

1 **Differential side-effects of *Bacillus thuringiensis* bioinsecticide on non-target *Drosophila* flies**

2 Aurélie **Babin**, Marie-Paule **Nawrot-Esposito**, Armel **Gallet**, Jean-Luc **Gatti**^{§*}, Marylène **Poirié**[§]

3

4 Université Côte d'Azur, INRAE, CNRS, ISA, France

5

6 ***Corresponding author**

7 Institut Sophia Agrobiotech, INRAE, 400 route des chappes, 06903 Sophia Antipolis, France.

8 E-mail : jean-luc.gatti@inra.fr

9

10 **§Co-last authors**

11

12 **Running title:** *Bt* bioinsecticide and *Drosophila* flies

13

14 **Keywords:** *Drosophila*, *D. melanogaster*, *D. suzukii*, *Bacillus thuringiensis*, biopesticide, non-
15 intentional effects, development, longevity, fitness

16

17 **Word count:** main text: 5 284, abstract: 207

18 **References:** 90

19 **Figures:** 7 figures and 1 table

20 **Supplementary information**

21 **Abstract**

22 Biopesticides based on *Bacillus thuringiensis* (*Bt*) spores and toxins are alternate pest management
23 solutions widely used to control insect pests. Their increasing use could lead to accumulation in the
24 environment, hence leading to chronic exposure of non-target organisms. Here, we tested for
25 potential non-intentional side-effects of chronic exposure to *Bt* biopesticide on larvae of non-target
26 *Drosophila* species present in *Bt*-treated areas. Doses up to those recommended for field application
27 ($\leq 10^6$ CFU/g of fly medium) had no effect on the fly development, whereas doses 10 to 100-fold
28 higher (10^7 - 10^8 CFU/g) increased developmental time and decreased adult emergence rates in a
29 dose-dependent manner and with varying effect amplitudes for all the species and strains tested. For
30 all them, all larvae died before pupation at the highest dose tested (10^9 CFU/g). Focusing on *D.*
31 *melanogaster*, delayed development and reduced emergence resulted from stage-dependent larval
32 mortality, and fitness-related traits of adult flies emerging from surviving *Bt* biopesticide exposure
33 were moderately increased. The effects of *Bt* biopesticide seemed to result from the spores/cleaved
34 toxins synergy, and possibly additives. While recommended doses had no effect on non-target
35 *Drosophila* species, misuse or local accumulation of *Bt* bioinsecticides in the environment could
36 have non-intentional side-effects on fly populations with potential implications for their associated
37 communities.

38

39

40 **Introduction**

41 The world's population is expected to reach more than 9 billion people by 2050,^[1] increasing the
42 demand for agricultural resources. This requires to improve pest management, especially insects
43 that cause more than 30% of agricultural losses.^[2] Nowadays, their management largely relies on
44 conventional chemical insecticides. However, their use and efficiency have been considerably
45 reduced due to the emergence of pests' resistance, development of secondary pests, adverse side-
46 effects on non-target species (natural enemies of pests, pollinators),^[3,4] and more generally the
47 impacts on biodiversity and human health (e.g. neurological disorders, functional impairment of
48 reproduction, cancers).^[5-8] Developed as a more specific and safer alternative, biopesticides
49 represent less than 5% of the pesticide market, the large majority being microbial insecticide
50 formulations based on viable spores and toxins of the bacterium *Bacillus thuringiensis* (*Bt*) (over
51 400 registered formulations).^[4,9]

52 *Bt* is a Gram-positive endospore-forming bacterium that synthesizes a wide range of toxins
53 with different chemical structures, modes of action and biological targets. The most abundant and
54 studied are Cry δ -endotoxins encoded by genes located on large plasmids and produced as
55 parasporal crystalline inclusions during the stationary growth phase.^[10,11] *Bt* produces other
56 insecticidal toxins, the Cyt (cytolytic δ -endotoxins) and Vip (secreted Vegetative Insecticidal
57 Proteins) that synergize their effects with Cry toxins, virulence factors such as β -exotoxins (or
58 thuringiensin), a secreted nucleotide toxic for almost all tested life forms thus prohibited in
59 commercial formulations,^[12] and anti-fungal factors.^[13,14] *Bt* subspecies and strains can differ in
60 their plasmid number and in the synthesized toxins cocktail responsible for their biological activity,
61 which was used to delineate potential target insects.^[15] For instance, *Bt* subsp. *kurstaki* (*Btk*)
62 produces the 5 Cry toxins, Cry1Aa, Cry1Ab, Cry1Ac, Cry2Aa and Cry2Ab,^[10,16] while *Bt* subsp.
63 *israelensis* (*Bti*) produces a combination of Cry4Aa, Cry4Ba, Cry10Aa, and Cry11Aa,^[17,18] both
64 strains being commercially used. The different toxin cocktails produced by some *Bt* subspecies can
65 also be detrimental to non-insect organisms such as nematodes, protozoa, and even molluscs.^[15]

66 The bioinsecticide formulations based on spores and toxin crystals of *Btk* and *Bti* are the most
67 sprayed in organic and conventional farming, and natural areas (e.g. forests, swamps) to deal with
68 larvae of Lepidopteran pests and Dipteran larvae of mosquitoes and black flies, respectively. It is
69 generally accepted that once ingested by insect larvae, the toxin crystals are dissolved by the midgut
70 alkaline pH, releasing ~130 kDa pro-toxins that are then processed by digestive proteases into
71 smaller, soluble, active toxin fragments of ~ 60-70 kDa.^[19,20] Active toxins bind to specific
72 receptors of midgut epithelial cells, eliciting pores formation in the cell membrane, cell lysis and
73 gut epithelium disorganization.^[21,22] This allows gut bacteria, including *Bt*, to colonize the
74 hemocoel, and leads to rapid septicaemia and death.^[23,24]

75 Numerous impact studies of field application rates and acute intoxications have shown that *Bt*
76 bioinsecticides are safe or have limited impact on non-target vertebrates and invertebrates, and
77 associated species communities.^[25,26] However, the increasing use of biopesticides based on *Bt*
78 spores and toxins has recently raised concern^[27] and led to the assessment of their potential effects
79 on non-target species, such as auxiliary insects of biological control,^[28] pollinators,^[29] and species
80 and species communities which simply share their habitat with *Bt*-targeted insect pests.^[30-32] Yet,
81 there is growing evidence of direct and indirect cross-effects of *Bt* bioinsecticides and toxins across
82 insect species and orders, or even across phyla, suggesting that *Bt* targeting is only partly
83 specific.^[33,34] Data showed that almost all of the applied *Btk* dose was still present on the leaves
84 surface 72 hours after spraying,^[35] its amount returning close to environmental levels only 28 days
85 after treatment.^[36] Finally, *Bt* spores can survive in the soil and on different supports for months and
86 even years after application.^[37-41] *Bt* formulations contain also numerous compounds to protect
87 spores and crystals and aggregate them into a wettable form, surfactants to facilitate spraying and
88 dispersion on plants, and phagostimulants.^[42,43] Nevertheless, since toxin crystals, and to a much
89 lesser extent spores,^[44] are somewhat sensitive to abiotic conditions (e.g. UV, pH, rainfall), repeated
90 sprayings with a minimum delay of 3 to 8 days is often recommended over the period of pest
91 occurrence to achieve the required pest control level^[43,45] (<http://www.certiseurope.fr>,

92 <http://www.certisusa.com>). All these can potentially lead to *Bt* accumulation in the environment,
93 thus raising the rarely addressed issue of potential side-effects of chronic exposure (*i.e.* continuous
94 and increasing exposure dose for an extended period) of non-target species to doses unexpectedly
95 above the recommended spraying doses.

96 Diptera are worldwide distributed insects, most of which are not targets for *Bt* and its toxins.
97 This is the case of the genus *Drosophila*, represented by ~ 1 500 described species,^[46] including the
98 model organism *D. melanogaster*. In the field, most of these flies feed and reproduce mainly on
99 ripening or rotting/fermenting fruits and are therefore naturally present in areas treated with *Bt* such
100 as orchards, vineyards and gardening areas. Unable to disperse between food patches, early
101 developmental stages of *Drosophila* eat intensively and grow exponentially,^[47] and may thus ingest
102 high doses of *Bt* bioinsecticides that have accumulated during the treatment periods. Surprisingly,
103 despite the presence of many *Drosophila* species in *Bt*-treated areas, their role in the decomposition
104 of organic matter, and the ease of study of some species, only a few studies have focused on these
105 flies. However, most of them suggested susceptibility to *Btk*, but they used mainly late 3rd instar
106 larvae preparing for pupation, which do not feed much, and used *Bt* preparations, especially field
107 isolates, that possibly contained highly toxic β -exotoxins, which are not authorized in commercial
108 *Bt* formulations.^[48-55] So far, no study addressed the effects of chronic exposure to *Bt* formulations
109 containing spores and toxin crystals but no β -exotoxins, on developing stages of these Dipterans
110 that are present in *Bt*-treated areas.

111 Here, we have tested the chronic side-effects of different commercial formulations of *Btk*
112 (devoid of β -exotoxins) and, to a lesser extent of *Bti*, with doses starting from the maximum
113 recommended spraying dose up to 1 000 times this dose (*i.e.* below acute intoxication doses used in
114 most studies). We mainly focused on developmental traits (developmental time, emergence rate),
115 firstly using the wild-type *D. melanogaster* CantonS. The spore-forming Gram-positive *Bacillus*
116 *subtilis* and the *Btk* strain (4D22), devoid of Cry toxin genes and thus of crystals, were used as non-
117 pathogenic controls. We also analysed two fitness-related traits of adult flies (male and female

118 longevity, offspring number) after entire development in presence of *Btk* formulation. To test for
119 effects specific to the fly genetic background, developmental traits upon exposure to *Btk*
120 formulation were measured on four other *D. melanogaster* strains. Finally, we extended further
121 these development experiments to seven other *Drosophila* species. We have chosen six species in
122 addition to *D. melanogaster*, including cosmopolitan species, which can co-occur in the field^[56-60]
123 and are usually present in *Bt*-treated areas, and the invasive *D. sukuzii* that can now co-exist with
124 the six species in the areas it has recently invaded. This aims at providing a first-step in the
125 exploration of potential implications of chronic exposure to *Btk* formulation in terms of species
126 competition and species community composition and dynamics.

127

128 **Material and methods**

129 *Commercial formulations, Bacillus productions and Colony Forming Unit measurement*

130 The tested commercial brands of *Bacillus thuringiensis kurstaki* (*Btk*; serotype 3a, b, c^[61]) were
131 Delfin[®] A and B (strain SA-11; wettable granules, Valent BioSciences, AMM 9200482, 32,000
132 UI/mg) and Scutello DF (a Dipel[®] sub-brand; strain ABTS-351; wettable granules, Biobest[®], AMM
133 2010513, 540g/kg). The commercial brand of *Bacillus thuringiensis israelensis* (*Bti*; strain HD-14;
134 serotype 14^[61]) was VectoBac[®] WG (wettable granules, Bayer, AMM 2020029, 3000 UTI/mg). For
135 each formulation, the number of viable spores (expressed as Colony Forming Units (CFU) per mg
136 of granules) was estimated using serial dilutions of a suspension on LB agar plates and counting of
137 bacterial colonies after overnight incubation at 30°C. CFU estimations were 5×10^7 CFU/mg for *Btk*
138 Delfin[®] A; 2.5×10^7 CFU/mg for *Btk* Delfin[®] B; 2.2×10^7 CFU/mg for *Btk* Scutello DF; 6×10^7
139 CFU/mg for *Bti* VectoBac[®]. Of note, our CFU estimations fell in those appended on the commercial
140 packaging, between 10^{13} and 5×10^{13} CFU/kg (32 000 UI/mg; <http://www.certiseurope.fr>). No
141 change in CFU estimations occurred during the time frame of the experiments. Manufacturer-
142 recommended doses for Delfin[®] range from 0.15 to 1.5 kg/ha depending on the crop type. Based on
143 our CFU estimations, this corresponds to recommended doses of 7.5×10^4 to 7.5×10^5 CFU/cm² of

144 Delfin[®] A, and 3.75×10^4 to 3.75×10^5 CFU/cm² of Delfin[®] B for each spraying in the field. For
145 Scutello DF, recommended doses range from 0.1 to 1 kg/ha, which are equivalent to 2.2×10^4 to
146 2.2×10^5 CFU/cm². Vectobac[®] WG is used at 0.125 to 1 kg/ha, equivalent to 7.5×10^4 to 6×10^5
147 CFU/cm².

148 The acrystilliferous (Cry toxin-free) *Btk* 4D22 strain (depleted for the toxin-encoding
149 plasmids^[62]) obtained from the Bacillus Genetic Stock Center (<http://bgsc.org>; Columbus USA),
150 and a *Drosophila* non-pathogenic *Bacillus subtilis* (gift from Dr. E. Bremer, University of Marburg,
151 Germany; A. Brun-Barale, pers. comm.) were grown at 30°C in the sporulation-specific medium
152 PGSM (Bactopectone[®] 7.5 g, KH₂PO₄ 3.4 g, K₂HPO₄ 4.35 g, glucose 7.5 g, PGSM salts 5 mL,
153 CaCl₂ 0.25 M, distilled water qsp 1L, pH 7.2; PGSM salts: MgSO₄·7H₂O, MnSO₄·H₂O,
154 ZnSO₄·7H₂O, FeSO₄·7H₂O) for about 14 days for sporulation to occur. Following elimination of
155 vegetative cells (1h at 70 °C), spore pellets were collected after centrifugation (4,500 rpm, 20 min,
156 4 °C), washed with sterile water, and lyophilized. CFU numbers were counted for each preparation
157 as described above.

158

159 *Fly stocks*

160 The four tested strains of *Drosophila melanogaster* (phylogenetic subgroup: melanogaster) were the
161 standard wild-type Canton S (Bloomington Drosophila Centre) used as a reference strain, the wild-
162 type Nasrallah strain from Tunisia (strain 1333, Gif-sur-Yvette), the double mutant standard strain
163 YW1118 (white and yellow mutations; gift from Dr. B. Charroux, IBD, Marseille-Luminy), and a
164 recently field-collected strain (caught in Southern France in 2013) that we named “Sefra”. For
165 *Drosophila* species comparison, we included 6 species of the *Drosophila* subgenus, *D. simulans*
166 (strain 1132; phylogenetic subgroup: melanogaster), *D. yakuba* (strain 1880; phylogenetic subgroup:
167 melanogaster), *D. hydei* (phylogenetic subgroup: hydei) and *D. sukukii* (phylogenetic subgroup:
168 immigrans) (both kindly provided by Dr. R. Allemand, LBBE, University Lyon 1), *D. immigrans*
169 (phylogenetic subgroup: immigrans), *D. subobscura* (phylogenetic subgroup: obscura), and one

170 species of the *Dorsilopha* subgenus, *D. busckii* (all three species collected in South-East of France
171 in Spring 2015).

172 All strains and species were maintained at controlled densities (150-200 eggs/40 ml of fly
173 medium) under standard laboratory conditions (25°C or 20°C for recently collected species, 60 %
174 relative humidity, 12:12 light/dark cycle), on a high-protein/sugar-free fly medium (10 % cornmeal,
175 10 % yeast, 0 % sugar). The *D. melanogaster* Canton S strain was also reared on a standard low-
176 protein/sugar-free fly medium (8 % cornmeal, 2 % yeast, 2.5 % sugar) to test for the influence of
177 the medium composition on *Btk* exposure effects.

178

179 *Intoxication method and dose-response assay*

180 Commercial formulations and laboratory spore productions were suspended and diluted in buffer to
181 perform dose-response assays with doses from 10^5 to 10^9 CFU/g of fly medium. All doses were
182 prepared in 100 μ l and homogenized thoroughly with the fly medium (100 μ l/g). *Drosophila* eggs
183 and larvae were collected from stock vials at the suitable developmental stage and transferred
184 carefully to the intoxication vials and dishes, then maintained under standard laboratory conditions
185 until a) the emergence of adults, or, in the larvae survival tests, b) until a given developmental stage
186 was reached from the egg, and c) for 24h. Control groups of individuals were transferred on fly
187 medium homogenized with the same volume of buffer.

188

189 *Development-related traits and larval survival*

190 To evaluate emergence rates and developmental times upon intoxication throughout the entire
191 development, precise numbers of eggs from mass oviposition were transferred to intoxication vials
192 containing fly medium mixed with doses of *Bt* formulations or bacteria productions and let to
193 develop under standard laboratory conditions until the fly emergence. Eggs without chorion and
194 transparent eggs were discarded. The initial number of eggs was adjusted depending on the species
195 biology and the vial size: 20 eggs for 2 g of fly medium in small vials (\varnothing 3.3 cm, surface \sim 8.5 cm²,

196 0.24 g/cm²) for tests with *D. melanogaster* Canton S, 50 eggs for 6 g of fly medium for comparison
197 of *D. melanogaster* strains and *Drosophila* species in wider vials (Ø 4.6 cm, surface ~16 cm², 0.37
198 g/cm²) except for *D. hydei*, *D. suzukii* and *D. immigrans* for which 30 eggs were transferred on 6 g
199 of fly medium. Numbers and sex of emerging flies were recorded once a day until the day the pupae
200 of the next generation should form. From these data, the emergence rate (proportion of emerged
201 flies from the initial eggs; ER), the mean developmental time (mean number of days for completion
202 of development; DT), and the sex-ratio (proportion of male flies; SR) were calculated for each
203 intoxication vial.

204 For the larval survival tests, 20 eggs or larvae from a 4-hour mass oviposition at the indicated
205 developmental stage, were transferred to small dishes containing 1 g of fly medium (Ø 3 cm,
206 surface ~7 cm²) homogenized with increasing doses of Delfin[®] A. Surviving larvae were counted at
207 the indicated developmental stage, or after 24-hour intoxication, to calculate the proportion of
208 surviving larvae. For the test from the egg, eggs which did not hatch were not included in the
209 counting. As a control measurement, we measured the pH of the fly medium in the presence of the
210 dose range of *Bt* formulations (see Supplementary Information S4).

211

212 *Adult fitness-related traits*

213 For the longevity and offspring number tests, males and females emerged from several rearing vials
214 for each dose of Delfin[®] A were pooled when aged 2 days. Groups of 15 males and 15 females were
215 transferred into vials with fresh fly medium without formulation. Fly medium was renewed every 3-
216 4 days. After each fly transfer to fresh food, discarded maintenance vials were incubated under
217 standard laboratory conditions for the offspring to develop. Mortality and sex of dead flies were
218 recorded daily until the last fly died. Offspring numbers were counted from the first emergence
219 until pupae of the next generation appeared. The tests were repeated twice. Due to the variation in
220 the duration of the two longevity experiments, offspring numbers of each vial were summed to
221 obtain a total offspring number per dose of Delfin[®] A for each experiment.

222

223 *Dialysis and Cry toxin analysis*

224 A suspension of 2×10^{10} CFU of Delfin[®] A was dialyzed against PBS (KH₂PO₄ 1.06 mM,
225 Na₂HPO₄(2H₂O) 3mM, NaCl 154 mM, qsp distilled water, pH 7.2), at 250 rpm, 4°C overnight,
226 using an 8-10 kDa MW cut-off membrane (ZelluTrans, Roth[®]). The CFUs of the dialyzed
227 suspension and the effects on emergence rate (ER) and developmental time (DT) were analysed as
228 described above. The dialyzed suspension was also subject to a 12.5 % SDS-PAGE and compared
229 to the non-dialyzed suspension after silver staining. The presence of Cry1A pro-toxins, activated
230 toxins and toxin fragments was probed by Western-blot using an in-house anti-Cry1A rabbit
231 polyclonal antibody.

232

233 *Data analysis*

234 Data on development traits (emergence rate ER and developmental time DT), sex-ratio (SR),
235 survival of larval stages and offspring number were analysed with mixed effect models including
236 the dose of *Btk* formulation/spore production, the *D. melanogaster* strain, the *Drosophila* species or
237 the developmental stage as fixed effects, and replicate (plus the experiment when necessary) as
238 random effects (for ER data, data were analysed with bias-corrected models with replicate as fixed
239 effect to allow pairwise comparisons; similar results obtained with models including replicate as
240 random effect). ER, SR and survival of larval stages were analysed with generalized linear models,
241 with binomial distribution and logit link. DT and offspring number were analysed with linear
242 models. DT were transformed into developmental rates (1/developmental time) to fulfil the
243 assumptions of the analysis of variance (homoscedasticity and residuals normality). Adult longevity
244 data were analysed with proportional hazard Cox regression models including fly sex and dose of
245 *Btk* formulation as fixed effects, and replicates as a random effect. For all the data sets, the main
246 fixed effects and their interactions were tested with log-likelihood ratio tests. *Post hoc* pairwise
247 comparisons were made for pairs of *D. melanogaster* strains, formulation/spore treatments, and

248 between the control dose and the other doses. All the analyses were performed in R^[63] using the
249 packages lme4,^[64] brglm,^[65] multcomp,^[66] survival,^[67] and coxme.^[68]

250

251 **Results**

252 ***Btk* formulations adversely impact the development of *D. melanogaster*.**

253 The wild-type Canton S strain of *D. melanogaster* was used to evaluate the dose-dependent effect
254 of the commercial *Btk* formulation Delfin[®] A on the emergence rate (ER, proportion of emerged
255 flies from the initial egg pool) and developmental time (DT, mean number of days from egg to adult
256 emergence). Eggs were transferred on a standard low-protein/high-sugar fly medium containing
257 Delfin[®] A at doses ranging from 5×10^5 CFU/g of medium (mean equivalent of the maximum
258 recommended doses for one field application; see Methods and Supplementary information S1) to
259 10^9 CFU/g (~ 1,000 times the recommended dose). To check for specific effects of *Btk* formulations
260 and the respective role of *Btk* spores and Cry toxins, we tested the same dose range of the
261 commercial *Bti* formulation Vectobac[®] targeting mosquitoes that contains different Cry toxins,^[22]
262 of the Cry-free strain *Btk* 4D22, and of the *Drosophila* non-pathogenic spore-forming *Bacillus*
263 *subtilis*.

264 Developmental traits (ER and DT) of exposed and non-exposed control flies were similar at
265 doses up to 10^7 CFU/g of Delfin[®] A (Fig. 1a-b; Table 1). At higher doses, both ER and DT were
266 affected in a dose-dependent manner: ER was reduced by 17% at 5×10^7 CFU/g (although not
267 statistically significant), up to 100% at 10^9 CFU/g, dose at which no individual reached the pupal
268 stage. The lethal dose 50 (LD50) was estimated between 5×10^7 and 10^8 CFU/g (Fig. 1a). DT was
269 increased of about 0.5 day at 5×10^7 CFU/g (+4% compared to controls), up to 1.5 days (+14%) at
270 10^8 CFU/g (Fig. 1b; Table 1). The sex-ratio at emergence (SR, proportion of males) was strongly
271 biased towards males at the highest dose at which complete development occurred (10^8 CFU/g),
272 with 58% more males compared to controls (Supplementary information S2). Because addition of
273 *Btk* formulation could modify parameters of the fly medium and thus contribute to these effects, we

274 checked the pH of the medium: the presence of formulation and its dose had no effect on it
275 (Supplementary information S4).

276 We observed no change in ER using the same dose range of the *Btk* Cry-free strain 4D22 (Fig.
277 1a, 1e; Table 1) and the non-pathogenic *Bacillus subtilis* (Fig. 1a, Table 1). Addition of *Bti*
278 Vectobac[®] did not affect ER up to 10⁸ CFU/g but reduced it by 89% at 10⁹ CFU/g (~2,000 times the
279 highest recommended dose for field application; Fig. 1a; Table 1; Supplementary information S1).
280 DT varied with the dose of *Btk* 4D22, the differences being mainly between doses but not with the
281 control. DT increased by ~1.5 days at the highest dose of Vectobac[®] (Fig. 1b; Table 1) and showed
282 a similar trend with *B. subtilis* ($P = 0.06$; Fig. 1b; Table 1). None of these three treatments
283 influenced dramatically the SR, the slight decrease in male proportion for most of the Vectobac[®]
284 doses being due to the higher average sex-ratio for the control dose compared to those for the two
285 other treatments (Supplementary information S2).

286 To test whether these effects are generic to *Btk* formulations, the fly development was
287 evaluated on two other formulations, Delfin[®] B (same brand) and Scutello DF (brand Dipel[®]), at the
288 critical doses 10⁸ and 10⁹ CFU/g. As Delfin[®] A, these formulations contain spores and Cry toxins
289 such as Cry-1A as pro-toxins of ~130 kDa, activated toxins of ~60-70 kDa, but also as smaller
290 fragments^[20] (Fig. 1e, red asterisks). ER remained unchanged at 10⁸ CFU/g whereas no individual
291 reached pupation at 10⁹ CFU/g on Delfin[®] B and very few individuals reached the adult stage on
292 Scutello DF[®], DT being increased by more than 2 days (Fig. 1c-d; Table 1). No significant bias in
293 SR was observed for either formulation (Supplementary information S2).

294

295 **Adverse effects of *Btk* formulation strongly impact the early development.**

296 Larval stages were assessed for their susceptibility to *Btk* formulation in two independent and
297 complementary dose-response tests of survival to Delfin[®] A, at doses ranging from 10⁵ to 10⁹
298 CFU/g of high protein/sugar free medium (this medium is used to rear fly species which are
299 difficult to rear in the lab (see below) and is less limiting for the development of early larval stages).

300 We focused on the 1st and 2nd larval instars, during which growth is exponential,^[47] so that larvae
301 are most heavily exposed to the bioinsecticide. In the first test, the cumulative survival was
302 measured by counting alive late 1st and 2nd instar larvae which have been exposed to Delfin[®] A
303 from the egg stage. Larval survival was not influenced at 10⁷ CFU/g, whereas it decreased for both
304 larval instars above that dose to reach up to 37% mortality at 10⁹ CFU/g (Fig. 2a). Reduced survival
305 tended to occur at a lower dose when cumulative survival was measured later in the development,
306 *i.e.* 10⁹ for late 1st instar larvae and 10⁸ CFU/g for 2nd instar larvae (Fig. 2a; Table 1). For both
307 instars, larvae surviving 10⁹ CFU/g were noticeably smaller and less active than those surviving
308 lower doses. In emergence assays with planned exposure from the egg to the adult stage, none of
309 these individuals reached the pupal stage (see results above). In the second test, larval survival was
310 measured after early 1st and 2nd instar larvae had been exposed for 24 hours to Delfin[®] A. Survival
311 of 1st instar larvae decreased by 36% on 10⁹ CFU/g whereas that of 2nd instar larvae did not change
312 (Fig. 2b, Table 1).

313

314 **Developmental exposure to *Btk* formulation does not strongly influence fitness-related traits**
315 **in adults.**

316 Long-term consequences on flies of exposure to *Btk* formulation throughout the development were
317 evaluated on two fitness-related traits, longevity and total offspring number. Traits were measured
318 on a *Btk*-free low-protein/high-sugar medium after individuals had completed their development on
319 the same fly medium but in presence of selected doses of Delfin[®] A: 5×10⁶ CFU/g, which had no
320 impact on development, and 5×10⁷ and 10⁸ CFU/g, which caused moderate and strong development
321 alterations, respectively (see Fig. 1a).

322 Adult longevity was analysed in two independent sets of experimental replicates on groups of
323 15 females and 15 males held together. Despite large variation between the two sets of experimental
324 replicates (Table 1), the longevity of adults reared on 5×10⁶ CFU/g of Delfin[®] A was similar to that
325 of non-exposed controls (Fig. 3). Males and females which developed on the two higher doses

326 showed a moderate longevity benefit, higher in females for 10^8 CFU/g (Fig. 3a-b, d-e; Table 1).
327 Males generally survived better than females (Table 1) but their longevity benefit of developing on
328 10^8 CFU/g was only observed in the second experiment (Fig. 3b, e).

329 The female offspring number - the sum of offspring produced by the 15 females of each fly
330 group during the longevity experiment - varied depending on both the experiment and the Delfin[®] A
331 dose (Table 1). In the 1st experiment, adults from larvae reared on 10^8 CFU/g had fewer offspring
332 compared to control adults and to adults developed on the other doses whereas the total offspring
333 number varied regardless of the *Btk* dose in the 2nd experiment (Fig. 3c, f, Table 1).

334

335 ***Btk*-formulation dose-dependent alterations of development are not specific to the *D.***
336 ***melanogaster* Canton S strain.**

337 Dose-dependent effects of *Btk* formulation on the development were tested on three additional *D.*
338 *melanogaster* strains: the wild-type Nasrallah (strain 1333), the wild-type Sefra population reared in
339 the laboratory for 4 years, and the double mutant YW1118. The emergence rates (ER) and
340 developmental times (DT) were measured on a high-protein/sugar-free medium (rearing medium of
341 these strains) mixed with Delfin[®] A doses ranging from 10^5 to 10^9 CFU/g. To allow the comparison
342 with previous results with Canton S flies on low-protein/high sugar fly medium, Canton S was also
343 reared and tested on the high-protein/sugar-free medium along with the other strains.

344 None of the fly strains was impacted at doses up to 10^7 CFU/g, whereas ER was strongly
345 reduced and DT was increased at higher doses for all the strains (Fig. 4a-b, Table 1), with no
346 individual reaching the pupal stage at 10^9 CFU/g (LD50 between 10^8 and 10^9 CFU/g). At 10^8
347 CFU/g, the magnitude of effects on Canton S flies was lower than that observed on the low-
348 protein/high-sugar medium. At this dose, ER varied between strains, the largest reduction being
349 observed for Sefra (Table 1). We observed no dose-dependent bias in SR (Supplementary
350 information S3).

351

352 ***Btk* formulation also affects other *Drosophila* species, including the invasive pest *D. suzukii*.**

353 The ER and DT were analysed for seven other *Drosophila* species from different phylogenetic
354 clades at doses of Delfin[®] A from 10⁵ to 10⁹ CFU/g of high-protein/sugar-free medium (rearing
355 medium of all the species). Tested species were *D. simulans* (*D. melanogaster* sister species), the
356 African *D. yakuba*, *D. subobscura*, *D. immigrans*, *D. hydei*, and the invasive *D. suzukii*, all
357 belonging to the *Drosophila* subgenus, and *D. busckii* from the *Dorsilopha* subgenus. For all the
358 species, doses up to 10⁶ CFU/g of Delfin[®] A had no effect on ER and DT whereas all individuals
359 failed to reach the pupal stage and no fly emerged at 10⁹ CFU/g (Fig. 5-6). Amplitudes of
360 development alterations at 10⁷ and 10⁸ CFU/g varied between species (Fig. 5-6; Table 1). All
361 species were affected at 10⁸ CFU/g as was *D. melanogaster* (see Fig. 4a for comparison). *D.*
362 *simulans* and *D. busckii* had unchanged ER, but DT was slightly increased for *D. simulans*
363 (although slightly reduced at 10⁷ CFU/g; similar results with a Japanese strain, data not shown) and
364 strongly increased for *D. busckii* (by 20%, *i.e.* ~ 4 days) (Fig. 5-6, Table 1). *D. yakuba* ER and DT
365 were similar to those of *D. melanogaster*, with an LD50 around 10⁸ CFU/g and a moderate DT
366 increase of ~ 1 day (Fig. 5-6, Table 1; similar results with a strain from Sweden, data not shown).
367 The ER of *D. hydei* and *D. subobscura* were very low at 10⁸ CFU/g (LD50 below this dose), with a
368 high DT (Fig. 5-6; Table 1), while *D. immigrans* did not survive. No *D. suzukii* individual emerged
369 at 10⁸ CFU/g and development was already moderately impacted at 10⁷ CFU/g (Fig. 5-6). No dose-
370 dependent bias in SR was detected for either species (Supplementary information S5).

371

372 **Development alterations may result from a synergy between formulation components.**

373 *Bt* spores and toxins can represent more than half the weight of commercial formulations (85% for
374 Delfin[®], <http://www.certisusa.com>), with up to about 10% of insecticidal protein toxins within this
375 fraction, mainly Cry pro-toxins and activated toxins^[69] (see Fig. 1e). The remaining weight consists
376 of various compounds such as residues of culture medium and various additives including
377 surfactant, anti-foaming agents, etc...^[25,43] It has been shown that, for some products, additives can

378 be more harmful in some cases than the active ingredient,^[70] we explored the role of small
379 diffusible molecular weight components of Delfin[®] A in the alterations of ER and DT of *D.*
380 *melanogaster* Canton S. For that, we mixed a 10 kDa dialyzed suspension of Delfin[®] A at 10⁷, 10⁸,
381 and 10⁹ CFU/g with low-protein/high-sugar medium. ER and DT were unaffected by the presence
382 of the dialyzed suspension at the 10⁷ CFU/g dose, whereas no individual reached the adult stage (no
383 pupation) with the suspension at the 10⁹ CFU/g dose (Fig. 7a; Table 1). At 10⁸ CFU/g, ER was not
384 modified but DT increased by ~ 1 day, only in one of the two experiments, partially reproducing the
385 changes observed without dialysis (Fig. 7a-b; see also Fig. 1a-b, Table 1; 3 independent
386 experiments for ER, 2 independent experiments for DT).

387 Cry1A profiles of dialyzed Delfin[®] A suspensions, like those of the non-dialyzed ones,
388 comprised 130-kDa pro-toxins and 60-70 kDa activated toxins, but also showed toxin degradation
389 as evidenced by additional smaller fragments of activated toxins (Fig. 7c). The respective roles of
390 *Btk* toxin fragments and spores in the alterations of *D. melanogaster* development were further
391 explored through experiments of dialysis such as those previously described, followed by
392 successive centrifugations to eliminate most of the spores and toxin crystals. Despite variation
393 between experiments, ER was strongly affected only in one of the three experiments while DT was
394 always significantly increased when flies were reared in presence of centrifuged supernatants that
395 contained a limited range of Cry 1A toxin fragments (Supplementary information S6).

396

397 **Discussion**

398 The increasing use of bioinsecticides based on *Bacillus thuringiensis* (*Bt*) raises concern about their
399 potential non-intentional side-effects on non-target organisms, and biodiversity in general, due to
400 their partially specific targeting,^[33,34,71] persistence in the environment,^[35,36,38,40,41,44,72] and the
401 requirement of and recommendations for repeated spraying to reach the desired pest control
402 level.^[43] Especially, side-effects of chronic exposure on non-target organisms, including insects
403 present on treated areas, remain under-evaluated. Here, we have tested the side-effects of ingestion

404 of *Bt* bioinsecticide formulations (mainly made of *Bt kurstakii* strains (*Btk*) but also of *Bt israelensis*
405 (*Bti*)) throughout the entire development of several non-target species of *Drosophila* flies which are
406 naturally present in treated areas. While formulation doses up to those recommended for each field
407 sprayings ($\leq 10^6$ CFU/g of medium) had no effect on *Drosophila* development, mortality and/or
408 developmental delay occurred markedly from doses only 10 times and 50 times higher than the
409 maximum recommended dose of the main tested *Btk* formulation for *D. suzukii* (10^7 CFU/g) and the
410 *D. melanogaster* strains (5×10^7 CFU/g), respectively. Development alterations were already strong
411 at these doses, suggesting the occurrence of alterations already at lower doses, i.e. between 10^6 and
412 10^7 CFU/g for *D. suzukii*, and between 10^7 and 5×10^7 CFU/g for *D. melanogaster* strains. Accurate
413 analyses would be needed to verify this possibility. Besides, all the tested species, except *D.*
414 *simulans*, were strongly affected at 10^8 CFU/g, and no (or extremely limited) fly development
415 occurred at the highest tested dose (10^9 CFU/g), equivalent to 1000 times the maximum
416 recommended dose but far below common acute intoxication doses used in many studies.^[5]
417 Recommended doses are those for each spraying on a homogeneous and dry zone without covering
418 areas. In the field, recommended repeated sprayings of stabilized formulation and rainfall washouts
419 following spraying may increase *Bt* spores and toxins presence in both space and time. While the
420 highest dose tested here (10^9 CFU/g) would be hardly reached in the field as those for acute
421 intoxications, the minimal doses at which the fly development was markedly impacted and the
422 lower ones from which changes in development appeared may be readily obtained. Furthermore, the
423 minimal quantity of *Bt* formulation inducing development alterations may be even lower since a
424 single *Drosophila* larva is unlikely to process 1g of medium given its size and feeding rate. Our data
425 also evidenced a window of *Btk* susceptibility during larval development, with the ingestion during
426 the 1st larval instar accounting for a large part of the observed detrimental effects on the
427 development (see below in the discussion).

428 When testing for non-intentional generic effects of *Bt* formulations, similar patterns of
429 development alterations were observed but shifted to higher doses with two other *Btk* formulations

430 and a formulation of *Bti*: there was no effect on *D. melanogaster* development at the doses up to 10^8
431 CFU/g but a strong detrimental effect at the highest dose tested, 10^9 CFU/g. The three *Btk*
432 formulations tested, based on two different bacterial strains (see Methods), had similar profiles of
433 Cry1A protoxins and activated toxins, but they differed in their efficient spore contents, formulation
434 type, and likely additives, which may account for the observed variation in the half-lethal dose. The
435 *Bti* formulation, widely used against Dipteran Nematocera insects (e.g. mosquitoes, black flies),^[73]
436 impacted *D. melanogaster* development only at the highest dose tested. The impacts of *Btk*
437 formulations on *D. melanogaster* development are consistent with growing evidence suggesting a
438 partly specific targeting of *Bt*.^[33,74] Until recently, it has been generally accepted that the mode of
439 action of *Bt* after ingestion by insects relies on key steps of specific binding of proteolyzed *Bt* toxins
440 to receptors of midgut epithelial cells, defining targets for each *Bt* subspecies.^[15,19,22] Several
441 primary and secondary types of toxin receptors, including cadherin-like proteins, aminopeptidases,
442 GPI-anchored alkaline phosphatases,^[10] and more recently the ATP dependent binding cassette
443 reporter C2,^[75] have been identified in Lepidoptera and Diptera mosquitoes. Focusing on the action
444 of *Btk* targeting Lepidoptera, no Lepidoptera cadherin-like Cry receptor orthologues were found in
445 *Drosophila*,^[75] supporting the idea that these flies would not be affected by the spraying of *Btk*
446 formulation. The existence of other types of Cry receptors in *Drosophila* flies would explain the
447 observed developmental impacts but remains to be investigated. In addition, the substantial amounts
448 of active Cry1A toxin fragments in *Btk* formulations could compensate for the possible lack of
449 solubilization of protoxin crystals in the fly midgut and proteolytic activation of toxins by fly gut
450 proteases, both required for Cry activity in insect larvae.^[19] Other toxins synthesized by *Btk* and
451 present in the formulations could also play a role in the observed cross-order activity as some, such
452 as Cry2A, have an insecticidal effect on both Lepidoptera and Diptera.^[76]

453 The lack of effect of ingestion of *Bacillus subtilis* or *Btk* Cry-free 4D22 on the development
454 of *D. melanogaster* excludes that development alterations result from severe disruption of digestion
455 and nutrient uptake/competition in the presence of high spore/bacteria loads in the larval gut

456 throughout development. It supports the idea of a synergistic action of *Btk* spores and Cry toxins,
457 consistent with the models of *Bt* action on insect larvae in which toxins first breach the gut
458 epithelium, allowing the gut content, including *Bt* spores, to colonize the hemocoel.^[19,22-24] The
459 partially reproduced mortality rate and delayed development in dialysis experiments may further
460 indicate that low diffusible molecular weight compounds in *Btk* formulations (e.g., culture media
461 residues, salts, additives) may contribute to these development alterations. This is supported by the
462 lack of impact on *D. melanogaster* development of the ingestion of laboratory-produced spores and
463 Cry toxins of a *Btk* 4D1 strain (or HD1, a reference strain used as a control strain here, not used in
464 commercial formulations) used without additives, even at the highest dose 10^9 CFU/g (additional
465 information S7; Fig. S7a, b). The *Btk* 4D1 culture contained few active Cry toxins and smaller toxin
466 fragments, in contrast to commercial *Btk* formulations (Fig. S7e), supporting the possible
467 contribution of these toxin fragments to the cross-order activity of *Btk* formulations on *Drosophila*.

468 As previously reported for *D. suzukii* exposed to laboratory-produced *Btk* cultures,^[53]
469 mortality of *D. melanogaster* during development on *Btk* formulation occurred early in
470 development. First and second instars larvae are likely highly exposed due to their high feeding rate
471 and their exponential growth.^[77] As the observed larval mortality was only about 40% at the highest
472 dose (10^9 CFU/g) (Figure 2), while none of the individuals reached the pupal stage at this dose
473 (Figure 4), the remaining mortality likely occurred during the third larval stage, maybe due to a
474 delayed action of *Btk* spores and toxins. Interestingly, alterations of the development (mortality and
475 delayed emergence) mimicked those typically generated by nutritional stress conditions in insect
476 larvae.^[78,79] Accordingly, the development alterations were partially rescued on a protein rich fly
477 medium, probably through compensatory protein intake, as in other arthropod species.^[79-81] In
478 addition, the sex ratio of flies was strongly biased towards males after development on the dose of
479 *Btk* formulation affecting fly emergence (10^8 CFU/g) and under low protein conditions. This
480 highlights the importance of nutritional conditions in *Btk* impacts on *Drosophila* development, with
481 sex-specific differences in larval susceptibility to environmental stressors, here the accumulation of

482 *Btk* formulation, under protein restriction conditions as reported previously in *D. melanogaster*.^[82]

483 The development on sublethal doses of *Btk* formulation did not dramatically affect the
484 longevity of *D. melanogaster* adults and the lifetime offspring number. Developmental exposure to
485 *Btk* doses that slightly and strongly reduced the likelihood of reaching the adult stage even gave to
486 males and females a dose-dependent longevity benefit, in addition to the male higher longevity
487 observed in mixed-sex populations,^[83] and slightly increased the offspring number (although not
488 significantly). Surviving the exposure to *Btk* formulation throughout the development has likely
489 selected for fitter individuals. This is similar to the increased longevity of adult insects that have
490 survived developmental nutritional stress,^[84,85] or are resistant to environmental stressors.^[83]

491 The origin of *Drosophila* (species and population/strain) influenced the magnitude of the
492 development alterations induced by *Btk* formulation. For *D. melanogaster*, all the strains tested
493 were susceptible to *Btk* formulation with both mortality and delayed development at the same dose,
494 but with variation in the effect magnitudes. This suggests potential population-specific differences
495 in susceptibility to *Btk* formulation accumulation in the environment, and hence potential spatial
496 and temporal heterogeneity of *Btk* spraying impacts for each *Drosophila* species. Among the other
497 seven species tested, differences in susceptibility to *Btk* formulation, in terms of effect magnitude
498 and type of development alteration (mortality and/or developmental delay), occurred between
499 *Drosophila* species, regardless of their phylogenetic distances. For the *Drosophila* subgenus, *D.*
500 *simulans* was less susceptible than its sister species *D. melanogaster*, whereas the African *D.*
501 *yakuba* experienced similar impacts on the development as *D. melanogaster*. The three species *D.*
502 *immigrans*, *D. subobscura* and *D. hydei* were similarly more susceptible than *D. melanogaster*, but
503 with slight differences in effect magnitudes. The phylogenetically distant *D. busckii* (*Dorsilopha*
504 subgenus) was the least affected of all the species tested in terms of developmental mortality, but its
505 development was strongly delayed. The five species *D. melanogaster*, *D. simulans*, *D. hydei*, *D.*
506 *immigrans*, and *D. busckii* belong to the guild of cosmopolitan domestic *Drosophila* species, *D.*
507 *subobscura* is sub-cosmopolitan species, and *D. busckii* is an opportunistic frugivorous species.^[86]

508 All these species coexist frequently and compete on the same discrete and ephemeral rotting fruit
509 patches, with seasonal variations in the composition of the fly community.^[56-59] At the community
510 level, differences in the species susceptibility to accumulation of *Btk* formulation could modify
511 larval competition conditions and lead to additional local and temporal variations in *Drosophila*
512 communities' composition. The potential side-effects of *Bt* sprays on non-target *Drosophila*
513 communities would be hardly predictable as they would depend on spatial patterns of *Bt*
514 accumulation. A formal mesocosm study of *Drosophila* community dynamics under exposure to
515 *Btk* formulation, at least under semi-field conditions, would allow assessing the consequences of *Bt*
516 accumulation on species competition and community composition.

517 As for the other species, the presence of *Btk* formulation impacts the development of the
518 invasive *D. suzukii*, as recently reported by Cossentine et al.,^[53] this species being the most
519 susceptible here with effects already clearly detectable at 10 times the recommended spraying dose.
520 Compared with the other seven species that live on rotten fruits, *D. suzukii* poses a threat to fruit
521 production because it feeds and lays eggs in healthy ripening soft fruits^[87-89] and hence colonizes
522 orchards and vineyards earlier during the fruit season. Despite its oviposition mode with a saw-like
523 ovipositor inserted inside ripening soft fruits,^[87] the exposure of *D. suzukii* larvae may be quite
524 similar to that of larvae of other *Drosophila* species laying on the surface of fermenting fruits or
525 rotting fruit flesh. Indeed, the saw-like ovipositor likely carries *Bt* spores and toxins from the
526 surface of treated fruits when piercing the fruit skin, and additional *Bt* may then enter the fruit
527 through the drilled holes. From an ecological point of view, the greater susceptibility of *D. suzukii*
528 to the accumulation of *Btk* formulation in the environment might mitigate the potential ecological
529 burden of its invasion for local communities of *Drosophila* frugivorous species in orchards.
530 Alternatively, as *D. suzukii* attacks on fruits can accelerate their decomposition by microorganisms,
531 its higher susceptibility to *Btk* could reduce the number of fruits made suitable for other *Drosophila*
532 species.

533 In conclusion, our results showed that at recommended *Bt* doses, no visible effects on

534 *Drosophila* was observed, but repeated sprayings and subsequent accumulation of *Btk* biopesticide
535 above these doses can potentially impact these non-target insects. The magnitude of this impact
536 possibly depends on both the formulation used and the insect species. Although our study was
537 carried out under controlled laboratory conditions which may dramatically differ from natural
538 conditions encountered in the field (temperature, pH, humidity, food availability, photoperiod,
539 predator/parasite/pathogen pressures, etc...), standard laboratory strains and flies derived from wild
540 populations recently collected exhibited similar patterns of development alterations, suggesting our
541 results may not be specific to laboratory-influenced genetic backgrounds. Recent studies have
542 reported similar adverse side-effects due to repeated sprayings of the *Bti* formulation, directly on
543 non-target organisms,^[31] and indirectly on predators via food webs.^[90] These studies and the data
544 presented here highlight that pest control with *Bt* bioinsecticides should be done with caution in the
545 field to avoid, or at least limit, potential non-intentional side-effects on non-target organisms and
546 hence on biodiversity. At last, *D. melanogaster*, a model species in many research fields, could also
547 serve as a study model to identify the mechanisms underlying these side-effects and/or the potential
548 emergence of resistance to these biopesticides.

549

550 **Acknowledgements**

551 We thank Xiao Han, Jingru Li and Abir Oueslati for help with preliminary experiments, L.
552 Kremmer, C. Rebuf and O. Magliano for providing and rearing flies and help in preparing fly
553 medium, A. Brun-Barale for the *Bacillus subtilis* spores, D. Pauron for preparation of Cry1A toxin,
554 Hugo Mathé-Hubert for advice on statistical analyses, and M. Amichot for helpful discussions. The
555 Cry1A antibody was produced in collaboration with the INRA-PFIE platform (Nouzilly, France).

556

557 **Financial Supports**

558 This work was supported by the French National Agency for Research (ANR-13-CESA-0003-001
559 ImBio), the European Union's Seventh Framework Program for research, technological

560 development and demonstration under grant agreement No. 613678 (DROPSA), the "Investments
561 for the Future" LABEX SIGNALIFE (ANR-11-LABX-0028), the INRA Plant Health Department
562 (to MPNE and JLG), the CNRS (to AG), and the University Nice Côte d'Azur (to MP).

563

564 **Author contributions**

565 AB, MPNE, AG, JLG and MP designed the experiments. AB performed the experiments with
566 contributions of MPNE. AB performed the statistical analyses. AB, JLG, and MP wrote the
567 manuscript with contributions from all the authors.

568

569 **Additional information:**

570 **Supplementary information**

571

572 **Competing interest:** The authors declare no competing interests.

573

574 **References**

- 575 1. United Nations, Department of Economic and Social Affairs, Population Division 2015. World
576 Population Prospects 2015 – Data Booklet (ST/ESA/ SER.A/377).
- 577 2. Pimentel, D., Burgess, M. 2014. Environmental and economic costs of the application of
578 pesticides primarily in the United States. In: Integrated Pest Management. Dordrecht: Springer
579 Netherlands. pp. 47–71.
- 580 3. Devine, G.J., Furlong, M.J. 2007. Insecticide use: contexts and ecological consequences.
581 *Agriculture and Human Values*, 24(3): 281–306.
- 582 4. Sanchis, V., Bourguet, D. 2008. *Bacillus thuringiensis*: applications in agriculture and insect
583 resistance management. A review. *Agronomy for Sustainable Development*, 28(1): 11–20.
- 584 5. WHO report 2007. WHO specifications and evaluations for public health pesticides: *Bacillus*
585 *thuringiensis* subspecies *israelensis* strain AM65-52. World Health Organization, Geneva.

- 586 6. Baldi, I., Cordier, S., Coumoul, X., Elbaz, A. 2013. Pesticides, effets sur la santé. Expertise
587 collective. INSERM. Available at: <http://ipubli-inserm.inist.fr/handle/10608/1>.
- 588 7. Gilden, R.C., Huffling, K., Sattler, B. 2010. Pesticides and health risks. *Journal of Obstetric,*
589 *Gynecologic & Neonatal Nursing*, 39(1): 103–110.
- 590 8. Rizzati, V., Briand, O., Guillou, H., Gamet-Payraastre, L. 2016. Effects of pesticide mixtures in
591 human and animal models: an update of the recent literature. *Chemico-Biological Interactions*,
592 254: 231–246.
- 593 9. Lacey, L.A., Grzywacz, D., Shapiro-Ilan, D.I., Frutos, R., Brownbridge, M., Goettel, M.S. 2015.
594 Insect pathogens as biological control agents: Back to the future. *Journal of Invertebrate*
595 *Pathology*, 132(C): 1–41.
- 596 10. Adang, M.J., Crickmore, N., Jurat-Fuentes, J.L. 2014. Diversity of *Bacillus thuringiensis* crystal
597 toxins and mechanism of action. In: Dhadialla, T.S. and Gill, S.S., *Advances in Insect Physiology*
598 vol. 47, Insect midgut and insecticidal proteins. Oxford Academic Press, pp. 39–87.
- 599 11. Crickmore, N. 2017. *Bacillus thuringiensis* toxin classification. In: Fiuza, L., Polanczyk, R.,
600 Crickmore, N. (eds) *Bacillus thuringiensis* and *Lysinibacillus sphaericus*. Springer.
- 601 12. WHO report 1999. Guideline specification for bacterial larvicides for public health use. WHO
602 document WHO/CDS/CPC/WHOPES/99.2. World Health Organization, Geneva.
- 603 13. Bravo, A., Pacheco, S., Gomez, I., Garcia-Gomez B., Onofre, J., Soberon, M. 2017. Insecticidal
604 proteins from *Bacillus thuringiensis* and their mechanism of action. In: Fiuza, L.M., Polanczyk,
605 R.A., Crickmore, N. (eds). *Bacillus thuringiensis* and *Lysinibacillus sphaericus*, characterization
606 and use in the field of biocontrol. pp. 53-66.
- 607 14. Rabinovitch, L., Vivoni, A.M., Machado, V., Knaak, N., Berlitz, D.L., Polanczyk, R.A., Fiuza,
608 L.M. 2017. *Bacillus thuringiensis* characterization: morphology, physiology, biochemistry,
609 pathotype, cellular, and molecular aspects. In: Fiuza, L.M., Polanczyk, R.A., Crickmore, N. (eds).
610 *Bacillus thuringiensis* and *Lysinibacillus sphaericus*, characterization and use in the field of
611 biocontrol. pp. 1-18.

- 612 15. Palma, L., Muñoz, D., Berry, C., Murillo, J., Caballero, P. 2014. *Bacillus thuringiensis* toxins:
613 an overview of their biocidal activity. *Toxins*, 6(12): 3296–3325.
- 614 16. Ben-Dov, E., Zaritsky, A., Dahan, E., Barak, Z., Sinai, R., Manasherob, R., Khamraev, A.,
615 Troitskaya, E., Dubitsky, A., Berezina, N., Margalith, Y. 1997. Extended screening by PCR for
616 seven cry-group genes from field-collected strains of *Bacillus thuringiensis*. *Applied and*
617 *Environmental Microbiology*, 63(12): 4883–4890.
- 618 17. Ben-Dov, E., Nissan, G., Pelleg, N., Manasherob, R., Boussiba, S., Zaritsky, A. 1999. Refined,
619 circular restriction map of the *Bacillus thuringiensis* subsp. *israelensis* plasmid carrying the
620 mosquito larvicidal genes. *Plasmid*, 42(3): 186-191.
- 621 18. Berry, C., O’Neil, S., Ben-Dov, E., Jones, A.F., Murphy, L., Quail, M.A., Holden, M.T.G.,
622 Harris, D., Zaritsky, A., Parkhill, J. 2002. Complete sequence and organization of pBtoxis, the
623 toxin-coding plasmid of *Bacillus thuringiensis* subsp. *israelensis*. *Applied and Environmental*
624 *Microbiology*, 68(10): 5082-5095.
- 625 19. Bravo, A., Gill, S.S., Soberon, M. 2007. Mode of action of *Bacillus thuringiensis* Cry and Cyt
626 toxins and their potential for insect control. *Toxicon*, 49: 423-435.
- 627 20. Wei, J., Liang, G., Wang, B., Zhong, F., Chen, L., Khaing, M.M., Zhang, J., Guo, Y., Wu, K.,
628 Tabashnik, B.E. 2016. Activation of Bt protoxin Cry1Ac in resistant and susceptible cotton
629 bollworm. *PLoS ONE*, 11(6): e0156560.
- 630 21. Schnepf, E., Crickmore, N., Van Rie, J., Lereclus, D., Baum, J., Feitelson, J., Zeigler, D.R.,
631 Dean, D.H. 1998. *Bacillus thuringiensis* and its pesticidal crystal proteins. *Microbiology and*
632 *Molecular Biology Reviews*, 62(3): 775–806.
- 633 22. Bravo, A., Likitvivatanavong, S., Gill, S.S., Soberon, M. 2011. *Bacillus thuringiensis*: a story of
634 a successful bioinsecticide. *Insect Biochemistry and Molecular Biology*, 41(7): 423–431.
- 635 23. Obata, F., Kitami, M., Inoue, Y., Atsumi, S., Yoshizawa, Y., Sato, R. 2009. Analysis of the
636 region for receptor binding and triggering of oligomerization on *Bacillus thuringiensis* Cry1Aa
637 toxin. *FEBS Journal*, 276: 5949-5959.

- 638 24. Caccia, S., Di Lelio, I., La Storia, A., Marinelli, A., Varricchio, P., Franzetti, E., Banyuls, N.,
639 Tettamanti, G., Casartelli, M., Giordana, B., Ferre, J., Gigliotti, S., Ercolini, D., Pennacchio, F.
640 2016. Midgut microbiota and host immunocompetence underlie *Bacillus thuringiensis* killing
641 mechanism. *Proceedings of the National Academy of Sciences of the United States of America*,
642 113(34): 9486-9491.
- 643 25. Glare, T.R., O'Callaghan, M. 2000. *Bacillus thuringiensis*: Biology, ecology and safety. John
644 Wiley & Sons, UK. pp. 350.
- 645 26. Rubio-Infante, N., Moreno-Fierros, L. 2016. An overview of the safety and biological effects of
646 *Bacillus thuringiensis* Cry toxins in mammals. *Journal of Applied Toxicology*, 36: 630-648.
- 647 27. EFSA Panel on Biological Hazards (BIOHAZ) 2016. Risks for public health related to the
648 presence of *Bacillus cereus* and other *Bacillus* spp. including *Bacillus thuringiensis* in foodstuffs.
649 *EFSA Journal*, doi: 10.2903/j.efsa.2016.4524.
- 650 28. Amichot, M., Curty, C., Benguettat-Magliano, O., Gallet, A., Wajnberg, E. 2016. Side effects of
651 *Bacillus thuringiensis* var. *kurstaki* on the hymenopterous parasitic wasp *Trichogramma chilonis*.
652 *Environmental Science and Pollution Research*, 23: 3097-3103.
- 653 29. Renzi, M.T., Amichot, M., Pauron, D., Tchamitchian, S., Brunet, J.-L., Kretzschmar, A., Maini,
654 S., Belzunces, L.P. 2016. Chronic toxicity and physiological changes induced in the honey bee
655 by the exposure to fipronil and *Bacillus thuringiensis* spores alone or combined. *Ecotoxicology*
656 *and Environmental Safety*, 127: 205-213.
- 657 30. Caquet, T., Roucaute, M., Le Goff, P., Lagadic, L. 2011. Effects of repeated field applications of
658 two formulations of *Bacillus thuringiensis* var. *israelensis* on non-target saltmarsh invertebrates
659 in Atlantic coastal wetlands. *Ecotoxicology and Environmental Safety*, 74: 1122-1130.
- 660 31. Duguma, D., Hall, M.W., Rugman-Jones, P., Stouthamer, R., Neufeld, J.D., Walton, W.E. 2015.
661 Microbial communities and nutrient dynamics in experimental microcosms are altered after the
662 application of a high dose of *Bti*. *Journal of Applied Ecology*, 52: 763-773.
- 663 32. Lagadic, L., Schäfer, R.B., Roucaute, M., Szöcs, E., Chouin S., de Maupeou, J., Duchet, C.,

- 664 Franquet, E., Le Hunsec, B., Bertrand, C., Fayolle, S., Francès, B., Rozier, Y., Foussadier, R.,
665 Santoni, J.-B., Lagneau, C. 2016. No association between the use of Bti for mosquito control and
666 the dynamics of non-target aquatic invertebrates in French coastal and continental wetlands.
667 *Science of the Total Environment*, 553: 486-494.
- 668 33. Venter, H.J., Bøhn, T. 2016. Interactions between Bt crops and aquatic ecosystems: A review.
669 *Environmental Toxicology and Chemistry*, 35(12): 2891–2902
- 670 34. van Frankenhuyzen, K. 2017. Specificity and cross-order activity of *Bacillus thuringiensis*
671 pesticidal proteins. In: Fiuza, L.M., Polanczyk, R.A., Crickmore, N. (eds). *Bacillus thuringiensis*
672 and *Lysinibacillus sphaericus*, characterization and use in the field of biocontrol. pp. 127-172.
- 673 35. Bizzarri, M.F., Bishop, A.H. 2008. The ecology of *Bacillus thuringiensis* on the phylloplane:
674 colonization from soil, plasmid transfer, and interaction with larvae of *Pieris brassicae*.
675 *Microbial Ecology*, 56(1): 133-139.
- 676 36. Raymond, B., Wyres, K.L., Sheppard, S.K., Ellis, R.J., Bonsall, M.B. 2010. Environmental
677 factors determining the epidemiology and population genetic structure of the *Bacillus cereus*
678 group in the field. *PloS Pathogens*, 6(5): e1000905.
- 679 37. Hendriksen, N.B., Hansen, B.M. 2002. Long-term survival and germination of *Bacillus*
680 *thuringiensis* var. *kurstaki* in a field trial. *Canadian Journal of Microbiology*, 48(3): 256–261.
- 681 38. Duchet, C., Tetreau, G., Marie, A., Rey, D., Besnard, G., Perrin, Y., Paris, M., David, J.-P.,
682 Lagneau, C., Desprès, L. 2014. Persistence and recycling of bioinsecticidal *Bacillus*
683 *thuringiensis* subsp. *israelensis* spores in contrasting environments: evidence from field
684 monitoring and laboratory experiments. *Microbial Ecology*, 67(3): 576–586.
- 685 39. Hung, T.P., Truong, L.V., Binh, N.D., Frutos, R., Quiquampoix, H., Staunton, S. 2016a.
686 Persistence of detectable insecticidal proteins from *Bacillus thuringiensis* (Cry) and toxicity after
687 adsorption on contrasting soils. *Environmental Pollution*, 208: 318–325.
- 688 40. Hung, T.P., Truong, L.V., Binh, N.D., Frutos, R., Quiquampoix, H., Staunton, S. 2016b. Fate of
689 insecticidal *Bacillus thuringiensis* Cry protein in soil: differences between purified toxin and

- 690 biopesticide formulation. *Pest Management Science*, 72: 2247–2253.
- 691 41. Enger, K.S., Mitchell, J., Murali, B., Bridsell, D.N., Keim, P., Gurian, P.L., Wagner, D.M. 2018.
- 692 Evaluating the long-term persistence of *Bacillus* spores on common surfaces. *Microbial*
- 693 *Biotechnology*, 56: 3073, 12.
- 694 42. Couch, T.L., 2000. Industrial fermentation and formulation of entomopathogenic bacteria. In:
- 695 Charles, J.-F., Delecluse, A., Nielsen-LeRoux, C. (Eds.), *Entomopathogenic Bacteria: From*
- 696 *Laboratory to Field Application*. Kluwer Academic Publishers, Dordrecht, The Netherlands, pp.
- 697 297–316.43. Brar, S.K., Verma, M., Tyagi, R.D., Valéro, J.R. 2006. Recent advances in
- 698 downstream processing and formulations of *Bacillus thuringiensis* based biopesticides. *Process*
- 699 *Biochemistry*, 41(2): 323–342.
- 700 44. Setlow, P. 2014. Spore resistance properties. *Microbiology Spectrum*, 2(5): TBS-0003-2012.
- 701 45. European Food Safety Authority 2012. Conclusion on the peer review of the pesticide risk
- 702 assessment of the active substance *Bacillus thuringiensis* subsp. *Kurstaki* (strains *ABTS 351*, *PB*
- 703 *54*, *SA 11*, *SA 12*, *EG 2348*). *EFSA Journal*, 10(2): 2540.
- 704 46. Bächli, G. 1999-2008. TaxoDros: the database on the taxonomy of Drosophilidae. Available at:
- 705 <http://www.taxodros.uzh.ch>.
- 706 47. Tennessen, J.M., Thummel, C.S. 2011. Coordinating growth and maturation - Insights from
- 707 *Drosophila*. *Current Biology*, 21(18): R750–757.
- 708 48. Benz, G., Perron, J.M. 1967. The toxic action of *Bacillus thuringiensis* “exotoxin” on
- 709 *Drosophila* reared in yeast-containing and yeast-free media. *Experientia*, 23(10): 871–872.
- 710 49. Saadoun, I., Al-Moman, F., Obeidat, M., Meqdam, M., Elbetieha, A. 2001. Assessment of toxic
- 711 potential of local Jordanian *Bacillus thuringiensis* strains on *Drosophila melanogaster* and *Culex*
- 712 sp (Diptera). *Journal of Applied Microbiology*, 90: 866-872.
- 713 50. Khyami-Horani, H. 2002. Toxicity of *Bacillus thuringiensis* and *B. sphaericus* to laboratory
- 714 populations of *Drosophila melanogaster* (Diptera: Drosophilidae). *Journal Basic Microbiology*,
- 715 42(2): 105-110.

- 716 51. Obeidat, M. 2008. Toxicity of local *Bacillus thuringiensis* isolates against *Drosophila*
717 *melanogaster*. *World Journal of Agricultural Sciences*, 4(2): 161-167.
- 718 52. Obeidat, M., Khymani-Horani, H. and Al-Momani, F. 2012. Toxicity of *Bacillus thuringiensis*
719 β -exotoxins and δ -endotoxins to *Drosophila melanogaster*, *Ephestia kuhniella* and human
720 erythrocytes. *African Journal of Biotechnology*, 11(46):10504-10512.
- 721 53. Cossentine, J., Robertson, M., Xu, D. 2016. Biological activity of *Bacillus thuringiensis* in
722 *Drosophila suzukii* (Diptera: Drosophilidae). *Journal of Economic Entomology*, 0(0): 1-8.
- 723 54. Biganski, S., Jehle, J.A., Kleepies, R.G. 2017. *Bacillus thuringiensis* serovar *israelensis* has no
724 effect on *Drosophila suzukii* Matsumura. *Journal of Applied Entomology*, 142: 33-36.
- 725 55. Haller, S., Romeis, J.X.R., Meissle, M., 2017. Effects of purified or plant-produced Cry proteins
726 on *Drosophila melanogaster* (Diptera: Drosophilidae) larvae. *Scientific Reports*, 1-11.
- 727 56. Shorrocks, B. 1991. Competition on a divided and ephemeral resource: a cage experiment.
728 *Biological Journal of the Linnean Society*, 43: 211-220.
- 729 57. Benado, M., Brncic, D. 1994. An eight-year phenological study of a local drosophilid
730 community in Central Chile. *Journal of Zoological Systematics and Evolutionary Research*, 32:
731 51-63.
- 732 58. Nunney, L. 1996. The colonization of oranges by the cosmopolitan *Drosophila*. *Oecologia*, 108:
733 552-561.
- 734 59. Mitsui, H., Kimura, M.T. 2000. Coexistence of drosophilid flies: aggregation, patch size
735 diversity and parasitism. *Ecological Research*, 15: 93-100.
- 736 60. Withers, P., Allemand, R. 2012. Les drosophiles de la region Rhône-Alpes (Diptera,
737 Drosophilidae). *Bulletin de la Société entomologique de France*, 117(4) : 473-482.
- 738 61. Zeigler, D.R. 1999. *Bacillus* genetic stock center catalog of strains, 7th edition. Part 2: *Bacillus*
739 *thuringiensis* and *Bacillus cereus*. pp. 58.
- 740 62. Gonzales, J.M.Jr, Brown, B.J., Carlton, B.C. 1982. Transfer of *Bacillus thuringiensis* plasmids
741 coding for δ -endotoxin among strains of *B. thuringiensis* and *B. cereus*. *Proceedings of the*

- 742 *National Academy of Sciences of the United States of America*, 79: 6951-6955.
- 743 63. R Development Core Team 2008. R: A language and environment for statistical computing. R
744 Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL
745 <http://www.R-project.org>.
- 746 64. Bates, D., Maechler, M., Bolker, B., Walker, S. 2015. Fitting linear mixed-effects models using
747 lme4. *Journal of Statistical Software*, 67(1): 1-48.
- 748 65. Kosmidis I (2017). *brglm: Bias Reduction in Binary-Response Generalized Linear Models*. R
749 package version 0.6.1, <http://www.ucl.ac.uk/~ucakiko/software.html>.
- 750 66. Horton, T., Bretz, F., Westfall, P. 2008. Simultaneous inference in general parametric models.
751 *Biometrical Journal*, 50(3): 346-363.
- 752 67. Therneau, T.M., Grambsch, P.M. 2000. Modeling survival data: extending the Cox model.
753 Springer, New York, ISBN 0-387-98784-3.
- 754 68. Therneau, T.M. 2015. coxme: Mixed Effects Cox Models. R package version 2.2-5.
755 <https://CRAN.R-project.org/package=coxme>
- 756 69. Koch, M.S., Ward, J.M., Levine, S.L., Baum, J.A., Vicini, J.L., Hammond, B.G. 2015. The food
757 and environmental safety of *Bt* crops. *Frontiers in Plant Science*, 6: 283.
- 758 70. Bradberry, S.M., Proudfoot, A.T., Vale, J.A. 2004. Glyphosate poisoning. *Toxicological*
759 *Reviews*, 23(3): 159-167.
- 760 71. de Souza Machado, A.A., Zarfl, C., Rehse, S., Kloas, W. 2017. Low-dose effects:
761 nonmonotonic responses for the toxicity of a *Bacillus thuringiensis* biocide to *Daphnia magna*.
762 *Environmental Science & Technology*, 51: 1679-1686.
- 763 72. Vettori, C., Paffetti, D., Saxena, D., Stotzky, G., Giannini, R. 2003. Persistence of toxins and
764 cells of *Bacillus thuringiensis* subsp. *kurstaki* introduced in sprays to Sardinia soils. *Soil Biology*
765 *and Biochemistry*, 35(12): 1635-1642.
- 766 73. Becker, N. 2000. Bacterial control of vector-mosquitoes and black flies. In: Charles, J.-F.,
767 Delécluse, A., Nielsen-Roux, C. (ed.). Entomopathogenic bacteria: from laboratory to field

- 768 application. Kluwer Academic Publishers. p. 383-398.
- 769 74. van Frankenhuyzen K. 2013. Cross-order and cross-phylum activity of *Bacillus thuringiensis*
770 pesticidal proteins. *Journal of Invertebrate Pathology*, 114(1): 76–85.
- 771 75. Stevens, T., Song, S., Bruning, J.B., Choo, A., Baxter, S.W. 2017. Expressing a moth abcc2
772 gene in transgenic *Drosophila* causes susceptibility to Bt Cry1Ac without requiring a cadherin-
773 like protein receptor. *Insect Biochemistry and Molecular Biology*, 80: 61-70.
- 774 76. George, Z., Crickmore, N. 2012. *Bacillus thuringiensis* applications in agriculture. In: E.
775 Sansinenea (ed.). *Bacillus thuringiensis* biotechnology. p. 392.
- 776 77. Santos, M., Borash, D.J., Joshi, A., Bounlutay, N., Mueller, L.D. 1997. Density-dependent
777 natural selection in *Drosophila*: evolution of growth rate and body size. *Evolution*, 51(2): 420-
778 432.
- 779 78. Nepoux, V., Haag, C.R., Kawecki, T.J. 2010. Effects of inbreeding on aversive learning in
780 *Drosophila*. *Journal of Evolutionary Biology*, 23: 2333-2345.
- 781 79. Vantaux, A., Ouattarra, I., Lefèvre, T., Dabiré, K.R. 2016. Effects of larvicidal and larval
782 nutritional stresses on *Anopheles gambiae* development, survival and competence for
783 *Plasmodium falciparum*. *Parasites & Vectors*, 9: 226.
- 784 80. Moret, Y., Schmid-Hempel, P. 2000. Survival for immunity: the price of immune system
785 activation for bumblebee workers. *Science*, 290(5494): 1166-1168.
- 786 81. Kutzer, M.A., Armitage, S.A.O. 2016. The effect of diet and time after bacterial infection on
787 fecundity, resistance, and tolerance in *Drosophila melanogaster*. *Ecology and Evolution*, 6(13):
788 4229-4242.
- 789 82. Andersen, L.H., Kristensen, T.N., Loeschcke, V., Toft, S., Mayntz, D. 2010. Protein and
790 carbohydrate composition of larval food affects tolerance to thermal stress and desiccation in
791 adult *Drosophila melanogaster*. *Journal of Insect Physiology*, 56: 336-340.
- 792 83. Khazaeli, A.A., Curtsinger, J.W. 2000. Genetic analysis of extended lifespan in *Drosophila*
793 *melanogaster* III. On the relationship between artificially selected and wild stocks. *Genetica*, 109:

794 245-253.

795 84. Rion, S., Kawecki, T.J. 2007. Evolutionary biology of starvation resistance: what we have
796 learned from *Drosophila*. *Journal of Evolutionary Biology*, 20(5): 1655-1664.

797 85. Burger, J.M.S., Buechel, S.D., Kawecki, T.J. 2010. Dietary restriction affects lifespan but not
798 cognitive aging in *Drosophila melanogaster*. *Aging Cell*, 9: 327-335.

799 86. Atkinson, W., Shorrocks, B. 1977. Breeding site specificity in the domestic species of
800 *Drosophila*. *Oecologia*, 29(3): 223-232.

801 87. Walsh, D.B., Bolda, M.P., Goodhue, R.E., Dreves, A.J., Lee, J., Bruck, D.J., Walton, V.M.,
802 O'Neal, S.D., Zalom, F.G. 2011. *Drosophila suzukii* (Diptera: Drosophilidae): invasive pest of
803 ripening soft fruit expanding its geographic range and damage potential. *Journal of Integrated*
804 *Pest Management*, 2(1): DOI: 10.1603/IPM10010.

805 88. Delbac, L., Rusch, A., Rouzes, R., Ravidat, M.-L., Launes, S., Thiéry, D. 2014. *Drosophila*
806 *suzukii* est-elle une menace pour la vigne? *Phytoma*, 679 : 16-21.

807 89. Poyet, M., Eslin, P., Héraude, M., Le Roux, V., Prévost, G., Givert, P., Chabrierie, O. 2014.
808 Invasive host for invasive pest: when the Asiatic cherry fly (*Drosophila suzukii*) meets the
809 American black cherry (*Prunus serotina*) in Europe. *Agricultural and Forest Entomology*, 1-9,
810 DOI: 10.1111/afe012052.

811 90. Poulin, B., Lefebvre, G., Paz, L. 2010. Red flag for green spray: adverse trophic effects of *Bti*
812 on breeding birds. *Journal of Applied Ecology*, 47: 884-889.

813 **Table 1.** Results of statistical analyses to assess the effect of the dose of formulation/spore
814 production and its interaction with the treatment, the larval instar, the experiment, the sex, the fly
815 strain and the fly species when appropriate. See figures for *post hoc* comparisons of the doses with
816 the control dose.
817

Source of variation/Data	χ^2 / Deviance	d.f.	P value
Development on <i>Btk</i> Delfin[®] A, <i>Btk</i> 4D22, <i>Bti</i> Vectobac[®], <i>Bacillus subtilis</i>			
<i>Emergence rate</i>			
Dose × Treatment	285.7	20	< 0.0001
Dose for each treatment:			
- Delfin [®] A	237.5	6	< 0.0001
- 4D22	7.0	7	0.40
- Vectobac [®]	165.8	5	< 0.0001
- <i>B. subtilis</i>	1.9	6	0.93
<i>Developmental time</i>			
Dose × Treatment	220.8	19	< 0.0001
Dose for each treatment:			
- Delfin [®] A	68.8	6	< 0.0001
- 4D22	16.08	7	0.024
- Vectobac [®]	37.5	6	< 0.0001
- <i>B. subtilis</i>	13.5	7	0.060
Development on <i>Btk</i> Delfin[®] B and Scutello DF (dose effect)			
<i>Emergence rate</i>			
- Delfin [®] B	151.2	2	< 0.0001
- Scutello DF	105.1	2	< 0.0001
<i>Developmental time</i>			
- Delfin [®] B	2.5	1	0.12
- Scutello DF	30.9	2	< 0.0001
Role of formulation components in the development alterations (dialysis)			
Dose effect			
<i>Emergence rate</i>	459.8	3	< 0.0001
<i>Developmental time</i>	13.7	2	0.0011
Survival of larval stages on Delfin[®] A			
<i>Cumulative survival</i>			
Dose × Larval instar	16.2	5	0.0063
Dose for each instar:			
- late 1 st instar	87.4	5	< 0.0001
- late 2 nd instar	25.7	5	0.0001
<i>24-hour survival</i>			
Dose × Larval instar	15.9	5	0.007
Dose for each instar:			
- late 1 st instar	55.9	5	< 0.0001
- late 2 nd instar	3.76	5	0.58
Adult fitness-related traits after development on Delfin[®] A			
<i>Longevity</i>			
Experiment	20.1	1	< 0.0001
<i>- 1st experiment:</i>			
Dose	12.3	3	0.0065
Sex	35.0	1	< 0.0001
(e ^β coefficient males vs females ± se: 0.55 ± 0.16)			
Dose × Sex	20.4	3	0.00014
<i>Sexes analyzed separately</i>			
- females	12.0	3	0.0073
(e ^β coefficients vs control ± se: 5×10 ⁶ : 1.05 ± 0.17, 5×10 ⁷ : 0.71 ± 0.16, 10 ⁸ : 0.60 ± 0.21)			
- males	20.4	3	0.00014
(e ^β coefficients vs control ± se: 5×10 ⁶ : 0.80 ± 0.16, 5×10 ⁷ : 0.66 ± 0.16, 10 ⁸ : 1.53 ± 0.18)			

818

819

Source of variation/Data	χ^2 / Deviance	d.f.	P value
Adult fitness-related traits after development on Delfin[®] A			
<u>- 2nd experiment:</u>			
Dose	16.5	3	0.00090
Sex	31.5	1	< 0.0001
(e ^β coefficient males vs females ± se: 0.45 ± 0.22)			
Dose × Sex	0.69	3	0.88
<u>Sexes analyzed separately</u>			
- females	13.2	3	0.0043
(e ^β coefficients doses vs control ± se: 5×10 ⁶ : 0.92 ± 0.22, 5×10 ⁷ : 0.63 ± 0.21, 10 ⁸ : 0.51 ± 0.21)			
- males	7.01	3	0.072
(e ^β coefficients doses vs control ± se: 5×10 ⁶ : 1.02 ± 0.22, 5×10 ⁷ : 0.70 ± 0.22, 10 ⁸ : 0.64 ± 0.22)			
<u>Total numbers of offspring</u>			
Dose × Experiment	28.1	3	< 0.0001
<u>Dose for each experiment:</u>			
- 1 st experiment	26.3	3	< 0.0001
- 2 nd experiment	4.1	3	0.25
Development of other strains of <i>D. melanogaster</i> on Delfin[®] A (including Canton S)			
<u>Emergence rate</u>			
Dose × Fly strain	105.5	15	< 0.0001
Dose for each fly strain:			
- Canton S	588.6	5	< 0.0001
- Nasrallah	745.3	5	< 0.0001
- Sefra	900.7	5	< 0.0001
- YW1118	636.9	5	< 0.0001
<u>Developmental time</u>			
Dose × Fly strain	9.3	12	0.68
Dose for each fly strain:			
- Canton S	40.3	4	< 0.0001
- Nasrallah	18.0	4	0.0012
- Sefra	27.2	4	< 0.0001
- YW1118	28.9	4	< 0.0001
Development of other <i>Drosophila</i> species on Delfin[®] A			
<u>Emergence rate</u>			
Dose × Fly species	538.2	30	< 0.0001
Dose for each species:			
- <i>D. simulans</i>	461.0	5	< 0.0001
- <i>D. yakuba</i>	750.7	5	< 0.0001
- <i>D. hydei</i>	596.8	5	< 0.0001
- <i>D. immigrans</i>	726.3	5	< 0.0001
- <i>D. subobscura</i>	729.6	5	< 0.0001
- <i>D. sukukii</i>	725.0	5	< 0.0001
- <i>D. busckii</i>	586.0	5	< 0.0001
<u>Developmental time</u>			
Dose × Fly species	59.9	22	< 0.0001
Dose for each species:			
- <i>D. simulans</i>	25.9	4	< 0.0001
- <i>D. yakuba</i>	34.7	4	< 0.0001
- <i>D. hydei</i>	11.5	4	0.022
- <i>D. immigrans</i>	6.01	3	0.11
- <i>D. subobscura</i>	68.8	4	< 0.0001
- <i>D. sukukii</i>	11.7	3	0.0085
- <i>D. busckii</i>	58.8	4	< 0.0001

821 **Figure legends**

822 **Figure 1. Development of *D. melanogaster* Canton S flies on *Btk* and *Bti* commercial**
823 **formulations. (a)** Emergence rate (mean \pm s.e.m.) and **(b)** developmental time (mean \pm s.e.m.) of
824 20 initial eggs on increasing doses of *Btk* Delfin[®] A (red triangles), the Cry-free *Btk* 4D22 (open
825 lozenges), the mosquito-targeting *Bti* Vectobac[®] (green squares), and the non-pathogenic *Bacillus*
826 *subtilis* (light grey circles). For Vectobac[®] and *B. subtilis*, $N = 4-7$ per dose; for Delfin[®] A and *Btk*
827 4D22, $N = 9-12$ for the control, $N = 3$ for $5 \cdot 10^5$ and 10^9 , $N = 4-9$ for 10^6 , $N = 7-14$ from $5 \cdot 10^6$ to 10^8 .
828 **(c)** Emergence rate (mean \pm s.e.m.) and **(d)** developmental time (mean \pm s.e.m.) on increasing doses
829 of the two *Btk* formulations Delfin[®] B (dark red circles) and Scutello DF (orange squares). $N = 4$
830 replicates of 20 eggs per dose and formulation, except for controls and 10^8 CFU/g of Delfin[®] B (9-
831 10 replicates of 20 eggs). Results of *post hoc* comparisons of each dose to the control: $\cdot 0.05 < P < 0.1$;
832 $* 0.01 < P < 0.05$; $** 0.001 < P < 0.01$; $*** P < 0.001$. **(e)** Immunoblotting with an anti-Cry1A polyclonal
833 antibody on proteins from a suspension of laboratory-produced spores of Cry-free *Btk* 4D22, the
834 three *Btk* formulations Delfin[®] A, B, Scutello DF, and a suspension of laboratory-produced Cry1A
835 toxins. Red asterisks indicate the Cry protoxins (~ 130 kDa) and the activated fragments (~ 60 kDa
836 and ~ 70 kDa).

837

838 **Figure 2. Survival of *D. melanogaster* Canton S larval stages on increasing doses of *Btk***
839 **Delfin[®] A. (a)** Proportion of surviving larvae (mean \pm s.e.m.) upon *Btk* exposure from the egg to
840 late 1st instar (open lozenges) and late 2nd instar (black triangles). **(b)** Proportion of surviving larvae
841 (mean \pm s.e.m.) upon 24-hour *Btk* exposure of early 1st instar larvae (open lozenges) and 2nd instar
842 larvae (black triangles). $N = 5-7$ replicates of 20 individuals per dose. Results of *post hoc*
843 comparisons of each dose with the control: $* 0.01 < P < 0.05$; $** 0.001 < P < 0.01$; $*** P < 0.001$.

844

845

846 **Figure 3. Fitness-related traits of adults (longevity and total offspring number) after**
847 **development on *Btk Delfin*[®] A.** (a, d) Female longevity (mean survival fraction over time \pm s.e.m.),
848 (b, e) Male longevity (mean \pm s.e.m.), and (c, f) total offspring number (mean \pm s.e.m.), measured
849 on individuals that developed without *Btk* (blue items) and on 5×10^6 CFU/g of *Btk Delfin*[®] A (green
850 items), 5×10^7 CFU/g (red items), and 10^8 CFU/g (dark red items). Data from 2 experiments (a-c,
851 experiment 1; d-f, experiment 2). For each trait, $N = 3$ -5 replicates of 15 males and 15 females per
852 dose in experiment 1, $N = 3$ replicates of 15 males and 15 females in experiment 2. Results of *post*
853 *hoc* comparisons of each dose with the control: * $0.01 < P < 0.05$; ** $0.001 < P < 0.01$; *** $P < 0.001$.

854

855 **Figure 4. Development of four *D. melanogaster* strains on increasing doses of *Btk Delfin*[®] A.** (a)
856 Emergence rate (mean \pm s.e.m.), (b) Developmental time (mean \pm s.e.m.) of the strains Canton S
857 (blue lozenges), Nasrallah (yellow triangles), Sefra (green squares), and YW1118 (red circles). $N =$
858 4 groups of 50 eggs per dose and fly strain for each trait. Results of *post hoc* comparisons of each
859 dose to the control: $\cdot 0.05 < P < 0.1$; * $0.01 < P < 0.05$; ** $0.001 < P < 0.01$; *** $P < 0.001$.

860

861 **Figure 5. Emergence rate of seven *Drosophila* species on increasing doses of *Btk Delfin*[®] A.**
862 Mean emergence rate (\pm s.e.m.). $N = 4$ replicates of 50 eggs per dose for *D. simulans*, *D. yakuba*, *D.*
863 *subobscura*, and *D. busckii*, $N = 4$ replicates of 30 eggs per dose for *D. hydei*, *D. suzukii*, and *D.*
864 *immigrans*. Results of *post hoc* comparisons of each dose with the control: $\cdot 0.05 < P < 0.1$; *
865 $0.01 < P < 0.05$; ** $0.001 < P < 0.01$; *** $P < 0.001$.

866

867 **Figure 6. Developmental time of seven *Drosophila* species on increasing doses of *Btk Delfin*[®] A.**
868 Mean developmental time (\pm s.e.m.). $N = 4$ replicates of 50 eggs per dose for *D. simulans*, *D.*
869 *yakuba*, *D. subobscura*, and *D. busckii*, $N = 4$ replicates of 30 eggs per dose for *D. hydei*, *D. suzukii*,
870 and *D. immigrans*. Results of *post hoc* comparisons of each dose with the control: * $0.01 < P < 0.05$;
871 ** $0.001 < P < 0.01$; *** $P < 0.001$.

872 **Figure 7. Evaluation of the role of small molecular weight components of *Btk* Delfin[®] A**
873 **(dialysis; membrane cut-off: 8-10 kDa) in the altered development of *D. melanogaster* Canton**
874 **S. (a) Emergence rate (mean \pm s.e.m.) and (b) developmental time (mean \pm s.e.m.) on increasing**
875 **doses of dialyzed Delfin[®] A. $N = 3$ experiments of 4 replicates with 20 eggs per dose for the**
876 **emergence rate, $N = 2$ experiments of 4 replicates per dose for the developmental time. Results of**
877 ***post hoc* comparisons of each dose with the control: \cdot $0.05 < P < 0.1$; $*$ $0.01 < P < 0.05$; $**$**
878 **$0.001 < P < 0.01$; $***$ $P < 0.001$. (c) Anti-Cry1A probed immunoblot of non-dialyzed (ND) and**
879 **dialyzed (D) suspensions showing the decrease in the amount of \sim 130 kDa protoxins and the**
880 **increase in that of \sim 60/70 kDa activated toxins after dialysis.**

Figure 1

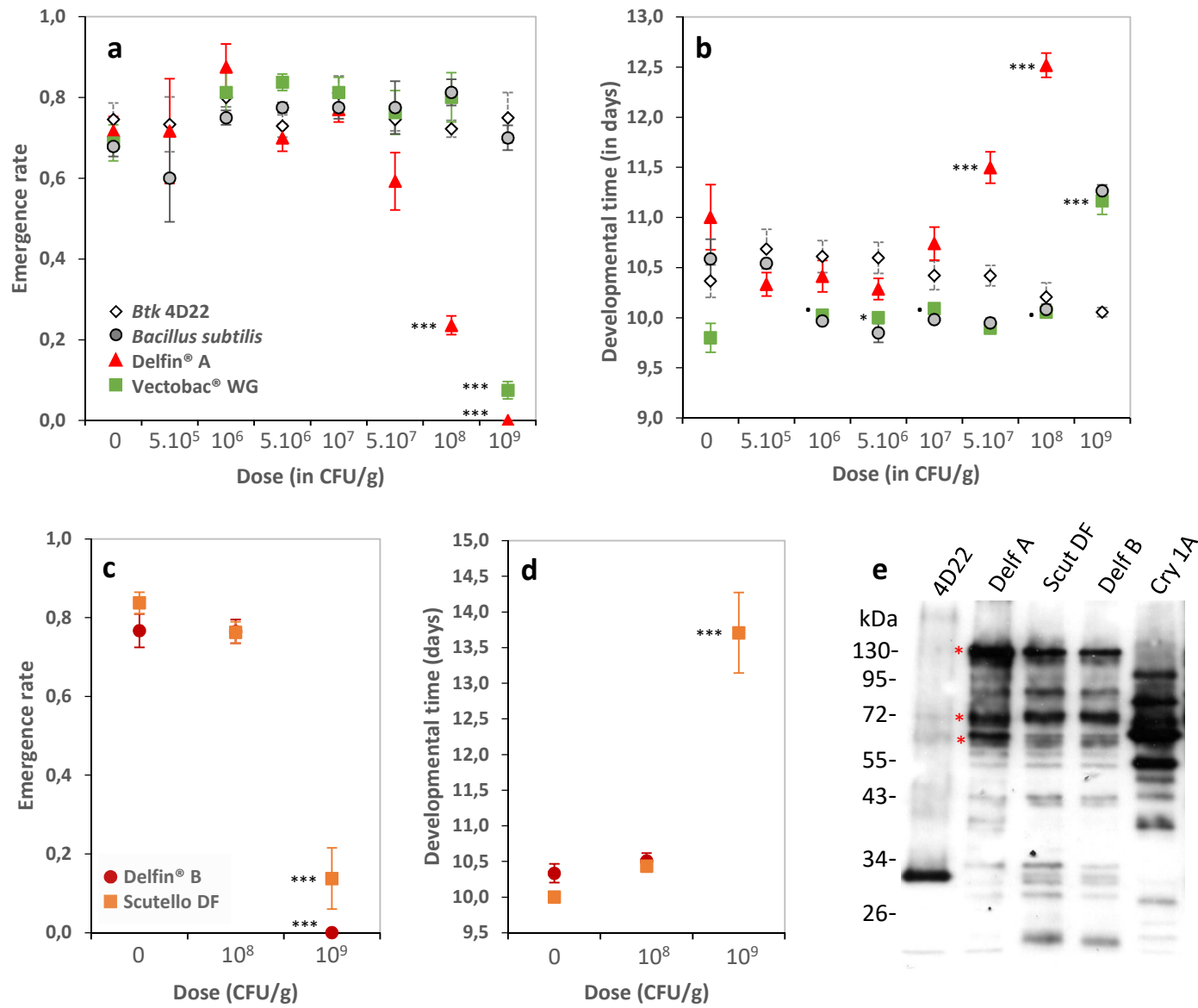


Figure 2

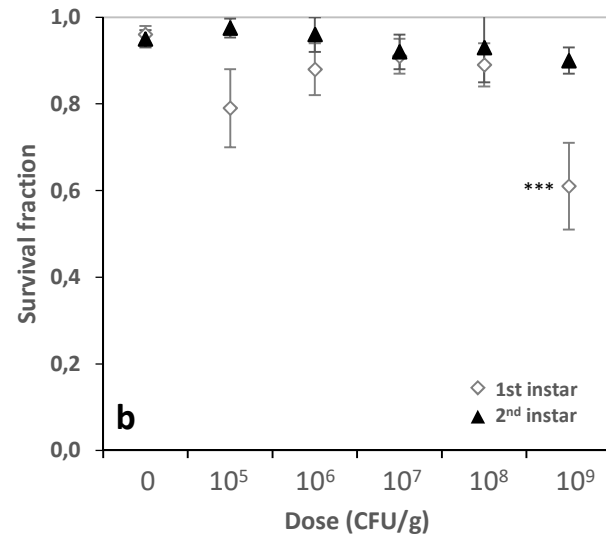
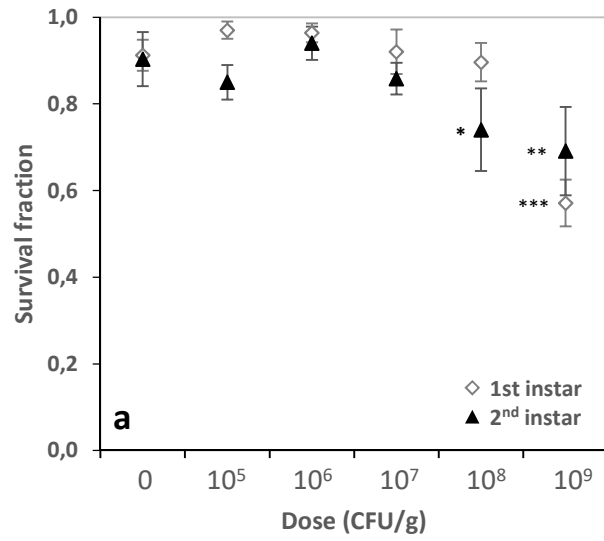


Figure 3

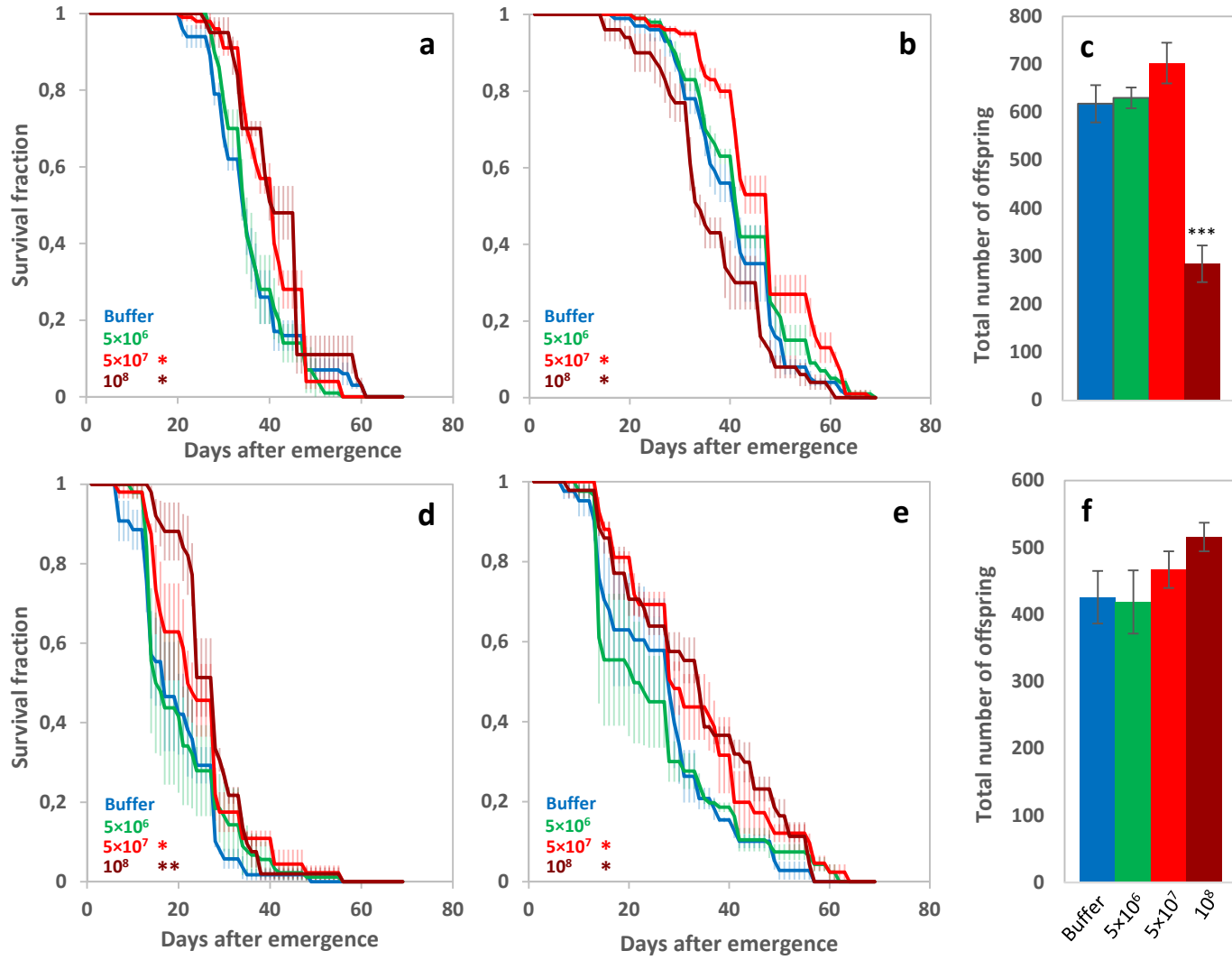


Figure 4

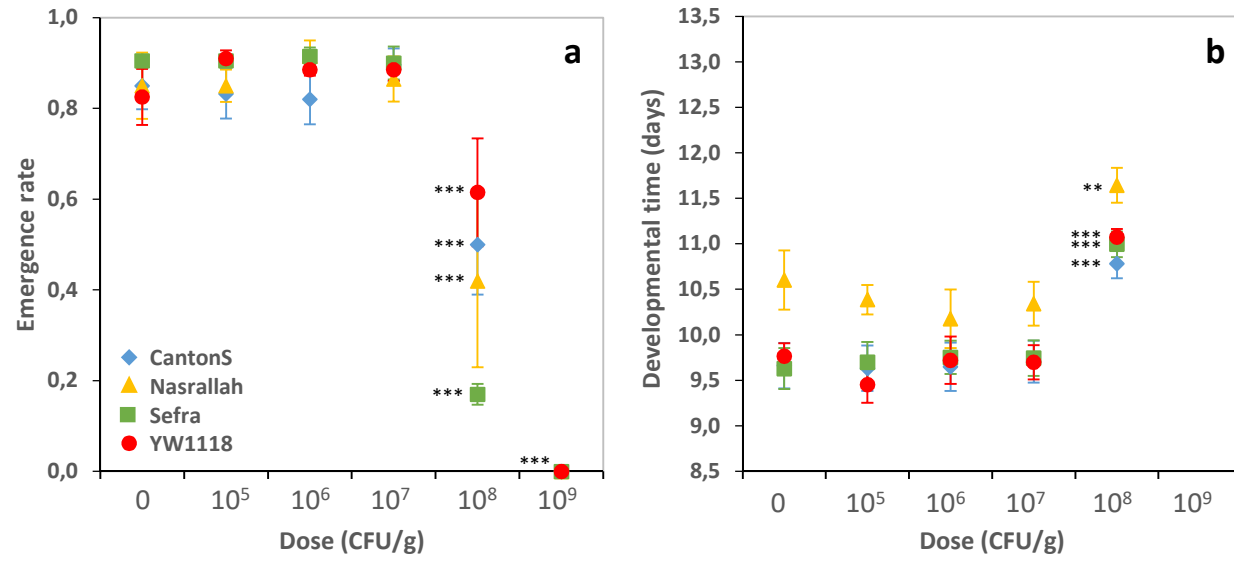


Figure 5

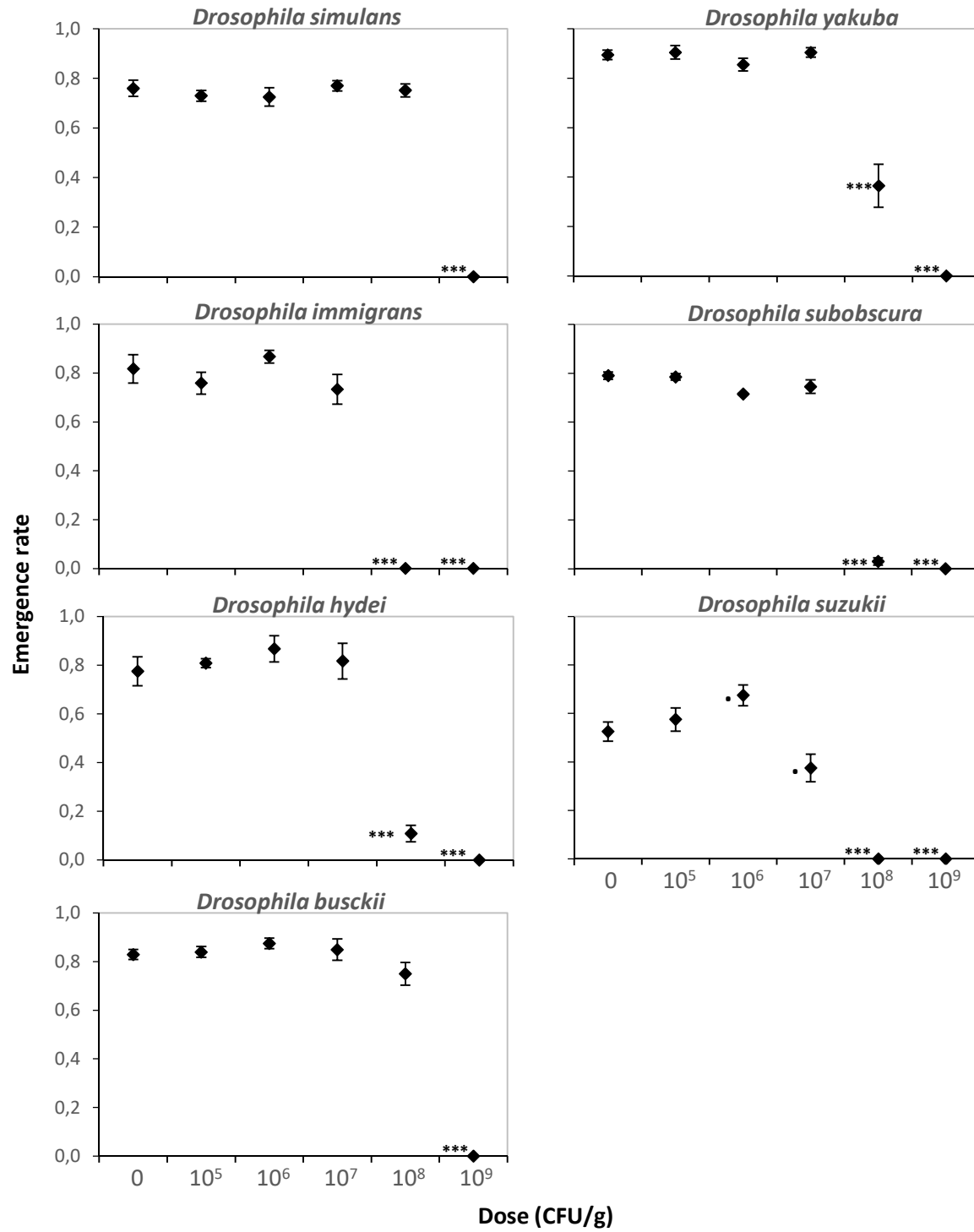


Figure 6

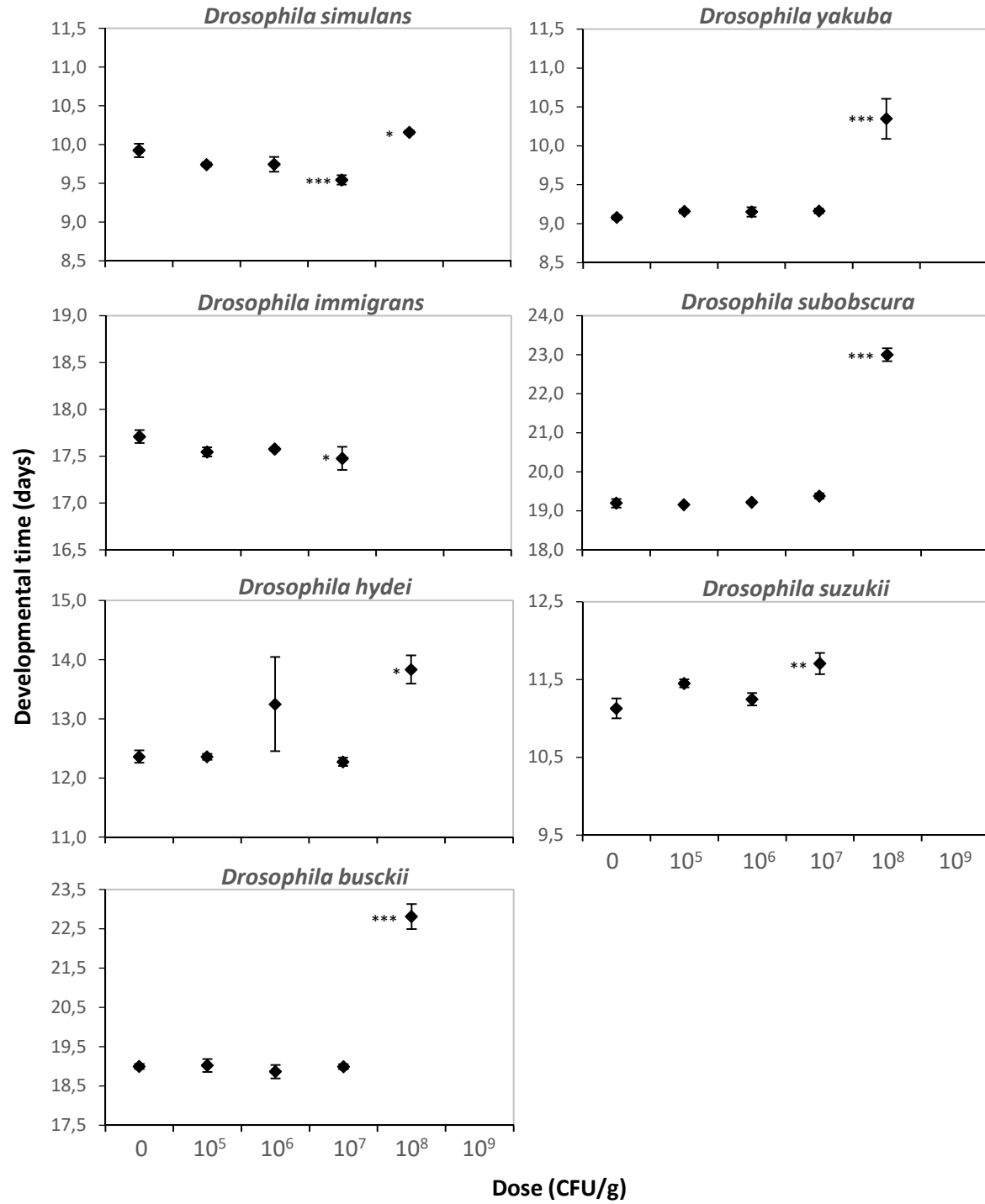


Figure 7

