Carbon stocks of coastal seagrass in Southeast Asia may be far lower than anticipated when accounting for black carbon

John B. Gallagher[1, 2], Chuan Chee Hoe[3], Yap Tzuen-Kiat[4], and Farahain W[3]

[1] Centre for Marine and Coastal Studies, Universiti Sains Malaysia, 11800, Penang, Malaysia

[2] Institute for Marine and Antarctic Studies, University of Tasmania, 7000, Tasmania, Australia

[3] Faculty of Science and Natural Resources, Universiti Malaysia Sabah, 88400, Sabah, Malaysia

[4] Borneo Marine Research Institute, Universiti Malaysia Sabah, 88400, Sabah, Malaysia

Valuing sedimentary 'blue carbon' stocks of seagrass meadows requires exclusion of allochthonous recalcitrant forms of carbon, such as black carbon (BC). Regression models constructed across a Southeast Asian tropical estuary predicted that carbon stocks within the sandy meadows of coastal embayments would support a modest but not insignificant amount of BC. We tested the prediction across three coastal meadows of the same region: one patchy meadow located close to a major urban centre and two continuous meadows contained in separate open embayments of a rural marine park; all differed in fetch and species. The BC/total organic carbon (TOC) fractions in the urban and rural meadows with small canopies were more than double the predicted amounts, $28 \pm 1.6\%$ and $36 \pm 1.5\%$ ($\pm 95\%$ confidence intervals), respectively. The fraction in the rural large-canopy meadow remained comparable to the other two meadows, $26 \pm 4.9\%$ ($\pm 95\%$ confidence intervals) but was half the amount predicted, likely due to confounding of the model. The relatively high BC/TOC fractions were explained by variability across sites of BC atmospheric supply, an increase in loss of seagrass litter close to

26   the exposed edges of meadows, and sediment resuspension across the dispersed patchy

27   meadow.

Introduction

The effect of anthropogenic emissions of CO_2 on the climate highlights the importance

of quantifying and managing existing sedimentary organic carbon stocks. Lately, focus has

turned to 'blue carbon' reservoirs – seagrass, saltmarsh, mangrove and macro-algae ecosystems

(1, 2). Together, these store half the ocean's total organic carbon (TOC), despite occupying

less than 2% of the area (3). This disproportionate contribution results from their high rates of

primary production and ability to enhance and stabilise accumulation of deposited litter and

detritus and prevent it from remineralising, ostensibly back to CO_2 (4, 5). Of the four, seagrass

ecosystems are best placed to augment organic carbon stocks. Seagrass canopy can trap soils

eroded from adjacent landscapes, which can account for as much as 50% of total sedimentary

organic stocks (6) that would otherwise be remineralised across the continental shelf (7).

Recently, the traditional mass balance approach to calculating blue carbon storage has

been challenged (8). It is argued that recalcitrant organic carbon produced outside an ecosystem

does not require protection from remineralisation. Consequently, such deposits cannot be

counted as a service in the mitigation of greenhouse gas emissions. Black carbon (BC) is an

example of an 'allochthonous recalcitrant'; it is formed during the incomplete combustion of

biomass and fossil fuels, for which Southeast Asia is a global hotspot (9). However, the BC

content of coastal seagrass sediments is unknown and thus too is the extent of bias in stock

estimates. Nonetheless, predictions of ~18 ± 3% (±95% confidence interval) have been made

using a Southeast Asian (Sabah, Malaysia) estuarine TOC-BC regression model (8). The

model's parameters (i.e. large BC intercept) indicate that high BC/TOC fractions are likely in

relatively low organic content sandy sediments of coastal meadows in regions where

atmospheric BC is pervasive over that of soil erosion (i.e. a small regression slope). The slope

is determined by the accumulated BC within the soil profile and is tempered by the edaphic

organic content of the soil itself (8). Taken together, such regressions give first-order predictive

estimates of meadow BC/TOC.

The accuracy of estuarine model's BC/TOC coastal predictions around this region,

while moderate, may be confounded by failure to take into account the relatively efficient

retention of seagrass litter within estuaries. In more exposed coastal systems, most of the litter

across a meadow or close to its exposed boundary can be lost to shelf waters (10–12). It is also

possible that increased turbulence closer to the exposed edges or across more patchy seagrass

seascapes may lead to localised sediment resuspension. This may further exacerbate the

remineralisation of more labile organic fractions, thereby amplifying the importance of BC to

the TOC stock.

This study examines the above estuarine predictions and likely controlling factors (i.e.

litter loss and sediment resuspension) across coastal meadows in Southeast Asia (Sabah,

Malaysia). We measured the variance in TOC, BC, canopy and sedimentary variables along

transects across and between three subtidal (0.5–1 m) seagrass meadows, two relatively pristine

and one impacted by surrounding developments.

Materials and Methods

Transects were set behind each other and perpendicular to the boundary exposed to

prevailing monsoons. Although replication of the meadows within the relatively pristine

environs of the Tun Mustapha Marine Park, at Limau-Limauan (LL) and Bak-Bak (BB), was

not possible, we controlled for good coverage. Confounding effects were reduced by selecting

extreme examples of canopy size (13) covariant with extremes of wind fetch (14). LL supported

Enhalus acoroides as a large climax species, and BB supported a mix of smaller canopy pioneer

species, including *Cymodocea rotundata* and *Halodule pinifolia* (supplemental Figure S1). The

third meadow, chosen for its sparse, patchy configuration, was located near the University of

Malaysia's Outdoor Development Centre beach (OD), which is under pressure from land

development across the urban environs of Sepanggar Bay (15). The area has a relatively

moderate wind fetch and canopy size. It sustains meadows comprising *Cymodocea serrulata,*

Halodule uninervis and *Halophila ovalis* and isolated small stands of *E. acoroides*, which

assisted in reducing confounding comparisons with the rural systems (supplemental Figure S1).

Details of sediment collection, measurements of canopy and sediment parameters and TOC

and BC analysis are found in the electronic supplementary information.

Results

Meadow parameters

The meadow canopies exhibited clear height differences, attributable to species

characteristics and relative wind fetch (supplemental Figure S1), in the order of LL >> BB <

OD (Table 1). Differences in canopy coverage between meadows LL and BB in the relatively

pristine marine park and OD in the relatively polluted Sepanggar Bay were as expected (LL ≈

BB >> OD, Table 1). Furthermore, significant differences in both height and coverage between

transect pairs was only apparent for OD, reflecting its sparse, patchy configuration.

The most striking difference in sediment characteristics was between LL, where the

longer *Enhalus* canopy species supported higher silt-clay fractions and a significantly smaller

mean particle size, and BB and OD, which were sandier (Table 1). However, these

dissimilarities do not reflect inter-meadow variability between transect pairs, which showed

significant differences only for OD (Table 2).

Differences in both TOC and BC content between meadows (LL << OD > BB) were

inverted compared to expectations regarding percentage of slit-clay content (i.e. LL >> OD <

BB; Table 1). Interestingly, it was noted that the mean particle sizes within OD's patchy stands

were much greater than within adjacent bare patches of the meadow: 8.1±2.4 µm and

107.5±19.0 µm (±95% CI), respectively, but with no statistical significance between their TOC

medians ($P = 0.53$; see supplemental Table S1). For LL and BB but not OD, the outer transect

showed, on average, lower values for both TOC and BC, a difference which became greater

and increasingly significant across LL (Figure 2a, Table 2).

Black carbon to TOC variability

Overall, BC represented a significantly larger fraction of TOC in all three meadows

than in the upper silty-mud and lower silty-sand sediments of the Salut–Mengkabong estuary

(Figure 1). For the urban and rural small-canopy meadows (OD and BB), BC was 28±1.6%

and 36±1.5% (±95% confidence intervals) of TOC, respectively (Table 1). The BC/TOC values

of the large-canopy meadow (LL) were comparable, 26±4.9% (±95% CI; Table 1), despite

large variability. This variability appears to originate with two high TOC and BC outliers

(Figure 2b) located at the start of each transect (see supplemental Table S1). Furthermore, the

larger BC/TOC fractions between the estuary and coastal regions were also reflected in their

adjacent bare sediments.

In all cases, the variability was part of a linear response. For BB and LL, the relationship

between TOC and BC was identical in terms of both their proportional response and the

locations of the intercept close to the origin (Figure 2b). In contrast, BC content in OD was

relatively invariant with increases in TOC [$R^2 = 0.16$, slope = 0.06±0.08 (±95% CI)], resulting

in a BC intercept significantly greater than zero [0.23%±0.168 (±95% CI)]. Furthermore, their

BC/TOC estimates were approximately half than the estuarine model could have predicted (8)

from their mean TOC for the relatively small-canopy meadows, BB and OD [BB: model

17.4±2.7%; measured 36.7±0.8% (±95% CI); OD: [model 13.0±1.8%; measured 30.0±0.8%

(±95% CI)], and more than double for the large-canopy meadow, LL [LL model 46.4±9.4%;

measured 26.1±2.5% (±95% CI)].

Discussion

All the coastal meadow sediments supported greater BC/TOC fractions than those

found in the nearby estuary. Comparisons with the estuarine model predictions, however, were

mixed. As for BB and OD, we expected the BC/TOC fractions to be larger than predicted for

open embayments for a corresponding mean TOC and consistent with a greater loss of the

labile fraction (see Introduction). This was not the case for the larger canopy *Enhalus sp.* of

LL. This disparity may be the result of estuarine model confounding. Unlike the clear waters

and sandier environs of LL, the estuarine *Enhalus sp.* meadow was found solely rooted in

highly organic sediments in the upper and relatively turbid parts of the estuary (8).

The mechanism by which BC/TOC fractions have seemingly been elevated within

coastal embayments can best be distinguished via comparisons of intra-meadow transects in

relation to sedimentary parameters, supported by patterns consistent with the relative canopy

sizes. As expected, the more exposed outer transects for both LL and BB supported lower

sedimentary TOC. Considering that the transect pairs have nearly identical sedimentary

parameters, falling rates of resuspension and subsequent oxidation closer to the meadows'

exposed edges is not a likely explanation. Neither do the contributions of BC from adjacent

bare sediments explain the data patterns. The net deposition of these sediments is likely to be

greater across the inner transects, as turbulence is increasingly attenuated (17). This should

reduce rather than increase the TOC of the inner transects. Thus, a model that describes

increases in BC with TOC, converging towards a positive TOC intercept close to its origin is

more consistent with additional organic carbon not associated with BC such as soil washout.

How much litter and soil make up these meadow's sediments is unknown and beyond the scope

of this study. The alternative explanation is reduction in the loss of deposited litter away from

the exposed edge due to greater attenuation of currents by the canopy. This contention is further

supported, though a weak inference, by a larger difference and increasing significance in the

TOC variance for the larger canopy species of LL.

In contrast, the intra-transect comparison in the patchy OD meadow shows little

evidence of attenuation in TOC or BC across the meadow. Nevertheless, there is evidence of

resuspension, as implied by the increased particle size and decreased silt-clay fraction away

from the exposed edge. Additionally, particle size increased within patches over adjacent

unvegetated areas. This pattern suggests that a sparse, patchy configuration allows sufficient

space for the seagrass leaves to add to turbulence and resuspension (17). For the urban patchy

OD meadow, BC's relatively invariant response with TOC, and its high content is consistent

with; 1) atmospheric deposition at the meadow scale or greater (8); and 2) the suggested greater

levels of turbulence and resuspension that can result in a greater loss of litter and oxidation of

the more labile fractions. Although, the loss of the labile fraction may have been tempered by

a high edaphic TOC contents, the result of a more eutrophic polluted environ. Taken together,

the atmospheric addition of relatively pure BC to the urban meadow over BC supplied and

diluted by soil organics to the rural meadows is a likely reason for the consistently greater

BC/TOC fractions. A difference that was likely compounded by the patchy state of the urban

system over that of the more contiguous rural meadows.

Conclusion

The study demonstrated that allochthonous recalcitrant BC may represent a major

fraction of the sediment TOC content of coastal seagrass meadows, and its underestimation

may lead to significant overestimates of the carbon sink services of such meadows. Within

rural contiguous meadows, the size of seagrass canopy species appears to explain a higher than

predicted BC contribution to sedimentary TOC. In contrast, in the urban meadow under land

development pressure, BC content is largely independent of canopy parameters and TOC

variability, and this meadow supports the highest and most consistent BC/TOC fractions.

Clearly, more work is needed within Southeast Asian BC hotspots, both rural and urban, to

understand the full extent of BC stock bias. We thus recommend moving away from simple

mass balance estimation approaches towards one that includes stability and origin via

incorporation of the concept of allochthonous recalcitrance, whether of BC or other recognised

types.

Supporting data and figures and methods can be found in the electronic supplementary

material. All the authors assisted in fieldwork and analysis of the samples. J.B.G. led the

writing. C.C.H. compiled the supplemental material and the statistical analyses set within the

tables. All authors approved the final version of the manuscript and agree to be accountable for

all aspects of the manuscript. We declare we have no competing interests. Funding was

provided by Sabah Park's Tun Mustapha Park Scientific Expedition (SDK0006-2017), and

Ministry of Science Technology and Innovation (FRGS0424-SG-1/2015).

Figure 1. BC/TOC fractions for seagrass sediments of upper (SMU) and lower (SML) Salut–

Mengkabong Estuary and the coastal seagrass meadows at ODEC beach (OD), Bak-Bak (BB)

and Limau-Limauan (LL). BSM, BOD and BLB indicate the fraction from bare sediments

outside the meadows within the estuary, at OD, and at LL and BB combined, respectively. The

box plot shows the median, 25% and 75% quartiles, 95% confidence limits and outliers. Data

for SMU and SML were compiled from supplementary material (5) in accordance with the

Open Access licence http://creativecommons.org/licenses/by/4.0/.

Figure 2. (**a**) Total organic carbon (TOC) and black carbon (BC) content of meadows at ODEC

beach (OD), Bak Bak (BB) and Limau Limauan (LL) averaged across inner and outer meadow

transects. Error bars represent 95% confidence limits. Significant differences between transect

TOC and BC means were found for LL ($P < 0.001$, $P < 0.001$) and BB ($P < 0.07$, $P < 0.001$).

A power analysis (80%) indicated that determining TOC between BB transects would require

another two or three samples to resolve differences in the mean from a possible Type 1 error.

(**b**) Ordinary least squares regressions for BC with TOC across the three meadows. The

probability of the regression intercepts and slopes between BB and LL are significantly similar

(ANCOVA $P > 0.38$) and not significantly similar to OD (ANCOVA $P < 0.0001$).

Table 1. Mean and standard errors of total organic carbon (TOC) and black carbon (BC)

surface sediment contents (as dry mass) and seagrass canopy and sediment parameters for three

Sabahan seagrass meadows.

Seagrass Meadow	n	TOC (%)	BC (%)	Coverage (%)	Canopy Height (cm)	Particle Size (µm)	Silt-Clay Fraction (%)
ODEC	22	1.07 ± 0.04	0.29 ± 0.01	10 ± 3	5.8 ± 1.7	110.22 ± 3.97	8.98 ± 0.96
Bak-Bak	22	0.71 ± 0.02	0.26 ± 0.01	42 ± 6	3.8 ± 0.6	132.30 ± 1.83	4.09 ± 0.18
Limau-Limauan	22	0.30 ± 0.04	0.08 ± 0.02	33 ± 5	11.5 ± 1.3	74.82 ± 4.65	17.14 ± 2.95

212　**Table 2:** Clay/silt fraction, seagrass coverage, canopy height and mean particle size of each studied transect. Transects labelled T1 correspond to

213　the inner transect, whereas T2 corresponds to the outer transect. †: Calculated using a Mann-Whitney rank sum test as the data were not

214　normally distributed according to the Shapiro-Wilk normality test ($P < 0.050$); central tendency and measure of spread are the median and the

215　interquartile range. *: Calculated using a t-test as the data were normally distributed according to the Shapiro-Wilk normality test ($P \geq 0.050$);

216　central tendency and measure of spread are mean and standard deviation.

Transect Site	Clay/Silt Fraction (%)			Seagrass Coverage (%)			Canopy Height (cm)			Mean Particle Size (µm)		
	Central Tendency	Measure of Spread	P	Central Tendency	Measure of Spread	P	Central Tendency	Measure of Spread	P	Central Tendency	Measure of Spread	P
T1 OD	5.240†	1.480†	<0.001†	13.000†	13.000†	0.333†	2.630†	17.670†	0.162†	126.028*	8.408*	<0.001*
T2 OD	12.880†	3.070†		0.000†	13.000†		0.000†	9.430†		94.414*	10.377*	
T1 BB	3.990*	0.977*	0.602*	63.000†	50.000†	0.129†	4.787*	3.353*	0.117*	136.010†	11.100†	0.101†
T2 BB	4.186*	0.747*		38.000†	63.000†		2.909*	1.800*		128.630†	19.920†	
T1 LL	14.920†	3.520†	0.555†	37.909*	26.950*	0.359*	37.909*	26.950*	0.359*	68.382*	19.004*	0.172*
T2 LL	14.750†	11.410†		28.909*	16.855*		28.909*	16.855*		81.253*	23.396*	

References

1.   McLeod E, Chmura GL, Bouillon S, Salm R, Björk M, Duarte CM, et al. A blueprint

for blue carbon: Toward an improved understanding of the role of vegetated coastal habitats in

sequestering CO2. Frontiers in Ecology and the Environment. 2011;9(10):552-60.

2.   Hill R, Bellgrove A, Macreadie PI, Petrou K, Beardall J, Steven A, et al. Can

macroalgae contribute to blue carbon? An Australian perspective. Limnology and

Oceanography. 2015;60(5):1689-706.

3.   Duarte CM, Middelburg JJ, Caraco N. Major role of marine vegetation on the oceanic

carbon cycle. Biogeosciences. 2005;2(1):1-8.

4.   Christianen MJA, van Belzen J, Herman PMJ, van Katwijk MM, Lamers LPM, van

Leent PJM, et al. Low-Canopy Seagrass Beds Still Provide Important Coastal Protection

Services. PLoS ONE. 2013;8(5):e62413.

5.   Kristensen E, Ahmed SI, Devol AH. Aerobic and anaerobic decomposition of organic

matter in marine sediment: Which is fastest? Limnology and Oceanography. 1995;40(8):1430-

7

6.   Kennedy H, Beggins J, Duarte CM, Fourqurean JW, Holmer M, Marba N, et al.

Seagrass sediments as a global carbon sink: Isotopic constraints. Global Biogeochemical

Cycles. 2010;24, GB4026.

7.   Cathalot C, Rabouille C, Tisnérat-Laborde N, Toussaint F, Kerhervé P, Buscail R, et

al. The fate of river organic carbon in coastal areas: A study in the RhÔne River delta using

multiple isotopic (δ13C, δ14C) and organic tracers. Geochimica Et Cosmochimica Acta.

2013;118:33-55.

8.   Chew ST, Gallagher JB. Accounting for black carbon lowers estimates of blue carbon

storage services. Scientific reports. 2018;8(1):2553.

9.   Gaveau DLA, Salim MA, Hergoualc'h K, Locatelli B, Sloan S, Wooster M, et al. Major

atmospheric emissions from peat fires in Southeast Asia during non-drought years: evidence

from the 2013 Sumatran fires. Scientific reports. 2014;4:6112.

10.   Chiu S-H, Huang Y-H, Lin H-J. Carbon budget of leaves of the tropical intertidal

seagrass Thalassia hemprichii. Estuarine, Coastal and Shelf Science. 2013;125:27-35.

11.   Ricart AM, Dalmau A, Pérez M, Romero J. Effects of landscape configuration on the

exchange of materials in seagrass ecosystems. Marine Ecology Progress Series. 2015;532:89-

100.

12.   Portillo E. Relation between the type of wave exposure and seagrass losses (Cymodocea

nodosa) in the south of Gran Canaria (Canary Islands — Spain). Oceanological and

Hydrobiological Studies. 2014;43(1):29-40.

13.   Oksanen L. Logic of experiments in ecology: is pseudoreplication a pseudoissue?

Oikos. 2001;94(1):27-38.

14.   Kilminster K, McMahon K, Waycott M, Kendrick G, Scanes P, McKenzie L, et al.

Unravelling complexity in seagrass systems for management: Australia as a microcosm.

Science of The Total Environment. 2015;534:97-109.

15.   Freeman AS, Short FT, Isnain I, Razak FA, Coles RG. Seagrass on the edge: Land-use

practices threaten coastal seagrass communities in Sabah, Malaysia. Biological Conservation.

2008;141(12):2993-3005.

16.   Fonseca MS, Cahalan JA. A preliminary evaluation of wave attenuation by four species

of seagrass. Estuarine, Coastal and Shelf Science. 1992;35(6):565-76.

17.   van Katwijk MM, Bos AR, Hermus DCR, Suykerbuyk W. Sediment modification by

seagrass beds: Muddification and sandification induced by plant cover and environmental

conditions. Estuarine, Coastal and Shelf Science. 2010;89(2):175-81.