

17 **Abstract:** Recycling effluent has become a bottleneck and an environmental risk associated with
18 the regular production of bauxite via flotation and the sustainable development of the aluminum
19 industry in China. To find a practical direction for biotreatment, the bacterial and archaeal
20 communities in recycling effluents containing concentrate and tailings from bauxite flotation
21 plants were investigated by a metagenomic sequencing method in association with the evaluated
22 geochemical properties. The results showed that *Paracoccus*, *Desulfomicrobium*,
23 *Exiguobacterium*, *Tindallia*, *Ercella* and *Anoxynatronum* were the primary bacterial genera and
24 *Methanothrix*, *Methanobacterium*, *Nitrososphaera* and *Methanosarcina* were the dominant
25 archaeal genera. Upon combining the microbial diversity and the geochemical properties of the
26 two sample types, the microbial community containing *Desulfomicrobium*, *Paracoccus*, *Tindallia*,
27 *Methanobacterium*, *Methanothrix* and *Nitrososphaera* was better adapted to the biodegradation of
28 flotation collectors, and the microbial community consisting of *Paracoccus*, *Exiguobacterium*,
29 *Methanothrix* and *Methanobacterium* was more efficient at hydrolyzed polyacrylamide (HPAM)
30 biodegradation. In addition, a large proportion of unclassified OTUs has indicated that recycling
31 effluent is a worthy resource for isolating new strains from the Firmicutes phylum.

32

33 **1 Introduction**

34 China has been the world's leading producer of metallurgical-grade alumina over
35 the last 16 years(1). The operational yield of the Chinese alumina industry in 2017
36 exceeded 69 million tons, with a year-on-year growth of 13.32%, which resulted in
37 the rapid depletion of high A/S (mass ratio of alumina to silica) bauxite as well as a
38 great challenge to Chinese bauxite resources(2). Bauxite desilication by flotation,

39 which could be used for low A/S bauxite that is found in large reserves in China,
40 became an important technical method for the sustainable development of the
41 aluminum industry in China(3, 4)

42 A large amount of water is needed for the bauxite desilication process, and a large
43 quantity of water can be saved by water recycling through flocculation and the rapid
44 settlement of the pulp(5). However, the production index becomes worse with the
45 increased circulation time of the circulating water; the periodic replacement of
46 circulating water was the primary method for solving this problem in bauxite flotation
47 plants, resulting in a problem of recycling effluents. These recycling effluents
48 contained high concentrations of ions and the residues of beneficiation reagents, such
49 as collectors and flocculants(6, 7). The periodic discharge of recycling effluents
50 would have potential environmental risks, but little study has been devoted to this
51 approach. However, with the rapid economic growth and social development in China,
52 the associated problems of water contamination and shortages of water resources have
53 become serious bottlenecks that challenge and limit sustainable development at the
54 regional as well as national levels(8, 9). Finding an efficient and safe method for
55 treating the recycling effluents, such as biological methods, is an urgent task.

56 Because of the importance of bauxite desilication by flotation for the alumina
57 industry of China and with the increased use of water resources, the microbial
58 communities in the recycling effluents from desilication processing were revealed and
59 linked with the chemical and geographical conditions to create effective advice for the
60 prevention and treatment of these problems. The compositions, diversity and

61 dynamics of the given microorganisms may interfere in the efficiency and stability of
62 wastewater treatment plants(10). Therefore, studies about the microbial community in
63 these systems have been focused on the functions and performance of wastewater
64 treatment(11). In this study, recycling effluents from the desilication process were
65 collected before being discharged at the bauxite flotation plant in Henan province, and
66 the community structures were assayed using 16S rRNA genes by high-throughput
67 sequencing method.

68 **2 Materials and methods**

69 **2.1 Site description and sample collection**

70 Samples of recycling effluent were collected from the bauxite flotation plant
71 located in Henan province. Recycling effluents of concentrates and tailings were
72 collected from a circulating cistern (Fig 1), and they were named HNC and HNT,
73 respectively.

74 **Fig 1** Flow chart of desilication process and sample collection

75 Water samples were collected in July of 2018. Approximately 25 L of recycling
76 effluent from each circulating cistern was collected, and the temperature and pH of
77 each sample were detected at the same time. The samples were then filtered through a
78 0.22 μm hyperfiltration membrane with a vacuum pump. The sediments on the
79 membranes were then immediately transferred to a tube and stored at -20°C for further
80 molecular analysis. The filtered water samples were sent for chemical analysis.

81 **2.2 Geochemical analysis**

82 We measured the concentrations of 28 elements, including Al, As, Ca, Cr, Cu, Fe,

83 K, Mn, Na, Pd, Zn and other elements, by Inductively Coupled Plasma Mass
84 Spectroscopy (ICP-MS, G8403A, Japan).

85 **2.3 DNA extraction, PCR and sequencing**

86 The DNA in the water samples was extracted from 200 mg of sediment with an
87 E.Z.N.A. Soil DNA Kit (OMEGA, USA). The DNA integrity was detected by agarose
88 gel electrophoresis. Accurate quantitative control of the genomic DNA added to the
89 PCR reaction system was achieved after the determination of the genomic DNA with
90 a Qubit 3.0 DNA detection kit.

91 The V3-V4 region(12)of the 16S rRNA genes was amplified with the primer pair
92 341F (5' -CCCTACACGACGCTCTTCCGATCTG (barcode) CCTACGGGNGGC
93 WGCAGA-3') and 806R (5' -
94 GACTGGAGTTCCTTGGCACCCGAGAATTCCAGA
95 CTACHVGGGTATCTAATCC-3') combined with the universal primers on the
96 MiSeq sequencing platform.

97 The PCR reactions of the first round had a total volume of 30 μ l, which contained
98 10-20 ng of genomic DNA, 15 μ l of 2 \times Taq master Mix, 1 μ l of Bar-PCR primer F (10
99 μ M), 1 μ l of primer R (10 μ M), and the remaining volume of H₂O. The cycling
100 conditions of the first round were 94°C for 3 min, 5 cycles of 94°C for 30 s, 45°C for
101 20 s, and 65°C for 30 s, 20 cycles of 94°C for 20 s, 55°C for 20 s, and 72°C for 30 s,
102 and a final extension at 72°C for 5 min. The PCR reactions of the second round had a
103 total volume of 30 μ l, and they contained 20 ng of genomic DNA, 15 μ l of 2 \times Taq
104 master Mix, 1 μ l of primer F (10 μ M), 1 μ l of primer R (10 μ M), and the remaining

105 volume of H₂O. The cycling conditions of the second round were 95°C for 3 min, 5
106 cycles of 94°C for 20 s, 55°C for 20 s, and 72°C for 30 s, and a final extension at 72°C
107 for 5 min.

108 The PCR products were purified by Agencourt AMPure XP (Beckman, USA) and
109 quantified accurately by Qubit 3.0 DNA detection kit. The PCR products of each
110 sample were mixed to a final concentration of 20 pmol and then sequenced with a
111 MiSeq benchtop sequencer (Illumina, San Diego, USA) for 2×300 bp paired-ends
112 sequencing.

113 **2.4 Sequence data processing and statistical analysis**

114 The sequenced reads with perfect matches to barcodes were split into sample
115 libraries, and the chimeras were high-quality trimmed using Usearch. The OTUs were
116 classified at a 97% similarity level and rarefaction analyses were conducted using
117 Mothur(13). The curves from the rarefaction analysis and rank abundance analysis
118 were marked using the originally detected OTUs from R Project 3.2.

119 The sequences were first filtered through a BLAST search against the RDP
120 database(14), with a common confidence threshold of 80–90%, and the satisfactory
121 sequences were phylogenetically assigned to taxonomic classifications via an RDP
122 naïve Bayesian classifier and Bergey's taxonomy at a 0.03 distance(15). A
123 dendrogram was constructed with data derived from a hierarchical cluster analysis of
124 the beta diversity distance matrix and an unweighted pair group method with
125 arithmetic mean (UPGMA) cluster analysis(16). Bray-Curtis ordination was used to
126 calculate the distance between samples. A phylogenetic tree was constructed with

127 Python software and functional gene compositions were calculated with PICRUSt
128 1.0.0 software(17). All the other analyses were performed in R 3.2 with the vegan
129 packages v. 2.0-10.

130 **3 Results and discussion**

131 **3.1 Geochemical properties of samples**

132 The plant from which the samples were collected is located in Guanyinsi village,
133 Henan province, China. The two samples were collected outdoors, from the exposed
134 cisterns, and the colors of the two samples were light yellow with some
135 suspended substances. The temperature of them both was approximately 31°C, and
136 the pH values were 9.37 and 9.28 for the HNC and HNT samples, respectively. The
137 detectable ion concentrations are listed in Table 1, and they showed similar
138 concentrations in the two samples. The samples contained a very high level of Na
139 from the addition of NaCO₃ during bauxite flotation processing. The concentrations
140 of Al, K and Ca were at a relatively higher level. In addition, the concentrations of
141 Mg, Fe, Zn, Mo and Ba were at a relatively lower level. Some ions, such as Ti, As,
142 Pb, Sb, B, Mn, and Se, showed very low concentrations, while some ions, such as Cu,
143 Ni, Cd, Ag and others, were not detected. The organics in the samples were the
144 residuals of collectors (such as hydroxamic acid, oleic oil and others) and flocculants
145 (primarily HPAM).

146 In all, the two samples were both taken from the bauxite flotation process. They
147 had similar ion concentrations, pH values, temperatures and flocculant
148 concentrations, but they had different concentrations of collectors for the

149 chemical adsorption on concentrate and no effect on the tailings.

150 **Table 1** Geochemical properties of the water samples (mg/L)

Sample	B	Na	Mg	Al	K	Ca	Mn	Fe	Zn	Ba
HNC	0.17	1410.96	1.96	6.27	13.89	54.88	0.05	1.46	1.87	1.09
HNT	0.15	1237.35	1.17	5.75	21.70	39.92	0.03	1.69	1.43	0.96

Sample	Ti	As	Se	Mo	Pb	Sb	F ⁻	NO ₃ ⁻	SO ₄ ⁻	Cl ⁻
HNC	0.11	0.01	0.01	0.78	0.01	0.01	6.84	5.17	93.21	121.89
HNT	0.10	0.01	0.01	0.71	0.01	0.01	5.33	1.36	69.15	94.92

151

152 **3.2 Analyses of OTUs**

153 Both sequencing data sets were separated from one another to investigate the core
154 microbiome in recycling effluents of the concentrate and tailings. Furthermore, a
155 differentiation between bacteria and archaea was made to avoid the loss of the
156 archaeal OTUs due to their low sequence reads. More than 35000 high-quality
157 sequences with at least 1600 OTUs for bacteria as well as more than 27000
158 high-quality sequences with at least 320 OTUs for archaea were used for the
159 microbial community composition analysis of the two samples after filtering out the
160 chimeras. The asymptotic level of the Shannon rarefaction curves, at a 97% similarity
161 level, suggested that the number of OTUs were sufficient for this study (Fig 2). The
162 Shannon diversity index of bacterial communities was significantly higher than that of
163 the archaeal communities, which suggested that the bacterial communities were more
164 varied.

165 **Fig 2** Curves from the Shannon rarefaction analysis (a-bacteria, b-archaea)

166 Rank-abundance curves are used to interpret two aspects of sample diversity,
167 namely the richness and evenness of the species in the sample. Fig 3 shows that the
168 samples had low diversity index for the narrow change along the lateral axis while the
169 two sites had a low degree of uniformity for both archaeal and bacterial compositions
170 because of the steep shapes of the curves. The rarefaction curves combined with the
171 rank abundance curves suggested that the work was able to reveal the real microbial
172 community compositions of the samples.

173 **Fig 3** Rank abundance curves based on OTU ranks (a-bacteria, b-archaea)

174 A Venn diagram could reflect the similarity and overlap of the OTU composition
175 between samples visually (Fig 4). The samples in this study had low similarity rates in
176 both archaeal and bacteria communities at the OTU level. The bacterial OTUs were
177 more abundant than archaeal OTUs in the two samples, and the microbial
178 composition of the HNC were more abundant than that of the HNT. For both the
179 archaeal and bacterial communities, the common OTUs both detected in the two
180 samples were at a low level for each total OTU number, at 4.09% and 2.57%,
181 respectively. That finding revealed the significant diversity between the HNC and
182 HNT samples that might be caused by the different properties of concentrate and
183 tailings, although both of them were involved in the water cycle of the bauxite
184 flotation process.

185 **Fig 4** Similarity and overlap analysis of OTUs by Venn diagram (a-bacteria,

186 b-archaea)

187 3.3 Analyses of microbial communities

188 To identify effective biological methods for the prevention and treatment of the
189 recycling effluents problem, it is vital to comprehensively understand the involved
190 microbial communities in terms of taxonomic compositions, similarity and diversity.
191 A taxonomic composition analysis is one of the most frequently used bioinformatics
192 analyses for microbial communities(18).

193 (1) Bacterial communities

194 As illustrated in Fig 5, the dominant bacterial genus was unclassified in the two
195 samples accounting for 38.47% and 45.75% for HNC and HNT, respectively.
196 Interestingly, almost all of the unclassified OTUs fell into the Firmicutes phylum,
197 which suggested a new source for seeking new species in the Firmicutes phylum. For
198 the HNC sample, with the exception of the unclassified bacterial genus,
199 *Desulfomicrobium*, *Paracoccus* and *Tindallia* were the top three predominant genera,
200 accounting for 29.62%, 10.17% and 8.01%, respectively, followed by
201 *Exiguobacterium*, *Dethiosulfatibacter*, *Nitrincola*, *Anoxynatronum* and *Ercella* with a
202 proportion over 1%. For the HNT sample, except for the unclassified bacterial genus,
203 *Paracoccus* and *Exiguobacterium* were the top two predominant genera, accounting
204 for 21.34% and 17.99%, respectively, followed by *Ercella*, *Tindallia*, *Pirellula*,
205 *Anoxynatronum* and *Ercella*, with a proportion between 1% to 3%.

206 **Fig 5** Taxonomic distribution of bacterial OTU abundance at the phylum level

207 According to the taxonomic comparison results, the dominant species were
208 selected and the evolutionary relationships and abundance differences in the dominant

209 microorganisms in the sequenced environmental samples were understood from the
210 phylogenetic system (Fig 6). According to the phylogenetic analysis, 99.03% of the
211 bacterial OTUs fell into the following six phylogenetic divisions: Proteobacteria,
212 Firmicutes, Synergistetes, Planctomycetes, Actinobacteria and Acidobacteria.
213 Proteobacteria and Firmicutes were the dominant phyla while the others contributed
214 little to the bacterial community structures.

215 Within the Proteobacteria phylum, which accounted for 35.44% of the bacterial
216 communities, *Paracoccus* belonged to Alphaproteobacteria (mostly found in HNC)
217 and *Dethiosulfatibacter* belonged to Desulfomicrobium (mostly found in HNT) were
218 the major genera. Firmicutes accounted for 59.36% of all the OTUs, but more than
219 half of that OTUs were classified into unknown genera, and 10.17% of all the OTUs
220 fell into the *Exiguobacterium* genus while 5.16% fell into *Tindallia*.

221 **Fig 6 Phylogenetic compositions and community structures of bacteria**

222 *Paracoccus* was found in both samples and played an important role in the
223 bacterial community structures. HPAM, hydroxamic acid and oleic oil were the
224 primary organic pollutants in recycling effluents, and the *Paracoccus* genus emerged
225 as a versatile species for the bioremediation of various recalcitrant pollutants, such as
226 pyridine(19), dimethylformamide(20), N-methylpyrrolidone(21), and others.
227 *Paracoccus* was also reported to not only reduce nitrate under an aerobic atmosphere
228 but also to convert ammonium to nitrogen gas via hydroxylamine, nitrite, nitrate and
229 nitrous oxide, which indicated that another nitrogen removal pathway immediately
230 accessible via hydroxylamine without any accumulation of nitrate and nitrite could be

231 responsible for this phenomenon(22). Hydroxylamine from organic pollutants and
232 nitrate from recycling effluents were the primary nitrogen source for microbial growth
233 as well as biodegradation actions.

234 Although several OTUs were detected in the HNT sample, *Desulfomicrobium*
235 was in the majority in the HNC. *Desulfomicrobium* is gram-negative and is commonly
236 found in aquatic environments with high levels of organic matter as well as in sewage
237 sludge as a major community member of extreme oligotrophic habitats(11). Reports
238 revealed the good bioremediation potential of the reductive process performed by
239 *Desulfomicrobium*, and the degradation pathway was that sulfate and other oxidized
240 sulfur compounds serve as terminal electron acceptors and are reduced to H₂S(23).
241 The two samples in this study had similar sulfate concentrations as well as
242 temperature and pH, and the diversity in *Desulfomicrobium* between two samples
243 might be due to that residue of hydroximic acid and oleic oil mostly in concentrate
244 pulp while little of that was associated with tailing pulp for collectors, which were
245 compatible with concentrate(24).

246 Contrary to the distribution of *Desulfomicrobium*, there was little
247 *Exiguobacterium* in the HNC but quite a lot in the HNT. The genus *Exiguobacterium*,
248 which belongs to the Bacilli class, is characterized by the following features: the cells
249 are gram-positive, non-endospore-forming, facultatively anaerobic and
250 oxidase-negative (25). *Exiguobacterium* has been reported to promote plant growth
251 and the degradation of environmental pollutants, to be halotolerant(26, 27) and to be
252 able to produce highly effective proteolytic enzymes(28). The HNT sample had a pH

253 value of approximately 9.28, which provided an environment for the bacteria
254 belonging to the *Exiguobacterium* genus to show their applicability to the biological
255 neutralization of alkaline wastewater(25). The HNC also had a pH value of 9.37 but
256 little *Exiguobacterium*, perhaps because of the inhibition of carbon sources (such as
257 hydroximic acid and oleic oil) and the competition from *Desulfomicrobium*.

258 *Tindallia*, *Anoxynatronum* and *Ercella* were all detected in the two samples as
259 auxiliary members, and they were all reported to be alkaliphilic and good for potential
260 use in treating alkaline wastewater(29-31).

261 (2) Archaeal communities

262 Archaea, which were originally found in extreme environments, were known to
263 be ubiquitous and crucial partners in numerous microbiomes, either in environmental
264 settings or associated with holobionts(32), and they were no exception in this research.
265 Fig 7 reveals that *Methanobacterium* and *Methanothrix* were the dominant archaeal
266 genera in the two samples, accounting for 46.67% and 30.24%, 75.13% and 19.99%
267 for HNC and HNT, respectively. *Nitrososphaera* was also abundant in HNC, with a
268 proportion of 18.62%, while it contributed only 4.17% to HNT. *Methanosarcina* was
269 found in both samples and accounted for a low proportion of the total.

270 **Fig 7** Taxonomic distribution of archaeal OTU abundance at the phylum level

271 According to the phylogenetic analyses (Fig 8), almost all of the archaeal OTUs
272 fell into the three archaeal phyla Euryarchaeota, Thaumarchaeota and Planctomycetes.
273 Euryarchaeota was dominant and Planctomycetes was seldom identified.
274 Euryarchaeota contains methane-forming, extremely halophilic, sulfate-reducing, and

275 extremely thermophilic sulfur-metabolizing species(33). Although there was a similar
276 distribution in the two samples at the phylum level, *Methanobacterium* was more
277 abundant in the HNC sample and *Methanotherix* was ascendant in the HNT at the
278 genus level under the Euryarchaeota phylum.

279 **Fig 8 Phylogenetic compositions and community structures of archaea**

280 *Methanotherix*, a genus of Methanosarcinales that are typical acetoclastic
281 methanogenic archaea(34), made up more than half of all the OTUs, while
282 *Methanobacterium*, a genus of Methanobacteriales that could reduce CO using
283 molecular hydrogen (hydrogenotrophic) and produce methane(35), accounted for
284 approximately one third of all the OTUs. These two primary genera belong to the
285 methanogenic archaea and made up more than 85% of the archaeal communities. The
286 methanogens were the initially and most widely recognized archaea, and they were
287 considered to play a prominent role in the degradation of complex organic compounds
288 by a consortia of anaerobic microorganisms. The methanogens are dependent on
289 sodium ions for methane formation(36) and the sodium ion concentrations were at a
290 high level (more than 1200 mg/L, as shown in Table 1) in the samples, which
291 provided an appropriate environment for the high abundance of methanogenic
292 archaea.

293 It is unclear how archaea communicated or structurally interacted with their
294 environment. Thus, the methanogens certainly represented the keystone species, but
295 the archaeal diversity beyond the methanogens must not be neglected. *Nitrososphaera*
296 and *Methanosarcina* are the remaining genera found in these samples, and

297 unclassified genera made up less than 0.1%, which verified the low diversity of
298 archaea again. *Nitrososphaera*, which is one of the primary phylotypes of
299 ammonia-oxidizing archaea (AOA)(37), had also been directly linked to active
300 nitrification; it was found that the biotransformation of micropollutants were involved
301 in the co-metabolism of ammonia-oxidizing microorganisms(38, 39).
302 *Methanosarcina* is one of the few methanogens shown to participate in direct
303 interspecies electron transfer (DIET) by directly receiving electrons to reduce CO₂
304 into CH₄(40), and it appeared to be of crucial importance to the anaerobic digestion
305 process(41).

306 (3) Microbial community diversity between two samples

307 Although the syntrophic and competitive relationships between archaea and
308 bacteria remained completely unclear, a consortia of bacterial and archaeal
309 communities played an important role in the degradation of complex organic
310 compounds(36). The ways in which archaea and bacteria define their niche within a
311 complex setting could be revealed by extensive and in-depth studies on microbial
312 community diversities in different environments.

313 The similarity of the pH values, water temperature, ion concentrations and types
314 of organic pollutants between HNC and HNT gave an intimation of the similarity in
315 components in both the archaeal and bacterial communities. The diversity in the
316 concentrations of collectors was the primary reason for the diversity of distributions in
317 the microbial composition. The primary bacterial compositions of both samples
318 consisted of *Desulfomicrobium*, *Paracoccus*, *Tindallia*, *Methanobacterium*,

319 *Methanotherix* and *Nitrososphaera*, which were all reported as alkaliphilic bacteria for
320 biological neutralization(29, 30, 42, 43). The primary archaeal compositions in both
321 samples consisted of (the genera *Methanotherix*, *Methanobacterium* and
322 *Methanosarcina*) and AOA (genus *Nitrososphaera*), which appear to be part of a key
323 strategy to adapt to a complex environment(32).

324 According to the sample HNC result, the community primarily made up of
325 *Desulfomicrobium*, *Paracoccus*, *Tindallia*, *Methanotherix*, *Methanobacterium* and
326 *Nitrososphaera* seemed better adapted to the biodegradation of the flotation collectors.
327 Flotation collectors are small molecular organic materials, and they could be degraded
328 by bacteria more easily. The sulfate reduction and ammonia oxidization tasks were
329 performed by *Desulfomicrobium* and *Nitrososphaera*, while other genera played a
330 supporting role.

331 According to the result for the HNT, the community primarily made up of
332 *Paracoccus*, *Exiguobacterium*, *Methanotherix* and *Methanobacterium* was considered
333 to be more efficient at biodegrading the flocculant. HPAM, the flocculant in the
334 samples, was a bio-refractory polymer compound with amide groups and a carbon
335 backbone that could be the energy source for microorganisms. In these habitats and
336 with flocculant as a substrate, *Exiguobacterium* produced lactate, acetate and formate,
337 and these products were converted into methane by *Methanotherix* and
338 *Methanobacterium*. *Paracoccus*, the dominant bacterial genus, could grow under
339 aerobic and denitrifying conditions, and it showed great potential at hydrogen sulfide
340 removal from synthetic biogas(44, 45). *Exiguobacterium* was reported to grow

341 through the utilization of molecules containing acyl groups(46). These two bacterial
342 genera supported methanogenesis by methanogens, which was important at the end of
343 the anaerobic digestion chain for biomass conversion.

344 **4 Conclusions**

345 The archaeal and bacterial communities in recycling effluents from bauxite
346 flotation plant were revealed, and the diversity between the recycling effluent of
347 concentrate and that of tailings was studied. Of the unclassified OTUs at the genus
348 level, almost all fell into the Firmicutes phylum, and they
349 accounted for approximately 40 percent of all the OTUs. Therefore, this recycling
350 effluent was a worthy resource for the isolation of new strains in the Firmicutes
351 phylum. At the genus level, *Paracoccus*, *Desulfomicrobium*, *Exiguobacterium*,
352 *Tindallia*, *Ercella* and *Anoxynatronum* were the primary members of the bacterial
353 community, while *Methanothrix*, *Methanobacterium*, *Nitrososphaera* and
354 *Methanosarcina* dominated the archaeal community. The biodegradation of the
355 pollutants in the recycling effluents involved the synergism of bacteria and archaea.
356 Combined with the geochemical properties of recycle effluents, the microbial
357 community containing *Desulfomicrobium*, *Paracoccus*, *Tindallia*, *Methanobacterium*,
358 *Methanothrix* and *Nitrososphaera* was more adept at the biodegradation of flotation
359 collectors, such as hydroximic acid and oleic oil. The microbial community consisting
360 of *Paracoccus*, *Exiguobacterium*, *Methanothrix* and *Methanobacterium* was more
361 efficient at the biodegradation of HPAM. The results of this study provided a practical
362 orientation on the biotreatment of recycling effluents from the bauxite flotation

363 process.

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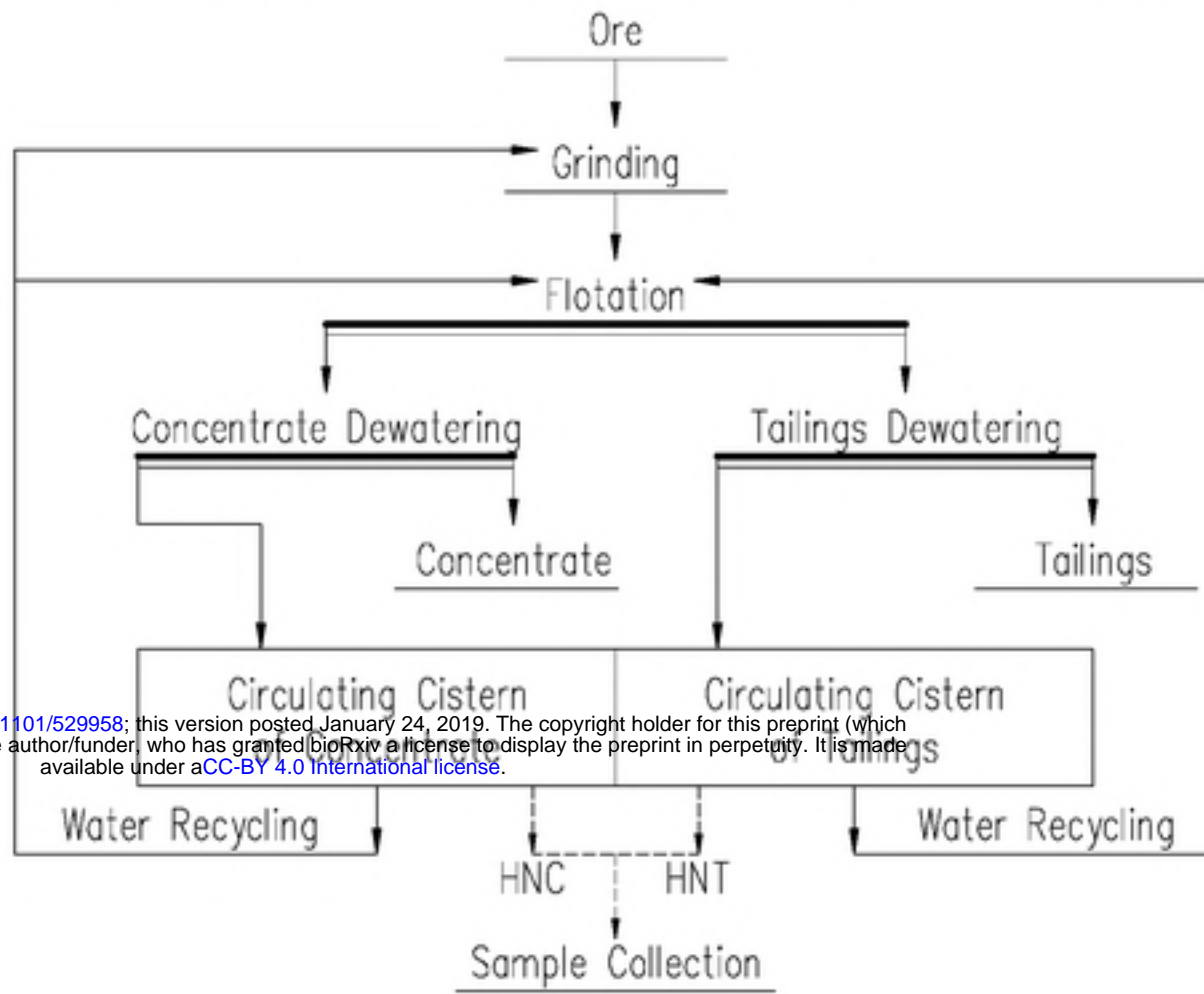
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Fig 1 Flow chart of desilication process and sample collection



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Fig 2 Curves from the Shannon rarefaction analysis (a-bacteria,
b-archaea)

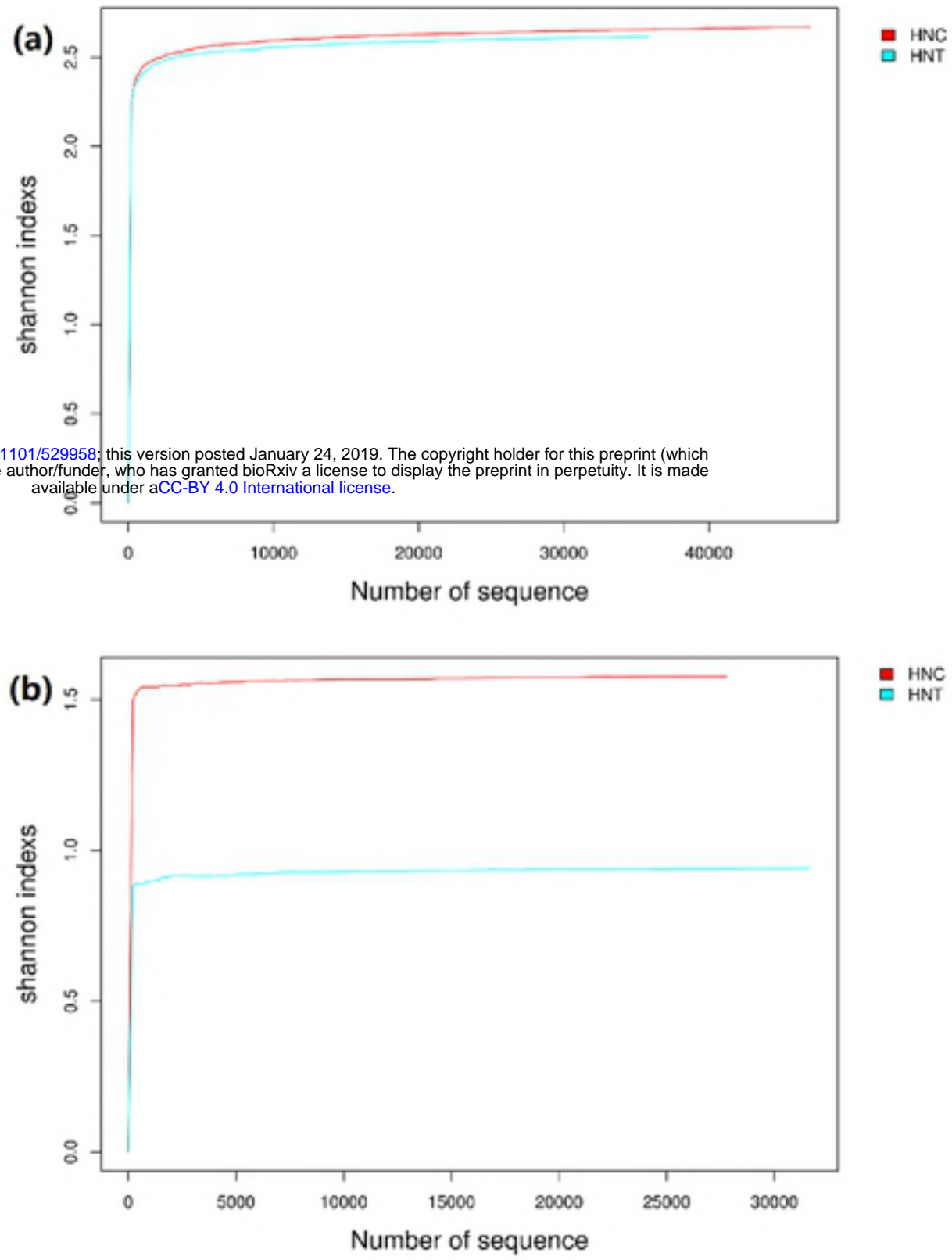


Fig 3 Rank abundance curves based on OTU ranks (a-bacteria, b-archaea)

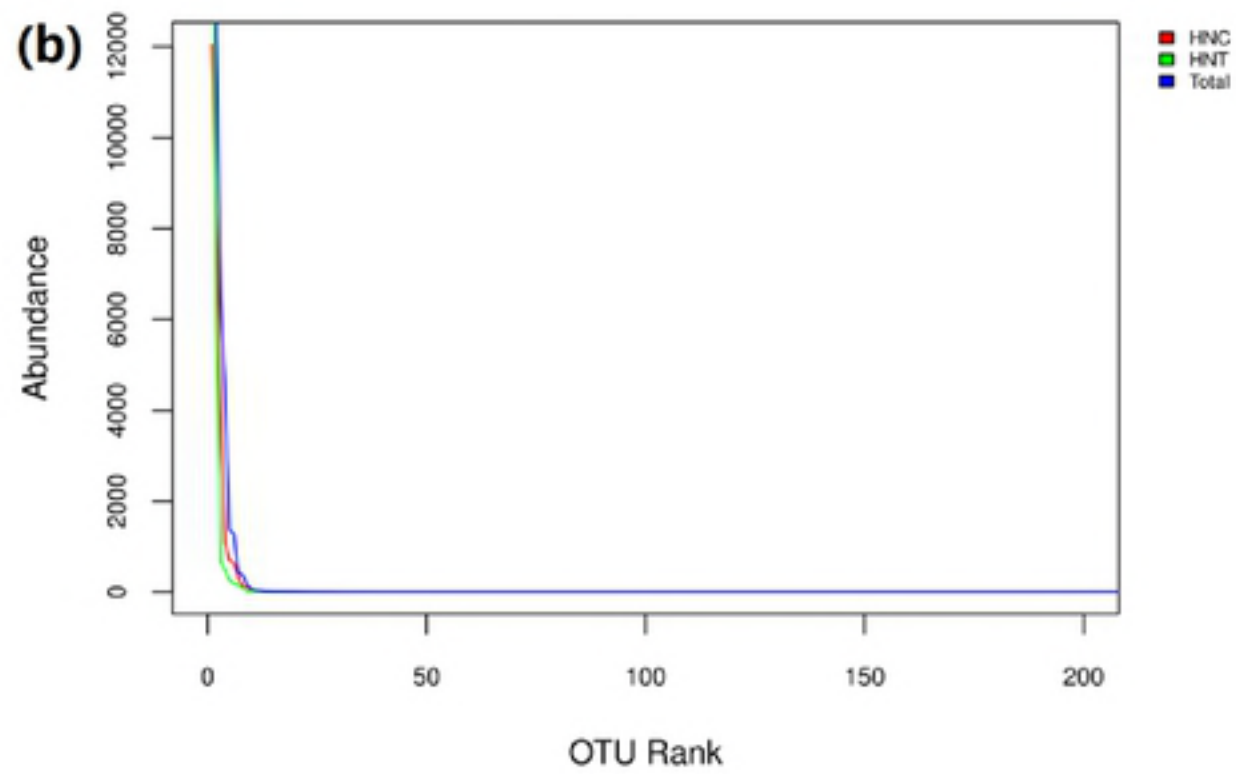
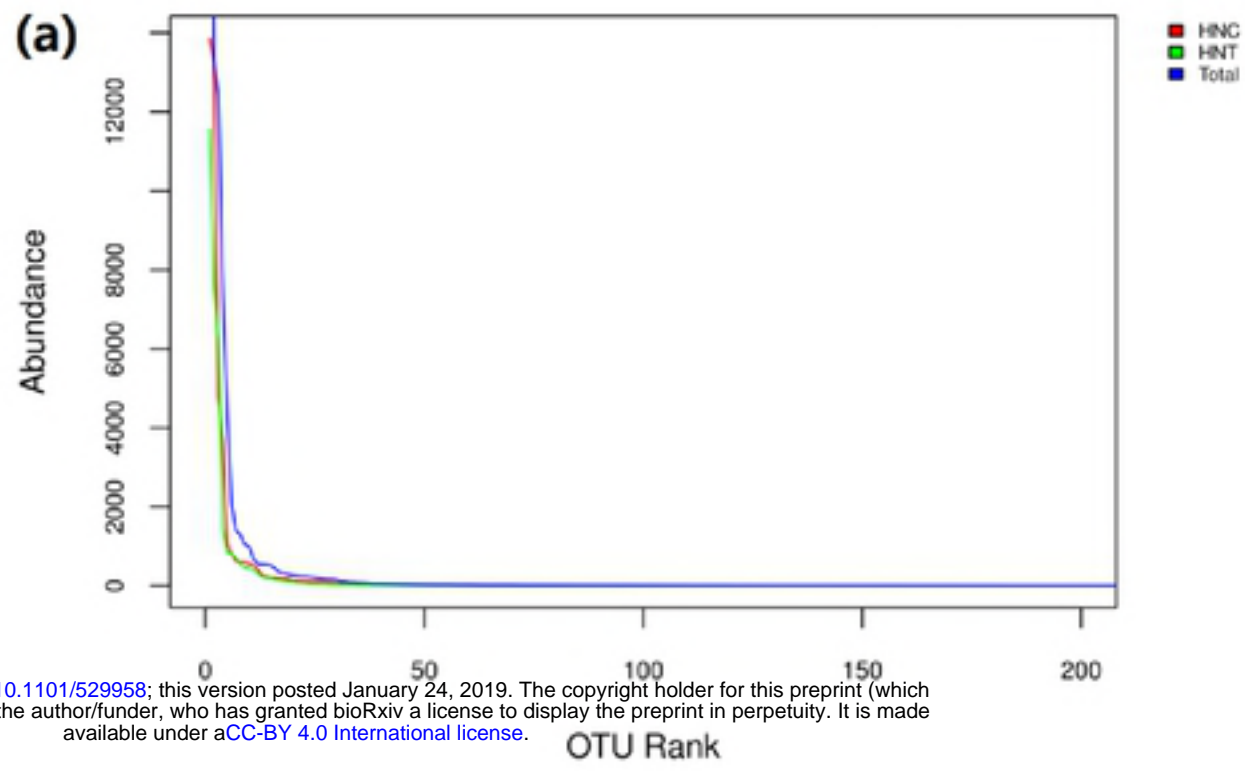
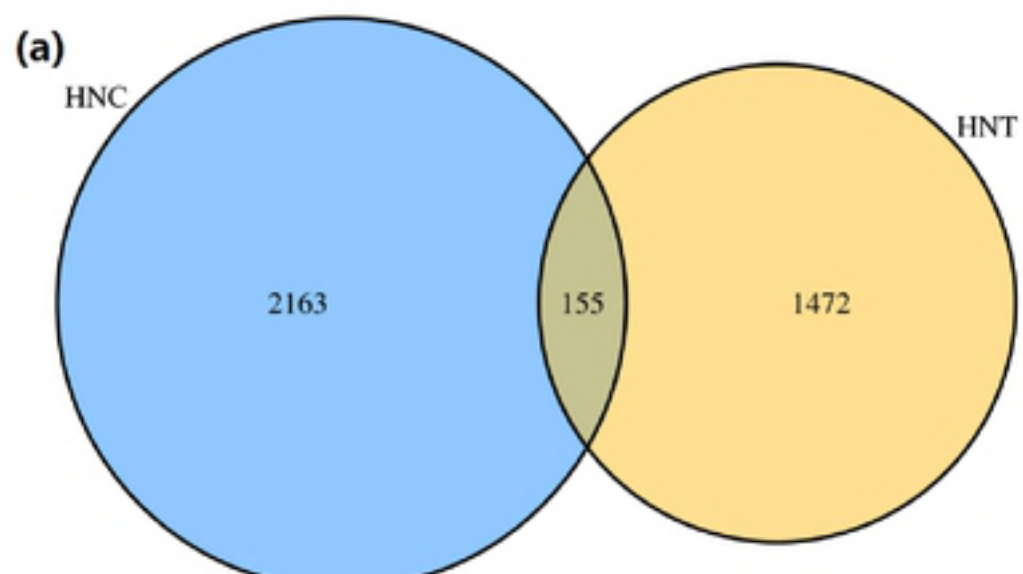


Fig. 4 Similarity and overlap analysis of OTUs by Venn diagram

(a-bacteria, b-archaea)



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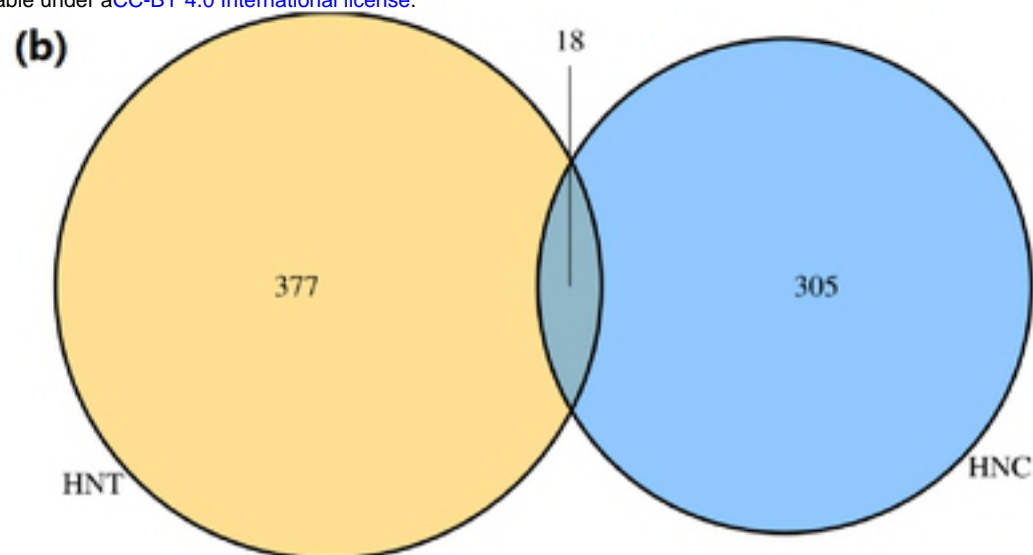


Fig 5 Taxonomic distribution of bacterial OTU abundance at the phylum level

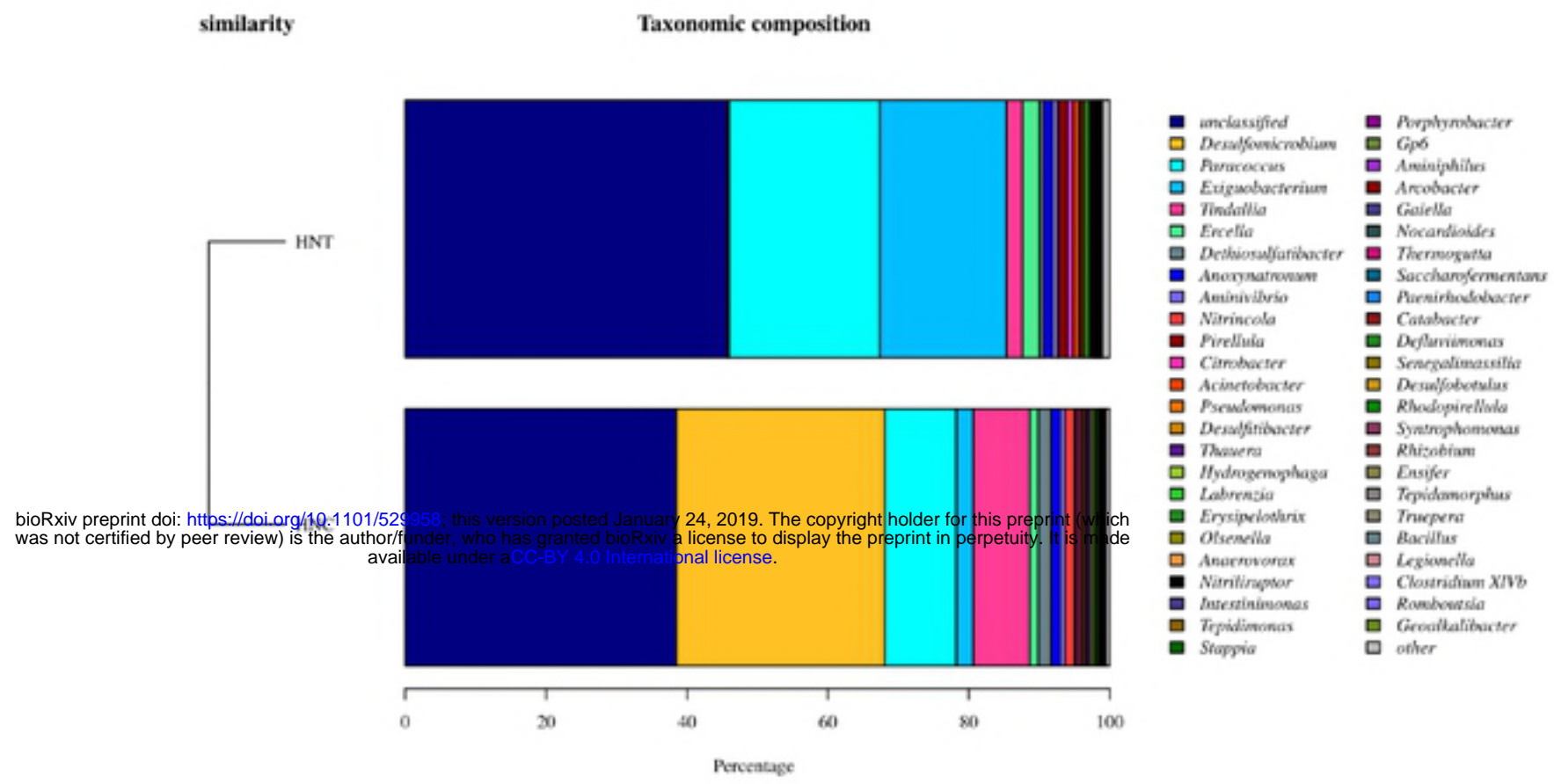


Fig 6 Phylogenetic compositions and community structures of bacteria

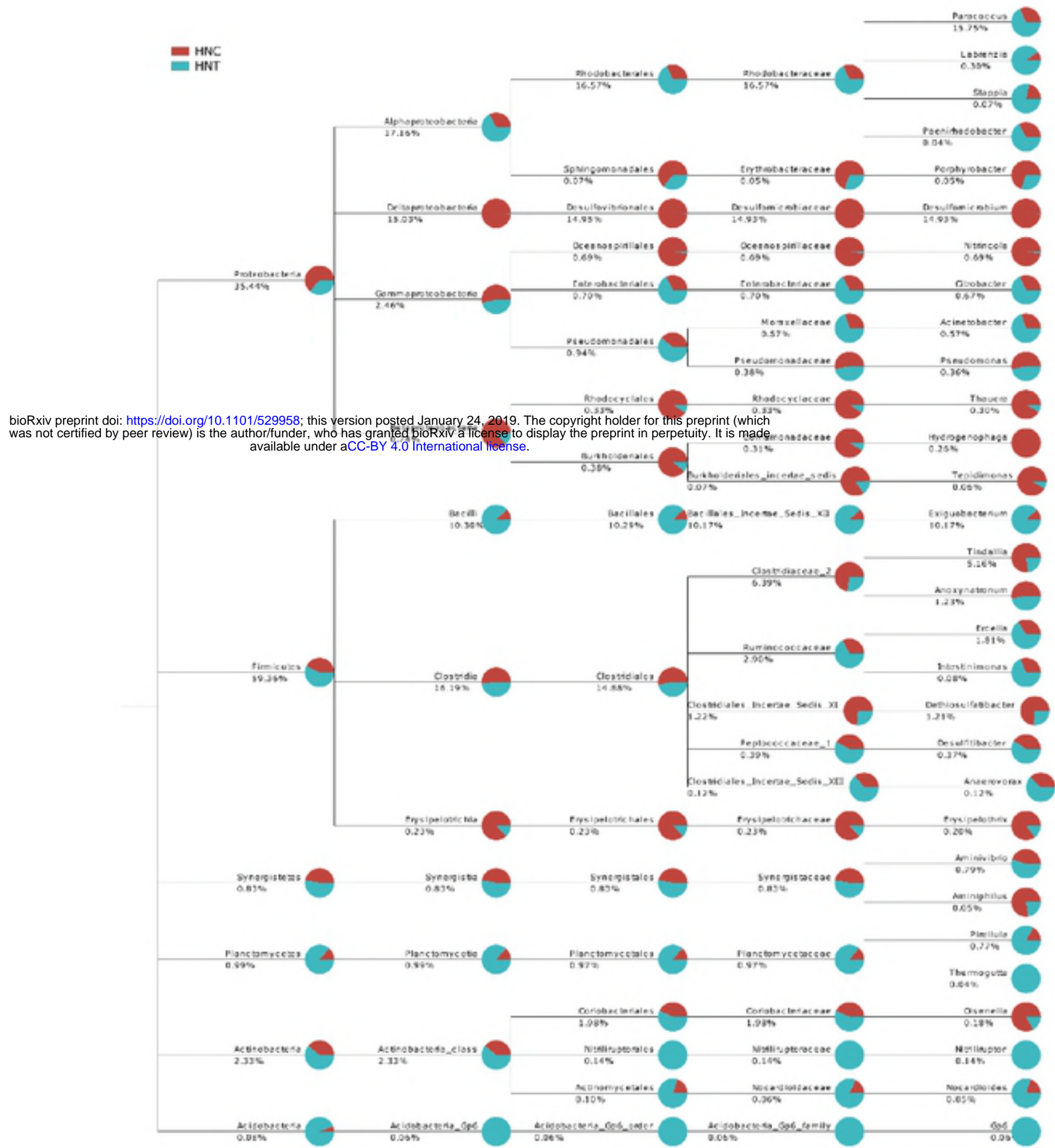


Figure 6

Fig 7 Taxonomic distribution of archaeal OTU abundance at the phylum level

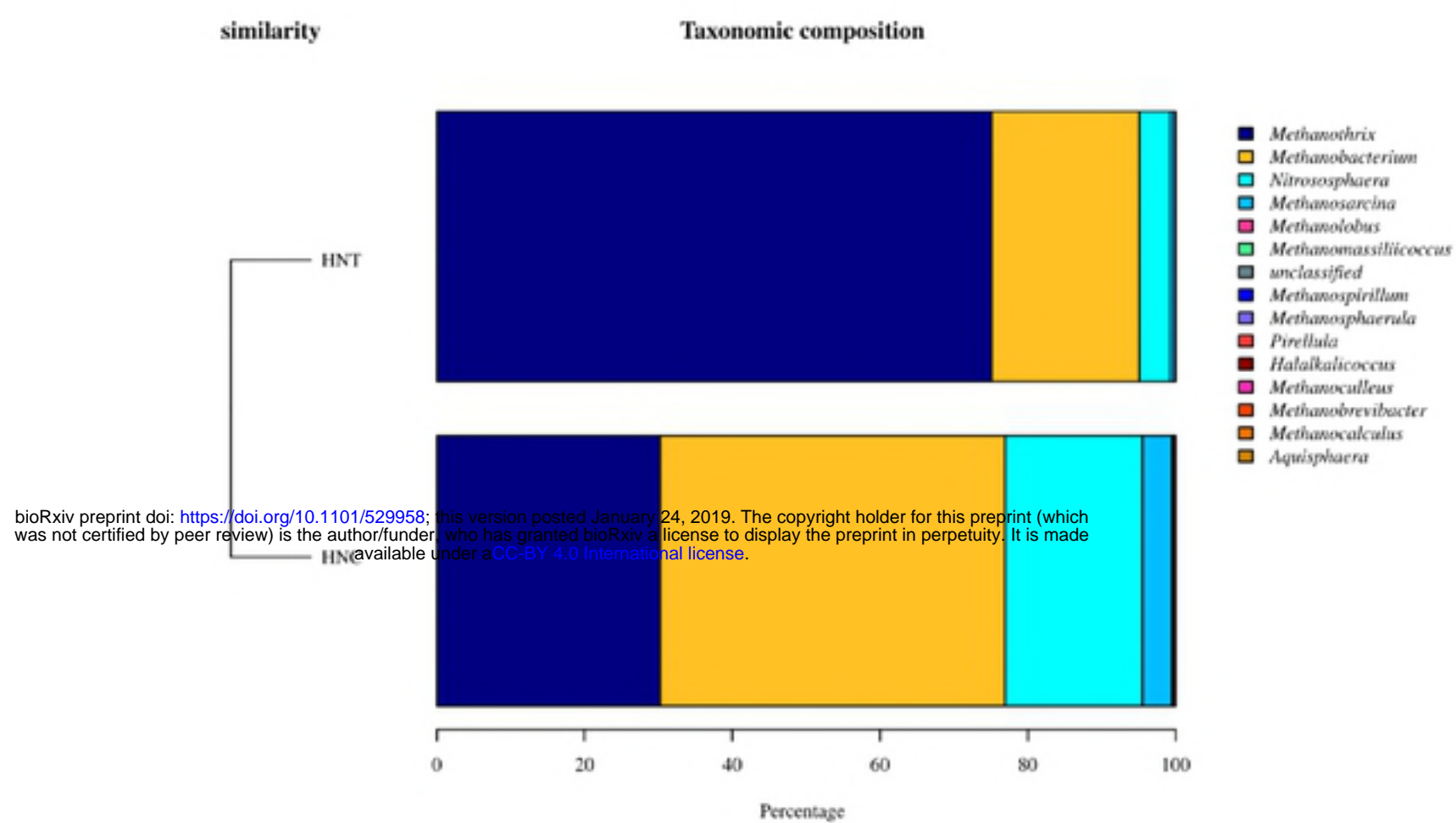


Fig 8 Phylogenetic compositions and community structures of archaea

