment tended to increase the relative

abundance of Firmicutes in all age categories but was only significant for the youngest

(Mann-Whitney-Wilcoxon test, $U = 0$, $p < 0.01$) and intermediate (Mann-Whitney-

Wilcoxon test, $U = 2$, $p < 0.05$) age categories. At the class level, endospore enrichment

changed the abundance of Bacilli and Clostridia, but in opposing directions. It increased

the relative abundance of Bacilli across all age categories, growing more pronounced in

the older samples (youngest: 5.8%, intermediate: 9.3%, oldest: 18.6%). These data were

significant for each age category (Mann-Whitney-Wilcoxon test, $U = 2$, $p < 0.05$, Figure

3B). In contrast, endospore enrichment significantly decreased the relative abundance of

Clostridia in the oldest age category (10%) (Mann-Whitney-Wilcoxon test, $U = 2$, $p <$

0.05, Figure 3B). Treatment did not significantly change the relative abundance of

Clostridia in the youngest and intermediate samples. These trends were driven by the

families *Planococcaceae*, *Thermoactinomycetaceae*, *Bacillaceae*, and *Paenibacillaceae*

for Bacilli and the family *Clostridiaceae* for Clostridia (Table 2).

Endospore enrichment increased the relative abundance of Actinobacteria in the

youngest age category from 22.4% to 32.8% (Mann-Whitney-Wilcoxon test, $U = 0$, $p <$

0.01, Figure 3A). This increase was driven by the families *Micrococcaceae* within the

Actinomycetales, as well as the *Gaiellaceae* and *Solirubrobacteraceae* (Table 2). There

were no significant differences in Actinobacteria relative abundance due to endospore

enrichment in the intermediate and oldest samples.

Chlamydiae relative abundance increased in the youngest age category from 0.7%

to 2.0% as a result of endospore enrichment (Mann-Whitney-Wilcoxon test, $U = 0$, $p <$

0.01, Figure 3A). This trend was driven by an increase in the relative abundance of the

family *Chlamydiaceae*. All other major taxa including Proteobacteria,

Alphaproteobacteria, Deltaproteobacteria, Bacteroidetes, Acidobacteria, Chloroflexi, and

Planctomycetes decreased in relative abundance due to the endospore enrichment (Table

S1).

Depletion of DNA from Dead Cells via Propidium Monoazide Treatment. To

determine if DNA from dead cells biases estimates of taxonomic relative abundance from

the whole community we used a propidium monoazide treatment to deplete dead cell

DNA. Depletion did not significantly change the relative abundance of taxa at the

phylum, class, order, or family levels for any of the age categories (Mann-Whitney-

Wilcoxon test, $p > 0.05$). Actinobacteria were consistently less abundant in depleted

samples compared with non-depleted controls. Unlike the results from endospore-

enrichment experiments, propidium monoazide treatment did not change alpha diversity

measurements—there were no significant differences in depleted samples compared with

non-depleted controls (Kruskal Wallis, $p > 0.05$, Figure 2D & S2).

To confirm propidium monoazide treatment successfully depleted DNA from

dead and membrane-compromised cells we determined 16S rRNA gene copy number in

treated and untreated samples using qPCR. Depletion decreased copy number for the

youngest (~24%, Student's t-test, $p > 0.05$), intermediate (~62%, Student's t-test, $t(16) =$

4.11, p < 0.01), and the oldest samples (~58%, Student's t-test, t (26) = 5.45, p < 0.01,

Figure 2C). 16S rRNA gene copy number varied significant across ages for the depleted

samples (Kruskal Wallis, X^2 (2) = 25.05, p < 0.001) and control samples (Kruskal Wallis,

X^2 (2) = 31.25, p < 0.001). The oldest samples had significantly fewer 16S rRNA gene

copies compared to the youngest (Dunn's test, p < 0.01) and intermediate (Dunn's test, p

< 0.01) for both the depleted and control samples (Figure 2C).

16S rRNA gene-based community analysis. To explore how experimental treatments

influenced overall community structure we compared 16S rRNA gene sequences from

19K, 27K, and 33K old permafrost from treated and control samples (n = 5 replicate

samples x 3 age categories x 4 treatments = 60 samples total). Comparisons of diversity

between samples revealed that samples clustered by age regardless of treatment (Figure

S3A). When comparing the samples within each age category, endospore enriched

samples consistently clustered separately from controls and DNA-depleted samples

(Figure S3B-D). Phyla that differed significantly based on age and treatment (endospore

enrichment and depletion of DNA from dead cells) using a linear mixed effect model

included Proteobacteria, Actinobacteria, Firmicutes, Bacteroidetes, Chloroflexi,

Acidobacteria, Planctomycetes, and Chlamydiae (FDR corrected p < 0.01).

Overall, alpha diversity estimates were highest in the youngest category and

lowest in the intermediate age (Table S2). All soil chemistry measurements correlated

significantly with non-metric multidimensional scaling vectors (PerMANOVA, p < 0.05,

Figure S4). Age and ice content were the strongest drivers of community structure

(PerMANOVA, R^2 = 0.49, F = 181.35, p < 0.001; PerMANOVA, R^2 = 0.26, F = 91.98, p

< 0.001). Other measurements had a small but significant effect on community structure:

DOC (PerMANOVA, $R^2 = 0.04$, F = 13.13, p < 0.001), C:N (PerMANOVA, $R^2 = 0.03$, F

= 12.27, p < 0.001), Total C (PerMANOVA, $R^2 = 0.02$, F = 6.63, p < 0.01), and Total N

(PerMANOVA, $R^2 = 0.01$, F = 3.43, p < 0.05).

DISCUSSION

In this study we used three complimentary methods—two of which have never

been applied to permafrost—to distinguish between live, dead, and dormant cell types:

live/dead staining, endospore enrichment, and depletion of DNA from dead cells. We

present evidence that endospore-forming organisms observed in DNA-based studies are

not always dormant. Instead, among potential endospore-forming organisms the tendency

to exist as an endospore was taxa dependent. Under the stressful conditions associated

with long-term interment in permafrost, members of class Bacilli were more likely to

exist as endospores in increasingly ancient permafrost while members of Clostridia were

more likely to remain in a non-dormant state. This trend grew more pronounced as

sample age increased. For Bacilli, this trend was driven by the families *Planococcaceae*,

Thermoactinomycetaceae, *Bacillaceae*, and *Paenibacillaceae* which are predominantly

aerobes (58). For Clostridia, this trend was driven by the family *Clostridiaceae* which

decreased by over 9% in the endospore-enriched samples compared to non-enriched

controls from the oldest aged category. Most members of the family *Clostridiaceae* are

obligate anaerobes, with metabolic strategies ranging from fermentation to anaerobic

respiration to chemolithoautotrophy (58). The limited oxygen availability in permafrost

could be a stressor that contributes to endospore formation for groups of Bacilli while

anaerobic and metabolically diverse Clostridia are able to persist in a non-dormant state.

These data suggest endospore formation is a viable survival strategy for Bacilli against

the conditions of ancient permafrost, at least for the timescales observed in this study.

However, over increasing timescales endospores can accumulate DNA damage. In the

absence of active repair machinery, this damage may become lethal (13) (19) (21).

Further investigations into even older permafrost may show decreases in endospore

forming taxa in favor of organisms that can actively repair DNA damage as has been

suggested previously (19).

The presence of endospore-forming Firmicutes is highly variable across Arctic

permafrost though they have been shown to increase in relative abundance over geologic

timescales (16). Tuorto et al (2014) used stable isotope probing to identify active

community members and found that Firmicutes were not among those replicating their

genomes at subzero temperatures with the caveat that stable isotope probing only

identifies taxa that actively replicate their genomes. The study would not have identified

cells that were metabolically active but not dividing during the experimental timeframe.

(10). Other studies demonstrate that endospore-forming taxa are active. For example,

Hultman et al (2015) found the abundance of transcripts from Firmicutes exceeded

expected levels based on the abundance of DNA sequence reads (4).

Though our endospore treatment was designed to enrich for true endospores it

also enriched for two other phyla—Actinobacteria and Chlamydiae—but only in the

youngest age category. Actinobacteria were previously found to resist this treatment

perhaps due to their ability to form dormant and spore-like structures (59). In our

samples, members of families *Micrococcaceae*, *Solirubrobacteraceae*, and *Gaiellaceae*

were overrepresented in our endospore-enriched group compared to non-enriched

controls. Members of these families can survive radiation, starvation, and extreme

desiccation (21) (58), which may also confer the ability to resist lysozyme treatment.

Unexpectedly, Chlamydiae were also more abundant in the endospore-enriched

samples compared to non-enriched controls from the youngest age category. This was

driven exclusively by the family *Chlamydiaceae*, which are not known to have resting

states (58). All genera within this class are obligate intracellular symbionts of

Acanthamoeba. Acanthamoeba are a genus of single-celled eukaryotes commonly found

in freshwater and soil. They exist in free-living forms and as stress-resistant dormant

cysts (60) which could account for the increase in the relative abundance of Chlamydiae

in the endospore-enriched samples. Several studies have found intact and viable

Acanthamoeba cysts in Holocene and Pleistocene permafrost (61-62). Though we did not

sequence eukaryotic marker genes, it is possible that there are Acanthamoeba in our

samples.

Depleting DNA from dead cells did not significantly alter microbial community

composition suggesting that sequencing 16S rRNA marker genes from total DNA

provides a reasonable representation of community structure in permafrost. This result

would be expected when death rate and the rate of degradation of dead cells is

proportional among taxa (63). Though depletion experiments did not bias estimates of

relative abundance, it reduced 16S rRNA gene copy number by ~48%, suggesting that

~1/2 of DNA from permafrost is 'relic DNA.' Our estimate of relic DNA is higher than

values taken from temperate soil where ~40% of DNA is relic (3).

One potential limitation of using the propidium monoazide treatment to deplete

DNA from dead cells is that soil particles can prevent light penetration and limit efficacy.

We attempted to mitigate this concern by extracting biomass from soil particles,

performing intermittent mixing during treatment, and increasing PMA concentration for

particle-rich samples (39). We also confirmed the efficacy of the experiment by spiking

samples with dead *E. coli* cells and showed that even in the presence of soil particles,

treatment removed 66% of added exogenous DNA.

Direct cell counts and analysis of 16S rRNA amplicons from permafrost soils

showed that age and ice content were the most important drivers of cell abundance and

microbial diversity, similar to a prior study conducted in the Permafrost Tunnel with

samples collected directly adjacent ours (16). While endospore enrichment significantly

altered microbial community structure within each age group compared to non-enriched

samples, the effects of age and ice content were still much stronger predictors of beta

diversity among samples from all ages.

Microbial cell counts ranged from 2.3×10^7 to 4.7×10^7 cells gdw^{-1} and were

consistent with previous studies from Arctic and sub-Arctic permafrost (1) (14-15) (64-

65), but were one to two orders of magnitude lower than is commonly observed in

temperate soils (66-68). Among the three age categories the intermediate had the highest

number of live, dead, and total cells and the highest proportion of live cells. 16S rRNA

gene copy number from total community DNA and from samples depleted of dead DNA

using propidium monoazide showed the highest counts in the intermediate age category

and the lowest counts in the oldest category, similar to cell count data derived from

live/dead staining. The oldest samples also had a markedly higher proportion of live cells

than the youngest samples, though the data narrowly missed the significance cutoff. The

high proportion of live cells in the intermediate and oldest samples may be attributable to

the higher levels of DOC measured in those samples. It also may be due to increased

reliance on necromass as labile carbon becomes less available in older samples. This is

supported by metagenomic evidence from Permafrost Tunnel microbes, showing an

increase in the abundance of genes related to scavenging detrital biomass in older

permafrost (16).

Though cell counts were highest in the intermediate age category, alpha diversity

was lower compared with the youngest and oldest samples. The decrease in alpha

diversity may reflect the high ice content in those cores, which has been shown

previously to decrease diversity in permafrost (42) (69). The observation that cell counts

are greatest in the intermediate samples is consistent with a previous publication from the

Permafrost Tunnel chronosequence, though counts here are greater by an order of

magnitude. This is likely due to increased cell recovery as a result of our biomass

extraction protocol, which used sonication (rather than vortexing) to separate cells from

soil particles (16).

The Live/Dead staining approach uses membrane permeability as a proxy for

viability (70) and has been used extensively in environmental samples (71-73) including

permafrost (2) (16). One potential drawback to using this approach is that live cells can

incorrectly stain as dead, particularly under dark anoxic conditions (74). We suggest this

is unlikely to impact our samples because permafrost is a generally stable environment in

which membrane potentials are well maintained (75) and we stained under aerobic

conditions in the light. However, if our treatment affected membrane potentials, this

would result in an underestimate of the number of live cells.

This study builds on previous work aimed at understanding how microbial

communities adapt to the extreme conditions of permafrost over geologic timescales (16).

We used two strategies never before applied to permafrost (endospore-enrichment and

depletion of DNA from dead organisms) to show that Firmicutes represented in DNA

based studies are not always dormant. We found that preserved DNA from dead

organisms does not bias DNA sequencing results, since its removal using propidium

monoazide did not significantly alter microbial community composition. We confirmed

that both microbial cell counts and diversity decreased between our youngest and oldest

samples and were primarily controlled by sample age and ice content. We also found that

despite the multiple treatments used, differences in community composition among age

categories were robust. There are changes within age categories caused by treatments but

it suggests future studies interrogating microbial communities among diverse permafrost

types may similarly show strong patterns driven by soil physiochemical properties,

including age. Age-driven survival strategies and community structure identified in this

and a companion study from the Permafrost Tunnel (16) may be common across other

types of permafrost.

Expanding investigations to older permafrost samples and permafrost representing

different biogeochemical properties will be essential to building a model for how

microorganisms in permafrost survive for geologic timescales. The question of dormancy

is a key building block of this model. While dormancy appears to be a survival strategy

for microbes not well adapted to life in permafrost (e.g. Bacilli, which are commonly

aerobes or facultative anaerobes), those that appear to be better adapted (e.g. Clostridia,

which are typically anaerobes and have a diverse suite of metabolic strategies to draw

upon) are less likely to resort to dormancy. This suggests an increase in endospore

formation among maladapted taxa upon entrance to the late Pleistocene, but that

metabolic activity may be increasingly necessary in older permafrost. Expanding the age

range of permafrost and conducting further in-depth studies of current samples (such as

quantitative measurements of endospore markers compared with vegetative cells

markers), will be crucial. If life exists on cryogenic bodies in the solar system, it must

have persisted for longer time scales than exist in Earth's permafrost. Thus, a

chronosequence-based approach to understanding survival may allow us to extrapolate to

more ancient cryoenvironments.

Finally, permafrost communities may have an important role in the biogeochemical

cycling of elements and greenhouse gas production following thaw. Our data suggest

DNA-based studies provide a reasonable representation of the taxonomic reservoir

present and poised to decompose soil carbon upon thaw. It also demonstrates that the

taxonomic diversity and the number of cells is governed by soil physicochemical

characteristics of permafrost soil. Communities with less diversity and fewer cells may be

slower to respond to thaw, which has implications for the speed at which carbon is

processed and released into the atmosphere. Our data suggests that the reservoir (both the

number of cells and the diversity) of microbes is controlled by age, carbon, and moisture

content and highlights the need for detailed understanding of how physicochemical

properties shape microbial communities across permafrost environments.

REFERENCES

1.   Steven B, Léveillé R, Pollard WH, Whyte LG. 2006. Microbial ecology and biodiversity in permafrost. Extremophiles 10:259–267.

2.   Hansen AA, Herbert RA, Mikkelsen K, Jensen LL, Kristoffersen T, Tiedje JM, Lomstein BA, Finster KW. 2007. Viability, diversity and composition of the bacterial community in a high Arctic permafrost soil from Spitsbergen, Northern Norway. Environ Microbiol 9:2870–2884.

3.   Carini P, Marsden PJ, Leff JW, Morgan EE, Strickland MS, Fierer N. 2016. Relic DNA is abundant in soil and obscures estimates of soil microbial diversity. bioRxiv 043372.

4.   Hultman J, Waldrop MP, Mackelprang R, David MM, McFarland J, Blazewicz SJ, Harden J, Turetsky MR, McGuire AD, Shah MB, VerBerkmoes NC, Lee LH, Mavrommatis K, Jansson JK. 2015. Multi-omics of permafrost, active layer and thermokarst bog soil microbiomes. Nature 521:208–212.

5.   Coolen MJL, Orsi WD. 2015. The transcriptional response of microbial communities in thawing Alaskan permafrost soils. Front Microbiol 6.

6.   Soina VS, Vorobiova EA, Zvyagintsev DG, Gilichinsky DA. 1995. Preservation of cell structures in permafrost: A model for exobiology. Adv Space Res 15:237–242.

7.   V Dmitriev V, E Suzina N, G Rusakova T, A Gilichinskii D, I Duda V. 2001. Ultrastructural Characteristics of Natural Forms of Microorganisms Isolated from Permafrost Grounds of Eastern Siberia by the Method of Low-Temperature Fractionation. Dokl Biol Sci Proc Acad Sci USSR Biol Sci Sect Transl Russ 378:304–6.

8.   Soina VS, Mulyukin AL, Demkina EV, Vorobyova EA, El-Registan GI. 2004. The Structure of Resting Bacterial Populations in Soil and Subsoil Permafrost. Astrobiology 4:345–358.

9.   Rivkina EM, Friedmann EI, McKay CP, Gilichinsky DA. 2000. Metabolic Activity of Permafrost Bacteria below the Freezing Point. Appl Environ Microbiol 66:3230–3233.

10.   Tuorto SJ, Darias P, McGuinness LR, Panikov N, Zhang T, Häggblom MM, Kerkhof LJ. 2014. Bacterial genome replication at subzero temperatures in permafrost. ISME J 8:139–149.

11.   Bakermans C, Tsapin AI, Souza-Egipsy V, Gilichinsky DA, Nealson KH. 2003. Reproduction and metabolism at − 10°C of bacteria isolated from Siberian permafrost. Environ Microbiol 5:321–326.

12.   Mykytczuk NCS, Foote SJ, Omelon CR, Southam G, Greer CW, Whyte LG. 2013. Bacterial growth at −15 °C; molecular insights from the permafrost bacterium

Planococcus halocryophilus Or1. ISME J 7:1211.

13.   Nicholson WL, Munakata N, Horneck G, Melosh HJ, Setlow P. 2000. Resistance
of Bacillus Endospores to Extreme Terrestrial and Extraterrestrial Environments.
Microbiol Mol Biol Rev 64:548–572.

14.   Jansson JK, Taş N. 2014. The microbial ecology of permafrost. Nat Rev
Microbiol 12:414–425.

15.   Steven B, Briggs G, McKay CP, Pollard WH, Greer CW, Whyte LG. 2007.
Characterization of the microbial diversity in a permafrost sample from the Canadian
high Arctic using culture-dependent and culture-independent methods. FEMS Microbiol
Ecol 59:513–523.

16.   Mackelprang R, Burkert A, Haw M, Mahendrarajah T, Conaway CH, Douglas
TA, Waldrop MP. 2017. Microbial survival strategies in ancient permafrost: insights
from metagenomics. ISME J.

17.   Deng J, Gu Y, Zhang J, Xue K, Qin Y, Yuan M, Yin H, He Z, Wu L, Schuur
EAG, Tiedje JM, Zhou J. 2015. Shifts of tundra bacterial and archaeal communities along
a permafrost thaw gradient in Alaska. Mol Ecol 24:222–234.

18.   Gittel A, Bárta J, Kohoutová I, Mikutta R, Owens S, Gilbert J, Schnecker J, Wild
B, Hannisdal B, Maerz J, Lashchinskiy N, Čapek P, Šantrůčková H, Gentsch N,
Shibistova O, Guggenberger G, Richter A, Torsvik VL, Schleper C, Urich T. 2014.
Distinct microbial communities associated with buried soils in the Siberian tundra. ISME
J 8:841–853.

19.   Willerslev E, Hansen AJ, Rønn R, Brand TB, Barnes I, Wiuf C, Gilichinsky D,
Mitchell D, Cooper A. 2004. Long-term persistence of bacterial DNA. Curr Biol 14:R9–
R10.

20.   Setlow P. 1995. Mechanisms for the prevention of damage to DNA in spores of
Bacillus species. Annu Rev Microbiol 49:29–54.

21.   Johnson SS, Hebsgaard MB, Christensen TR, Mastepanov M, Nielsen R, Munch
K, Brand T, Gilbert MTP, Zuber MT, Bunce M, Rønn R, Gilichinsky D, Froese D,
Willerslev E. 2007. Ancient bacteria show evidence of DNA repair. Proc Natl Acad Sci
104:14401–14405.

22.   Hamilton TD, Craig JL, Sellmann PV. 1988. The Fox permafrost tunnel: A late
Quaternary geologic record in central Alaska. GSA Bull 100:948–969.

23.   Douglas TA, Fortier D, Shur YL, Kanevskiy MZ, Guo L, Cai Y, Bray MT. 2011.
Biogeochemical and geocryological characteristics of wedge and thermokarst-cave ice in
the CRREL permafrost tunnel, Alaska. Permafr Periglac Process 22:120–128.

24.   Katayama T, Kato T, Tanaka M, Douglas TA, Brouchkov A, Fukuda M, Tomita

690 F, Asano K. 2009. Glaciibacter superstes gen. nov., sp. nov., a novel member of the
691 family Microbacteriaceae isolated from a permafrost ice wedge. Int J Syst Evol Microbiol
692 59:482–486.

693 25. Katayama T, Tanaka M, Moriizumi J, Nakamura T, Brouchkov A, Douglas TA,
694 Fukuda M, Tomita F, Asano K. 2007. Phylogenetic Analysis of Bacteria Preserved in a
695 Permafrost Ice Wedge for 25,000 Years. Appl Environ Microbiol 73:2360–2363.

696 26. Douglas TA, Mellon Mt. Sublimation of terrestrial permafrost and the
697 implications for ice-loss landforms on Mars. Nat Commun In review.

698 27. Amalfitano S, Fazi S. 2008. Recovery and quantification of bacterial cells
699 associated with streambed sediments. J Microbiol Methods 75:237–243.

700 28. Portillo MC, Leff JW, Lauber CL, Fierer N. 2013. Cell Size Distributions of Soil
701 Bacterial and Archaeal Taxa. Appl Environ Microbiol 79:7610–7617.

702 29. Eichorst SA, Strasser F, Woyke T, Schintlmeister A, Wagner M, Woebken D.
703 2015. Advancements in the application of NanoSIMS and Raman microspectroscopy to
704 investigate the activity of microbial cells in soils. FEMS Microbiol Ecol 91.

705 30. Lindahl V, Bakken LR. 1995. Evaluation of methods for extraction of bacteria
706 from soil. FEMS Microbiol Ecol 16:135–142.

707 31. Poté J, Bravo AG, Mavingui P, Ariztegui D, Wildi W. 2010. Evaluation of
708 quantitative recovery of bacterial cells and DNA from different lake sediments by
709 Nycodenz density gradient centrifugation. Ecol Indic 10:234–240.

710 32. Kallmeyer J, Smith DC, Spivack AJ, D'Hondt S. 2008. New cell extraction
711 procedure applied to deep subsurface sediments. Limnol Oceanogr Methods 6:236–245.

712 33. Kepner RL, Pratt JR. 1994. Use of fluorochromes for direct enumeration of total
713 bacteria in environmental samples: past and present. Microbiol Rev 58:603–615.

714 34. Epstein SS, Rossel J. 1995. Enumeration of sandy sediment bacteria: search for
715 optimal protocol. Oceanogr Lit Rev 9:759.

716 35. Holmsgaard PN, Norman A, Hede SC, Poulsen PHB, Al-Soud WA, Hansen LH,
717 Sørensen SJ. 2011. Bias in bacterial diversity as a result of Nycodenz extraction from
718 bulk soil. Soil Biol Biochem.

719 36. Wunderlin T, Junier T, Paul C, Jeanneret N, Junier P. 2016. Physical Isolation of
720 Endospores from Environmental Samples by Targeted Lysis of Vegetative Cells. J Vis
721 Exp.

722 37. Nocker A, Cheung C-Y, Camper AK. 2006. Comparison of propidium monoazide
723 with ethidium monoazide for differentiation of live vs. dead bacteria by selective removal
724 of DNA from dead cells. J Microbiol Methods 67:310–320.

38.   Nocker A, Sossa-Fernandez P, Burr MD, Camper AK. 2007. Use of Propidium
Monoazide for Live/Dead Distinction in Microbial Ecology. Appl Environ Microbiol
73:5111–5117.

39.   Heise J, Nega M, Alawi M, Wagner D. 2016. Propidium monoazide treatment to
distinguish between live and dead methanogens in pure cultures and environmental
samples. J Microbiol Methods 121:11–23.

40.   Bae S, Wuertz S. 2009. Discrimination of Viable and Dead Fecal Bacteroidales
Bacteria by Quantitative PCR with Propidium Monoazide. Appl Environ Microbiol
75:2940–2944.

41.   Gilichinsky D a., Wilson G s., Friedmann E i., Mckay C p., Sletten R s., Rivkina
E m., Vishnivetskaya T a., Erokhina L g., Ivanushkina N e., Kochkina G a.,
Shcherbakova V a., Soina V s., Spirina E v., Vorobyova E a., Fyodorov-Davydov D g.,
Hallet B, Ozerskaya S m., Sorokovikov V a., Laurinavichyus K s., Shatilovich A v.,
Chanton J p., Ostroumov V e., Tiedje J m. 2007. Microbial Populations in Antarctic
Permafrost: Biodiversity, State, Age, and Implication for Astrobiology. Astrobiology
7:275–311.

42.   Steven B, Pollard WH, Greer CW, Whyte LG. 2008. Microbial diversity and
activity through a permafrost/ground ice core profile from the Canadian high Arctic.
Environ Microbiol 10:3388–3403.

43.   Niederberger TD, Steven B, Charvet S, Barbier B, Whyte LG. 2009. Virgibacillus
arcticus sp. nov., a moderately halophilic, endospore-forming bacterium from permafrost
in the Canadian high Arctic. Int J Syst Evol Microbiol 59:2219–2225.

44.   Caporaso JG, Lauber CL, Walters WA, Berg-Lyons D, Huntley J, Fierer N,
Owens SM, Betley J, Fraser L, Bauer M, Gormley N, Gilbert JA, Smith G, Knight R.
2012. Ultra-high-throughput microbial community analysis on the Illumina HiSeq and
MiSeq platforms. ISME J 6:1621–1624.

45.   Caporaso JG, Kuczynski J, Stombaugh J, Bittinger K, Bushman FD, Costello EK,
Fierer N, Peña AG, Goodrich JK, Gordon JI, Huttley GA, Kelley ST, Knights D, Koenig
JE, Ley RE, Lozupone CA, McDonald D, Muegge BD, Pirrung M, Reeder J, Sevinsky
JR, Turnbaugh PJ, Walters WA, Widmann J, Yatsunenko T, Zaneveld J, Knight R. 2010.
QIIME allows analysis of high-throughput community sequencing data. Nat Methods
7:335–336.

46.   Edgar RC. 2010. Search and clustering orders of magnitude faster than BLAST.
Bioinformatics 26:2460–2461.

47.   Wang Q, Garrity GM, Tiedje JM, Cole JR. 2007. Naïve Bayesian Classifier for
Rapid Assignment of rRNA Sequences into the New Bacterial Taxonomy. Appl Environ
Microbiol 73:5261–5267.

48.   Claesson MJ, O'Sullivan O, Wang Q, Nikkilä J, Marchesi JR, Smidt H, Vos WM

de, Ross RP, O'Toole PW. 2009. Comparative Analysis of Pyrosequencing and a
Phylogenetic Microarray for Exploring Microbial Community Structures in the Human
Distal Intestine. PLOS ONE 4:e6669.

49.   Liu Z, DeSantis TZ, Andersen GL, Knight R. 2008. Accurate taxonomy
assignments from 16S rRNA sequences produced by highly parallel pyrosequencers.
Nucleic Acids Res 36:e120.

50.   Price MN, Dehal PS, Arkin AP. 2009. FastTree: Computing Large Minimum
Evolution Trees with Profiles instead of a Distance Matrix. Mol Biol Evol 26:1641–1650.

51.   Salter SJ, Cox MJ, Turek EM, Calus ST, Cookson WO, Moffatt MF, Turner P,
Parkhill J, Loman NJ, Walker AW. 2014. Reagent and laboratory contamination can
critically impact sequence-based microbiome analyses. BMC Biol 12:87.

52.   Thorsten Pohlert. 2014. The Pairwise Multiple Comparison of Mean Ranks
Package (PMCMR). R Package.

53.   Faith DP. 1992. Conservation evaluation and phylogenetic diversity. Biol Conserv
61:1–10.

54.   Lozupone C, Knight R. 2005. UniFrac: a New Phylogenetic Method for
Comparing Microbial Communities. Appl Environ Microbiol 71:8228–8235.

55.   McMurdie PJ, Holmes S. 2013. phyloseq: An R Package for Reproducible
Interactive Analysis and Graphics of Microbiome Census Data. PLOS ONE 8:e61217.

56.   Dixon P, Palmer MW. 2003. VEGAN, a package of R functions for community
ecology. J Veg Sci 14:927–930.

57.   José Pinheiro and Douglas Bates. 2017. Package "nlme"□: Linear and Nonlinear
Mixed Effects Models. R Package.

58.   1964. BERGEY'S MANUAL OF DETERMINATIVE BACTERIOLOGY (7th
ed.). Am J Public Health Nations Health.

59.   Wunderlin T, Junier T, Roussel-Delif L, Jeanneret N, Junier P. 2014. Endospore-
enriched sequencing approach reveals unprecedented diversity of Firmicutes in
sediments. Environ Microbiol Rep 6:631–639.

60.   Marciano-Cabral F, Cabral G. 2003. Acanthamoeba spp. as Agents of Disease in
Humans. Clin Microbiol Rev 16:273–307.

61.   Podlipaeva I, Shmakov LA, Gilichinskiĭ DA, Gudkov AV. 2006. [Heat shock
protein of HSP70 family revealed in some contemporary freshwater Amoebae and in
Acanthamoeba sp. from cysts isolated from permafrost samples]. Tsitologiia 48:691–694.

62.   Shmakova LA, Rivkina EM. 2015. Viable eukaryotes of the phylum Amoebozoa

797 from the Arctic permafrost. Paleontol J 49:572–577.

798 63. Lennon JT, Placella SA, Muscarella ME. 2017. Relic DNA contributes minimally
799 to estimates of microbial diversity. bioRxiv 131284.

800 64. Singh P, Singh SM, Singh RN, Naik S, Roy U, Srivastava A, Bölter M. Bacterial
801 communities in ancient permafrost profiles of Svalbard, Arctic. J Basic Microbiol n/a–
802 n/a.

803 65. Rivkina E, Gilichinsky D, Wagener S, Tiedje J, McGrath J. 1998.
804 Biogeochemical activity of anaerobic microorganisms from buried permafrost sediments.
805 Geomicrobiol J 15:187–193.

806 66. Christensen H, Hansen M, Sørensen J. 1999. Counting and Size Classification of
807 Active Soil Bacteria by Fluorescence In Situ Hybridization with an rRNA
808 Oligonucleotide Probe. Appl Environ Microbiol 65:1753–1761.

809 67. Pershina E, Valkonen J, Kurki P, Ivanova E, Chirak E, Korvigo I, Provorov N,
810 Andronov E. 2015. Comparative Analysis of Prokaryotic Communities Associated with
811 Organic and Conventional Farming Systems. PLOS ONE 10:e0145072.

812 68. Torsvik V, Øvreås L. 2002. Microbial diversity and function in soil: from genes to
813 ecosystems. Curr Opin Microbiol 5:240–245.

814 69. Gilichinsky DA. 2002. Permafrost Model of Extraterrestrial Habitat, p. 125–142.
815 *In* Astrobiology. Springer, Berlin, Heidelberg.

816 70. Bernardini J. N., La Duc M. T., Diamond R., Verceles J. 2012. Fluorescence-
817 Activated Cell Sorting of Live Versus Dead Bacterial Cells and Spores. NASA Tech
818 Briefs 22.

819 71. Leuko S, Legat A, Fendrihan S, Stan-Lotter H. 2004. Evaluation of the
820 LIVE/DEAD BacLight Kit for Detection of Extremophilic Archaea and Visualization of
821 Microorganisms in Environmental Hypersaline Samples. Appl Environ Microbiol
822 70:6884–6886.

823 72. Bianciotto V, Minerdi D, Perotto S, Bonfante P. 1996. Cellular interactions
824 between arbuscular mycorrhizal fungi and rhizosphere bacteria. Protoplasma 193:123–
825 131.

826 73. Biggerstaff JP, Le Puil M, Weidow BL, Prater J, Glass K, Radosevich M, White
827 DC. 2006. New methodology for viability testing in environmental samples. Mol Cell
828 Probes 20:141–146.

829 74. Kirchhoff C, Cypionka H. 2017. Propidium ion enters viable cells with high
830 membrane potential during live-dead staining. J Microbiol Methods.

831 75. Ponder MA, Thomashow MF, Tiedje JM. 2008. Metabolic activity of Siberian

832   permafrost isolates, Psychrobacter arcticus and Exiguobacterium sibiricum, at low water
833   activities. Extremophiles 12:481–490.

ACKNOWLEDGEMENTS

This work was supported by the National Aeronautics and Space Administration (NASA) (grant numbers NNX15AM12G and NNH15AB58I). AB acknowledges support from the California State University, Northridge Biology Department Graduate Student Tuition Waiver. We thank Steven Escalante, Tara Mahendrarajah David Romero, and Archana Srinivas for assistance with permafrost subsampling.

FIGURE CAPTIONS

Figure 1. Experimental strategy overview. Live, dead, and dormant cell counts were conducted by separating cells from soil using a Nycodenz density centrifugation, staining with a Live/Dead differential stain or DAPI, and counting via fluorescence microscopy. For the endospore enrichment and dead cell depletion experiments, we separated biomass from soil using a gravity separation technique. To deplete DNA from dead organisms, cell mixtures were treated with propidium monoazide and then exposed to light, causing cross-links with DNA not enclosed by an intact cell envelope or spore coat. The cross-links inhibit downstream PCR amplification. To enrich for endospores, cell mixtures were exposed to lysozyme, heat, and DNAse, which lyses vegetative cells and degrades DNA. In both the endospore enrichment and dead cell depletion experiments, the 16S rRNA gene was amplified and used for downstream analysis.

Figure 2. (A) Direct cell counts as determined by cell staining and fluorescent microscopy. Live/dead and total (DAPI) counts. (B) Proportion of live cells as

determined by direct counts of live cells (stained with SYTO 9) and total cells (stained

with DAPI). (C) 16S rRNA gene copy number in samples depleted of dead DNA using

propidium monoazide and non-depleted controls. qPCR of the V4 region of the 16S

rRNA gene in samples showed that average copy number decreased for all age groups in

the propidium monoazide treated group (DNA depleted samples) compared to the

untreated group (non-depleted controls) and that there was a decrease in viable DNA

across the three age categories. (D) Phylogenetic diversity index compared across age

categories and treatment types. Values show the average of five replicate cores and error

bars show the standard error of the mean. Significant p-values tested by Dunn's post-hoc

test are indicated (* $p < 0.05$, ** $p < 0.01$).

Figure 3. (A) Phylum-level changes in relative abundance due to endospore enrichments.

(B) Class-level changes in relative abundance due to endospore enrichments. Values

show averages of five replicate coress and error bars show standard error of the mean.

Significant p-values tested by the Mann-Whitney-Wilcoxon test are indicated (* $p < 0.05$,

** $p < 0.01$).

Table 1. Permafrost physicochemical characteristics across the three time periods. Values

are averages of five replicates plus/minus one standard error of the mean. Statistical

differences were tested using a Kruskal Wallis test and a Dunn's post hoc test with p-

values corrected using the False Discovery Rate. Significant p-values ($p < 0.05$) are

show in bold.

Table 2. Average percent difference in relative abundance between the endospore-

enriched samples and non-enriched controls across the three age categories. A negative

value shows underrepresentation in the endospore-enriched samples compared to the non-

enriched controls while a positive value shows overrepresentation. Values show averages

of five replicate cores.

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by separating cells from soil using a Nycodenz density centrifugation, staining with a Live/Dead

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29   **Table 1.** Permafrost physicochemical characteristics across the three time periods. Values are averages of five replicates plus/minus

30   one standard error of the mean. Statistical differences were tested using a Kruskal Wallis test and a Dunn's post hoc test with p-values

31   corrected using the False Discovery Rate. Significant p-values ($p < 0.05$) are show in bold.

	19K	p-value (19K vs 27K)	27K	p-value (27K vs 33K)	33K	p-value (33K vs 19K)	Kruskal Wallis $X^2(2)$	p-value
Ice Content (%)	27.70 ± 1.73	**0.001**	50.01 ± 4.16	0.077	35.30 ± 0.50	0.077	12.5	**0.002**
Total C (%)	1.64 ± 0.18	**0.003**	3.47 ± 0.19	0.229	3.06 ± 0.08	0.061	10.82	**0.004**
Organic C (%)	1.62 ± 0.16	**0.014**	2.99 ± 0.30	0.724	2.76 ± 0.06	**0.020**	9.5	**0.009**
Total N (%)	0.16 ± 0.02	**0.013**	0.30 ± 0.01	0.943	0.29 ± 0.01	**0.013**	9.45	**0.009**
C/N Ratio	10.45 ± 0.34	**0.013**	11.70 ± 0.11	0.944	10.62 ± 0.17	**0.013**	9.38	**0.009**
pH	7.32 ± 0.04	0.203	7.46 ± 0.05	**0.004**	6.88 ± 0.10	0.084	10.26	**0.006**
EC (dS/m)	0.39 ± 0.02	**0.003**	0.87 ± 0.04	0.072	0.46 ± 0.03	0.179	11.18	**0.004**
DOC (ppm)	3141 ± 162	**0.009**	9338 ± 1649	0.524	5776 ± 212	**0.029**	9.78	**0.008**

Table 2. Average percent difference in relative abundance between the endospore-enriched samples and non-enriched controls across the three age categories. A negative value shows underrepresentation in the endospore-enriched samples compared to the non-enriched controls while a positive value shows overrepresentation. Values show averages of five replicate cores.

Taxa	Family	19K (%)	U-value	27K (%)	U-value	33K (%)	U-value
Actinobacteria	*Micrococcaceae*	2.4 ± 0.5 **	0	0.0 ± 0.0	9.5	0.0 ± 0.0	11
	Solirubrobacteraceae	1.5 ± 0.4	3	0.6 ± 0.3	4	0.1 ± 0.1	5
	Gaiellaceae	1.0 ± 0.3 *	1	-0.2 ± 0.1	8	0.1 ± 0.1	8.5
Bacilli	*Planococcaceae*	2.4 ± 0.9	4	-8.5 ± 1.0 *	1	4.0 ± 2.6	8
	Thermoactinomycetaceae	0.4 ± 0.1*	0	20 ± 2.5 *	2	7.8 ± 2.2	6
	Bacillaceae	0.7 ± 0.1 *	0	-2.5 ± 1.2	8	1.0 ± 0.4	9
	Paenibacillaceae	2.5 ± 0.4 *	0	0.4 ± 1.6	12	5.6 ± 2.7	4
Clostridia	*Clostridiaceae*	3.1 ± 0.4 **	0	-2.5 ± 1.5	5	-9.2 ± 1.2 **	0

(Mann-Whitney-Wilcoxon test, ** $p < 0.01$, * $p < 0.05$).