1 Performance in a novel environment subject to ghost competition.

- 2 <u>Running title:</u> Ghost competition in novel environment
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- 4 Karen Bisschop^{1,2}, Frederik Mortier², Dries Bonte^{$2\ddagger$}, Rampal S. Etienne^{$1\ddagger$}
- 5 ‡ Joint last authorship
- 6 ¹ Groningen Institute for Evolutionary Life Sciences, University of Groningen, P.O. Box
- 7 11103, NL-9700 CC Groningen, The Netherlands
- 8 ² TEREC (Terrestrial Ecology Unit), Department of Biology, Ghent University, Karel
- 9 Lodewijk Ledeganckstraat 35, B-9000 Ghent, Belgium
- 10
- 11 Corresponding author: Karen Bisschop, <u>kbisschop.evo@gmail.com</u>

12 ABSTRACT

13 Background. A central tenet of the evolutionary theory of communities is that competition impacts evolutionary processes such as local adaptation. Species in a community exert a selection 14 pressure on other species and may drive them to extinction. We know, however, very little about 15 16 the influence of unsuccessful or ghost species on the evolutionary dynamics within the community. 17 Methods. Here, we studied the long-term influence of a ghost competitor on the performance of a 18 more successful species using experimental evolution. We transferred the spider mite 19 20 Tetranychus urticae onto a novel host plant under initial presence or absence of a competing

- 21 species, the congeneric mite *T. ludeni*.
- 22 **Results.** The latter species unintentionally went extinct soon after the start of the experiment, but
- 23 we nevertheless completed the experiment and found that the initial density of this ghost
- competitor positively affected the performance (i.e. fecundity) of the more successful species.
- 25 This effect on *T. urticae even* lasted for at least 25 generations.
- 26 **Discussion.** Our study supports the hypothesis that early experienced selection pressures can
- 27 exert a persistent evolutionary signal on species' performance in novel environments.
- 28
- 29 Keywords: interspecific competition, intraspecific competition, experimental evolution, local
- 30 adaptation, spider mites, *Tetranychus urticae*

31 INTRODUCTION

32 Species are facing a continuously changing world that they can possibly cope with in various ways, such as through phenotypic plasticity or by tracking their favoured habitat. If these 33 34 solutions are not possible, evolutionary rescue by genetic adaptation may eventually allow 35 persistence (Lindsey et al., 2013). One factor influencing local adaptation is competition, which can occur among con- or heterospecifics. Niche overlap is usually larger within than among 36 species (Bolnick, 2001; Svanbäck & Bolnick, 2007); therefore, it is expected that competition 37 between conspecifics has a larger impact on evolution than competition at the interspecific level. 38 39 40 Interspecific competition is also known to influence local adaptation, but the effect is still largely unpredictable (Rice & Knapp, 2008; Alzate et al., 2017; Zhao et al., 2018). First, heterospecific 41 42 competitors might modify the selection pressure exerted by the abiotic environment, enhancing or 43 limiting genetic adaptation to the novel environment (Osmond & de Mazancourt, 2013). Classical examples of enhanced genetic adaptation are seen in adaptive radiations of three-spined 44 45 sticklebacks or fast character displacements in Darwin finches or Myzomelid honeyeaters 46 (Diamond et al., 1989; Schluter, 1994; Reznick & Ghalambor, 2001). Previously, we found that additional selection pressure exerted by a congeneric species facilitated adaptation of the focal 47 species to a novel environment under high dispersal from a maladapted ancestral population 48 (Alzate et al., 2017). Adaptation to the novel environment can also be reduced by interspecific 49 50 competition when there is, for instance, a trade-off between traits responsible for adaptation to the competing species and to the novel environment (Siepielski et al., 2016). 51 Furthermore, interspecific competition can create new niches or change the current environment 52

53 for species to adapt to. Species may use waste products or adapt to plants with modified defences

caused by coexisting individuals (Sarmento et al., 2011; Lawrence et al., 2012). These new 54 55 niches will subsequently create opportunities for adaptive shifts to novel environmental conditions. This illustrates that competition and facilitation can jointly shape evolution, making it 56 difficult to study the consequences of interspecific competition alone. 57 58 As a last scenario, interspecific competition can hinder the process of local adaptation by 59 restricting resource availability and hence decrease effective population size. The resulting 60 increased probabilities of genetic drift will then decrease the evolutionary potential and hence the 61 chance of local adaptation (Lawrence et al., 2012; Osmond & de Mazancourt, 2013; Zhao et al., 62 2018). 63 64 While inferior competitors are expected to eventually go extinct, they may coexist with the more successful competitors for many generations (Holmes & Wilson, 1998; Lankau, 2011). These 65 66 early and non-persisting interactions may leave a strong signature on the future community dynamics (Law & Daniel Morton, 1996; Miller, TerHorst & Burns, 2009; Mallon et al., 2018), 67 because they have the possibility to induce large habitat modifications or evolutionary changes in 68 69 the more successful species. Historical contingency (i.e. the influence of the arrival time of a certain species in a community; Fukami 2015) in terms of limitations imposed by so-called ghost 70 71 species (Hawkes & Keitt, 2015), may thus have a strong impact on the eco-evolutionary 72 trajectories of populations and communities, in the same way as successful species do (Fukami, 73 2015). The role of competition intensity of an inferior species prior to its extinction on the 74 ecological and evolutionary dynamics of persisting species is still largely unknown, however. 75

Here we present results from an evolutionary experiment with two related spider mite species 76 77 adapting to a novel host. Both species were placed alone or together on a new host plant and we wanted to verify how this interspecific competition affects local adaptation. The two competitors 78 were supposed to be competitively similar, but the experiment demonstrated that this was not the 79 80 case: both species could only temporarily coexist. This provided us the unique opportunity to 81 investigate the influence of ghost competition on adaptation. More precisely, we explored 82 whether we could detect long-term evolutionary effects on performance (measured as fecundity) due to differences in initial selection pressures caused by this ghost competitor. We chose the 83 average of the initial population size of the unsuccessful species during the first month of 84 85 coexistence as an indication for the initial selection pressure. These differences in population sizes arose naturally and can be attributed to selection, as well as drift and founder effects. 86 T. ludeni showed a lower fecundity than T. urticae on bean and cucumber in the control 87 populations, which may explain their rapid extinction. Still, the ghost species T. ludeni showed an 88 effect on the surviving species T. urticae, because the eventually achieved strength of adaptation 89 of T. urticae increased with the initial density of T. ludeni. We therefore suggest that ghost 90 91 competition is an underestimated process for adaptation and may lead to differences in long-term 92 local adaptation.

93 MATERIAL AND METHODS

94 Study species

95 We used two species of the family Tetranychidae (Acari, Arachnida): *Tetranychus urticae* Koch,

96 1836, and *T. ludeni* Zacher, 1913. These herbivorous mite species are highly suitable for

97 evolutionary experiments due to their small body sizes, their possibility to maintain large

98 populations in the lab, and short generation times (Zhang, 2003).

99 For this study, we used inbred populations of *T. urticae* from Bitume and colleagues (2013). Each

100 population originated from two adult females from the LS-VL line (Van Leeuwen, Stillatus &

101 Tirry, 2004) and was afterwards kept at low population densities. The LS-VL line was collected

102 from roses in October 2000 (Ghent, Belgium). After this initial collection, all populations were

103 maintained on bean plants (*Phaseolus vulgaris*, Prelude).

104 Two populations of *T. ludeni* were used: the Tl Alval (Lisbon, Portugal) and Tl CVM (Lourinhã,

105 Portugal). Both populations were sampled early autumn 2013 from common morning-glory and

afterwards maintained on bean plants (*P. vulgaris*, Prelude). The founder populations were 160

107 and 300 individuals for Tl Alval and Tl CVM respectively. Our evolutionary experiment started

108 in September 2015, implying that *T. urticae* and *T. ludeni* had been under laboratory conditions

109 for about fifteen and two years, respectively.

110

For this study, we chose to subject the inbred lines of *T. urticae* (Bitume et al., 2013) to further inbreeding by mother-son mating for one more generation prior to the experiments. This resulted in the creation of 13 isofemale lines. It may sound counterintuitive to use inbred populations for an evolutionary experiment that mainly uses standing genetic variation (note that no spider mites were added during our experiment), but in this way we could generate genetically similar

replicates and control for putative initial drift effects by differences in starting genetic variation. 116 117 We deemed this more important than potential inbreeding effects, because no effects of inbreeding on genetic trait variation were found in these and other lines (Van Petegem et al., 118 2018; Bonte unpub. results). We additionally created six isofemale lines for T. ludeni (coming 119 120 from Tl Alval and Tl CVM). We wanted to create 13 lines for this species as well, but were unsuccessful due to low fertility or early mortality. The stock Tl Alval and Tl CVM populations 121 122 were placed on bean plants (four two-weeks-old plants) and are from here on referred to as the 123 control population of T. ludeni. A control population of T. urticae was also created from the 124 created 13 isofemale lines (four mites per line) on bean plants (four two-weeks-old plants). All populations were kept in a climate-controlled room $(25^{\circ}C - 30^{\circ}C, 16.8 \text{ L:D})$. 125 126 Experimental set-up 127 128 At the beginning of the actual experiment, the isofemale lines were placed on novel host islands (two three-weeks-old cucumber plants, *Cucumis sativus* Tanja, per island) with or without 129 130 heterospecifics. After the first week, two fresh three-weeks-old cucumber plants were added to 131 create the island size of the experiment. Afterwards, the islands were weekly refreshed by replacing the two oldest plants with two new three-weeks-old cucumber plants. In this way, 132 sufficient time was provided for a generation of spider mites to develop on the new plants, while 133 134 allowing the population to move toward the fresh leaves. Hence, while the removed old plants may have contained mites or unhatched eggs, we chose for this refreshment procedure to 135 136 maintain natural movement dynamics. It is for instance known that especially young fertilised 137 females disperse more (Li & Margolies, 1993) and dispersive individuals may differ in their body condition or performance compared to sedentary individuals (Bonte et al., 2014; Dahirel et al., 138

2019). This refreshment procedure may have caused an extra competitive pressure if one species
was more dispersive or delayed its dispersal for avoiding competition, but we preferred to design
the experiment in a way that it resembled more the actual life strategy of spider mites

142 (colonisation with few founders followed by rapid growth).

143

144 The novel host islands were placed in boxes with yellow sticky paper (Pherobank) at the bottom 145 and Vaseline at the walls to avoid contamination between islands; this method is known to work 146 from previous research (Alzate et al., 2017; Alzate, Etienne & Bonte, 2019; Bisschop et al., 147 2019). Eight islands (or replicates) received both T. urticae and T. ludeni. Eight islands (or replicates) received only T. urticae and another eight islands (or replicates) received only T. 148 149 *ludeni*. Each island started with the same total population size and as similar as possible gene 150 pool. The group with both spider mite species received 26 adult females of T. urticae (two from 151 each of the isofemale lines) and 26 adult females of *T. ludeni* (resulting in 52 adult females). Twelve T. ludeni females per island came from the six isofemale lines and were supplemented 152 153 with 14 mites from its stock population, because of the lack of success for creating more 154 isofemale lines. The use of the outbred stock population of T. ludeni to supplement the populations provided an unanticipated opportunity. In this way, we could benefit from the larger 155 initial genetic variation of T. ludeni among replicates and hence differences in initial selection 156 157 pressures on T. urticae. The group with only T. urticae received four adult females from each of 158 the thirteen isofemale lines (resulting in 52 adult females). The last group with only T. ludeni 159 received four adult females from the six isofemale lines and was supplemented with 28 females 160 from its stock population. We started with a rather low population size to make it biologically relevant as natural populations usually colonise plants at small population sizes. All adult female 161

162	mites were equally distributed over the plants. In an ideal situation, the initial population sizes per
163	species, the total population density, and the island size should be kept equal, but this is of course
164	impossible. We chose for the same total population size and no differences among island sizes, as
165	it is known that differences in densities change both the intra- and interspecific competitive
166	pressure and that an increase in island size would change the adaptive potential of the treatment
167	(Alzate, Etienne & Bonte, 2019). We acknowledge that this necessity of differences in initial
168	population sizes might increase drift effects.
169	
170	The total experiment lasted for ten months, which is about 25 generations and long enough to
171	detect local adaptation (Gould, 1979; Fry, 1989; Magalhães et al., 2007, 2009; Bonte et al.,
172	2010). For logistical reasons the experiment was performed in two blocks with one month
173	difference, each block consisted of four replicates per treatment.
174	
175	Measurements
176	Every two weeks, the density of the spider mites in the evolutionary experiment was measured by
177	counting adult females on a square of 1 x 1 cm ² ; the first counting was done after two weeks. The
178	location of the square was right next to the stalk of the highest, fully grown leaf of the two newest
179	plants of each island. Both the abaxial as well as the adaxial side were measured and summed for
180	a total overview. The location on the leaf was chosen to standardise the measurements in time
181	and make them comparable. The populations of T. ludeni under competition with T. urticae went
182	extinct after about two months. To get an impression of its competitive pressure on the more
183	successful T. urticae populations while it was still present, we used the mean population density

Fecundity tests for the control populations on bean and for the experimental cucumber 185 186 populations were performed every two months to determine the level of adaptation. As the experimental populations of T. ludeni went extinct under competition, we obviously only have 187 results from fecundity tests on the control population of *T. ludeni*. We chose fecundity as proxy 188 189 of adaptation because previous research confirmed it to be the best predictor of adaptation 190 compared to survival or development (Magalhães et al., 2007; Alzate et al., 2017; Alzate, Etienne 191 & Bonte, 2019). Five adult females were sampled from each island and separately placed on a 192 bean leaf disc (17 x 27 mm²) for two generations of common-garden to standardise juvenile and 193 maternal effects (Magalhães et al., 2011; Kawecki et al., 2012). Bean discs were chosen because 194 this is a very suitable host plant and will not cause a change in allele frequencies of the evolved 195 lines (Magalhães et al., 2011). These leaf discs were placed in a petri dish on wet cotton wool and 196 surrounded with paper strip borders. Then, the fecundity of two quiescent deutonymph females 197 that originated from the same common-garden replicate was tested. One female was put on a bean leaf and one on a cucumber leaf (same set-up as for common garden) in a climate cabinet of 198 199 30°C under 16:8 L:D. Fecundity (number of eggs laid after six days) was measured based on 200 daily pictures taken. Females that drowned in the cotton before the sixth day were excluded from the analysis (this was 13.5% for the populations of *T. urticae* without *T. ludeni*, 15% for the 201 populations of T. urticae with the ghost competitor, and 10.5% for the populations of T. urticae 202 203 in the control treatment maintained on bean). The cucumber plants for the fecundity test after 204 four months did not grow for one of our two experimental blocks, so we were not able to test 205 fecundity at that time point. In total, the fecundity was measured for 974 females (exact sample 206 sizes per treatments are given in the electronic supplementary material Table S1).

207

208 <u>Statistical analysis</u>

209	We used general linear mixed models (GLMMs) with Negative Binomial distribution with log
210	link to account for overdispersion of the data. The variance was determined as $\mu * (1 + \mu/k)$ in
211	which μ is the mean and k is the overdispersion parameter (standard negative binomial
212	parametrisation).
213	
214	The dynamics and performance of the ghost competitor
215	We first studied the performance of the control populations that had been maintained on bean of
216	both species on bean and cucumber, using Generalized Linear Mixed Models (GLMM). The
217	dependent variable in the maximal model was fecundity (number of eggs after six days) and the
218	explanatory categorical variables were the plant species (bean or cucumber) and the mite species
219	(<i>T. urticae</i> or <i>T. ludeni</i>). Model selection was based on the lowest AICc and a Wald $\chi 2$ test was
220	performed on the maximal model to check the reliability of the model selection. Pairwise
221	comparisons were adjusted for multiple comparisons with Tukey's method.
222	
223	Signature of the ghost competitor on performance of T. urticae
224	We investigated the impact of the density of <i>T. ludeni</i> and of <i>T. urticae</i> at the onset of the
225	experiment (i.e., mean density during the first month) on the fecundity of <i>T. urticae</i> on its initial
226	and novel host plant. The explanatory variables in the maximal model were time (as categorical
227	variable; 2, 4, 6, 8 or 10 months), the density of <i>T. ludeni</i> (continuous variable), the density of <i>T.</i>
228	urticae (continuous variable), and the interaction between densities of both species and time. In
229	this way, we aimed to determine whether it was the own density or the density of the ghost
230	competitor that affected performance of <i>T urticae</i> . We compared this with an additional model
	11

with the total density (summing the density of *T. ludeni* and *T. urticae*), time, and their 231 232 interaction to find out whether fecundity was affected by the species' individual densities or just the total density. The random effects for all models were the island or replicate nested within the 233 experimental blocks. However, we ran into convergence problems with the results from the 234 235 assessment on cucumber as the random effect variance was estimated to be zero (Magnusson et al., 2018). As a consequence, we only used replicates as random variable. We chose a categorical 236 237 variable for time instead of a continuous one, because differences in quality of leaves at the different measurements were likely and we did not want to assume a linear response of 238 adaptation. Model selection was based on the lowest AICc and an additional Wald $\gamma 2$ test was 239 240 performed on the maximal model. Pairwise comparisons for the slopes and means were adjusted 241 for multiple comparisons with the Tukey's method.

242

243 We furthermore used the density assessed through time of the different spider mite populations to investigate differences in demography between the species with or without competitor. We 244 245 divided the density through time in the first two months, when T. ludeni was still present under 246 competition, and the last eight months. The dependent variable in the maximal model was the density (number of adult female mites per cm^2) and the explanatory variables the treatment (T. 247 urticae with and without competitor, and T. ludeni with and without competitor), time 248 249 (continuous variable; linear for the first two months, second-degree polynomial for the last eight months) and their interaction. The random effects were the different islands or replicates within 250 251 their experimental block. Model selection, Wald χ^2 test, and pairwise comparisons were 252 performed as explained above.

253

254 *Performance without interspecific competitor*

255	Because we were interested in the magnitude of the differences in performance due to the
256	presence of <i>T. ludeni</i> , we did a further analysis including also the control population of <i>T. urticae</i>
257	on bean and the populations of <i>T. urticae</i> on cucumber without <i>T. ludeni</i> . We investigated the
258	fecundity in function of the three different treatments (control of <i>T. urticae</i> on bean, the
259	populations of <i>T. urticae</i> on cucumber without <i>T. ludeni</i> , and those with <i>T. ludeni</i>) and time
260	(categorical variable; 2, 4, 6, 8, or 10 months), and their interaction. The islands were treated as
261	random effects and were nested within the two experimental blocks. Model selection, Wald χ^2
262	test, and pairwise comparisons were performed as explained above.
263	
264	The estimates provided in the tables are the raw and untransformed estimates for the fixed effects
265	of the final models (negative binomial distribution). All analyses were performed in R (version
266	3.6.0) with glmmTMB version 0.2.3 (Brooks et al., 2017), MuMIn version 1.43.6 (Barton, 2019),

emmeans version 1.3.5.1 (Lenth, 2019), and fitdistrplus version 1.0-14 (Delignette-Muller, 2015).

268 **RESULTS**

269 *The dynamics and performance of the ghost competitor*

270	In the competition treatments T. ludeni went extinct after about two months. While T. ludeni was
271	able to maintain a population on cucumber in the absence of a competing species, it reached a
272	significantly lower density than <i>T. urticae</i> (<i>t</i> ratio = 8.535 and <i>p</i> < 0.0001 for <i>T. urticae</i> under
273	ghost competition; <i>t</i> ratio = 9.168 and $p < 0.0001$ for <i>T. urticae</i> without competitor). This
274	suggests that the host plant itself is not a problem for <i>T. ludeni</i> , but that mainly the presence of <i>T</i> .
275	urticae hindered the survival of the population (Fig. 1a, electronic supplementary material Table
276	S2-S4).
277	The fecundity tests on the initial and novel host plant with mites from the control populations of
278	T. urticae and T. ludeni (which had been maintained on bean plants and had never been on
279	cucumber before) showed that <i>T. urticae</i> had a significantly lower fecundity on cucumber than on
280	bean (<i>t</i> ratio = -3.629 and <i>p</i> = 0.0025), and that <i>T. ludeni</i> laid significantly fewer eggs on both
281	bean (<i>t</i> ratio = -7.463 and $p < 0.0001$) and cucumber (<i>t</i> ratio = -6.177 and $p < 0.0001$) than <i>T</i> .
282	<i>urticae</i> . There was no difference in the performance of <i>T. ludeni</i> on bean or cucumber (<i>t</i> ratio = -
283	1.012 and $p = 0.7426$). This suggests that the fecundity of <i>T. ludeni</i> on the novel host was already
284	lower than that of <i>T. urticae</i> at the onset of the experiment (Fig. 1b, Table 1, electronic
285	supplementary material Table S2-4).
206	

286

287 <u>Signature of the ghost competitor on performance of T. urticae</u>

Throughout the evolutionary experiment, we measured the densities of the populations of both spider mite species. During the first month, the ghost competitor (*T. ludeni*) was still present and

the initial density calculated during the first month gave an indication of the pressure exerted by

the ghost species on T. urticae. We found that the initial density of the ghost competitor 291 292 positively affected the fecundity of T. *urticae* on the novel host plant (z value = 2.33 and p =0.0199). This effect emerged from the start of the experiment and remained stable through time 293 as the minimal model did not include time. Performance on the original host plant, bean, was not 294 295 related to initial ghost competitor density. Also, the density of T. urticae itself or the total initial 296 density was not related to performance on both bean and cucumber (Table 1; electronic supplementary material Table S2-S5). 297 298 The initial presence of the ghost competitor influenced the demography of *T. urticae* only 299 slightly. The populations with and without the ghost competitor reached similar equilibrium densities after the ghost competitor went extinct (t ratio = -0.666 and p = 0.7833), and the 300 301 increase during the growth phase did not differ significantly (t ratio = -0.666 and p = 0.7833). At the start no significant difference between densities were found, but after one month the density 302 303 of T. urticae without competition was temporarily significantly higher than the populations of T. *urticae* with *T. ludeni* present (*t* ratio = 3.005 and *p* = 0.0159) (Table 1; electronic supplementary 304 305 material Table S2-S4).

306

307 <u>Performance without interspecific competitor</u>

We additionally compared the performance of mites from a control population that was maintained on bean plants with mites adapting to cucumber where in both cases the interspecific competitor was replaced by conspecifics. The control population had a significantly lower fecundity on cucumber than the populations grown on cucumber (*t* ratio = -3.110 and *p* = 0.0056), which suggests local adaptation to the novel host plant for the latter group (Fig. 2b; Table 1; electronic supplementary information Table S2-S4).

314 DISCUSSION

315	The process of genetic adaptation to novel environmental conditions is typically studied and
316	understood from the perspective of the available genetic variation and selection pressures as
317	imposed by the environment. Because competing species are an intrinsic part of novel
318	experienced environmental conditions, they are known to mediate sometimes complex
319	evolutionary processes. Here we provide empirical evidence that initial competition between two
320	species can have a long-lasting effect on their performance in a novel environment.
321	
322	The unintentional rapid extinction of <i>T. ludeni</i> seems a logical consequence of the higher attained
323	fecundity of <i>T. urticae</i> on the novel host already at the onset of the experiment (Fig. 1b). This
324	higher fecundity and hence higher growth rate increased the chance for better establishment or
325	recovery after disturbance (Turcotte, Reznick & Hare, 2011, 2013). Also, populations from T.
326	urticae lived under a higher density than populations from T. ludeni (Fig. 1a). The density of T.
327	urticae on the measured surface was almost fifty percent more than the density of T. ludeni when
328	grown alone. This suggests that T. urticae has a higher resource efficiency than T. ludeni where
329	the extra energy can be allocated to a higher fecundity. Higher resource efficiency could for
330	instance arise from evolved detoxification mechanisms as often found between herbivores and
331	their hosts (Després, David & Gallet, 2007; Dermauw et al., 2018). After one month the density
332	of populations of T. urticae without heterospecific competition was higher than that of

populations with heterospecific competition, but this difference vanished together with the

extinction of the competitor. This probably means that the ghost competitor decreased the

available resources resulting in a lower population size for *T. urticae* (Fig. 1a).

We propose that differences in selection pressures exerted by the environment cause the 337 338 differences within the treatment under ghost competition, but possibly also between this treatment and the populations without competitor. 339 We have shown that the higher the density of the ghost competitor was, the higher the fecundity 340 341 of T. urticae was on the novel host plant (Fig 2b). We speculate that a higher selection pressure was exerted under a higher initial density of the ghost competitor. This selection pressure 342 eventually led to an increase in fecundity of the focal species. It is known that the competitor, T. 343 344 ludeni, can down-regulate plant defences (Godinho et al. 2016), but this cannot explain the 345 correlation between its higher density and the increased fecundity of the focal species even long 346 after the ghost competitor went extinct, because plants were refreshed weekly. 347 We found that populations of *T. urticae* without *T. ludeni* reached higher fecundity on the novel host plant than the control population maintained on the initial host plant, implying local 348 349 adaptation to the novel host (Fig. 2b). This difference in fecundity with the control population was not found for the populations that were initially under competition with *T. ludeni*. A likely 350 351 explanation for the lack of evidence for adaptation in the treatment under interspecific 352 competition would be the difference in initial population size and subsequent drift effects. Under

a scenario of strong drift effects we would expect large differences in performance between

replicates; therefore we plotted the fecundity results from the last measured month per replicate

355 (Fig. 3). Since the differences in fecundity assessed on cucumber among replicates are similar

356 with and without *T. ludeni* (and thus with high or low initial population size for *T. urticae*). This

357 indicates that strong drift can be largely discarded as a driver behind the observed evolutionary

358 dynamics.

359

The selection pressure of conspecifics was higher than heterospecifics, which may be because an 360 361 initial larger population sizes, but also due to a larger niche overlap (Bolnick, 2001; Svanbäck & Bolnick, 2007). Furthermore, intraspecific competition is known to push towards ecological 362 specialisation and may thus enhance adaptation to novel host plants (Silvertown, 2004). 363 364 Intraspecific effects are found to be particularly strong in those cases where communities are 365 affected by indirect interactions such as cascading effects (Des Roches et al., 2018). Many 366 examples are found where indirect interactions are induced after a herbivore attack (e.g. the release of volatiles to attract natural enemies, activation or production of toxins or defensive 367 structures; Zhang et al., 2009; Kant et al., 2015). Given that we used entire plants, such indirect 368 369 interactions between herbivores and their host plants likely took place in our experiment. 370 The history of species in a community can have an impact on interspecific interactions (Fukami, 371 372 2015). The magnitudes of such historical contingencies do, however, strongly differ among species and environments (Vannette & Fukami, 2014). Differences in historical contingency have 373 374 been put forward as an explanation for the fact that some populations can experience radiations, 375 whereas others from the same clade are not capable to achieve this under seemingly similar conditions (Seehausen, 2007). Our results suggest that increased interspecific competition leads 376 to higher selection pressures and thus improved performance (Fig. 2b). Our results coincide in 377 378 this respect with other empirical work demonstrating that increased competition with 379 heterospecifics increased local adaptation in bunchgrasses (Rice & Knapp, 2008). Similarly, 380 intraguild predation between lizard species increased the selection pressure and led to strong divergence in morphological adaptation as associated with niche specialisation (Stuart et al., 381 2014). 382

384	Nevertheless, we have to be careful with generalising our results. First, we chose small
385	populations sizes as they are more biologically relevant, but this may limit adaptation and
386	establishment (Del Castillo et al., 2011; Yates & Fraser, 2014). We also used populations that
387	have been maintained in the lab for many generations, probably leading to a decrease in genetic
388	variation compared to wild populations. The problems we encountered to create isofemale lines
389	for <i>T. ludeni</i> could be an indication of inbreeding depression. However, we are confident that our
390	results are robust as we could still provide evidence for local adaptation in the populations of T .
391	urticae without competitor (Fig. 2b). This suggests that the initial amount of genetic variation did
392	not limit <i>T. urticae</i> in our study.
393	Second, it is impossible to add a competitor without changing total population sizes, population
394	densities, or island sizes; all of these are affecting genetic variation and drift (Del Castillo et al.,
395	2011; Alzate, Etienne & Bonte, 2019). As it is known that larger populations usually contain
396	more genetic variation, we chose to standardise this by means of isofemale lines, knowing that
397	this might create differences in drift among treatments. One way to better disentangle the effects
398	of drift from those of selection with our small population size would have been to increase the
399	number of replicates which was difficult for logistical reasons.
400	Third, our experimental design is not strictly suitable to assess adaptation in the interspecific
401	competition treatment, as we did not keep a control mixed population on bean. Hence, we cannot
402	disentangle the effect of changes in fecundity due to competition (independent of the novel
403	environment) from the effect of competition on adaptation to the novel environment. Although
404	this means that we cannot detect adaptation in the treatment under ghost competition, we did find

405	a positive influence of the density of the ghost competitor on fecundity, meaning that initial
406	selection pressures substantially matter and providing evidence for eco-evolutionary dynamics.
407	
408	In conclusion, we did find indications for local adaptation in the populations without ghost
409	competition as the performance increased on the novel host compared to a control population.
410	Furthermore, we have shown the importance of initial selection pressures such as ghost
411	competition. Even when one species becomes extinct, the competition signature continues to
412	affect the adaptation process of the successful species. We thus provide experimental evidence on
413	the impact of ghost species on the long-term performance of populations colonizing new

414 environments.

415 ACKNOWLEDGEMENTS

- 416 We thank Viki Vandomme, Angelica Alcantara, Pieter Vantieghem, Katrien Van Petegem,
- 417 Stefano Masier, Matti Pisman, Mike Creutz, Hilde De Nil and Johan Bisschop for helping during
- the research experiments, and to Sarah Magalhães for providing the strains of *T. ludeni*. We thank
- the Terrestrial Ecology department of Ghent University and the Centre for Ecology, Evolution
- 420 and Environmental Changes of the University of Lisbon for the spider mite populations. R.S.E.
- 421 thanks the Netherlands Organisation for Scientific Research (NWO) for financial support through
- 422 a VICI grant (VICI grant number 865.13.00). K.B. thanks the Special Research Fund (BOF) of
- 423 Ghent University and the Ubbo Emmius sandwich program of the University of Groningen. DB
- and RSE are supported by the FWO network EVENET (W0.003.16N) and FWO research grant

425 G018017N.

- 426
- 427 The authors declare no conflicts of interest.

428 **REFERENCES**

- 429 Alzate A, Bisschop K, Etienne RS, Bonte D. 2017. Interspecific competition counteracts negative
- 430 effects of dispersal on adaptation of an arthropod herbivore to a new host. *Journal of*
- 431 *Evolutionary Biology* 30:1966–1977. DOI: 10.1111/jeb.13123.
- 432 Alzate A, Etienne RS, Bonte D. 2019. Experimental island biogeography demonstrates the
- 433 importance of island size and dispersal for the adaptation to novel habitats. *Global Ecol*
- 434 *Biogeogr* 28:238–247. DOI: 10.1111/geb.12846.
- 435 Barton K. 2019.MuMIn: multi-model inference. R package version 1.43.6. Available at
- 436 *https://cran.r-project.org/package=MuMIn*
- 437 Bisschop K, Mortier F, Etienne RS, Bonte D. 2019. Transient local adaptation and source–sink
- 438 dynamics in experimental populations experiencing spatially heterogeneous environments.
- 439 *Proceedings of the Royal Society B: Biological Sciences* 286:20190738. DOI:
- 440 10.1098/rspb.2019.0738.
- 441 Bitume E V., Bonte D, Ronce O, Bach F, Flaven E, Olivieri I, Nieberding CM. 2013. Density and
- 442 genetic relatedness increase dispersal distance in a subsocial organism. *Ecology Letters*
- 443 16:430–437. DOI: 10.1111/ele.12057.
- 444 Bolnick DI. 2001. Intraspecific competition favours niche width expansion in Drosophila
- 445 melanogaster. *Nature* 410:463–466. DOI: 10.1038/35068555.
- Bonte D, De Roissart A, Vandegehuchte ML, Ballhorn DJ, van Leeuwen T, de la Peña E. 2010.
- 447 Local adaptation of aboveground herbivores towards plant phenotypes induced by soil biota.
- 448 *PLoS ONE* 5:e11174. DOI: 10.1371/journal.pone.0011174.
- Bonte D, De Roissart A, Wybouw N, Van Leeuwen T. 2014. Fitness maximization by dispersal:
- 450 evidence from an invasion experiment. *Ecology* 95:3104–3111. DOI: 10.1890/13-2269.1.

451	Brooks ME, Kristensen K, van Benthem KJ, Magnusson A, Berg CW, Nielsen A, Skaug HJ,
452	Mächler M, Bolker BM. 2017. glmmTMB balances speed and flexibility among packages
453	for zero-inflated generalized linear mixed modeling. The R Journal 9:378–400. DOI:
454	10.3929/ETHZ-B-000240890.
455	Del Castillo RF, Trujillo-Argueta S, Sánchez-Vargas N, Newton AC. 2011. Genetic factors
456	associated with population size may increase extinction risks and decrease colonization
457	potential in a keystone tropical pine. Evolutionary Applications 4:574–588. DOI:
458	10.1111/j.1752-4571.2010.00177.x.
459	Dahirel M, Masier S, Renault D, Bonte D. 2019. The distinct phenotypic signatures of dispersal
460	and stress in an arthropod model: from physiology to life history. The Journal of
461	experimental biology 222:jeb203596. DOI: 10.1242/jeb.203596.
462	Delignette-Muller ML. 2015. fitdistrplus: an R package for fitting distributions. Journal of
463	Statistical Software 64:1–34. DOI: 10.18637/jss.v064.i04.
464	Dermauw W, Pym A, Bass C, Van Leeuwen T, Feyereisen R. 2018. Does host plant adaptation
465	lead to pesticide resistance in generalist herbivores? Current Opinion in Insect Science
466	26:25–33. DOI: 10.1016/j.cois.2018.01.001.
467	Després L, David JP, Gallet C. 2007. The evolutionary ecology of insect resistance to plant
468	chemicals. Trends in Ecology and Evolution 22:298–307. DOI: 10.1016/j.tree.2007.02.010.
469	Diamond J, Pimm SL, Gilpin ME, LeCroy M. 1989. Rapid evolution of character displacement in
470	Myzomelid honeyeaters. The American Naturalist 134:675–708. DOI: 10.1086/285006.
471	Fry JD. 1989. Evolutionary adaptation to host plants in a laboratory population of the
472	phytophagous mite Tetranychus urticae Koch. Oecologia 81:559–565. DOI:
473	10.1007/BF00378969.
	23

- 474 Fukami T. 2015. Historical contingency in community assembly: integrating niches, species
- 475 pools, and priority effects. *Annual Review of Ecology, Evolution, and Systematics* 46:1–23.
- 476 DOI: 10.1146/annurev-ecolsys-110411-160340.
- 477 Godinho DP, Janssen A, Dias T, Cruz C, Magalhães S. 2016. Down-regulation of plant defence
- in a resident spider mite species and its effect upon con- and heterospecifics. *Oecologia*
- 479 180:161–167. DOI: 10.1007/s00442-015-3434-z.
- Gould F. 1979. Rapid host range evolution in a population of the phytophagous mite *Tetranychus*
- 481 *urticae* Koch. *Evolution* 33:791–802. DOI: 10.1111/j.1558-5646.1979.tb04735.x.
- 482 Hawkes C V., Keitt TH. 2015. Resilience vs. historical contingency in microbial responses to
- 483 environmental change. *Ecology Letters* 18:612–625. DOI: 10.1111/ele.12451.
- Holmes EE, Wilson HB. 1998. Running from trouble: long-distance dispersal and the competitive
 coexistence of inferior species. *The American Naturalist* 151:578–586. DOI:
- 486 10.1086/286143.
- 487 Kant MR, Jonckheere W, Knegt B, Lemos F, Liu J, Schimmel BCJ, Villarroel CA, Ataide LMS,
- 488 Dermauw W, Glas JJ, Egas M, Janssen A, Van Leeuwen T, Schuurink RC, Sabelis MW,
- 489 Alba JM. 2015. Mechanisms and ecological consequences of plant defence induction and
- 490 suppression in herbivore communities. *Annals of Botany* 115:1015–1051. DOI:
- 491 10.1093/aob/mcv054.
- 492 Kawecki TJ, Lenski RE, Ebert D, Hollis B, Olivieri I, Whitlock MC. 2012. Experimental
- 493 evolution. *Trends in Ecology and Evolution* 27:547–560. DOI: 10.1016/j.tree.2012.06.001.
- 494 Lankau RA. 2011. Rapid evolutionary change and the coexistence of species. *Annual Review of*
- 495 *Ecology, Evolution, and Systematics* 42:335–54. DOI: 10.1146/annurev-ecolsys-102710-
- 496 145100.

- 497 Law R, Daniel Morton R. 1996. Permanence and the assembly of ecological communities.
- 498 *Ecology* 77:762–775. DOI: 10.2307/2265500.
- Lawrence D, Fiegna F, Behrends V, Bundy JG, Phillimore AB, Bell T, Barraclough TG. 2012.
- 500 Species interactions alter evolutionary responses to a novel environment. *PLoS Biology*
- 501 10:e1001330. DOI: 10.1371/journal.pbio.1001330.
- 502 Van Leeuwen T, Stillatus V, Tirry L. 2004. Genetic analysis and cross-resistance spectrum of a
- 503 laboratory-selected chlorfenapyr resistant strain of two-spotted spider mite (Acari:
- 504 Tetranychidae). *Experimental and Applied Acarology* 32:249–261. DOI:
- 505 10.1023/B:APPA.0000023240.01937.6d.
- Lenth R. 2019.emmeans: estimated marginal means, aka least-squares means. R package version

507 1.3.5.1. Available at https://cran.r-project.org/package=emmeans

- Li J, Margolies DC. 1993. Effects of mite age, mite density, and host quality on aerial dispersal
- 509 behavior in the twospotted spider mite. *Entomologia Experimentalis et Applicata* 68:79–86.
- 510 DOI: 10.1111/j.1570-7458.1993.tb01691.x.
- Lindsey HA, Gallie J, Taylor S, Kerr B. 2013. Evolutionary rescue from extinction is contingent
- on a lower rate of environmental change. *Nature* 494:463–467. DOI: 10.1038/nature11879.
- 513 Magalhães S, Blanchet E, Egas M, Olivieri I. 2009. Are adaptation costs necessary to build up a
- 514 local adaptation pattern? *BMC Evolutionary Biology* 9:182. DOI: 10.1186/1471-2148-9-182.
- 515 Magalhães S, Blanchet E, Egas M, Olivieri I. 2011. Environmental effects on the detection of
- adaptation. Journal of Evolutionary Biology 24:2653–2662. DOI: 10.1111/j.1420-
- 517 9101.2011.02388.x.
- 518 Magalhães S, Fayard J, Janssen A, Carbonell D, Olivieri I. 2007. Adaptation in a spider mite
- 519 population after long-term evolution on a single host plant. *Journal of Evolutionary Biology*

520 20:2016–2027. DOI: 10.1111/j.1420-9101.2007.01365.x.

- 521 Magnusson A, Skaug H, Nielsen A, Berg CW, Kristensen K, Maechler M, van Bentham K,
- 522 Bolker BM, Brooks ME. 2018. Troubleshooting with glmmTMB. Available at https://cran.r-
- 523 project.org/web/packages/glmmTMB/vignettes/troubleshooting.html
- 524 Mallon CA, Le Roux X, Van Doorn GS, Dini-Andreote F, Poly F, Salles JF. 2018. The impact of
- 525 failure: unsuccessful bacterial invasions steer the soil microbial community away from the
- 526 invader's niche. *ISME Journal* 12:728–741. DOI: 10.1038/s41396-017-0003-y.
- 527 Miller TE, TerHorst CP, Burns JH. 2009. The ghost of competition present. The American
- 528 *Naturalist* 173:347–353. DOI: 10.1086/596531.
- 529 Osmond MM, de Mazancourt C. 2013. How competition affects evolutionary rescue.
- 530 Philosophical transactions of the Royal Society of London B, Biological sciences
- 531 368:20120085. DOI: 10.1098/rstb.2012.0085.
- 532 Van Petegem K, Moerman F, Dahirel M, Fronhofer EA, Vandegehuchte ML, Van Leeuwen T,
- 533 Wybouw N, Stoks R, Bonte D. 2018. Kin competition accelerates experimental range
- expansion in an arthropod herbivore. *Ecology Letters* 21:225–234. DOI: 10.1111/ele.12887.
- 535 Reznick DN, Ghalambor CK. 2001. The population ecology of contemporary adaptations: what
- empirical studies reveal about the conditions that promote adaptive evolution. *Genetica*
- 537 112:183–198. DOI: 10.1023/A:1013352109042.
- 538 Rice KJ, Knapp EE. 2008. Effects of competition and life history stage on the expression of local
- adaptation in two native bunchgrasses. *Restoration Ecology* 16:12–23. DOI: 10.1111/j.1526-
- 540 100X.2007.00257.x.
- 541 Des Roches S, Post DM, Turley NE, Bailey JK, Hendry AP, Kinnison MT, Schweitzer JA,
- 542 Palkovacs EP. 2018. The ecological importance of intraspecific variation. *Nature Ecology*

543 *and Evolution* 2:57–64. DOI: 10.1038/s41559-017-0402-5.

- 544 Sarmento RA, Lemos F, Bleeker PM, Schuurink RC, Pallini A, Oliveira MGA, Lima ER, Kant
- 545 M, Sabelis MW, Janssen A. 2011. A herbivore that manipulates plant defence. *Ecology*
- 546 *Letters* 14:229–236. DOI: 10.1111/j.1461-0248.2010.01575.x.
- 547 Schluter D. 1994. Experimental evidence that competition promotes divergence in adaptive
- radiation. *Science* 266:798–801. DOI: 10.1126/science.266.5186.798.
- 549 Seehausen O. 2007. Evolution and ecological theory: chance, historical contingency and
- ecological determinism jointly determine the rate of adaptive radiation. *Heredity* 99:361–
- 551 363. DOI: 10.1038/sj.hdy.6801047.
- 552 Siepielski AM, Nemirov A, Cattivera M, Nickerson A. 2016. Experimental evidence for an eco-
- evolutionary coupling between local adaptation and intraspecific competition. *The American Naturalist* 187:447–456. DOI: 10.1086/685295.
- 555 Silvertown J. 2004. The ghost of competition past in the phylogeny of island endemic plants.

Journal of Ecology 92:168–173. DOI: 10.1111/j.1365-2745.2004.00853.x.

- 557 Stuart YE, Campbell TS, Hohenlohe PA, Reynolds RG, Revell LJ, Losos JB. 2014. Rapid
- evolution of a native species following invasion by a congener. *Science* 346:463–466. DOI:
 10.1126/science.1257008.
- Svanbäck R, Bolnick DI. 2007. Intraspecific competition drives increased resource use diversity
 within a natural population. *Proceedings of the Royal Society B: Biological Sciences*
- 562 274:839–844. DOI: 10.1098/rspb.2006.0198.
- 563 Turcotte MM, Reznick DN, Hare JD. 2011. The impact of rapid evolution on population
- 564 dynamics in the wild: experimental test of eco-evolutionary dynamics. *Ecology Letters*
- 565 14:1084–1092. DOI: 10.1111/j.1461-0248.2011.01676.x.

- 566 Turcotte MM, Reznick DN, Hare JD. 2013. Experimental Test of an Eco-Evolutionary Dynamic
- 567 Feedback Loop between Evolution and Population Density in the Green Peach Aphid. *The*
- 568 *American Naturalist* 181:S46–S57. DOI: 10.1086/668078.
- 569 Vannette RL, Fukami T. 2014. Historical contingency in species interactions: towards niche-
- 570 based predictions. *Ecology Letters* 17:115–124. DOI: 10.1111/ele.12204.
- 571 Yates MC, Fraser DJ. 2014. Does source population size affect performance in new
- 572 environments? *Evolutionary Applications* 7:871–882. DOI: 10.1111/eva.12181.
- 573 Zhang ZQ. 2003. Mites of greenhouses: identification, biology and control. Wallingford: CABI
- 574 Publishing. DOI: 10.1079/9780851995908.0000.
- 575 Zhang P-J, Zheng S-J, van Loon JJA, Boland W, David A, Mumm R, Dicke M. 2009. Whiteflies
- interfere with indirect plant defense against spider mites in Lima bean. *Proceedings of the*
- 577 *National Academy of Sciences* 106:21202–21207. DOI: 10.1073/pnas.0907890106.
- 578 Zhao XF, Buckling A, Zhang QG, Hesse E. 2018. Specific adaptation to strong competitors can
- 579 offset the negative effects of population size reductions. *Proceedings of the Royal Society B:*
- 580 *Biological Sciences* 285:20180007. DOI: 10.1098/rspb.2018.0007.

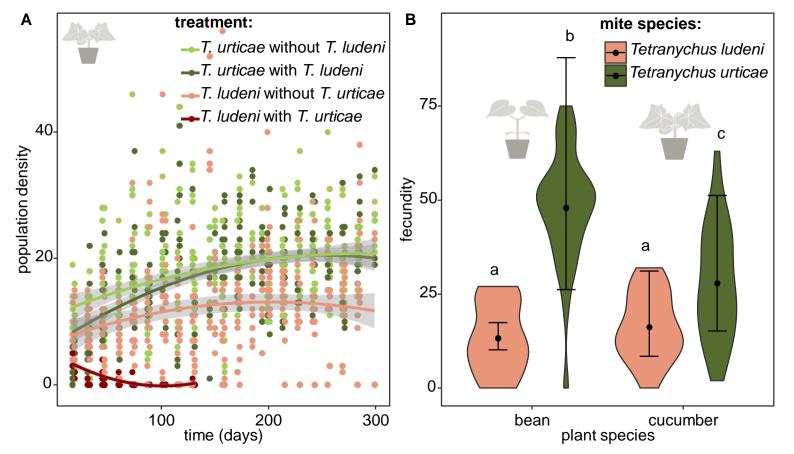
582 FIGURES

583 Figure 1: a) Overview of the population density for the different treatments. Population density of T. urticae (green dots) and T. ludeni (red dots) measured as the sum of the abaxial and adaxial 584 585 density (number of adult females/cm²) per island through time. The lighter colours correspond to 586 the populations in absence of the competing species and the darker to the treatment where both species are present. The grey zone shows the 95% confidence interval. b) Comparison of control 587 populations of *Tetranychus ludeni* and *T. urticae* on bean and cucumber. The fecundity of *T.* 588 ludeni is significantly lower than T. urticae, on both bean and cucumber. The violin plots show 589 590 the observed data, and the points and lines show the mean model estimates and their 95% 591 confidence interval, respectively.

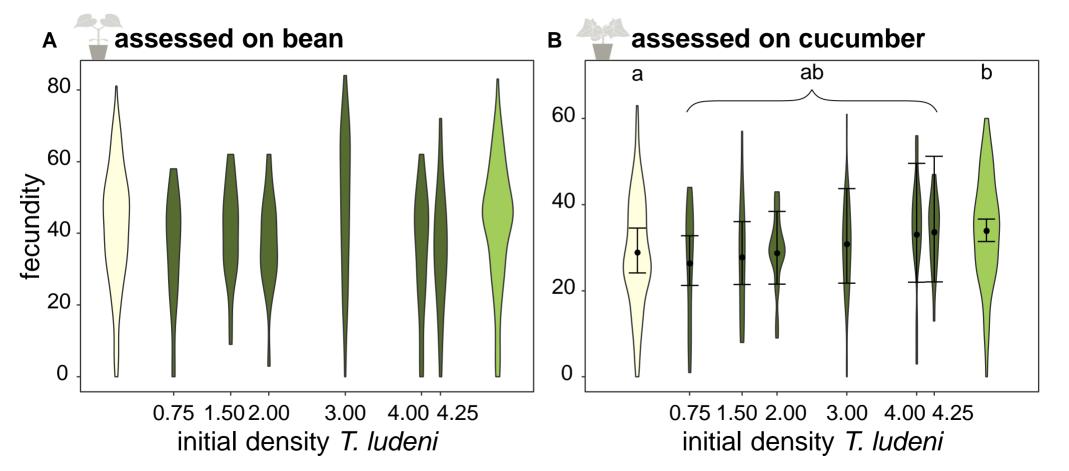
592

593 Figure 2: Fecundity affected by ghost competition. On the x-axis the initial density of T. ludeni 594 (number of adult females/cm²) and the different treatments (T. urticae from cucumber but without T. ludeni, T. urticae with ghost competition of T. ludeni, and the control population of T. urticae 595 from bean) are presented and on the y-axis the fecundity (number of eggs after six days) of T. 596 597 *urticae*. This fecundity is averaged over time as this variable was not included in the minimal 598 model. A significant interaction between the density of T. ludeni and the fecundity of T. urticae 599 was not found when the fecundity was assessed on (A) bean, but it was found when assessed on 600 (B) cucumber. Each violin plot is presenting the observed data, while the points and lines show 601 the means of the model estimate and their 95% confidence interval, respectively.

- **Figure 3:** Comparison of fecundity measured after ten months between replicates. The different
- replicates are given on the x axis and the fecundity (number of eggs after six days) on the y axis.
- The coloured points are the real measurements, while the black dot and lines present the mean,
- 605 minimum and maximum value respectively. The upper plots are measurements from *T. urticae* on
- bean with A) the treatment with *T. ludeni* and B) the treatment without *T. ludeni*. The lower plots
- are measurements from *T. urticae* on cucumber with C) the treatment with *T. ludeni* and D) the
- 608 treatment without *T. ludeni*.



treatment: T. urticae (control) T. urticae (comp.) T. urticae (no comp.)



T. urticae assessed on bean

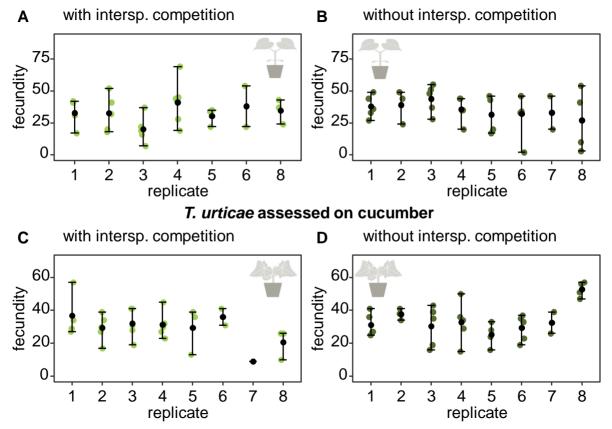


Table 1:

Summary of the final GLMM explaining reproductive performance.

	Estimate	SE	z value	<i>P</i> value	
The dynamics and performance of the g	ghost competito	or			
(Intercept) (T. ludeni on bean)	2.5867	0.1367	18.92	<2e-16	***
Cucumber	0.1990	0.1965	1.01	0.3114	
T. urticae	1.2838	0.1720	7.46	8.47e-14	***
Cucumber : T. urticae	-0.7407	0.2468	-3.00	0.0027	**
Signature of the ghost competitor on pe	rformance of 7	. urticae			
Fecundity assessed on bean					
(Intercept)	3.7098	0.0899	41.27	<2e-16	
Fecundity assessed on cucumber					
(Intercept)	3.2209	0.0883	36.47	<2e-16	***
Initial density T. ludeni	0.0693	0.0298	2.33	0.0199	*
Density during first two months					
(Intercept) (T. ludeni under comp.)	1.6678	0.3132	5.32	1.01e-07	***
T. urticae under comp.	-0.1136	0.3814	-0.30	0.7658	
T. urticae without comp.	0.2925	0.3744	0.78	0.4346	
<i>T. ludeni</i> without comp.	0.2669	0.3818	0.70	0.4845	
Time	-0.0275	0.0099	-2.79	0.0053	**
<i>T. urticae</i> comp. : time	0.0470	0.0117	4.00	6.26e-05	***
T. urticae no comp. : time	0.0454	0.0114	3.97	7.35e-05	***
<i>T. ludeni</i> no comp. : time	0.0351	0.0117	3.01	0.0027	**
Density during last eight months					
(Intercept) (<i>T. urticae</i> under comp.)	2.9066	0.0326	89.25	<2e-16	***
T. urticae without comp.	0.0307	0.0462	0.67	0.5054	
T. ludeni without comp.	-0.4023	0.0471	-8.54	<2e-16	***
Time	2.3787	0.5468	4.35	1.36e-05	***
Time ²	-0.9981	0.5520	-1.81	0.0706	•
Performance without interspecific comp	petitor				
(Intercept) T. urticae without comp.	3.5241	0.0395	89.23	<2e-16	***
T. urticae under comp.	-0.1119	0.0515	-2.17	0.0297	*
<i>T. urticae</i> control	-0.1605	0.0516	-3.11	0.0019	**