

1 **High levels of antibiotic resistance gene expression among** 2 **birds living in a wastewater treatment plant**

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19

20 **Short title:** Antibiotic resistance goes wild

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27 **Keywords:** Metatranscriptomics, Microbiome, Birds, Resistome, Antimicrobial resistance.

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30

31 **Abstract**

32 Antibiotic resistance is rendering common bacterial infections untreatable. Wildlife can
33 incorporate and disperse antibiotic resistant bacteria in the environment, such as water
34 systems, which in turn serve as reservoirs of resistance genes for human pathogens. We
35 used bulk RNA-sequencing (meta-transcriptomics) to assess the diversity and expression
36 levels of functionally active resistance genes in the microbiome of birds with aquatic
37 behavior. We sampled birds across a range of habitats, from penguins in Antarctica to ducks
38 in a wastewater treatment plant in Australia. This revealed 81 antibiotic resistance genes in
39 birds from all localities, including β -lactam, tetracycline and chloramphenicol resistance in
40 Antarctica, and genes typically associated with multidrug resistance plasmids in areas with
41 high human impact. Notably, birds feeding at a wastewater treatment plant carried the
42 greatest resistance gene burden, suggesting that human waste, even if it undergoes
43 treatment, contributes to the spread of antibiotic resistance genes to the wild. Differences in
44 resistance gene burden also reflected the birds' ecology, taxonomic group and microbial
45 functioning. Ducks, which feed by dabbling, carried a higher abundance and diversity of
46 resistance genes than turnstones, avocets and penguins, that usually prey on more pristine
47 waters. In sum, this study helps to reveal the complex factors explaining the distribution of
48 resistance genes and their exchange routes between humans and wildlife.

49

50 Introduction

51 Tons of antibiotics are used annually in clinical and agricultural settings worldwide. Food
52 animals alone consumed over 130,000 tons of antibiotics in 2013 (1), and antibiotic usage
53 by humans increased 65% between 2000 and 2015, reaching 34.8 billion defined daily
54 doses (2). The resulting proliferation and spread of bacteria that are resistant to antibiotics
55 poses a major health and economic threat (3). Genes for the production of antibiotics and
56 antibiotic resistance determinants are naturally found in some microbial species and their
57 presence in the environment is not necessarily an indication of human impact (4, 5).
58 However, the use of antibiotics in clinical and agricultural settings selects for bacteria
59 carrying resistance genes. When these genes are encoded in mobile elements, such as
60 plasmids and conjugative transposons, they can be readily transmitted via horizontal gene
61 transfer between environmental bacteria and clinically important pathogens (i.e. acquired
62 resistance genes). Multiple resistance genes can be present in a single mobile element and
63 the spread of plasmid-borne resistance has jeopardized the efficacy of many antibiotics,
64 including β -lactam drugs of last resort (6, 7).

65 Both the environment and wildlife are major sources and reservoirs of resistance
66 gene diversity (8, 9). The ecological niches and behavior of birds make them particularly
67 likely to transport antibiotic resistant bacteria. Migrating bird species transport pathogens
68 which may contain antibiotic resistance genes across large distances (10, 11). Birds also
69 serve as sensitive bioindicators of environmental contamination with antibiotic resistant
70 bacteria (10, 12-16). For instance, ESBL-producing *Escherichia coli* were found to occur
71 over 3 times more frequently in gulls than in humans in the same region (15). Bacteria
72 resistant to β -lactam and tetracycline drugs are commonly found in the gut microbiome of
73 birds, especially in scavenging and aquatic species, such as waterfowl, gulls and waders
74 (10, 12, 14, 17-20). Aquatic bird species likely acquire these genes through contact with
75 contaminated water. Human sewage is enriched in antibiotic-resistant bacteria, which are
76 only partially removed during the water treatment process (21-26). Birds in contact with
77 wastewater treatment influents or effluents could therefore be at increased risk of acquiring
78 these genes, although empirical data to support this idea are scarce (8).

79 While the majority of studies on birds were based on *in vitro* assessments of bacterial
80 cultures, the development of culture-independent sequencing techniques has substantially
81 expanded our knowledge of the environmental reservoir of resistance genes (7, 9, 27-32).
82 Among these techniques, sequencing the entire set of transcribed (i.e. expressed) genes via
83 'meta-transcriptomics' has rarely been used in the context of antibiotic resistance, despite its
84 advantages. In particular, use of meta-transcriptomics allows data to be obtained from the
85 entire microbial population, with a focus on functionally active genes. This is important
86 because genetic material is a metabolic burden and genes that are not essential tend to be

87 lost (33-36). In the absence of antibiotics, it is likely that resistance genes are regularly lost
88 by bacteria, either by large deletions or gradual deactivation (erosion). Other high-
89 throughput techniques, such as DNA-based metagenomics, cannot distinguish recently
90 deactivated resistance genes from their functional relatives. An alternative is to clone inserts
91 from environmental strains into cultivable vectors (e.g. *E. coli*), select for resistance *in vitro*
92 and then sequence their genomes (e.g. 17, 22). However, this approach can result in bias
93 towards genes present in organisms closely related to the cloning vector (7). Meta-
94 transcriptomics does not have this limitation as the transcripts of all microorganisms are
95 assessed using bulk RNA sequencing. To our knowledge, only two studies have used meta-
96 transcriptomics to report on the presence of resistance genes that are functionally active
97 under natural conditions in human and environmental samples (37, 38).

98 We used meta-transcriptomics to assess the diversity and abundance of antibiotic
99 resistance genes actively transcribed in the microbiome of water birds of Australia and
100 penguins in Antarctica. Birds were sampled across a range of habitats, from remote places
101 in Antarctica and Australia, beaches in Melbourne, the second largest city in Australia, to the
102 ponds of a waste water treatment plant (WWTP) processing half of Melbourne's sewage. We
103 specifically tested whether ducks from the WWTP harbor a higher diversity and abundance
104 of acquired resistance genes, as might be expected given their exposure to partially treated
105 human waste. Additionally, we explore possible associations between resistance gene
106 burden and intrinsic bird traits such as feeding behavior, taxonomic order and gut functional
107 profile (expression of metabolic pathways by the microbiome).

108

109 **Results**

110 Microbiome samples from 110 birds, grouped into 11 libraries (Table S1), contained
111 transcripts corresponding to 81 unique antibiotic resistance genes, previously associated
112 with phenotypic resistance to nine classes of antibiotics (Fig. 1, Table S2). These results
113 only include acquired resistance genes, which are most commonly spread among bacteria
114 via mobile genetic elements, and do not include resistance mediated by chromosomal
115 mutations (e.g. in housekeeping genes). Resistance to tetracyclines and phenicols
116 (chloramphenicol and florfenicol) was present in samples from all bird orders and in all
117 locations, except for one site in Antarctica where phenicol resistance was not detected.

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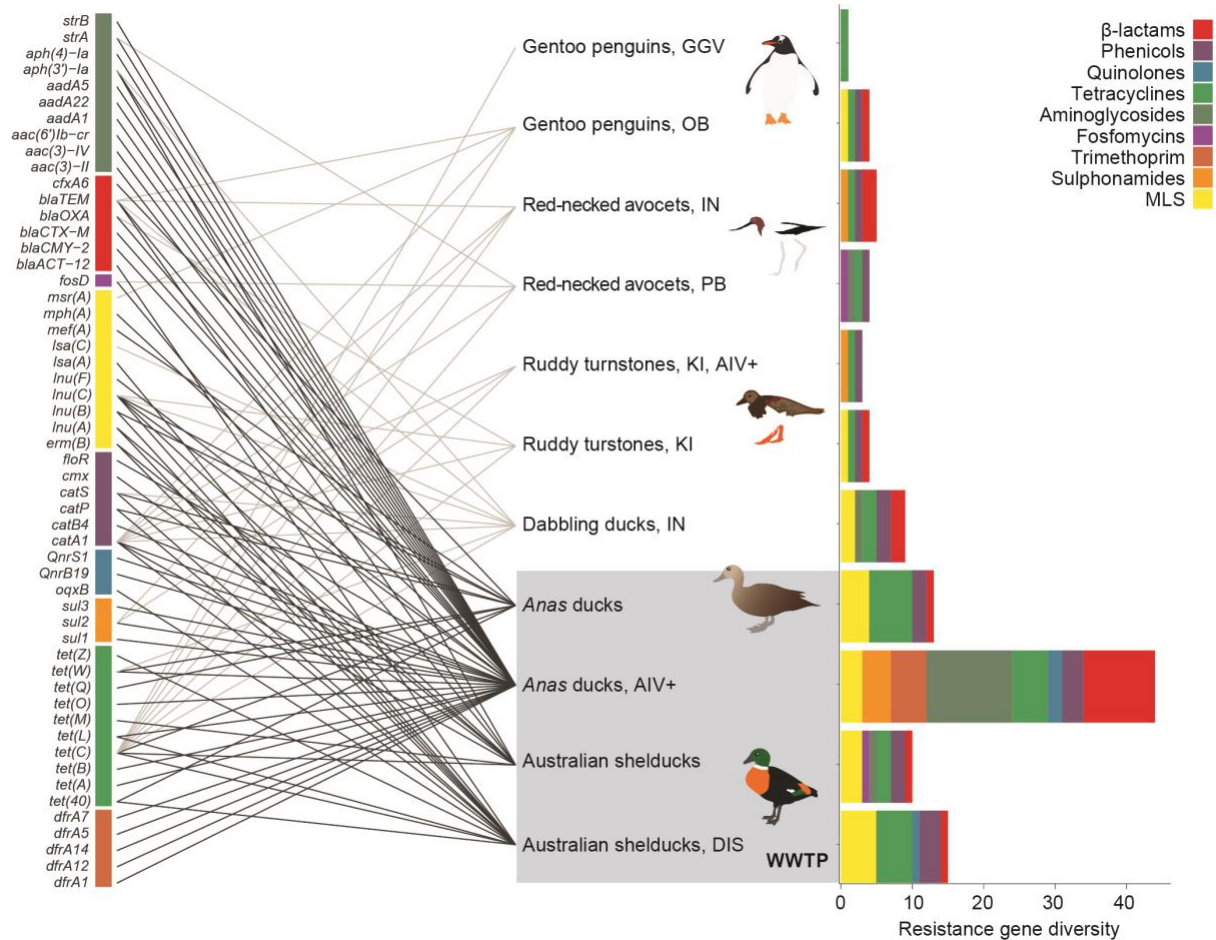
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125 **Figure 1.** Antibiotic resistance genes expressed in the microbiome of wild birds. The graph
 126 on the right shows the diversity of resistance genes observed in each library (containing a
 127 pool of 10 individual birds each), colored by the drug class to which these genes confer
 128 resistance. Closely related gene variants were merged into one category (see Table S2) for
 129 representation on the left side of the figure. Lines link genes to the libraries where they were
 130 found, and dark lines indicate the genes observed in the Wastewater treatment plant
 131 (WWTP) in Melbourne, Australia. PB = Western Port Bay, Melbourne area, Australia; KI =
 132 King Island, Bass Strait, Australia; IN = Innamincka reserve, Australia; OB = O'Higgins Base,
 133 Antarctica; GGV = Gabriel González Videla Base, Antarctica. Libraries of birds infected with
 134 avian influenza virus are indicated with 'AIV+', and the library of diseased birds is indicated
 135 with 'DIS'. MLS = Macrolides, Lincosamide and Streptogramin B resistance. Bird drawings:
 136 M. Wille.

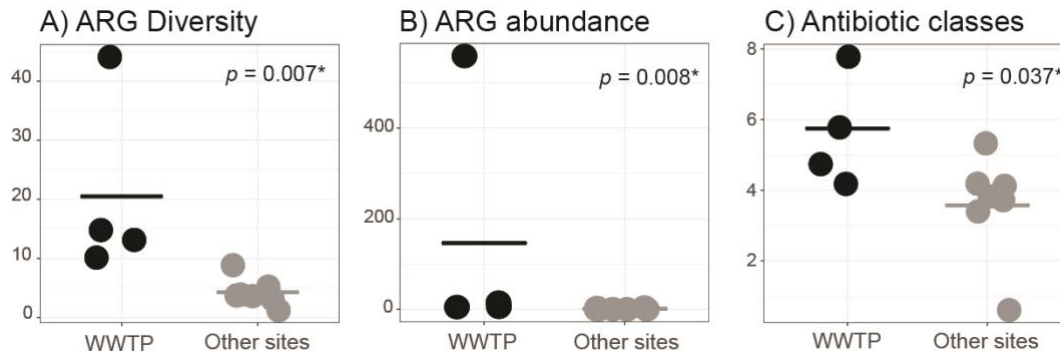
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139 *Anthropogenic impact*

140 Birds foraging at the partially treated lagoons of a wastewater treatment plant (the last stage
 141 of the wastewater treatment process, after aerating and decanting has taken place) had a
 142 significantly higher diversity and abundance of antibiotic resistant genes, as well as a
 143 significantly higher number of antibiotic classes against which these genes confer resistance
 144 (Kruskal-Wallis $p < 0.05$, Fig. 2). For simplicity, we refer to 'resistance gene burden' or
 145 'resistance load' as the resistance gene diversity, abundance (i.e. gene expression levels)
 146 and number of antibiotic classes to which these genes confer resistance. Most notably,

147 ducks (order Anseriformes) foraging at the WWTP harbored 86% of the resistance gene
148 diversity observed, most of which occurred exclusively at the WWTP (Figs. 1, 2, Table S2).
149
150



151
152 **Figure 2.** Diversity and abundance of antibiotic resistance genes (ARG) in birds foraging in
153 a wastewater treatment plant (WWTP) compared with birds from other sites in Australia and
154 Antarctica. Differences between groups were assessed with a Kruskal-Wallis test and p -
155 values are given. Resistance gene abundances were estimated based on a stably
156 expressed host gene.
157
158

159 When only ducks were considered in comparing the effects of wastewater, we
160 observed that those from the WWTP carried more resistance genes than ducks from the
161 remote Innamincka reserve, located in the interior of Australia (Fig. 1, Table S3): ducks from
162 the Innamincka reserve carried nine resistance genes, fewer than the number observed in
163 any library from the WWTP (average 20.5, +/-15.8 SD). The abundance of these genes was
164 also smallest in ducks from Innamincka (2.9, compared with an average of 146.1, +/- 275.2
165 SD in ducks from the WWTP). The number of antibiotic classes to which these genes confer
166 resistance did not differ substantially between sites (5 antibiotics in birds from Innamincka,
167 compared with 5.7, +/- 1.7 in birds from the WWTP).

168 To take into account potential confounding variables, we also analyzed libraries by
169 individual collection localities (Fig S1), and without including diseased birds or birds infected
170 with avian influenza in the analyses (Fig. S2). The results consistently indicated that birds
171 from the WWTP have a higher resistance gene burden than birds from other localities.
172 Importantly, an additional PCR-based assessment of the resistance genes in individual birds
173 from two libraries (n=20 samples) confirmed the results obtained using meta-transcriptomics:
174 we observed 68 resistance gene occurrences (amplifications) in samples from the WWTP
175 and 12 occurrences in other sites (Kruskal-wallis $p=0.0023$, Supplementary Materials, Fig
176 S3).

177 Samples from gentoo penguins (*Pygoscelia papua*) collected in two localities next to
178 research bases in Antarctica, contained five resistance genes in total, conferring resistance

179 against β -lactams (*bla*_{TEM}), tetracyclines (two variants of *tet*(C)), chloramphenicol (*catA1*)
180 and erythromycin (*msr*(A)) (Table S2). The erythromycin-resistance gene, which confers
181 resistance to Macrolides, Lincosamide and Streptogramin B, was observed in penguins only.
182 Penguins living near the research base with the largest human population (O'Higgins Base)
183 contained more antibiotic resistance genes (four genes - *bla*_{TEM}, *msr*(A), *catA1*, and *tet*(C)),
184 than those living next to the more remote Gabriel González Videla Base (one *tet*(C) gene).

185

186 *Host traits and functional context*

187 Our sampling design included birds from a range of habitats and species, which will
188 impact their microbiome and possibly their propensity to carry antibiotic resistance genes.
189 Shelducks and *Anas* ducks (Anseriiformes) feed by dabbling (filtering water). Turnstones
190 and avocets (Charadriiformes) commonly prey on invertebrates, and penguins
191 (Sphenisciformes) prey on fish. Dabbling ducks live in a range of habitats, including nutrient-
192 rich and heavily altered environments. The majority of ducks analyzed here were sampled at
193 the WWTP: 40 samples (4 libraries) at the WWTP and 10 samples (1 library) in a pristine
194 site. Turnstones and penguins on the other hand inhabit pristine habitats. Host taxonomic
195 order therefore serves as a proxy for the ecology of the birds analyzed here. Our results
196 indicated that ducks contained the greatest diversity and abundance of resistance genes,
197 while penguins contained the lowest resistance load (Fig. 1 and Fig. S4).

198 Host ecology is intrinsically linked to microbiome function. By investigating how
199 microbiomes functionally differ among bird orders and collection sites we can gain insights
200 into why some hosts harbor more resistance genes than others. We characterized the
201 metabolic pathways expressed by the microbial community (that is, their functional profile,
202 Table S4). Some of the metabolic pathways observed were produced by common human
203 pathogens (e.g. *E. coli*), but a large proportion of the metabolic products (91%) could not be
204 associated with particular bacterial genera (Table S4). Compared with the human gut, the
205 microbiome of wild animals is far less characterized, and it is expected that several bacterial
206 species remain undetected. Principal coordinate analyses showed that ducks (from
207 Innamincka reserve and from the WWTP) have a distinct microbial metabolism (i.e. set of
208 metabolic pathways) when compared with birds from other sites (Fig S5). We statistically
209 assessed the distinctiveness of functional profiles between sites and bird orders using
210 Random forest analysis, a machine learning approach based on classification trees that has
211 a suitably high discriminating power for use in microbial ecology (39). This analysis revealed
212 a clear distinction (zero out-of-bag classification error) in the functional profiles between
213 birds from the WWTP and other sites, and between Anseriformes and the two other bird
214 orders that comprised the data set (Charadriiformes and Sphenisciformes; Table S5).

215 The bird microbiome, and consequently its functional profile, can also be affected by
216 pathogens (40). We sampled birds with avian influenza virus infection and Newcastle
217 disease symptoms; potential associations between these infections and antibiotic resistance
218 are discussed in the Supplementary Materials.

219

220 Discussion

221 This study shows that clinically important and functional antibiotic resistance genes are
222 widespread, even in birds from areas as remote as Antarctica, and that the resistance gene
223 load is significantly higher in birds living in lagoons of a wastewater treatment facility.

224 Although resistance genes can be found in natural environments regardless of human
225 influence (4, 5), our results indicate that contact with human waste – even if it goes through
226 sewage treatment – appears to have a strong impact on the acquisition of antibiotic
227 resistance genes by avian wildlife.

228 The resistance genes observed here encompass the three major resistance
229 mechanisms of relevance to human infection: (i) drug inactivation, (ii) reduced influx of
230 antibiotics into bacterial cells or increased efflux from cells and (iii) alteration in, or
231 overexpression of, the antibiotic target (7, 41). The observed resistance genes conferred
232 resistance against nine classes of antibiotics (Fig. 1). This number is slightly higher than the
233 six classes of antibiotic resistance observed in humans, pigs, sponges and environmental
234 samples in another study using meta-transcriptomics (37). Among the most common were
235 genes conferring resistance to β -lactam drugs, which is one of the oldest and most widely
236 used antibiotic classes. Genes conferring resistance to aminoglycosides and tetracyclines
237 were also common, in agreement with studies reporting these genes in human-impacted
238 soils and sewage (22, 27, 29, 31).

239 Some of the resistance genes observed are particularly concerning for public health.

240 *bla*_{CTX-M} genes, observed exclusively in birds from the WWTP, play a key role in widely
241 disseminated and highly resistant *E. coli* and *Klebsiella pneumoniae* strains (42). A
242 fosfomycin resistance gene (*fosD*) was found in birds from metropolitan Melbourne (WWTP
243 and Western Port Bay). Fosfomycin was discovered over 40 years ago, it is uncommonly
244 used in humans, but the low resistance levels against this drug have led to a renewed
245 interest in its therapeutic use (43). One of the bird libraries from the WWTP contained a
246 florfenicol resistance gene, which was first observed in *Salmonella typhimurium* (44).
247 Florfenicol is restricted to livestock and veterinary use. It is possible that the presence of this
248 gene is due to the administration of florfenicol to pets and wildlife within the WWTP
249 catchment range. The florfenicol gene has been observed co-located with other resistance
250 genes in integrons and plasmids (44, 45). It is therefore also possible that this gene is found
251 in the WWTP due to co-selection with other genes. We also found resistance against

252 chemically synthesized antibiotic classes, such as quinolones and sulphonamides, which are
253 not expected to be widespread in the environment (unlike naturally produced antibiotics such
254 as penicillin, which is derived from fungi). Quinolone drugs can persist in the environment for
255 long periods (46) and despite being a synthetic drug, the origins of quinolone resistance
256 were traced back to aquatic bacterial species (47). Therefore it is perhaps unsurprising that
257 these genes are found in birds with aquatic behavior (also reported in 19). It is noteworthy,
258 however, that quinolone resistance was only observed in birds near the WWTP, suggesting
259 that these genes most likely derive from bacteria of human origin. One of the WWTP
260 libraries also contained the *aac(6)-Ib-cr* gene (100% identity with clinical isolates), which
261 confers resistance to quinolones and aminoglycosides and is often localized in multidrug
262 resistance plasmids. First reported in Shanghai in 2003, this gene has already been found in
263 several parts of the world, including in a recent report of multidrug-resistant *Salmonella* in
264 Australia (7, 48, 49).

265 The distinct ecological niche and microbiome functioning of the different bird species
266 analyzed here likely influences their acquisition of antibiotic resistant bacteria. Penguins and
267 avocets hunt small aquatic animals, while ducks filter water and sediments to trap plant and
268 animal material. It is possible that ducks ingest large amounts of bacteria while dabbling. In
269 addition, birds may have historical-evolutionary associations with particular microbial
270 species, resulting in a distinct microbiome composition and functioning across avian
271 taxonomic groups. Indeed, a metabarcoding study showed that bird taxonomy explained
272 most of the compositional variation in the microbiome of birds (50). Our functional analyses
273 also suggest that the different bird orders harbor microbial communities with distinct
274 metabolisms. The microbiome of Anseriformes (ducks) expressed genes encoding
275 significantly different metabolic pathways compared with other birds, while there was no
276 clear distinction among Charadriiformes and Sphenisciformes (Fig. S5, Table S5). It is
277 therefore plausible that the high resistance gene expression in ducks from the WWTP is
278 influenced by their distinct microbiome, which in turn reflects their ecological niche and
279 established host-microbe associations. In this scenario, bird traits modulate (amplifying or
280 diminishing) the human impact on the spread of resistance genes.

281 Migratory birds are of particular concern as they might spread antibiotic resistance
282 across large geographic distances in the same way that they disperse pathogens (9, 11, 51,
283 52). There are significant differences in gut microbiomes of migratory and resident red-
284 necked stints (*Calidris ruficolis*) and curlew sandpipers (*Calidris furringea*), although these
285 differences may be temporary (53, 54). Ruddy turnstones have a remarkable migration
286 habit, travelling between breeding areas in Siberia to non-breeding sites in Australia via
287 East-Asia, potentially acquiring and distributing resistant bacteria along the way. The
288 turnstones analyzed here carried resistance against several antibiotic classes, but the

289 diversity of genes within those classes was much smaller than in birds at the WWTP (Fig. 1).
290 *Anas* ducks travel hundreds of kilometers within Australia (55). It is possible that ducks from
291 the Innamincka reserve have been in sites of high human impact previously, resulting in the
292 higher load of resistance genes when compared with other birds from remote areas. It is
293 also plausible that ducks acquire resistant bacteria due to their feeding behavior and the
294 composition of their gut microbiome.

295 Despite their isolation, we found genes conferring resistance against four antibiotic
296 classes in penguins from Antarctica. Previous studies of antibiotic resistance in penguins
297 have produced contradictory results. In one, various tetracycline resistant bacteria were
298 isolated from the cloaca of penguins (56), while in another high levels of resistance against
299 multiple antibiotics were detected in penguin droppings (57). However, other studies have
300 reported that antibiotic resistant bacteria are rare in these animals (24, 58, 59). It is possible
301 that penguins acquire resistance genes from migratory fish and other prey or animals with
302 which they interact. As antibiotics are naturally produced by bacteria, it is also possible that
303 the resistance genes observed in the penguin microbiome occur in the environment
304 regardless of human influence. The possibility of some cross-library and/or environmental
305 contamination cannot be completely excluded. Nevertheless, the *bona fide* influence of
306 human activity is supported by the larger number of resistance genes adjacent to the more
307 populated O'Higgins Base compared with the much smaller González Videla Base.
308 Additionally, previous research shows higher antibiotic resistance levels near research
309 facilities compared to more pristine sites in Antarctica (24, 57). Increasing research
310 activities, tourism and limited sewage treatment (60) are therefore the most likely
311 explanation for the presence of antibiotic resistance in Antarctic penguins.

312 The bird microbiome expressed resistance against nine classes of antibiotics, even
313 though we putatively enriched libraries with resistant bacterial strains using only two classes
314 of antibiotics in the collection media (aminoglycoside and β -lactams, see Materials and
315 Methods). Acquired (horizontally transferred) resistance genes can be constitutively
316 expressed, in which case the presence of their transcripts is expected even without antibiotic
317 exposure. It is also possible that these resistance genes were acting against antibiotics
318 present in the environment and/or that these genes are co-transmitted with others that have
319 functions in addition to antibiotic resistance (e.g. metal resistance, 61). Variables related to
320 the ecology and geographic distribution of the different bird species could also play a role,
321 although the results based on individual collection sites and bird taxonomic group show that
322 these variables are unlikely to change the conclusion that birds from the WWTP carry the
323 highest diversity and abundance of resistance genes. Meta-transcriptomic studies
324 necessarily rely on reference databases, which limits the discovery of novel resistance
325 genes (30), and the database used here (ResFinder, 62) does not include resistance that

326 arises through *de novo* mutation in the bacterial genome (which would increase the
327 detection of false positives). Therefore, although we were limited to assessing acquired
328 resistance genes, these genes residing on mobile elements pose greater public health risk
329 as they can be transferred easily between bacteria (63). Potentially unequal RNA yields
330 across libraries represent an additional caveat. We observed no correlation between library
331 size or microbial mRNA reads and the diversity or abundance of resistance genes, indicating
332 that the higher number of resistance genes observed in the WWTP does not result from
333 unequal sequencing effort, or from failing to extract and sequence microbes from other sites
334 (Fig. S6). Considering our rather conservative analyses (see Materials and Methods), it is
335 possible that we underestimate the presence of some resistance genes that were not
336 expressed or were expressed at low abundance.

337 In sum, we show that ducks feeding on wastewater are particularly prone to harbor
338 bacteria with transcriptionally active antibiotic resistance genes. Ecological and functional
339 traits are likely intertwined in explaining the higher propensity of ducks to carry antibiotic
340 resistance genes. This study also contributes to the increasing literature reporting
341 widespread antibiotic resistance in birds, even in isolated areas like the Australian outback
342 and Antarctica. For antibiotic resistant bacteria, aquatic systems are major traffic routes
343 between wildlife and humans (8, 64). The resistance genes acquired by birds can be re-
344 introduced in the environment, possibly in other water systems (e.g. by migrating ducks) and
345 might re-infect humans directly via contact with contaminated water, or indirectly by the
346 introduction of these genes into the food chain (64). Investigating the mechanisms that
347 sustain the persistence and cycling of resistance genes in wild populations despite the
348 metabolic burden that these gene impose is a logical next step towards tackling antibiotic
349 resistance.

350

351 **Materials and Methods**

352

353 *Sampling*

354 Samples were collected as part of long-term avian influenza virus surveillance studies (65-
355 70). Ethics approvals, bird capture methods and sample handling are reported in the
356 Supplementary Materials. In short, cloacal and oropharyngeal swabs were collected using a
357 sterile-tipped applicator and placed in viral transport media (VTM, Brain-heart infusion broth
358 containing 2x10⁶ IU/l penicillin, 0.2mg/ml streptomycin, 0.5mg/ml gentamicin, 500U/ml
359 amphotericin B, Sigma). VTM is a standard buffer used in avian influenza surveys and has
360 the advantage of killing a portion of non-resistant bacterial strains. This step enriches meta-
361 transcriptomes libraries with antibiotic resistant bacteria and, consequently, increases the
362 sensitivity of the antibiotic resistance survey. As all samples were stored in VTM, there is no

363 reason to believe that this step would affect the abundance comparisons among libraries,
364 and thus, it is unlikely to bias the results. All birds in this study were apparently healthy, with
365 the exception of one library constructed from dead and dying shelducks with symptoms of
366 Newcastle Disease. Samples were assayed for avian influenza virus as previously described
367 (66). Samples were collected at sites with different levels of anthropogenic impact
368 (Supplementary Materials). Birds sampled at the WWTP were found in lagoons composed of
369 partially treated water (the final stage of wastewater treatment).

370

371 *RNA-sequencing and data processing*

372 RNA isolation procedures are detailed in the Supplementary Materials. Libraries were
373 composed of 10 conspecific bird samples pooled at equal concentrations. Paired-end
374 sequencing (100bp) was performed on a HiSeq2500 platform and the number of reads
375 obtained are reported in Table S1. Low quality reads, adapters, host reads and ribosomal
376 RNA were filtered out from the data set (Supplementary Materials).

377

378 *Resistance genes characterization*

379 The ResFinder reference database (62) was used in conjunction with the KMA program (71)
380 (downloaded in December 2017) to identify resistance genes in the meta-transcriptomic data
381 set. The ResFinder database currently contains 2255 resistance genes compiled from
382 published manuscripts and existing databases. KMA was preferred over other alignment
383 tools because it performs well in aligning short reads against highly redundant databases
384 and is able to resolve non-unique read matches by assessing and statistically testing global
385 alignment scores. To minimize the risk of false-positives and increase the minimum mapping
386 length allowed, only genes with a mapping coverage greater than 20% were considered in
387 the analyses, all of which had an alignment p -value $\ll 0.05$. The average length of the
388 resistance genes observed was 944bp – a gene with this length was only considered in the
389 downstream analyses if query reads overlapped by at least 189 bp (20% coverage). This
390 approach is highly conservative because it uses an aligner that yields a minimal number of
391 false positives (71), does not include housekeeping genes (which would increase the
392 occurrence of false positives), and defines resistance genes based on gene fragments (at
393 least 20% of the genes) rather than individual reads. The gene fragments analyzed here are
394 longer than the ones obtained via qPCR (generally 100bp amplicons), which are widely used
395 in AMR assessments of environmental samples and in diagnostic laboratories. One gene
396 (*bla*_{TEM-116}) was observed in all libraries but was removed from the analyses due to its
397 potential contaminant nature (72). It is possible that the data set contains other laboratory
398 contaminants, but the fact that one of the libraries contained only one resistance gene, and
399 that no other gene (except for *bla*_{TEM-116}) was found in all libraries, suggests that

400 contamination is unlikely. Genes conferring resistance to Macrolide, Lincosamide and
401 Streptogramin B were considered as one antibiotic class (MLS). Absolute read abundances
402 were estimated based on a stably expressed host gene and normalized for gene length
403 (Supplementary Materials). The number of antibiotic classes to which resistance was found,
404 the diversity (i.e. number of genes) and the abundance of resistance genes in each library
405 were classified into two bins ('WWTP' and 'Other', Fig. 2). Differences between WWTP and
406 other sites were tested with a Kruskal-Wallis test using the native *stats* R package (R Core
407 Team 73). The higher diversity of resistance genes in libraries from the WWTP was
408 validated with a PCR-based approach targeting resistance genes in individual birds from two
409 libraries (n=20, Supplementary Materials).

410

411 *Functional profiling*

412 The microorganism-based functional profile was inferred with HUMAnN2 (74)
413 (<http://huttenhower.sph.harvard.edu/humann2>), using the UniRef90 protein database as
414 reference (75). Community-level differences in expression of pathways between sites and
415 bird orders was visually assessed with Principal Coordinate Analysis based on an Euclidean
416 distance matrix with the *ape* R package (76) and further investigated with Random Forest
417 analysis, using 1000 trees, with the *randomForest* R package (77).

418

419 **Acknowledgments**

420 The sampling was supported by NIAID (HHSN266200700010C), ARC discovery grants (DP
421 130101935 and DP160102146) and the Instituto Chileno Antártico INACH (T 12-13 and T
422 27-10). The Melbourne WHO Collaborating Centre for Reference and Research on Influenza
423 is supported by the Australian Department of Health. ECH is funded by an ARC Australian
424 Laureate Fellowship (FL170100022). TCS is a Sydney Medical Foundation Fellow whose
425 work is also supported by the Sydney Medical School Foundation. JRI is supported by a
426 NHMRC Practitioner Fellowship (GNT1104232). We thank the High Performance Computing
427 team at Sydney University, the genomic facilities at the Westmead Institute for Medical
428 Research, the Centre for Integrative Ecology at Deakin University, the Victorian Wader
429 Study Group, and the logistic support from Melbourne Water, Innamincka Station and
430 Innamincka Regional Reserve. We also thank Sebastiaan van Hal, Sally Partridge,
431 Ali Khalid, Philip Clausen, Simeon Lisovski and Marta Ferenczi for their support.

432

433 **Authors' contributions**

434 VRM, MW and ECH designed the research. MW, DG-A and MK collected samples. MW
435 carried out DNA/RNA isolation and meta-transcriptome library preparation. VRM and J-SE
436 performed PCRs. VRM performed the data analyses. ACH, DG-A, MK, TCS and ECH

437 contributed with reagents and/or funds for research. All authors contributed to interpreting
438 data and manuscript writing. All authors gave final approval for publication.

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441 **References**

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