

1 **Extent of the annual Gulf of Mexico hypoxic zone influences**
2 **microbial community structure**

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24 **Running Title:**

25 Microbes in 2014 nGOM hypoxic zone

26

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29 ammonia-oxidizing archaea (AOA)

30 **Abstract**

31 Rich geochemical datasets generated over the past 30 years have provided fine-scale resolution
32 on the northern Gulf of Mexico (nGOM) coastal hypoxic (≤ 2 mg of O₂ L⁻¹) zone. In contrast,
33 little is known about microbial community structure and activity in the hypoxic zone despite the
34 implication that microbial respiration is responsible for forming low dissolved oxygen (DO)
35 conditions. Here, we hypothesized that the extent of the hypoxic zone is a driver in determining
36 microbial community structure, and in particular, the abundance of ammonia-oxidizing archaea
37 (AOA). Samples collected across the shelf for two consecutive hypoxic seasons in July 2013 and
38 2014 were analyzed using 16S rRNA gene sequencing, oligotyping, microbial co-occurrence
39 analysis and quantification of thaumarchaeal 16S rRNA and archaeal ammonia-monooxygenase
40 (*amoA*) genes. In 2014 Thaumarchaeota were enriched and inversely correlated with DO while
41 Cyanobacteria, Acidimicrobiia and Proteobacteria were more abundant in oxic samples
42 compared to hypoxic. Oligotyping analysis of *Nitrosopumilus* 16S rRNA gene sequences
43 revealed that one oligotype was significantly inversely correlated with dissolved oxygen (DO) in
44 both years and that low DO concentrations, and the high Thaumarchaeota abundances,
45 influenced microbial co-occurrence patterns. Taken together, the data demonstrated that the
46 extent of hypoxic conditions could potentially influence patterns in microbial community
47 structure, with two years of data revealing that the annual nGOM hypoxic zone is emerging as a
48 low DO adapted AOA hotspot.

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53 **Introduction**

54 Deoxygenation of the ocean is one of the primary consequences of global climate change [1,2],
55 with much attention directed towards coastal hypoxic zones (dissolved oxygen (DO)
56 concentrations below 2 mg L⁻¹ or 62.5 μmol/kg). Hypoxic zones are frequently referred to as
57 “dead zones” because they are inhospitable to macrofauna and megafauna; however,
58 microorganisms thrive in such environments [3]. Eutrophication-associated dead zones have
59 been reported in over 500 locations spanning the globe [4] and are predicted to increase in
60 number and size in the near future as a result of increasing greenhouse gas emissions [2]. The
61 northern Gulf of Mexico (nGOM) is the site of the second largest eutrophication-associated
62 coastal dead zone in the world, with bottom water hypoxia extending to over 20,000 km² [5] and
63 covering anywhere from 20% to 50% of the water column during the summer months [6]. The
64 nGOM hypoxic zone is influenced by the freshwater input and nutrient load from the Mississippi
65 (MI) and Atchafalaya (AR) Rivers [7], which results in a phytoplankton bloom, the biomass of
66 which is subsequently respired by aerobic microorganisms leading to low DO concentrations or
67 hypoxic zones [8,9]. Hypoxia in the nGOM has increased in severity during the summer months
68 in direct response to additional inorganic nitrogen loading in the Mississippi watershed,
69 beginning around the 1950s, after which the nitrate flux to the nGOM continental shelf tripled
70 [10–12]. Therefore, thermal warming, nutrient rich freshwater discharge and microbial
71 respiration culminate and result in an annually extensive nGOM hypoxic zone, that reached a
72 record size at 8,776 mi² (22,720 km²) in 2017 (LUMCON 2017,
73 <https://gulfhypoxia.net/research/shelfwide-cruise/?y=2017>, accessed 1/22/2018).

74

75 While the sequence of events that leads to the nGOM hypoxic zone are well documented [4,13–
76 16], efforts to understand how these chemical, biological, and physical factors influence
77 microbial community structure, abundances, and activity across environmental gradients in the
78 nGOM coastal shelf have only recently been undertaken [17–21]. King and colleagues [17]
79 reported that before the onset of hypoxia (March), Alpha- and Gamma- Proteobacteria,
80 Bacteroidetes and Actinobacteria are abundant in nGOM waters <100m, with Planctomycetes
81 and Verrucomicrobia being less abundant. Gillies and colleagues [20] sampled during the 2013
82 hypoxic event in late July, when hypoxic conditions historically and predictably prevail, and
83 reported that the high normalized abundance of a Thaumarchaeota (100% similar to the
84 ammonia-oxidizing archaea *Nitrosopumilus maritimus* [22]) Operational Taxonomic Unit (OTU)
85 (16S rRNA gene iTag data) and the absolute abundance of Thaumarchaeota 16S rRNA and
86 archaeal *amoA* gene copies (quantitative PCR (qPCR)) were significantly inversely correlated
87 with DO [20]. In this 2013 study, Thaumarchaeota comprised more than 40% of the microbial
88 community (iTag sequence data) in hypoxic samples [20]. In contrast, King and colleagues [17]
89 and Tolar and colleagues [23] showed Thaumarchaeota were present at depths <100m increasing
90 in abundance at depths >100m in the nGOM when conditions are oxic over the shelf. Bristow
91 and colleagues [19] sampled a similar geographic area in 2012 during a historically small
92 hypoxic zone (smallest recorded since 1988) and reported that thaumarchaeal abundances
93 increased with depth; reaching a maximum at 120 m. In their shallow, bottom water (15 m)
94 sample, Thaumarchaeota reached ~15% of the microbial community and the highest rate of
95 ammonia oxidation were reported at this single hypoxic station ([19]; Station 6, see their Fig
96 3A). Further, genomic reconstruction of two *Nitrosopumilus* genomes (79% and 96% complete)
97 assembled from shotgun metagenomic data from a subset of the 2013 hypoxic zone samples

98 [20], combined with complimentary shotgun metatranscriptomic data, revealed that the
99 *Nitrosopumilus* reported in that study were active in the 2013 hypoxic zone (Campbell and
100 Mason, unpublished). Specifically, transcripts from *amoABC* genes, with the same synteny as
101 observed in the genome of *N. maritimus* [24] were identified. These data suggest that
102 nitrification, particularly ammonia oxidation, was actively being carried out in the 2013 hypoxic
103 zone by Thaumarchaeota-- an aerobic process that would continue to draw down oxygen,
104 exacerbating hypoxic conditions across the shallow continental shelf [25,26].

105

106 In the nGOM hypoxic zone, the importance of AOA is only beginning to emerge; however,
107 AOA have been shown to be more abundant than ammonia-oxidizing bacteria (AOB) in
108 terrestrial and marine ecosystems, [27–31], and it is proposed that AOA can outcompete bacteria
109 in low energy conditions [32]. In other oxygen minimum zones (OMZs), several studies have
110 reported an increase in abundance of Thaumarchaeota in low DO samples [18,33–37]. The
111 abundance of archaeal *amoA* transcripts also increases in OMZs [35,36,38].

112

113 Therefore, to begin to fill in the knowledge gaps on microbial community structure and
114 specifically on AOA abundances in the nGOM hypoxic zone, we describe the microbial
115 communities in samples collected in July 2014 within and outside of the 13,080 km² hypoxic
116 zone, and compared these to Y14 results to the same thirty-three sites sampled during the Y13
117 hypoxic zone (15,120 km²) [20]. We expanded sampling in Y14 to include surface water and
118 more bottom water oxygen minimum zone samples, as compared to Y13. Our goal was to
119 determine the influence the extent of the hypoxic zone has on the overall microbial ecology and
120 in particular the abundance of AOA across hypoxic seasons in the nGOM. Further, we sought to

121 determine if the high AOA abundances observed in low DO conditions in this large coastal
122 hypoxic zone occur predictably and, if so, consider the potential ecological implications for an
123 annual increase of AOA in the hypoxic zone.

124

125 **Material and Methods**

126 **Sample collection**

127 The 2014 hypoxic zone was mapped over seven days, from 27 July to 2 August 2014 and
128 measured 13,080 km². At each of the 52 sites sampled, a sample was collected at the surface
129 (except for sites A'2, B9 and C7) (1 meter below sea level (mbsl); samples designated S for
130 surface; 47 samples total) and near the seafloor at the oxygen minimum zone (19 mbsl avg.
131 collection depth; samples designated B for bottom; 50 samples total). Samples were collected
132 from the surface of the Mississippi River at two sites (0mbsl; designated R2 and R4 for
133 Mississippi River) for a total of ninety-nine samples. For each of the ninety-nine samples
134 collected, temperature, depth, salinity (conductivity) and *in situ* chemistry were determined using
135 a conductivity-temperature-depth (CTD) instrument (SeaBird SBE32 5L bottle carousel).
136 Concentrations of DO, ammonium (NH₄), nitrite (NO₂) + nitrate (NO₃), phosphate (PO₄) and
137 chlorophyll a were determined. The sample location map and subsequent plot of DO data were
138 made with Ocean Data View [39].

139 **Oxygen, chlorophyll a and nutrients**

140 Oxygen concentrations were determined *in situ* with the CTD dissolved oxygen sensor. Oxygen
141 concentrations were verified using Winkler titrations [40] shipboard. Chlorophyll a samples were
142 concentrated on 25-mm Whatman GF/F filters from 500 ml⁻¹ L seawater and stored at -20°C.

143 Chlorophyll a was extracted using the methods described in the Environmental Protection
144 Agency Method 445.0, “In Vivo Determination of Chlorophyll a in Marine and Freshwater
145 Algae by Fluorescence” [41]; however, no mechanical tissue grinder or HCl were used.
146 Chlorophyll a concentrations were determined using a fluorometer with a chlorophyll a standard
147 (*Anacystis nidulans* chlorophyll a). For nutrients, 60 ml of seawater was filtered through 0.22-
148 µm Sterivex filters into two 30-ml Nalgene bottles and stored at -20°C. Nutrient concentrations
149 were determined by the marine chemistry lab at the University of Washington following the
150 WOCE Hydrographic Program using a Technicon AAII system
151 (<http://www.ocean.washington.edu/story/Marine+Chemistry+Laboratory>).

152 **Microbial sampling and DNA extractions**

153 From each station, up to 5 L of seawater were collected and filtered with a peristaltic pump both
154 at the surface and at the oxygen minimum. A 2.7-µM Whatman GF/D pre-filter was used and
155 samples were concentrated on 0.22-µM Sterivex filters (EMD Millipore). Sterivex filters were
156 sparged, filled with RNAlater and frozen. Samples were transported to Florida State University
157 on dry ice and stored at -80 until DNA extractions and purifications were carried out. DNA was
158 extracted directly off of the filter by placing half of the Sterivex filter in a Lysing matrix E
159 (LME) glass/zirconia/silica beads Tube (MP Biomedicals, Santa Ana, CA, USA) using the
160 protocol described in [20] which combines phenol:chloroform:isoamylalcohol (25:24:1) and bead
161 beating. Genomic DNA was purified using a QIAGEN (Valencia, CA, USA) AllPrep DNA/RNA
162 Kit and quantified using a Qubit2.0 Fluorometer (Life Technologies, Grand Island, NY, USA).

163 **16S rRNA gene sequencing and analysis**

164 16S rRNA genes were amplified from 10 ng of purified genomic DNA in duplicate using
165 archaeal and bacterial primers 515F and 806R, which target the V4 region of *Escherichia coli* in

166 accordance with the protocol described in [42,43], used by the Earth Microbiome Project
167 (<http://www.earthmicrobiome.org/emp-standard-protocols/16s/>), with a slight modification: the
168 annealing temperature was modified to 60 °C. Duplicate PCR reactions were combined and
169 purified using Agencourt AMPure XP PCR Purification beads (Beckman Coulter, Indianapolis,
170 IN) and sequenced using the Illumina MiSeq platform. Ninety-nine samples from 52 stations
171 were sequenced and analyzed. Raw sequences were joined using fastq-join [44] with the
172 *join_paired_ends.py* command and then demultiplexed using *split_libraries_fastq.py* with the
173 default parameters in QIIME version 1.9.1 [45]. Demultiplexed data matching Phi-X reads were
174 removed using the SMALT 0.7.6 akutils phix_filtering with the *smalt map* command [46].
175 Chimeras were removed using *vsearch -uchime_denovo* with VSEARCH 1.1.1 [47]. Neither
176 PhiX contamination nor chimeric sequences were observed. Sequences were then clustered into
177 operational taxonomic units (OTUs), which was defined as $\geq 97\%$ 16S rRNA gene sequence
178 similarity using SUMACLUST [48] and SortMeRNA [49] with *pick_open_reference.py -m*
179 *sortmerna_sumaclus*. Greengenes version 13.5 [50] was used for taxonomy. The resulting OTU
180 table was filtered to keep only OTUs that had 10 sequences or more (resulting in 8,959 OTUs),
181 and normalized using cumulative sum scaling (CSS) with metagenomeSeq [51] in R. These
182 sequences will be available in NCBI's SRA and on the Mason server at
183 <http://mason.eoas.fsu.edu>. While taxonomy was determined using Greengenes, taxonomy for
184 OTUs that were determined to be significantly (Wilcoxon) different between environments
185 (surface and bottom, hypoxic and oxic, Y13 and Y14 samples) were additionally verified beyond
186 the class level using SILVA ACT, version 132 alignment and classification [52] with the
187 confidence threshold set to 70%. To determine close relatives of specific OTUs of interest

188 NCBI's nucleotide blastn was used to search the Nucleotide collection using Megablast with
189 default parameters [53]. Pairwise sequence comparisons were also carried out using blastn.

190 **Statistics**

191 Multiple rarefactions and subsequently the alpha diversity metrics Shannon (H') [54], observed
192 OTUs (calculates the number of distinct OTUs, or richness) and equitability (Shannon
193 diversity/natural log of species richness; the scale is 0-1.0; with 1.0 indicating that all species are
194 equally abundant) were calculated in QIIME version 1.9.1 using *multiple_rarefactions.py*
195 followed by *alpha_diversity.py*. The Shapiro-Wilk test (*shapiro.test* function) was used to test
196 diversity values and environmental variables for normality in R. In R, the Wilcoxon Rank-Sum
197 test (*wilcox.test* function) with the application of Benjamini-Hochberg's (B-H) False Discovery
198 Rate (FDR) ($\alpha = 0.05$) [55] was used to test for significant differences in diversity values,
199 environmental variables, absolute abundance data (Thaumarchaeota 16S rRNA and *amoA* gene
200 copy numbers per L of seawater) and CSS normalized oligotype data between surface and
201 bottom samples, oxic and hypoxic samples and Y13 and Y14.

202

203 Statistically significant differences in all CSS normalized OTU abundances between surface and
204 bottom samples, hypoxic and oxic conditions and between Y13 and Y14 samples were
205 determined using the non-parametric Wilcoxon test in METAGENassist [56], with the B-H
206 correction for multiple tests. Prior to the Wilcoxon test, data filtering was carried out to remove
207 OTUs that had zero abundance in 50% of samples and the interquartile range estimate was used
208 to filter by variance to detect near constant variables throughout [57]. After quality filtering 528
209 OTUs remained out of 8,960 OTUs for Y14 surface and bottom samples, 623 OTUs remained

210 out of 7,924 OTUs for Y14 bottom only samples, and 724 OTUs remained out of 9,784 OTUs
211 for the same samples collected in Y13 and Y14.

212

213 Beta-diversity of CSS normalized data was examined using non-metric multidimensional
214 (NMDS) scaling with Bray-Curtis in R with the *metaMDS* function in the *vegan* package [58].

215 The *envfit* function in *vegan* was then used to fit vectors of environmental parameters onto the
216 ordinations with p-values derived from 999 permutations with the application of the B-H

217 correction for multiple tests using the *p.adjust* function (we defined corrected p-values ≤ 0.05 as
218 significant). Using the *vegan* package in RStudio, the non-parametric test *adonis* [59] was used

219 to test whether microbial community composition was significantly different between clusters of
220 samples (surface and bottom, hypoxic and oxic, year 2013 and 2014, east and west latitudes).

221 The *betadisper* function (*vegan*) was used to test for homogeneity of multivariate dispersion
222 among sample clusters for surface/bottom samples, hypoxic/oxic samples, Y13/Y14 samples and

223 east/west samples. *Betadisper* p-values were derived from 999 permutations using the *permutest*
224 function and a B-H correction for multiple tests was performed. Using the *psych* package [60] in

225 R, the nonparametric Spearman's rank order correlation coefficient (ρ) and p-values (B-H
226 correction) were determined for environmental variables and CSS normalized OTUs that were

227 significantly different (Wilcoxon) for Y14 surface and bottom samples, hypoxic and oxic
228 samples and Y14 and Y13 same samples.

229

230 Co-occurrence analysis between OTUs that were significantly (Wilcoxon) different between Y13
231 and Y14 was carried out by determining Spearman's correlation coefficients using the *psych*

232 package in R, similar to [61–64]. Spearman's correlation results were visualized in R for OTUs

233 that were significantly (corrected p-values ≤ 0.05) correlated with one or more of the three
234 Thaumarchaeota OTUs (4369009, 1584736 and 4073697).

235 **Oligotyping**

236 Shannon entropy (oligotyping) [65,66] analysis was carried out on all 16S rRNA gene sequences
237 identified as *Nitrosopumilus* to identify variability in specific nucleotide positions in this genus.
238 Scripts were used to format QIIME generated data for oligotyping using *q2oligo.py* and
239 *stripMeta.py* [67]. The QIIME command *filter_fasta.py* was used to obtain all *Nitrosopumilus*
240 16S rRNA gene sequences. All *Nitrosopumilus* 16S rRNA gene data was then analyzed using the
241 oligotyping pipeline version 0.6 for Illumina data [65,66] with the following commands, *o-trim*,
242 *o-pad-with-gaps* and *entropy-analysis*. All oligotype data was CSS normalized using
243 metagenomeSeq in R.

244 **Quantitative PCR**

245 Thaumarchaeal and bacterial 16S rRNA and archaeal *amoA* genes were quantified in duplicate
246 using the quantitative polymerase chain reaction (qPCR) assay. For each qPCR reaction 10 ng of
247 genomic DNA was used. Thaumarchaeota 16S rRNA genes were amplified using 334F and
248 554R with an annealing temperature of 59°C [68]. Bacterial 16S rRNA genes were amplified
249 using 1369F and 1492R with 56°C as the annealing temperature [68]. Archaeal *amoA* genes were
250 amplified using Arch-*amoA*-for and Arch-*amoA*-rev with 58.5°C as the annealing temperature
251 [29]. Standards (DNA cloned from our samples for Thaumarchaeota 16S rRNA and archaeal
252 *amoA* and *E. coli* for bacterial 16S rRNA) were linearized, purified, and quantified by
253 fluorometry. The reaction efficiencies for the standard curve were calculated from the slope of
254 the curve for all qPCR assays and were 91.1% for Thaumarchaeota 16S rRNA genes, 88.5% for

255 bacterial 16S rRNA genes and 85.3% for archaeal *amoA* genes. The qPCR data was converted to
256 gene copies L⁻¹ of seawater.

257

258 **Data deposition**

259 All sequences reported in this paper will be deposited into the NCBI sequence archive upon
260 article acceptance to the journal.

261

262 **Results**

263 **In situ chemistry and physical attributes of the 2014 hypoxic zone**

264 All environmental parameters measured, including DO concentrations in all ninety-nine surface
265 and bottom samples and the two surface samples from sites at the MI River mouth (R4 and R2)
266 collected in late July from Y14 are shown in Supplementary Table 1. In Y14, all environmental
267 variables (S1 Table) except chlorophyll a were significant between surface and bottom samples
268 (Wilcoxon with B-H corrected p-values; S2 Table). Average NO₂ + NO₃ concentrations were
269 higher in surface samples compared to bottom samples, while NH₄ and PO₄ were higher in
270 bottom samples.

271

272 Bottom water hypoxic conditions in 2014 were confined to the coast and shallower depths (S1
273 Fig), reaching a total area of 13,080 km². When looking at bottom samples only, all
274 environmental variables except NH₄, temperature and chlorophyll a were significantly different
275 between hypoxic and oxic samples (S2 Table). Average NO₂+NO₃ and PO₄ concentrations were
276 higher in hypoxic water samples compared to oxic samples, while average salinity

277 concentrations were higher in oxic samples. The average depth of the bottom water hypoxic
278 samples was 15m and the average depth of oxic samples was 25m.

279 **Microbial community composition across the shelf and with depth** 280 **in the 2014 hypoxic zone**

281 ITag sequencing of 16S rRNA genes was used to determine microbial community composition
282 across the shelf in the Y14 nGOM hypoxic zone, surface and bottom samples. This sequencing
283 effort resulted in 11.9 million reads and 8,959 number of OTUs (the full OTU table is included
284 as S3 Table). In the two MI River samples, Actinobacteria, Cyanobacteria and Proteobacteria
285 were the most abundant phyla (Fig 1A) (Thaumarchaeota relative abundances were low, with R4
286 relative abundances being 1.4% and R2 abundances <0.001%). Actinobacteria OTU4345058
287 (ac1) relative abundance was the highest (up to 22% relative abundance at site R4 and 0.7% at
288 the more saline R2 site), decreasing to 0.1% in surface samples near the mouth of the MI River
289 to < 0.001% to undetectable outside of the MI River in surface and bottom samples (avg. in
290 surface samples 3.54×10^{-4} and avg. in bottom samples 3.8×10^{-5}). Cyanobacteria (*Cyanobium*
291 PCC-6307) OTU404788 was the second most abundant OTU in river samples with a higher
292 abundance in the more saline R2 sample (up to 28% at site R2 and 3% at R4). The surface
293 microbial communities had similar dominant phyla to the MI River samples with Cyanobacteria
294 (avg. 33%) and Proteobacteria (avg. 32%), predominantly Gamma- and Alphaproteobacteria
295 being the most abundant, but had lower abundances of Actinobacteria (3%) than in the two river
296 samples (Fig 1A). In surface samples the average relative abundance of Thaumarchaeota, *N.*
297 *maritimus* OTU4369009 was 0.6%.

298

299 **Fig 1. Phyla level bar graph and boxplots of most abundant bacterial and archaeal groups.**

300 (A) Phyla level bar graph of relativized 16S rRNA gene iTag sequence data, in which only the
301 more abundant bacterial and archaeal groups are shown. Less abundant groups were summed
302 under “Other.” Samples are sorted from lowest to highest DO concentrations on the x-axis and
303 surface and bottom samples are differentiated by brackets with the two Mississippi (MI) River
304 samples on the far left. (B) Boxplots of most abundant classes for bottom samples plotted along a
305 DO gradient from lowest to highest DO concentrations (* indicates that taxonomy for these
306 phyla were further refined based on the OTU taxonomy in these groups. # indicates
307 that taxonomy was not resolved beyond phylum).

308

309 Bottom water (avg. collection depth was 19 m and avg. DO concentration was 2.26 mg L⁻¹)
310 samples were dominated by Proteobacteria (avg. 37%), primarily Gamma- Alpha- and
311 Deltaproteobacteria, as well as Thaumarchaeota (avg. 14%) (Figs 1A and B). The dominant
312 Thaumarchaeota, *N. maritimus* OTU4369009 had an average relative abundance of 13% in
313 bottom water samples. When plotting bottom water data along a DO gradient, several trends
314 emerged. Gammaproteobacteria, Alphaproteobacteria, Bacteroidetes, Cyanobacteria, and the
315 Actinobacteria Acidimicrobiia generally increased in relative abundance with higher DO (Fig
316 1B). Deltaproteobacteria, MGII Euryarchaeota and Planctomycetes generally increased in
317 abundance with decreasing DO (Fig 1B). In hypoxic samples, Thaumarchaeota were most
318 abundant, particularly at the lowest DO concentration, with decreasing abundances as DO
319 increased (Fig 1B). Of these taxa in the bottom samples, the normalized abundances of
320 Thaumarchaeota and Planctomycetes were significantly inversely correlated with DO
321 (Spearman’s ρ for Thaumarchaeota= -0.38 and Planctomycetes= -0.35, corrected p-values \leq
322 0.05).

323 **Microbial ecology and correlation analyses with environmental**

324 **variables across the 2014 hypoxic zone**

325 **Surface and bottom samples**

326 Statistical analysis of alpha diversity metrics revealed that microbial diversity (Shannon) was
327 significantly lower in the surface samples when compared with bottom samples (Wilcoxon with
328 B-H corrected p-values; S2 Table). Specifically, Shannon diversity indices averaged 5.97 in all
329 nGOM surface samples and 6.02 in the two surface river samples, as compared with 6.45 in
330 bottom samples. Richness (observed species) increased significantly with depth (avg. 329.92 in
331 surface samples and avg. 494.89 in bottom samples). A test of evenness (equitability) between
332 the surface and bottom sample types revealed that surface samples (avg. 0.72) were less even
333 than bottom samples (avg. 0.75).

334
335 Non-parametric statistical analysis (Wilcoxon) was then used to determine which OTUs were
336 responsible for the significant difference in species richness when comparing all surface and
337 bottom samples. Seventeen OTUs showed significant differences in their CSS normalized
338 abundances between surface and bottom samples (S2 Fig). Seven of these OTUs had higher
339 average CSS normalized abundances in bottom samples and were significantly inversely
340 correlated with DO (S2 Fig; Spearman's ρ and corrected p-values ≤ 0.05 in S4 and 5 Tables).

341 The dominant Thaumarchaeota, *N. maritimus* OTU4369009 was significantly inversely
342 correlated with DO and positively correlated with $\text{NO}_2 + \text{NO}_3$ and PO_4 (S4 and 5 Tables).

343

344 **Bottom water samples**

345 When analyzing bottom water samples alone, Shannon diversity, richness and evenness indices
346 for hypoxic versus oxic samples were not significantly different. However, comparison of the
347 bottom hypoxic and oxic samples, including environmental variables (Wilcoxon), revealed that
348 the CSS normalized abundances of 16 OTUs were significantly different depending on DO status
349 (S3 Fig). Of the 16 OTUs, six had higher CSS normalized abundances in hypoxic samples and
350 were significantly inversely correlated with DO (S3 Fig; Spearman's ρ and corrected p-values \leq
351 0.05 in S6 and 7 Tables). Thaumarchaeota, *N. maritimus* OTU4369009 comprised an average of
352 16% of the microbial community in hypoxic samples, with a peak abundance of 33% of the
353 microbial community in sample A'2 bottom, which had one of the lowest DO concentrations
354 (0.31 mg of O₂ L⁻¹), versus an average of 10% relative abundance in oxic samples. Further,
355 Thaumarchaeota, *N. maritimus* OTU4369009 was significantly inversely correlated with DO in
356 Y14 bottom samples and significantly inversely correlated with NH₄ and temperature in bottom
357 samples (S6 and 7 Tables). Thaumarchaeota, *N. maritimus* OTU4369009 was significantly
358 positively correlated with NO₂+NO₃, PO₄ and salinity in bottom samples (S6 and 7 Tables).
359 While this Thaumarchaeota OTU4369009 was abundant in hypoxic samples and inversely
360 correlated with DO, the normalized abundances were not significantly different between hypoxic
361 and oxic bottom samples (Wilcoxon).

362 **Microbial community organizational structure and drivers in the** 363 **2014 hypoxic zone**

364 To examine the primary drivers in structuring the 2014 microbial communities in the surface and
365 in the bottom water nGOM hypoxic zone, whole community 16S rRNA gene sequence data were
366 examined using Bray-Curtis distances with non-metric multidimensional scaling (NMDS). All
367 environmental variables shown as vectors were significantly correlated (corrected p-values \leq

368 0.05) with NMDS axes revealing the primary drivers in influencing the microbial community
369 structure to be DO, depth, NO₂+NO₃, and PO₄ (Fig 2 and S8 Table), with NO₂+NO₃ and PO₄,
370 decreasing with increasing distance from the MI river mouth. While the adonis test for all Y14
371 samples was significant, so too was beta-dispersion, suggesting non homogenous dispersion for
372 these sample clusters.

373

374 **Fig 2. NMDS ordination of normalized 16S rRNA gene iTag sequence data.**

375 **(A)** NMDS ordination of normalized 16S rRNA gene iTag sequence data for all Y14 surface and
376 bottom samples. Mississippi River samples (MI) are in red **(A)**, samples east of the Atchafalaya
377 River (AR) are in magenta and samples west of the AR are in purple. **(B)** NMDS ordination of
378 normalized 16S rRNA gene iTag sequence data of Y14 bottom samples only. Bubble sizes
379 represent DO concentrations e.g. larger bubbles indicate higher DO concentrations. All
380 environmental variables shown as vectors were significantly (corrected p-values ≤ 0.05)
381 correlated with an NMDS axis for both **(A)** and **(B)**.

382

383 To analyze beta diversity in Y14 bottom water hypoxic verses oxic conditions, surface samples
384 were excluded and bottom water samples were examined using NMDS ordination (Fig 2B). An
385 adonis test revealed distinct microbial communities in hypoxic samples as compared to oxic
386 water samples (Fig 2B) (adonis $R^2= 0.09$ and p-value ≤ 0.05 , beta-dispersion $F= 0.60$ p-value \geq
387 0.05). Bottom samples showed significant spatial clustering east and west of the AR (adonis $R^2=$
388 0.35 and p-value ≤ 0.05 , beta-dispersion $F= 0.01$ p-value ≥ 0.05) (Fig 2B).

389 **Taxonomic and functional gene abundances in the 2014 hypoxic**

390 **zone**

391 Bacterial and thaumarchaeal 16S rRNA and archaeal *amoA* gene copy numbers (qPCR) were
392 determined in surface and bottom water samples. Bacterial 16S rRNA gene copy numbers L⁻¹
393 were similar in the surface (avg. 4.10×10^7) and in the bottom water samples (avg. 3.73×10^7).
394 In contrast, thaumarchaeal 16S rRNA gene copy numbers L⁻¹ were significantly higher in bottom
395 water samples (avg. 2.18×10^7) compared to surface samples (avg. 2.19×10^6) as were *amoA*
396 copy numbers L⁻¹ (bottom avg. 3.12×10^7 and surface avg. 2.96×10^6) (Wilcoxon, S2 Table).
397 When comparing hypoxic and oxic samples in bottom only samples, thaumarchaeal 16S rRNA
398 gene copy numbers L⁻¹ were significantly higher in hypoxic water samples (avg. 3.28×10^7)
399 compared to oxic water samples (avg. 8.87×10^6) as were *amoA* copy numbers L⁻¹ (avg. hypoxic
400 4.65×10^7 and avg. oxic 1.32×10^7) (S2 Table In surface and bottom water samples, and in
401 bottom water only samples, thaumarchaeal 16S rRNA and *amoA* gene copy numbers L⁻¹ were
402 significantly positively correlated with each other, NO₂+NO₃ and PO₄ and significantly inversely
403 correlated with DO (S4-7 Tables). The ratio of Thaumarchaeota 16S rRNA:*amoA* gene copy
404 number L⁻¹ was one (avg.).

405 **Differences in the extent of hypoxia and microbial community** 406 **structure between years 2013 and 2014**

407 The total area of low oxygen in Y14 was 13,080 km², compared to 15,120 km² in Y13
408 (gulfhypoxia.net). Comparing the same hypoxic sites we sampled in Y13 and Y14 revealed that
409 DO in Y13 samples was lower (avg. 0.62mg/L) than Y14 (avg. 1.1mg/L) and the average depth
410 of the hypoxic zone was deeper in Y13 (17.7m) compared to Y14 (14.6m) (S4 Fig).
411 Ammonium, salinity and temperature were significantly different between the two hypoxic zones
412 (Wilcoxon, S2 Table) with average NH₄ concentrations being higher in Y13. In the Y13 hypoxic

413 zone, the average ammonium and salinity concentrations were higher, while in the Y14 hypoxic
414 zone (which was located at shallower depths) the average temperature was higher.

415
416 Alpha diversity, Shannon, observed (species richness) and equitability were not statistically
417 significantly different between the years. To look at beta diversity between the years, the bottom
418 samples collected at the same stations in Y13 and Y14 were examined using NMDS (S5 Fig),
419 with environmental variables that were significantly correlated with NMDS axes represented as
420 vectors (corrected p-values ≤ 0.05 , S8 Table). An adonis test revealed distinct microbial
421 communities in Y14 samples as compared to Y13 samples (S5A Fig) (adonis $R^2= 0.20$ and p-
422 value ≤ 0.05 , beta-dispersion $F= 1.07$ p-value ≥ 0.05). Whereas Y14 showed distinct east and
423 west clusters, Y13 did not (adonis $R^2= 0.09$ and p-value ≤ 0.05 , beta-dispersion $F= 10.47$ p-value
424 ≤ 0.05) (S5A Fig).

425
426 Analysis of normalized iTag sequence data of bottom samples collected at the same stations in
427 Y13 and Y14 revealed that the phyla Thaumarchaeota, Actinobacteria, Planctomycetes,
428 Euryarchaeota and SAR406 were greater in Y13 than Y14, whereas Cyanobacteria,
429 Proteobacteria and Bacteroidetes were greater in Y14. At the OTU level, 17 had significant
430 normalized abundances between Y13 and Y14 (Wilcoxon). Five of the 17 OTUs were more
431 abundant in Y14, whereas 12 OTUs were more abundant in Y13 (S6 Fig). Specifically,
432 Thaumarchaeota, *N. maritimus* OTU4369009 (S6 Fig), a Thaumarchaeota OTU4073697 (95%
433 similar to cultured representative *Nitrosopelagicus brevis* strain CN25) and Euryarchaeota, MGII
434 OTU3134564 had higher normalized abundances in Y13 than in Y14. Of the five OTUs that had
435 higher abundances in Y14, one was a Thaumarchaeota OTU1584736 (95% similar to cultured

436 representative *Nitrosopelagicus brevis* strain CN25). The two Thaumarchaeota that were 95%
437 similar to *Nitrosopelagicus brevis* strain CN25 had a one nucleotide (nt) base pair differentiation
438 at 82/250 (G-A) in the 16S rRNA gene sequence.

439

440 When comparing absolute abundance data in the same hypoxic sites sampled in both years,
441 thaumarchaeal 16S rRNA and archaeal *amoA* gene copy numbers (qPCR) per L averages were
442 significantly higher (S4 Fig and S2 Table) in Y14 (avg. 3.31×10^7 and 4.52×10^7) than in Y13
443 (avg. 7.25×10^6 and 6.87×10^6). In both years, copy number per L of each gene was significantly
444 inversely correlated with DO (Spearman's ρ and corrected p-values in S9-12 Tables).

445 **Shannon entropy analysis of *Nitrosopumilus* 16S rRNA gene**

446 **sequence data in the 2013 and 2014 hypoxic zones**

447 Due to the high abundances of *Nitrosopumilus* in hypoxic samples in Y13 and Y14, oligotyping
448 analysis was carried out to examine closely related *Nitrosopumilus* (all sequences annotated as
449 such) in relationship to environmental variables. The Y13 iTag data were not analyzed in this
450 way in our previous paper [20], so here we present CSS normalized OTU data oligotyping results
451 for both Y13 and Y14 in the same station samples ($n = 35/\text{year}$). The normalized abundance of
452 oligotype G (G in nt position 115/250) was significantly higher in abundance in hypoxic samples
453 and more abundant in Y13 compared to Y14 (Wilcoxon, S2 Table). Oligotype G was
454 significantly inversely correlated with DO in both Y13 and Y14 (S9-12 Tables). Conversely,
455 oligotype A (A in nt position 115/250) was lower in hypoxic samples in both years (Fig 3).
456 Oligotypes T or C at nt position 115/250 were lower, reaching maximal abundances of 6.45%
457 and 0.29%, respectively.

458

459 **Fig 3. Ocean Data View plots of DO and *Nitrosopumilus* G oligotype data.**

460 Plots of (A) DO concentrations for the same samples collected in Y13 (B) and Y14. (C)

461 Oligotype data for *Nitrosopumilus* G oligotype for the same samples collected in Y13 (C) and

462 Y14 (D).

463

464 **Thaumarchaeota AOA microbial species co-occurrence patterns in**
465 **the 2013 and 2014 hypoxic zones**

466 Similar to oligotyping analysis, species co-occurrence was not considered in our Y13 dataset.

467 We therefore evaluated species co-occurrences by determining Spearman's correlation

468 coefficients for the same Y13 and Y14 bottom water sites (n = 35/year). For this analysis, we

469 included only the 17 OTUs discussed above that were significantly different between the years

470 (Wilcoxon), which included the three Thaumarchaeota OTUs; 4369009, 1584736, 4073697 (S6

471 Fig). Of the 17 OTUs that were significantly different between Y13 and Y14 same station

472 samples, ten OTUs were significantly correlated at least once with one of the three

473 Thaumarchaeota OTUs in Y13 and/or Y14 hypoxic or oxic samples (Fig 4).

474

475 **Fig 4. Plots representing co-occurrence (Spearman's ρ) data for OTUs of interest.**

476 Co-occurrence (Spearman's ρ) of Thaumarchaeota (*N. maritimus* OTU4369009, *Ca.*

477 *Nitrosopelagicus* OTU1584736 and *Ca. Nitrosopelagicus* OTU4073697) and ten OTUs whose -

478 normalized abundances were significantly different between years (Wilcoxon test with B-H

479 corrected p-values) in the same Y13 and Y14 bottom water samples. Correlations that were

480 significant (corrected p-values ≤ 0.05) are indicated by an *.

481

482 **Discussion**

483 The Y14 hypoxic zone was slightly smaller in size (13,080 km²), closer to shore and
484 discontinuous as compared to the Y13 hypoxic zone (15,120 km²) (S4 Fig). In July 2014, wind
485 speeds reached between 10 to 20 knots blowing from the west, with higher than average wave
486 height (1.4 meters) reported (<http://www.wavcis.lsu.edu/>). Further, in Y14 there was an
487 unseasonably late (late July), above average Mississippi River discharge (avg. is $\sim 4 \times 10^5$ ft²/sec
488 compared to 5×10^5 ft²/sec in Y14), which resulted in nitrogen (NO₂ + NO₃) concentrations
489 reaching a near-record high (200 $\mu\text{mol L}^{-1}$) and an anomalously high phytoplankton biomass
490 (e.g., 118 $\mu\text{g/L}$ Chl *a* at the end of July) (LUMCON 2014, <http://bit.ly/2wHLSk1>, accessed
491 02/08/17) late in the hypoxic season (S1 Table and S1 Fig). These factors could have resulted in
492 the slight DO replenishment seen between 92° W and 92.38° W, specifically along transect G,
493 west of the Atchafalaya River (AR) (S1 Table and S1 Fig). Previous reports of wind out of the
494 west and southwest during the summer months have correlated with a smaller hypoxic zone,
495 which moves nutrient enhanced water masses to the east and to deeper waters, disrupting density
496 stratification [69,70]. Therefore, a plausible hypothesis is that the variability in wind forcing
497 resulted in the movement of water masses to the east [71]. This wind forcing could have
498 influenced the hypoxic area to the west of the Atchafalaya River, resulting in the variability in
499 the size and shape of the hypoxic zone between Y13 and Y14, while also potentially shaping the
500 microbial community as seen in the east west latitude clustering in Y14, but not in Y13 (S4 and 5
501 Figs).

502

503 Specifically, our data revealed a significant inverse correlation between AOA and DO in the
504 nGOM. However, in Y14 when the hypoxic zone was less extensive, and focused at the MI

505 River mouth the normalized abundances of AOA were lower than in Y13. In Y14, although
506 AOA were enriched in the hypoxic bottom water, after correcting for multiple comparisons, their
507 normalized abundances (iTag) were not significantly higher in hypoxic versus oxic samples. Yet
508 the absolute abundances of thaumarchaeal 16S rRNA and *amoA* genes were significantly higher
509 in Y14 than in Y13 (S4 Fig). This discrepancy between iTag and qPCR suggested that the iTag
510 primers did not discern the full breadth of Thaumarchaeota diversity in comparison to qPCR
511 primers. To reconcile the disparity between iTag and qPCR, samples that had high
512 Thaumarchaeota 16S rRNA and *amoA* gene copy number were further analyzed. Both MI River
513 samples had low normalized abundances of Thaumarchaeota (less than 1%), but up to 10^5 copies
514 of Thaumarchaeota 16S rRNA genes/L and 10^6 copies of *amoA* genes/L. These results could be
515 indicative of Thaumarchaeota that are introduced to the nGOM during freshwater input,
516 compared to Thaumarchaeota that are adapted to saline conditions introduced to the hypoxic
517 zone via nGOM seawater in the bottom layer. Therefore, the location and physical parameters of
518 the hypoxic zone could influence Thaumarchaeota abundance and subsequently overall microbial
519 diversity.

520

521 Further comparison of Y13 and Y14 revealed that the *Nitrosopumilus* oligotype G was both
522 annually abundant in the nGOM and was significantly inversely correlated with DO (Fig 3).
523 This oligotype was differentiated by 1 bp in all 16S rRNA gene sequences annotated as
524 *Nitrosopumilus*. Whether this polymorphism is ubiquitous in low DO adapted Thaumarchaeota
525 is not yet known. The data did suggest that its abundance was influenced by the severity of
526 bottom water hypoxia (expansive and deeper versus shallow and smaller), where abundances
527 were greater in Y13 (Fig 3). Thus, this *Nitrosopumilus* oligotype may be adapted to low oxygen,

528 coastal conditions. Oligotyping has revealed ecologically important sub-OTUs in human and
529 environmental microbiomes across environmental gradients [66,72–75]. For example, Sintes and
530 colleagues [74] identified two groups of *amoA* oligotypes that clustered according to high or low
531 latitude and subclustered by ocean depth, however little other oligotypic analysis has been
532 carried out on AOA, specifically *Nitrosopumilus*. Thus, oligotyping can reveal subtle nucleotide
533 variations within AOA. Whether a low DO adapted AOA oligotype that was dominant in coastal
534 hypoxic samples are ecologically consequential remains to be determined.

535
536 Microbial species co-occurrences relationships can reveal community patterns, and facilitate
537 hypothesis generation regarding abiotic influences on random and non-random patterns [61–64].
538 In our study, co-occurrence analysis of OTUs that were significantly different between bottom
539 water sites in Y13 and Y14 revealed that increasing normalized abundances of *N. maritimus*
540 OTU4369009, particularly in Y13, potentially perturbed co-occurrence relationships with other
541 OTUs, specifically MGII Euryarchaeota OTU3134564 and Deltaproteobacteria OTU837775
542 (putatively identified as a SAR324 by SILVA) (Fig 4). Previous studies reported that MGII have
543 an aerobic photoheterotrophic lifestyle [76–78] with higher abundances in the euphotic zone
544 compared to depths below the euphotic zone [28,79–81]. Whereas Thaumarchaeota, closely
545 related to the *N. maritimus*, increase in abundance with depth [17,18,28,79,82]. Therefore, if
546 high thaumarchaeal abundances in the hypoxic zone are not met with higher MGII abundances,
547 potential metabolic linkages would be decoupled, which could be ecologically significant. It has
548 been reported that some SAR324 have the ability to oxidize hydrogen sulfide [83] which could,
549 in theory, mean that when hypoxic conditions prevail and AOA continues to draw down DO,
550 SAR324 could oxidize sulfide that may flux from the sediments resulting in detoxification of

551 bottom water. During hypoxic conditions when *N. maritimus* OTU4369009 abundances are high,
552 co-occurrence with SAR324 abundances are weakened (Fig 4) and sulfide oxidation by SAR324
553 may not keep pace, resulting in a deteriorating environment beyond low DO.

554
555 The two Thaumarchaeota (OTU1584736 and OTU4073697) that were significantly different
556 between the years were significantly positively correlated with each other in both hypoxic and
557 oxic samples in both years (Fig 4). These two OTUs are 95% similar to the *N. Brevis* CN25 [84],
558 an AOA that has been shown to produce N₂O enrichment cultures [85], similar to *N. maritimus*
559 [86,87]. AOA have been reported to be the primary source of N₂O in the surface ocean [85] and
560 it has been shown that decreasing oxygen concentration could influence the production of N₂O
561 by AOA [85,86,88–90]. In the shallow nGOM water column, Walker JT and colleagues [91]
562 reported that nitrification was the primary source of N₂O during peak hurricane season,
563 consistent with the results of Pakulski (2000) [25]. Therefore, the three Thaumarchaeota OTUs
564 that we show to be abundant in the n GOM hypoxic zone of both years may contribute to N₂O
565 production, a potent greenhouse gas [92,93], which can flux from ocean to the atmosphere when
566 the water column is mixed by tropical storm activity [91].

567
568 Additionally, a decoupling between the two-step transformation of ammonium to nitrate (which
569 are carried out by distinct groups of microorganisms) has previously been reported (e.g. [94,95],
570 including in the nGOM hypoxic zone [19]. In our dataset nitrate oxidizing bacteria (NOB) were
571 undetectable, thus the annual increase in AOA in the nGOM, and the lack of co-occurrence with
572 NOB suggested that nitrite may accumulate, as shown by Bristow and colleagues [19] in this
573 expansive hypoxic zone. This metabolic decoupling, leading to NO₂ accumulation during the

574 summer months when the nGOM hypoxic zone develops, is consequential given the toxicity of
575 nitrite [96].

576

577 **Conclusion**

578 Collectively, this dataset supports that the nGOM hypoxic zone can serve as a hotspot for AOA
579 and the that the normalized abundance of Thaumarchaeota 16S rRNA (iTag) and the absolute
580 abundance (qPCR) of Thaumarchaeota 16S rRNA and archaeal *amoA* gene copy numbers can
581 reflect the extent of bottom water hypoxia. There are few reports describing abundant
582 Thaumarchaeota in shallow coastal environments; most of which represent polar environments
583 [97–101]. Our findings of an increase in Thaumarchaeota in the hypoxic nGOM are consistent
584 with several studies that have reported an increase in abundance of Thaumarchaeota in low
585 oxygen marine environments [18–20,33–36], however this is first known dataset that has
586 sampled the water column microbial community of the shelfwide nGOM hypoxic zone in two
587 consecutive years. Future studies that determine the ecological implications of an AOA hotspot,
588 their co-occurrences and the potential impact on biogeochemical cycling, such as the
589 contribution to N₂O production and their role in ocean deoxygenation, which is intensifying in a
590 changing climate [1,2,102–104] are needed.

591

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862 the past five decades. *Nature.* 2017 Feb 15;542(7641):335–9.

863

864 **Conflict of Interest Statement**

865 The authors declare no conflict of interest.

866

867 **Author Contributions**

868 LGC, NNR and JCT collected samples. LGC carried out DNA extractions, library preparation
869 and sequencing and statistical analysis of the data. OUM designed the project and did the
870 bioinformatics. LGC and OUM wrote the manuscript.

871

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881

882 **Supporting Information**

883 **S1 Fig. (A)** Sample map of the 52 stations sampled during the Y14 nGOM shelfwide cruise, in
884 which bottom water status is indicated (red circles indicates oxic stations while blue circles
885 indicates hypoxic stations). **(B)** DO concentrations from the bottom samples collected.

886 **S2 Fig. (A)** NMDS ordination of normalized 16S rRNA gene iTag sequence data for all samples
887 collected in year 2014 grouped by surface and bottom. The seventeen bubble plots represent the
888 same NMDS plot with normalized abundances of the OTUs that were statistically significantly
889 different (Wilcoxon) between surface and bottom samples depicted by bubble size, where larger
890 bubble size represents higher normalized abundances. The symbol * represents OTUs that were
891 statistically significantly inversely correlated with dissolved oxygen (Spearman correlation; B-H
892 corrected p-value ≤ 0.05).

893 **S3 Fig. (A)** NMDS ordination of normalized 16S rRNA gene iTag sequence data for the all

894 bottom samples collected in year 2014 where bubble size represents DO concentration. The other
895 sixteen NMDS bubble plots represent the normalized abundances of the OTUs that were
896 statistically significantly different (Wilcoxon) between hypoxic and oxic samples. Larger bubble
897 size represents higher normalized abundances. The symbol * represents OTUs that were
898 statistically significantly inversely correlated with dissolved oxygen while the other OTUs are
899 significantly positively correlated with DO (Spearman correlation; B-H corrected p-value \leq
900 0.05).

901 **S4 Fig.** Plots of DO concentrations, relative abundances of Nitrosopumilus
902 OTU4369009 16S rRNA genes (iTag), Thaumarchaeota 16S rRNA and amoA gene copy
903 number/L (qPCR) for bottom water samples collected at the same sites in Y13 and Y14.

904 **S5 Fig. (A)** NMDS ordination of normalized 16S rRNA gene iTag sequence data for the same
905 samples collected in Y13 and Y14 with environmental variables, qPCR data and oligotype data
906 shown as vectors. All environmental variables represented as vectors on the NMDS were
907 significantly correlated (corrected p-value \leq 0.05) with an NMDS axis. **(B)** The same NMDS
908 ordination depicting oxygen concentration (DO) as bubble size e.g. larger bubbles indicate
909 higher DO concentrations.

910 **S6 Fig. (A)** NMDS ordination of normalized 16S rRNA gene iTag sequence data for the same
911 samples collected in Y13 and Y14 where bubble size depicts oxygen concentration. The other
912 seventeen NMDS bubble plots represent the normalized abundances of the OTUs that were
913 statistically significantly different (Wilcoxon) between the years where larger bubble size
914 represents higher normalized abundances. The symbol * represents OTUs that were statistically
915 significantly inversely correlated with DO in Y14, and the symbol # represents a significant
916 inverse correlation with DO in Y13 (Spearman correlation; B-H corrected p-value \leq 0.05).

- 917 **S1 Table.** 2014 nGOM hypoxic zone sample metadata and in situ chemistry.
- 918 **S2 Table.** Wilcoxon B-H corrected p-values for diversity statistics and environmental variables
919 for all datasets.
- 920 **S3 Table.** Operational taxonomic unit (OTU) table for all samples collected in 2014.
- 921 **S4 Table.** P-values (B-H corrected) for Spearman's correlation coefficients for Y14 surface and
922 bottom samples.
- 923 **S5 Table.** Spearman's Correlation coefficients (ρ) for Y14 surface and bottom samples.
- 924 **S6 Table.** P-values (B-H corrected) for Spearman's correlation coefficients for Y14 bottom
925 samples only.
- 926 **S7 Table.** Spearman's Correlation coefficients (ρ) for Y14 bottom samples only.
- 927 **S8 Table.** Envfit B-H corrected p-values for NMDS ordinations.
- 928 **S9 Table.** P-values (B-H corrected) for Spearman's correlation coefficients for Y13 same
929 samples only.
- 930 **S10 Table.** Spearman's Correlation coefficients (ρ) for Y13 same samples only.
- 931 **S11 Table.** P-values (B-H corrected) for Spearman's correlation coefficients for Y14 same
932 samples only.
- 933 **S12 Table.** Spearman's Correlation coefficients (ρ) for Y14 same samples only.

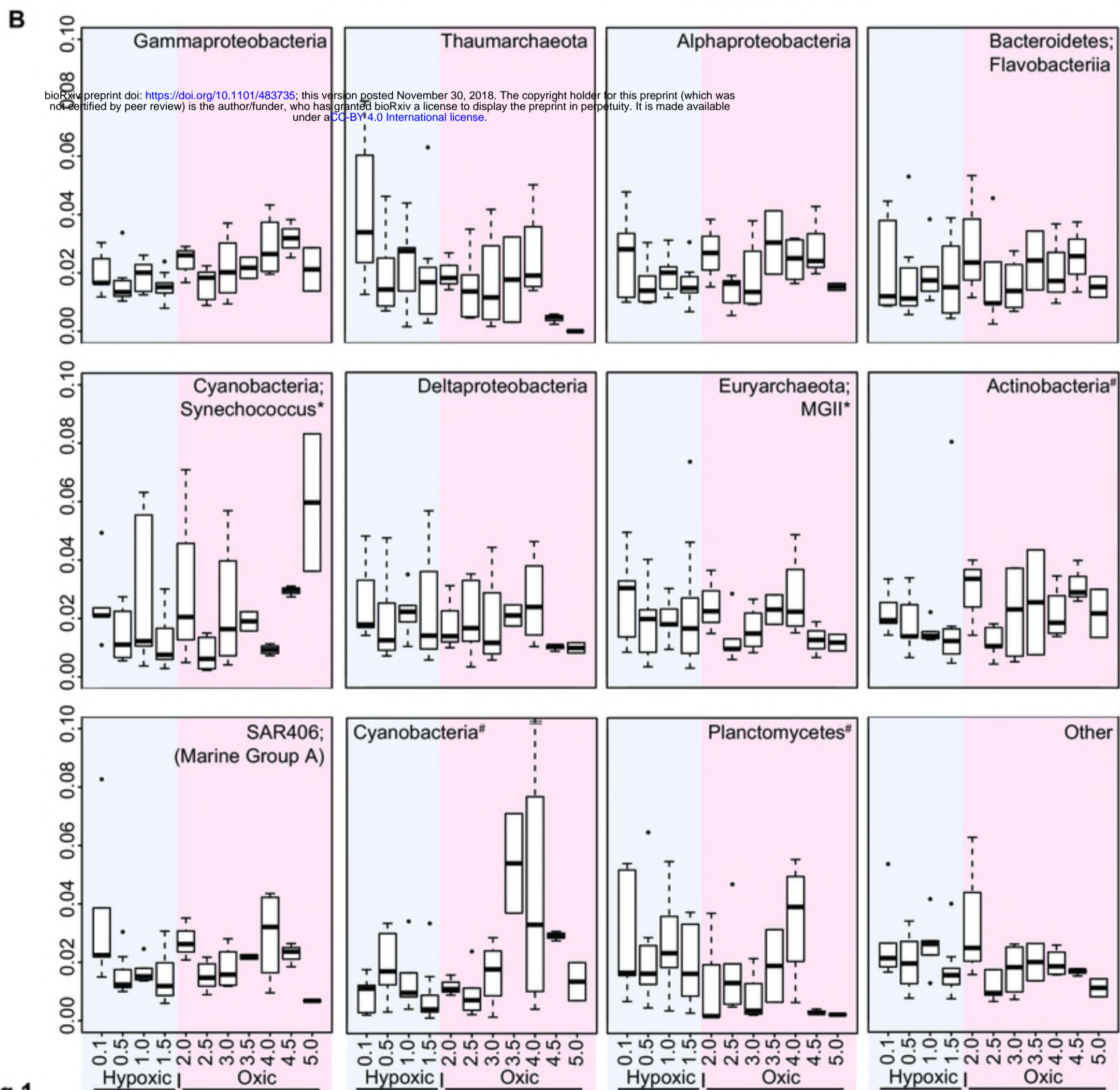
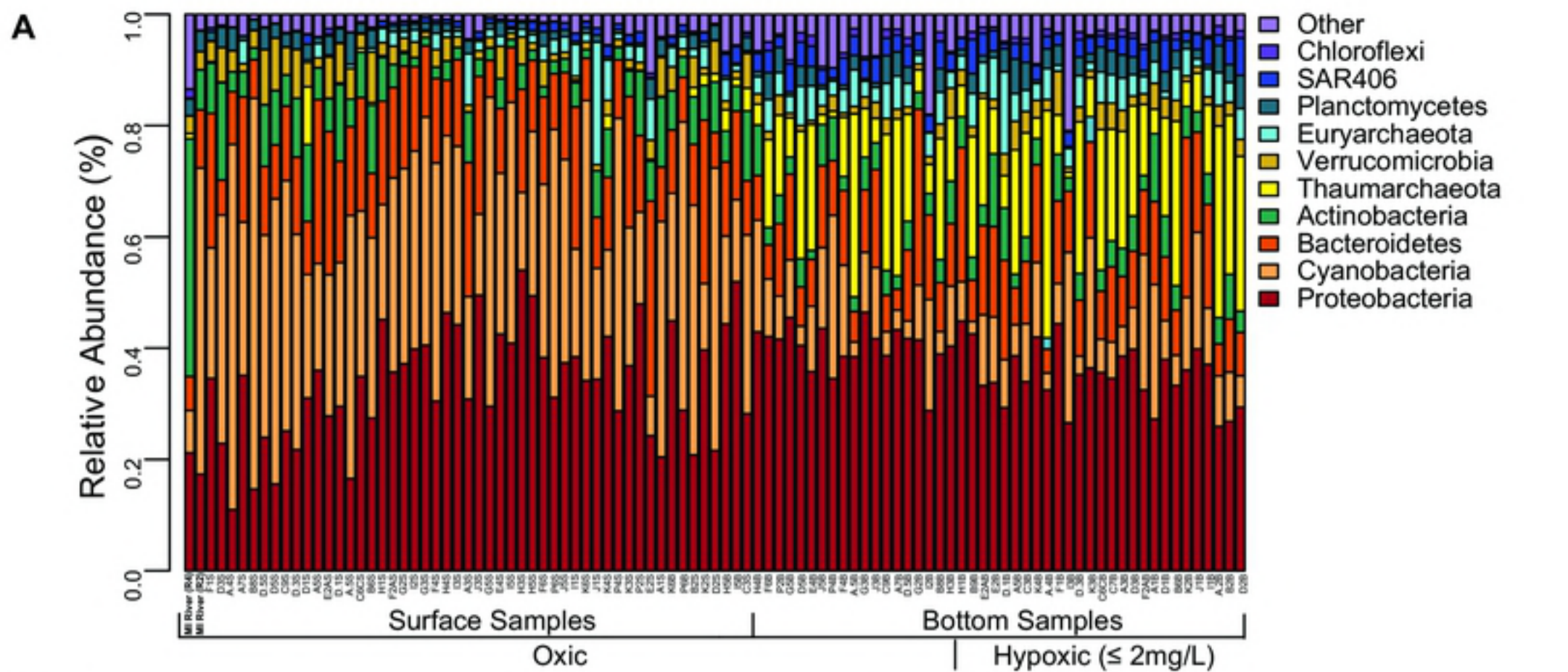


Fig 1

Fig 1

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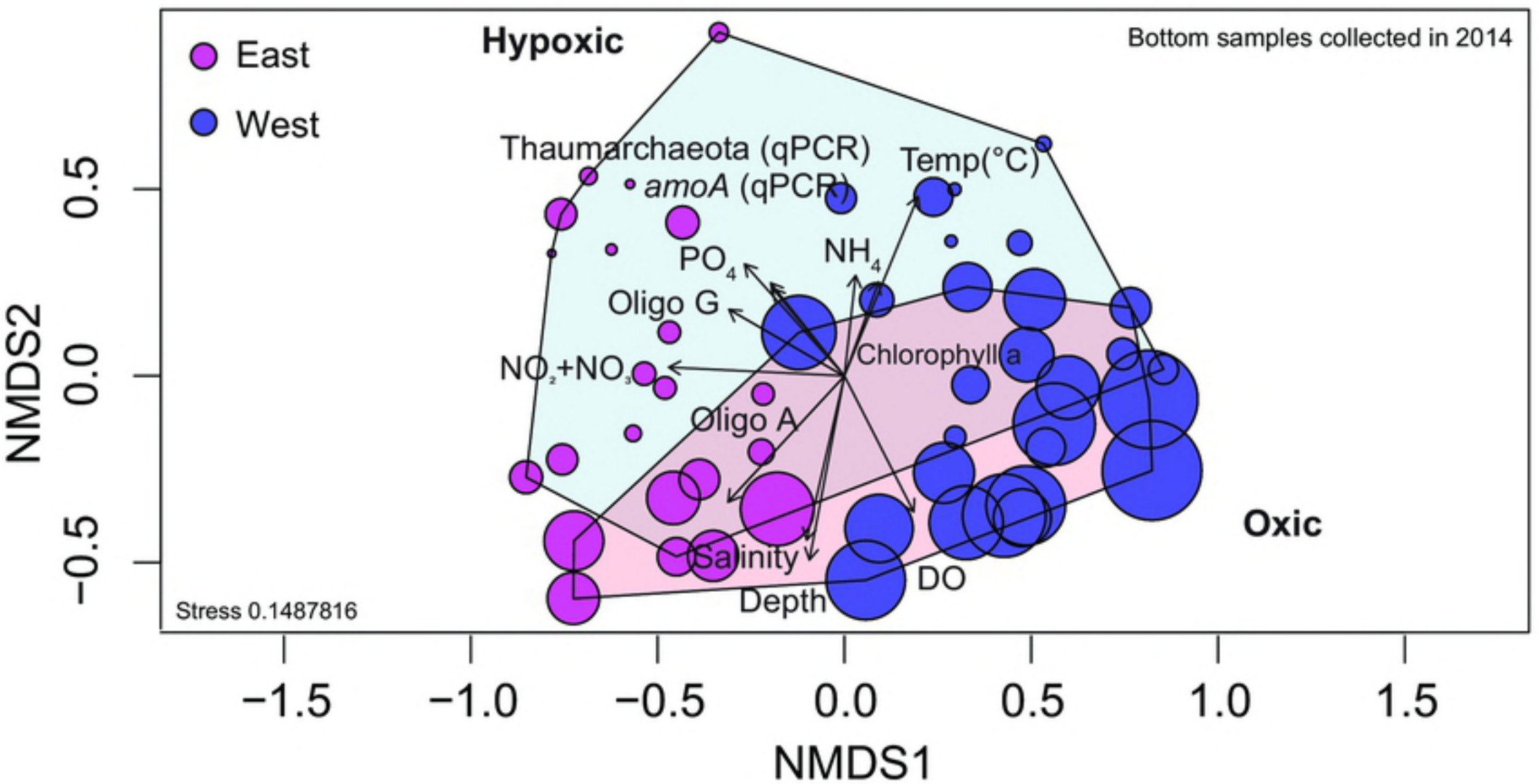
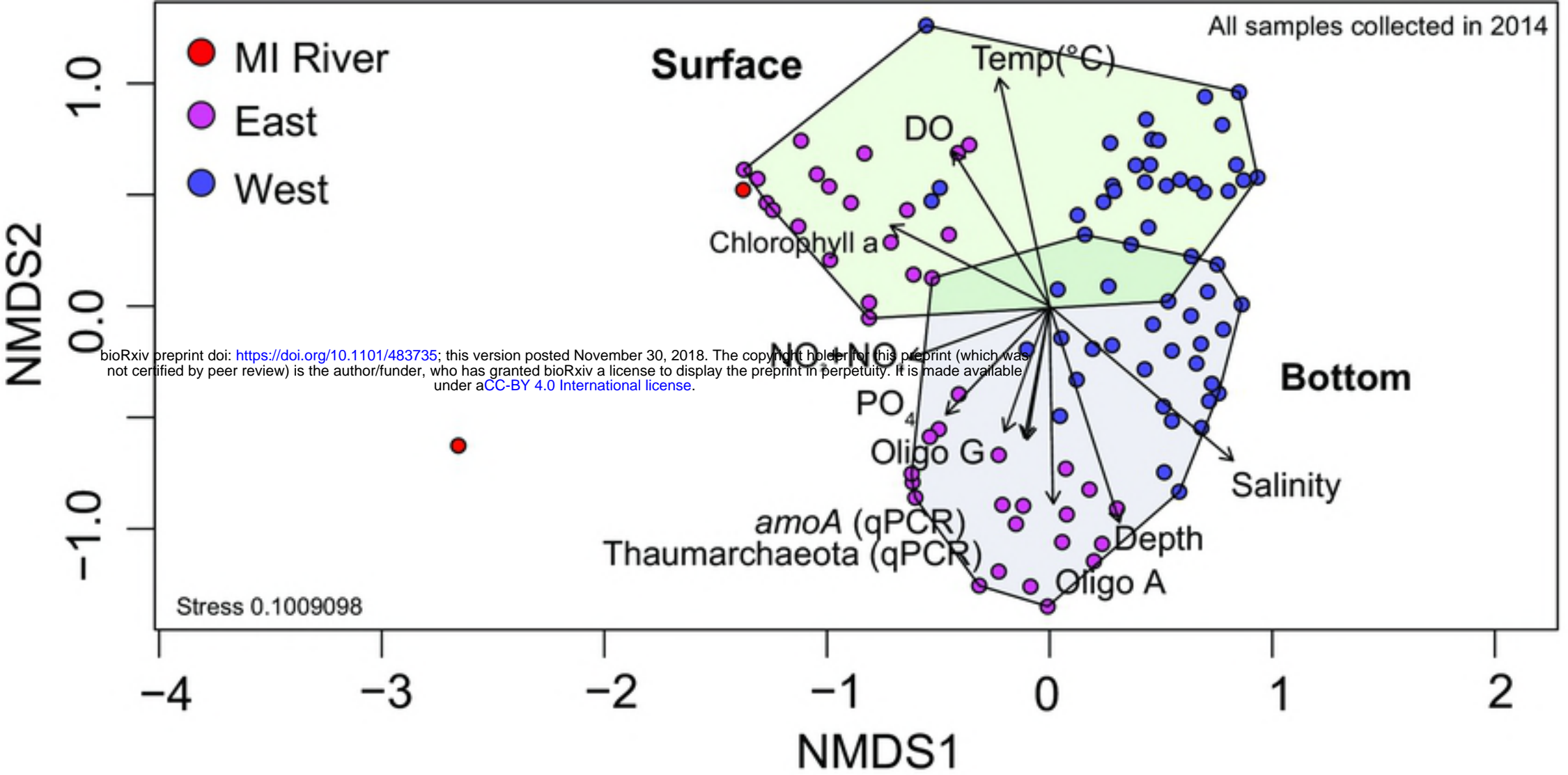


Fig 2

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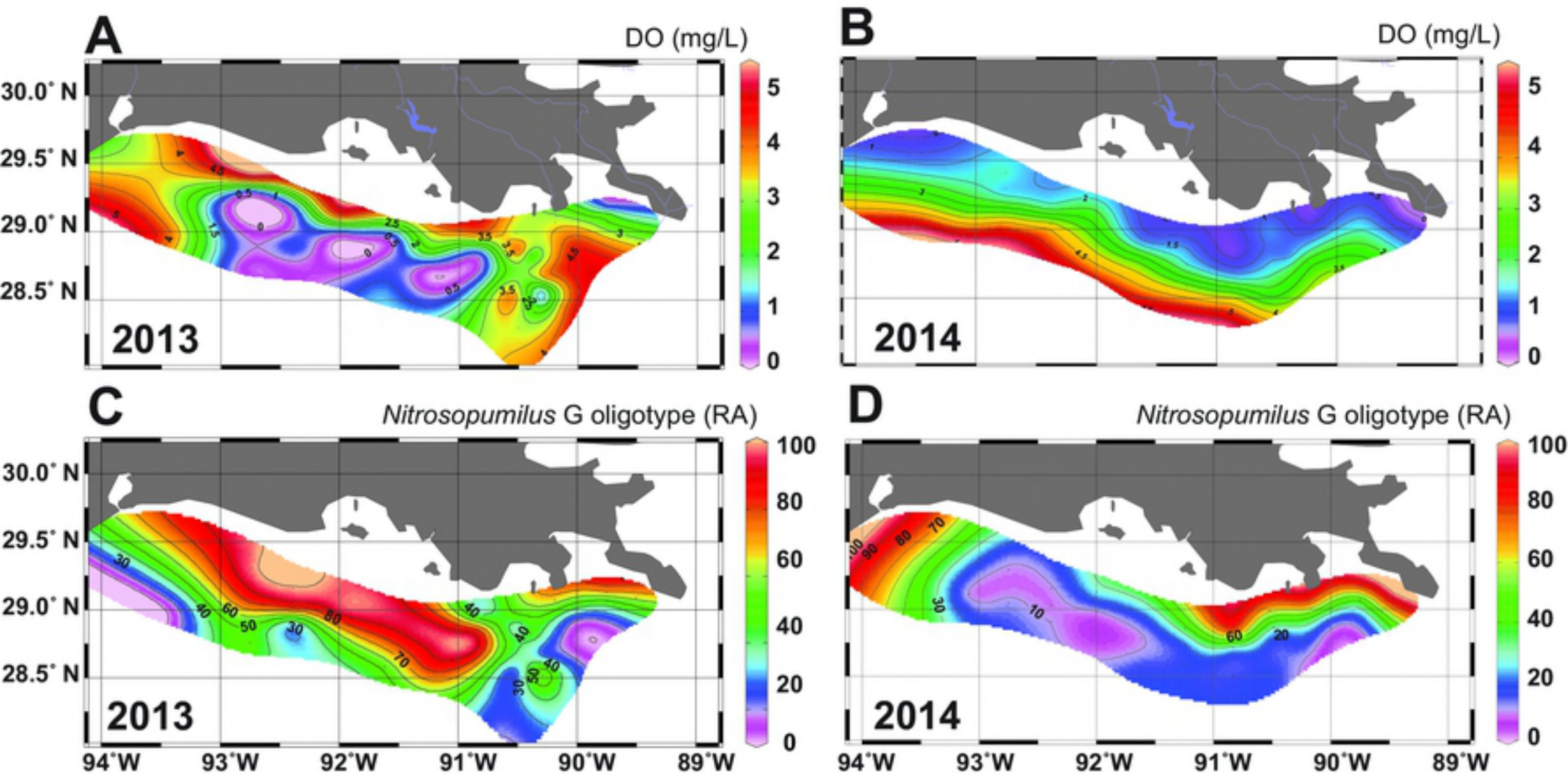


Fig 3

Fig 3

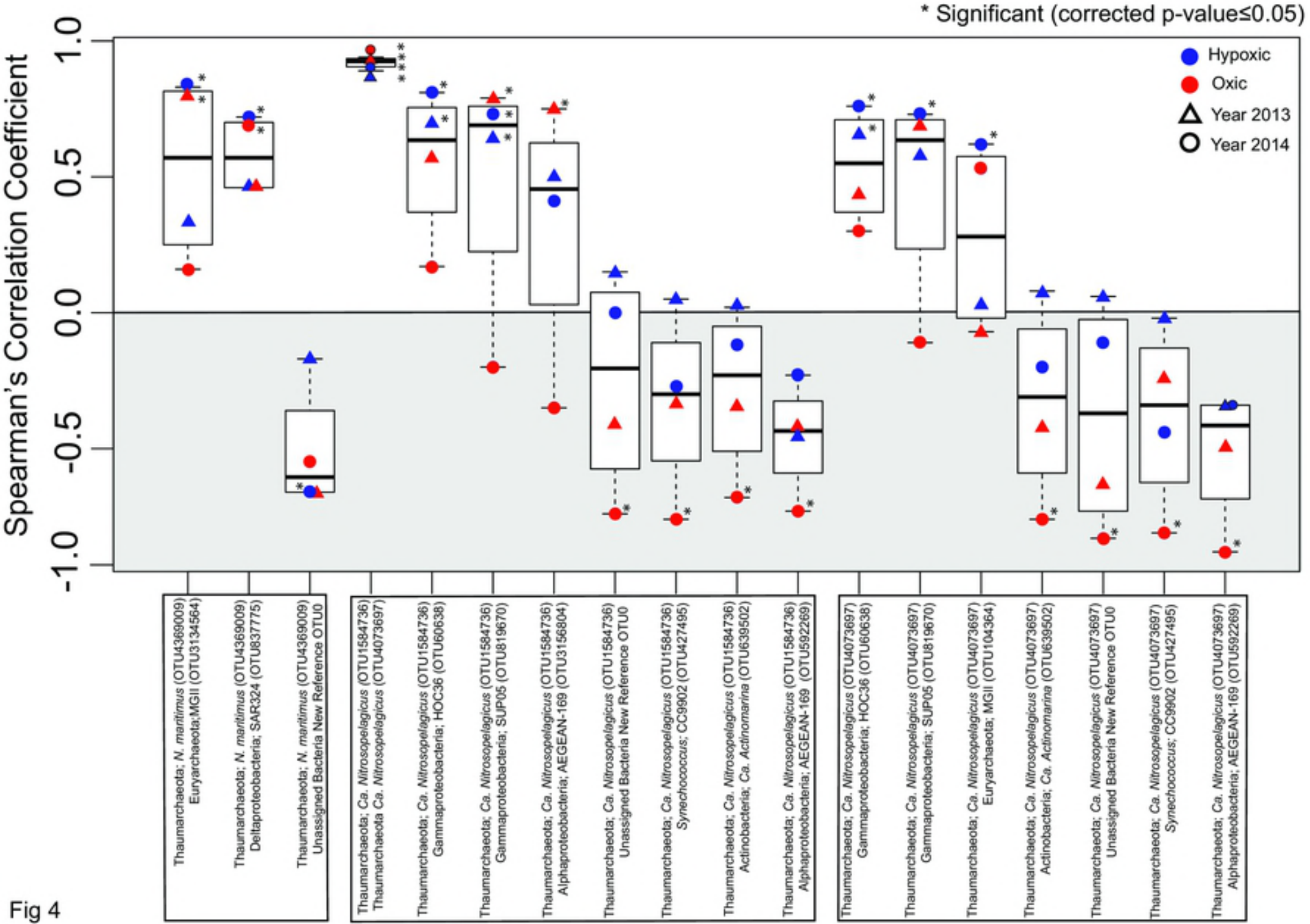


Fig 4

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