Habitat fragmentation compromises the population dynamic of the globally near-

threatened Straight-billed Reedhaunter (*Limnoctites rectirostris*)

Maycon S. S. Gonçalves[1,2], Priscila S. Pons[1,2], Felipe C. Bonow[2], Vinicius A.

Bastazini[3], José A. Gil-Delgado[1], Germán M. López-Iborra[4]

[1]Instituto Pró-Pampa (IPPampa), Laboratório de Ornitologia, Pelotas, Rio Grande do

Sul, Brazil

[2]Instituto Cavanilles de Biodiversidad y Biología Evolutiva, Universidad de Valencia,

Paterna, Valencia, Spain

[3]Centre for Biodiversity Theory and Modelling, Station d'Ecologie Théorique et

Expérimentale du CNRS, Moulis, France

[4]Departamento de Ecología - IMEM Ramon Margalef, Universidad de Alicante,

Alicante, Spain

Corresponding author: mayconsanyvan@gmail.com

Abstract

Understanding the consequences of habitat fragmentation to organism populations is

crucial to develop sound conservation polices. The Straight-billed Reedhaunter

(*Limnoctites rectirostris*) is a little known and threatened Passeriform that is highly

dependent Erygo wetlands patches. Here, we evaluated the effects of habitat

fragmentation on populations of the Straight-billed Reedhaunter, during the

construction of a water reservoir in southern Brazil. During eight months, we monitored

five Eryngo wetlands patches occupied (n=3) and no occupied (n=2) by Straight-billed

Reedhaunter individuals, collecting data on their temporal occupancy patterns and

registering new fragmentation events in formally continuous habitat patches. We

employed an information-theoretic approach in order to evaluate the consequences of

habitat fragmentation on the probabilities of patch occupancy, colonization and

extinction of populations of the Straight-billed Reedhaunter. Out of the three patches

occupied by Straight-billed Reedhaunter, two were not altered by construction activities

and their populations were present during the entire study period. After fragmentation

events, local extinction in one of the wetland patches was observed, and individuals

sporadically observed in two other initially unoccupied sites. The model in which

fragmentation affected only the extinction probability was the most plausible among the

set of candidate models. Fragmentation greatly increased the chance of local population

within patches. Our results indicate that the conservation of populations of the Straight-

billed Reedhaunter is highly dependent on continuous and unaltered wetland patches.

Keywords: inland wetlands, Eryngo, waterbird, habitat fragmentation, conservation,

reservoir, perturbation, extinction, colonization

Introduction

Continental wetlands are among the most threatened ecological systems in the

World (Davidson 2014). For South America, the habitat fragmentation of some regions

has reduced the surface of this ecosystem to less than 10% of its original area (Maltchik

et al. 2003; Guadagnin et al. 2005). One of the most interesting and little-known

continental aquatic systems in South America refers to those dominated by Eryngo *Eryngium pandanifolium* Cham. & Schltdl. (Apiaceae), popularly known as *gravatazais*. These wetlands are distributed in the form of patches in the grassland landscape and are periodically altered or destructed by many anthropic disturbances, such as agriculture, intensive livestock, intentional fires and drainage for dam construction (Irgang 1999; López-Lanús et al. 1999; Bencke et al. 2003; Volcan et al. 2014).

Collected for the first time in June 1833 by Charles Darwin (Steinheimer 2004), the Straight-billed Reedhaunter (*Limnoctites rectirostris*) is a South American aquatic passerine that strongly depends on the wetlands dominated by Eryngo (BirdLife International 2016). The Straight-billed Reedhaunter has no known migratory movements and its geographical distribution is limited to south Brazil (Fontana et al. 2008; Bencke et al. 2010), and to neighboring countries – Uruguay (Aldabe et al. 2009) and Argentina (Chebez et al 2011). Globally, this species is included within the "near-threatened" category and displays a decreasing population trend (BirdLife International, 2016). In national scale, although the Straight-billed Reedhaunter has been recently excluded from the list of threatened species of Brazil (MMA 2014), the scarce knowledge about its population size and life history, along with its reduced distribution and high habitat specificity has led to the inclusion in threatened species lists in Uruguay and Argentina, under "near-threatened" (Aldabe et al. 2009) and "threatened" (Chebez et al. 2011) categories, respectively.

Gravatazais present high biological, taxonomic and functional diversity, including the presence of many other threatened species, such as annual fishes, amphibians, birds and mammals (Fontana et al. 2003, Teixeira de Mello et al. 2011; Lanés et al. 2014). However, in southernmost Brazil, regions with high concentration of

this type of wetland are also characterized by low annual precipitation, and the

constructions of dams have been very important for the development of the cities and

local communities (Boschi et al. 2011). This is of particular concern as dams

significantly affect the population dynamic of terrestrial species (Kingsford 2000).

In this work, we aimed to identify the effects of habitat fragmentation on the

population dynamics of Straight-billed Reedhaunter. Specifically, we determined the

probabilities of occupation, colonization, and extinction associated with the habitat

changes promoted by the construction of a water reservoir. For this purpose,

gravatazais with and without Straight-billed Reedhaunter subpopulations were

monitored for the presence and absence of the species during the reservoir construction

activities. Given the close relationship of Straight-billed Reedhaunter to wetlands

dominated by Eryngo, we expected to find significant effects of habitat fragmentation

on the persistence of their subpopulations, which would reflect on the extinction and

colonization probabilities.

Methods

Study area

The study was conducted in the municipality of Bagé in southernmost Brazil (S

31º 17' 26,4' and W 54º 09' 32,7'). This region greatly represents the landscape of the

Pampa Biome (Rambo 1959; Overbeck et al. 2007). The study area is characterized by

large extensions of natural grassland, marked by the constant presence of livestock

(main economic activity in the region) and agricultural crops, such as corn, soy,

sorghum and exotic forest (IBGE 2013). Specifically, the place designated for the

reservoir construction covers an area of approximately 340 hectares (Figure 1). Inside

the construction area, an extensive riparian forest surrounds the most important

watercourse of the locality, called "Arvorezinha". Grasslands and drainage lines with

grass and low shrubs are the dominant cover inside the area, as well as patches of

wetlands dominated by sedge (*Eryngium pandanifolium*), reed (*Cyperus californicus*),

cattail (*Typha dominguensis*) and panic grass (*Panicum prionitis*).

Patch selection and population monitoring

Firstly, from satellite imagery, we identified all potential wetlands within the

study area. During a pilot inspection in December 2011, we excluded all sites without

Eryngo habitat. Five Eryngo wetlands patches were selected for research (Table 1). The

size of wetlands varied between 5444 m^2 and 34150 m^2. Of these, three wetlands were

confirmed with the presence of two individuals of Straight-billed Reedhaunter during

the pilot inspection. Posteriorly, between September 2012 and April 2013 (except for

November), we made seven monthly visits to the five selected wetlands. In this sense,

including the first observation performed during the pilot inspection, eight samplings

were performed for each wetland. Nests were recorded in three patches initially

occupied by the Straight-billed Reedhaunter – see Gonçalves et al. (2017) for more

information about breeding biology.

During each visit, we aimed to detect the presence/absence of Straight-billed

Reedhaunter and whether wetlands were fragmented. We considered fragmentation to

be any change in the natural structure of a wetland that implied a new a split of the

continuous patch. All samplings were made when weather conditions were favorable

(without rain and little wind). To facilitate the detection of Straight-billed Reedhaunte,

observations were always made by two researchers. "Playback" techniques were also

run on the edge of patches and the search time in each site varied between 1 and 2

hours.

Occupancy models

The presence and absence of Straight-billed Reedhaunter was correlated with the

habitat integrity (occurrence or not of one event of fragmentation or destruction of the

wetlands). Four parameters were tested: probability of occupation (ψ); probability of

colonization (γ); probability of extinction (ε) and probability of detection (p) (this last

was kept constant). From these four parameters, three models were tested and the results

of each parameter were transformed in probabilities. The models tested were: 1)

$\psi(.)\gamma(.)\varepsilon(.)p(.)$ – model which the probability of occupation, colonization, extinction and

detection are constants (.); 2) $\psi(.)\gamma(fragmentation)\varepsilon(fragmentation)p(.)$ – model which

the fragmentation affects the colonization and extinction; 3) $\psi(.)\gamma(fragmentation)\varepsilon(.)p(.)$

– model which the fragmentation affects only the probability of colonization; and 4)

$\psi(.)\gamma(.)\varepsilon(fragmentation)p(.)$ – model which the fragmentation affects only the

extinction. Analyzes were conducted with the Colext function in Unmarked Package

(Fisk and Chandler 2015), carried in R-Program v3.1.3 (R Development Core Team

2014). The Unmarked Package allows create and select the best models from numerous

criteria and, especially, allows estimate the occupation of rare species, such as the

Straight-billed Reedhaunter. The selection of models was based on the weight of the

Akaike Information Criterion (AIC – ω), which may also be interpreted as a probability.

Thus, the more plausible model was that defined by major AIC value (ω).

Results

The spatial-temporal dynamic of the Straight-billed Reedhaunter is presented in

Figure 2. Of these three wetlands, two were not altered by the reservoir construction and

the presence of individuals was constant throughout the study period. One of the

occupied patches (number 3, Figure 2) was partially destroyed during the breeding

period (Figure 3). During the next sampling periods, the species was absent in this

patch, and individuals were concomitantly observed in two other initially unoccupied

wetlands, including one already fragmented site (Fig. 2a-2d).

The model whose the fragmentation affected only the extinction showed the

highest plausibility, with 61% probability of extinction to be related to the habitat

alteration (Tab. 1). Model in which fragmentation affects both the colonization and

extinction was also highly plausible (ΔAIC < 2), although its probability was much

lower than the model in which the probability of extinction is a function of habitat

alteration (Tab. 1). Extinction probability of Straight-billed Reedhaunter tended to

increase drastically in fragmented wetlands (Tab. 2; Fig. 4).

Discussion

Our results highlight that the habitat fragmentation effect on Straight-billed

Reedhaunter populations is significant. Fragmentation increases the extinction

probability of the Straight-billed Reedhaunter and reduces, at least partially, its

colonization probability.

Habitat fragmentation results primarily in the local extinction of populations and

later in the species' regional and global extinction (Henle et al. 2004). Although species

are able to occupy fragmented landscapes when their daily life cycles include multiple

fragments (Redpath 1995), the fragmentation effects is stronger in those species with

specific ecological requirements (Swihart et al. 2003; DeVictor et al. 2008). This seems

to be the case of the Straight-billed Reedhaunter. After local extinction in one of the

wetlands, individuals were sporadically observed in two other initially unoccupied

patches, including a fragmented site. Although we cannot confirm that these individuals

are the same as those in the previously destroyed wetlands, the temporary occurrence of

the Straight-billed Reedhaunter in these sites could indicate an immediate attempt to

extend its territory, as a result of the habitat loss and displacement of competitive

individuals.

The size and stability of the populations in natural wetlands depend strongly on a

range of habitat and landscape factors, as well as on body size, morphology, behavior,

effects of niche breadth and the effect of geographic range boundaries (Redpath, 1995;

Marsh and Trenham 2000; Swihart et al. 2003). Stable subpopulations were observed

only in unaltered wetlands, which indicate that the species' ecological plasticity is

apparently low. Paradoxically, the species has been known to occupy wetlands in

widely altered landscapes, e.g., on the edge of roads and dams (Ricci and Ricci 1984;

Barbarskas and Fraga 1998). We recommend taking these observations cautiously as the

populations may have colonized these wetlands after the alteration of the landscape,

which favors the colonization of Eryngo due to artificial water concentrations by

construction of roads and dams. Hence further research is needed to elucidate the effects

of the area and habitat structure on the size of the subpopulations located in unchanged

natural patches.

Our results show the Straight-billed Reedhaunter populations has greater

stability in unaltered patches, which reduces the chances of permanence and

colonization in those patches subjected to fragmentation events. These results

demonstrate a more accurate view of this species' ecological plasticity and their

tolerance to habitat fragmentation, and contribute to lower the uncertainties of its degree

of threat at different scales. The Straight-billed Reedhaunter is globally near-threatened

(BirdLife International, 2016), but has been recently excluded from the list of

endangered species in Brazil (MMA, 2014) and Rio Grande do Sul State (Rio Grande

do Sul, 2014). Indeed, the number of records of this species has significantly increased

in recent years, especially in southernmost Brazil (Develey et al. 2008; ICMBIO, 2014;

Wikiaves, 2016). On other hand, it is important to note that the knowledge about many

aspects of its biology and ecology is completely lacking. In this work, we discuss the

fragmentation effect promoted by an activity that tends to drastically alter landscapes.

Additionally, these wetlands are commonly channeled for water subtraction – a practice

that also fragments wetlands and may have similar consequences to the disturbance

explored herein. Therefore, we strongly recommend understanding the relationship of

both the habitat structure and degree of connectivity of patches with the Straight-billed

Reedhaunter's spatial distribution by identifying how environmental and spatial

stochastic processes on different landscape scales influence the species' demographic

dynamics. Finally, the natural distribution of the *gravatazais* may impose the Straight-

billed Reedhaunter to be distributed as a metapopulation, and we end this article by

encouraging further research about not only the Straight-billed Reedhaunter, but also

about the vast biodiversity associated with wetlands dominated by Eryngo.

Acknowledgements

We are grateful to the Ecossis Soluções Ambientais for support while undertaking

fieldwork.

References

Aldabe, J., Rocca, P., Claramunt, S. 2009. *Uruguay*. Pág. 383 – 392 *in* Devenish, C.;
Díaz Fernández, D. F.; Clay, R. P.; Davidson, I. & Zabala, Y. I. (eds.). Important Bird
Areas Americas - Priority sites for biodiversity conservation. Quito, Ecuador: BirdLife
International.

Barbarskas, M. & Fraga, R. 1998. Actualizando la distribución de la Pajonalera Pico
Recto Limnoctites rectirostris en la provincia de Entre Ríos. Cotinga 10:79-81.

Bencke, G. A., Fontana, C. S., Dias, R. A., Maurício, G. N., Mahler Jr., J. K. (2003).
Aves. *In* Fontana CS, Bencke GA and Reis RE (ed) Livro vermelho da fauna ameaçada
de extinção no Rio Grande do Sul. Porto Alegre, Brasil (in Portuguese).

Bencke, G. A.; Dias, R. A.; Bugoni, L.; Agne, C. E.; Fontana, C. S.; Maurício, G. N. &
Machado, D. M. 2010. Revisão e atualização da lista das aves do Rio Grande do Sul,
Brasil. Iheringia, Sér. Zool. 100(4):519-556.

BirdLife International. 2016. *Limnoctites rectirostris*. The IUCN Red List of Threatened
Species 2016: e.T22702652A93885012. http://dx.doi.org/10.2305/IUCN.UK.2016-
3.RLTS.T22702652A93885012.en. Downloaded on 25 February 2018.

Boschi, R. A.; Oliveira, S. R. M. & Assad, E. D. 2011. Técnicas de mineração de dados
para análise da precipitação pluvial decenal do Rio Grande do Sul. Eng. Agríc.
Jaboticabal 31(6):1189-1201

Chebez, J. C.; Gasparri, B., Hansen Cier, M., Nigro, N. A., Rodríguez, L. 2011. Estado
de conservación de los tetrápodos de la Argentina. In: Porini, G. & Ramadori, D. (eds.).
Manejo de Fauna Silvestre en Argentina. Conservación de especies amenazadas.
Fundación de Historia Natural "Félix de Azara".

Davidson, N. C. (2014). How much wetlands has the world lost? Long-term and recent
trends in global wetland area. *Marine and Freshwater Research*, 65, 934–941.

Develey, P. F., Setubal, R. B., Dias, R. A., & Bencke, G. A. (2008). Conservação das
aves e da biodiversidade no bioma Pampa aliada a sistemas de produção animal. *Revista*
Brasileira de Ornitologia, *16*(4), 308-315.

Devictor, V., Julliard, R., & Jiguet, F. (2008). Distribution of specialist and generalist
species along spatial gradients of habitat disturbance and fragmentation. *Oikos*, *117*(4),
507-514.

Fontana, C. S., Rovedder, C. E., Repenning, M., Gonçalves, M. L. 2008. Estado atual do conhecimento e conservação da avifauna dos Campos de Cima da Serra do sul do Brasil, Rio Grande do Sul e Santa Catarina. Revista Brasileira de Ornitologia, 16(4):281-307.

Guadagnin, D. L., Peter, A. S., Perello, L. F. C., & Maltchik, L. (2005). Spatial and temporal patterns of waterbird assemblages in fragmented wetlands of Southern Brazil. *Waterbirds*, 28, 261–272.

IBGE, Instituto Brasileiro de Geografia e Estatística. (2013). Bagé – RS. Disponível em: <http://www.ibge.gov.br> Acesso em: 05.02.2015.

Kingsford, R. T. (2000). Ecological impacts of dams, water diversions and river management on floodplain wetlands in Australia. *Austral Ecology*, *25*(2), 109-127.

ICMBIO – Instituto Chico Mendes de Conservação da Biodiversidade. 2014. Plano de Ação Nacional para a Conservação dos Passeriformes Ameaçados dos Campos Sulinos e Espinilho.

Irgang, B. E. 1999. Comunidade de macrófitas aquáticas da Planície Costeira do Rio Grande do Sul – Brasil: um sistema de classificação. Universidade Federal do Rio Grande do Sul (in Portuguese).

Lanés, L. E. K., Gonçalves, A. C., Volcan, M. V. (2014). Discovery of endangered annual killifish Austrolebias cheradophilus (Aplocheiloidei: Rivulidae) in Brazil, with comments on habitat, population structure and conservation status. Neotropical Ichthyology, 12(1): 117-124

López-Lanús, B.; Di Giacomo, A. G.; Babarskas, M. 1999. Estudios sobre ecología y comportamiento de la Pajonalera Pico Recto *Limnoctites rectirostris* en la Reserva Otamendi, Buenos Aires, Argentina. Cotinga 12:61-63 (in Spanish).

Maltchik, L., Rolon, A. S., Guadagnin, D. L., Sternet, C. 2003. Wetlands of Rio Grande do Sul, Brazil: a classification with emphasis on plant communities. Acta Limnol. Bras. 16(2):137-151.

MacKenzie, D. I., Nichols, J. D., Lachman, G. B., Droege, S., Andrew Royle, J., & Langtimm, C. A. (2002). Estimating site occupancy rates when detection probabilities are less than one. *Ecology*, *83*(8), 2248-2255.

MacKenzie, D.I., Nichols, J.D., Hines, J.E., Knutson, M.G., and Franklin, A.B. 2003. Estimating site occupancy, colonization, and local extinction when a species is detected imperfectly. Ecology 84: 2200–2207.

Mackenzie, D. I. (2006). Modeling the probability of resource use: the effect of, and dealing with, detecting a species imperfectly. *Journal of Wildlife Management*, *70*(2), 367-374.

Marsh, D. M. & Trenham, P. 2000. Metapopulation Dynamics and Anphibian Conservation. Conservation Biology 15:40-49.

MMA – Ministério do Meio Ambiente. 2014. Lista Nacional Oficial de Espécies da
Fauna Ameaçada de Extinção. PORTARIA Nº 444, DE 17 DE DEZEMBRO DE 2014.
Brasil.

Overbeck, G. E., Müller, S. C., Fidelis, A., Pfadenhauer, J., Pillar, V. D., Blanco, C. C.,
... & Forneck, E. D. (2007). Brazil's neglected biome: the South Brazilian
Campos. *Perspectives in Plant Ecology, Evolution and Systematics*, *9*(2), 101-116.

Rambo, B. (1956). A fisionomia do Rio Grande do Sul. Porto Alegre: Livraria Selbach.

Redpath, S. M. (1995). Habitat fragmentation and the individual: tawny owls Strix
aluco in woodland patches. *Journal of Animal Ecology*, 652-661.

Teixeira de Mello, F., González-Bergonzoni, I. & Loureiro, M. (2011). Freshwater
Fishes of Uruguay. PPR-MGAP.

Volcan, M. V., Goncalves, A. C., & Lanés, L. E. K. (2014). Austrolebias quirogai
(Actinopterygii: Cyprinodontiformes: Rivulidae) in Brazil: occurrence, population
parameters, habitat characteristics, and conservation status. *Acta Ichthyologica et*
piscatoria, *44*(1), 37.

Ricci, J. J. and Ricci, F. (1984). Nidificación de la Pajonalera de Pico Recto Limnornis
rectirostris en Benavídez, Buenos Aires, Argentina. Honero 12:205-208.

Rio Grande do Sul. (2014). DECRETO N.º 51.797, DE 8 DE SETEMBRO DE 2014.
(http://www.al.rs.gov.br/filerepository/repLegis/arquivos/DEC%2051.797.pdf).

R Development Core Team (2016). R: a language and environment for statistical
computing. R Foundation for Statistical Computing, Vienna.

Steinheimer, F. D. 2004. Charles Darwin's bird collection and ornithological knowledge
during the voyage of H.M.S. "Beagle", 1931-1836. Journal of Ornithology, 145:300-
320.

Swihart, R. K., Gehring, T. M., Kolozsvary, M. B., & Nupp, T. E. (2003). Responses of
'resistant'vertebrates to habitat loss and fragmentation: the importance of niche breadth
and range boundaries. *Diversity and Distributions*, *9*(1), 1-18.

Wikiaves. 2016. Limnoctites rectirostris. Wikiaves – A enciclopédia de aves do Brasil.

Disponível em: www.wikiaves.com.br/arredio-do-gravata. Acesso em: 25.03.2016.

315
316
317

318 **Tables**

319 **Table 1**. Summary of model selection procedure for the Straight-billed Reedhaunter. The models are ordered from highest to lowest AIC values
320 (ω).

Model	Parameters	AIC	ΔAIC	ω
ψ(.)γ(.)ε(Fragmentation)p(.)	5	40.11	0	0.61
ψ (.)γ(Fragmentation)ε(Fragmentation)p(.)	6	42.07	1.96	0.23
ψ(.)γ(.)ε(.)p(.)	4	43.38	3.27	0.12
ψ (.)γ(Fragmentation)ε(.)p(.)	5	45.35	5.24	0.04

328   **Table 2.** Probabilities estimates (with 1 standard error in parentheses) for the Straight-billed Reedhaunter for the best model after the

329   fragmentation event. Subscript for extinction probabilities denote non-fragmented (ε_{nf}) and fragmented patches (ε_f).

Best Model	ψ	γ	ε_{nf}	ε_f	p
$\psi(.)\gamma()\varepsilon(\text{Fragmentation})p(.)$	0.60 (0.22)	0.12 (0.08)	0.06 (0.06)	0.67 (0.27)	1 (0)

330   E21

Fig. 1. Study area located in southernmost Brazil. The continuous line demarcates the approximated flooding area of the reservoir. Selected patches are numbered and marked. Image dates to April 2007 taken from Google Earth Pro (accessed on 12 June 2016).

Fig. 2. Schematic representation of the Straight-billed Reedhaunter's occupation

dynamics. A dotted line in Figure 2a marks the area affected by the construction

activities. The areas with the presence and absence of species are filled in gray and

white, respectively. The patches with gridlines represent the occurrence of a

fragmentation event. The dates of the samples were: 2a) December 2011; 2b) September

2012; 2c) October 2012; 2d) December 2012; 2e) January 2013; 2f) February 2013; 2g)

March 2013; and 2h) April 2013.

Fig. 3a-3c. 3a, Straight-billed Reedhaunter perched on *Eryngium pandanifolium*; 3b, Nest built on the stems of "Eryngo"; 3c, wetland partially destroyed by the reservoir construction activities. Photos: 3a, Christian Andretti; 3b and 3c, Priscila Pons.

355
356 **Fig. 4.** Extinction probability (95% Confidence Interval) of Straight-billed Reedhaunter
357 estimated for each sampling month in areas with and without fragmentation. Red:
358 fragmented patches; Grey: non-fragmented patches.

360