Title: Soil type shapes unique pathogen communities on nearby populations of a California

native bunchgrass

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Abstract

The role of infectious disease in regulating host populations is increasingly recognized,

but the environmental conditions that facilitate versus hinder pathogen-mediated population

regulation remain poorly understood. We compared the relative contributions of soil type and

pathogen community composition to foliar disease burden in a perennial bunchgrass species

found on two distinct soil types that support distinct plant communities in a California grassland.

We hypothesized that populations on different soil types would have significantly different

disease burdens caused by unique pathogen communities. To test this hypothesis, we compared

foliar disease burden and foliar fungal pathogen communities in nearby populations of *Stipa*

pulchra found in nonserpentine greenstone soil that is dominated by invasive Mediterranean

grasses, and in serpentine soil, a harsh soil high in heavy metals and low in essential nutrients

that supports a diverse community of native plant species. We analyzed the chemical makeup of

serpentine and nonserpentine plant tissue to understand potential impacts of soil chemistry on

plant health and pathogen community composition. We found that serpentine and nonserpentine

S. pulchra experienced consistent, low disease pressure caused by distinct communities of foliar

pathogens, and that serpentine plants, like the soil in which they grew, had elevated Ni and Mg

content and decreased C, N, Ca, and P content compared to nonserpentine plants. The results

imply that pathogens are unlikely to regulate the population dynamics of this native plant, and

that pathogen communities are structured either by plant community composition or tissue

chemistry. Local variation in soil type and annually variable conditions associated with high

species turnover in pathogen communities may create a refuge from disease outbreaks for *S.*

pulchra, contributing to the low disease burden observed on this and other Mediterranean

grassland species.

Introduction

How environmental heterogeneity affects species interactions, and ultimately population

growth, remains a fundamental ecological question. Infectious disease, which arises from

interactions between hosts, parasites, and the environment, is particularly likely to vary across

populations in differing environments. Both biotic factors, such as the phylogenetic structure of

the host community, and abiotic factors, such as elevation, have been associated with the

presence of specific diseases and the severity of their impacts on their hosts (Abbate and

Antonovics 2014, Parker et al. 2015). Additionally, mounting evidence suggests that specific

characteristics of pathogen community structures, including the identities and abundances of the

diseases present in a host population, mediate the outcomes of host – pathogen interactions

(Seabloom et al. 2015, Borer et al. 2016). Together, these observations highlight the importance

of understanding interactions between environmental variation, pathogen community

composition, and pathogen pressure on host populations to assess the potential for disease to

regulate host population dynamics. For plant hosts, such assessments are important both in

conservation and agricultural contexts because identifying populations at high risk for severe

disease outbreaks and applying preventative measures may prevent local species extinctions and

improve crop yields, respectively. While studies have frequently explored the interactions of

single host – pathogen pairs under differing abiotic conditions (e.g., Abbate and Antonovics

2014) or the effect of varying pathogen community composition on disease burden (e.g., Telfer

et al. 2010), relationships between abiotic variables and pathogen community structures remain

largely unexplored in natural systems (but see Spear 2017).

California grasslands are an ideal model system in which to study interactions between

environmental variation, pathogen community composition, and disease burden because these

plant communities are widespread, exist across a variety of environmental gradients, and include

distinct plant community subtypes (Eviner 2016). Here, we take advantage of the natural

occurrence of the native perennial bunchgrass *Stipa pulchra* on distinct soil types—serpentine,

hosting diverse, native-dominated grassland communities and a nonserpentine greenstone

(hereafter, nonserpentine), dominated by exotic plants—to understand the extent to which soil

type and time influence the landscape of disease encountered by a single plant species, including

foliar fungal pathogen community composition and disease severity.

Differences in plant community composition and soil chemistry suggest that *S. pulchra*

populations growing in distinct soil types are likely to host distinct fungal pathogen communities

and/or experience differing disease burdens, particularly if the phenotype of *S. pulchra* is

influenced by the soil in which it grows. *S. pulchra* naturally occurs in both nonserpentine soils

and in relatively rare serpentine soils that host grassland plant communities. These two soil types

differ in their chemistry and host distinct plant communities, resulting in a suite of environmental

differences that may influence pathogen infection, reproduction, and persistence (Huenneke et al.

1990, Harrison and Viers 2007). In contrast to the majority of California grasslands, which are

heavily invaded by exotic species, serpentine grasslands have a diverse array of native species

adapted to chemically and physically harsh soils: low in essential plant nutrients (N, P, K, Ca),

high in heavy metals that are potentially toxic to plants (Co, Ni, Cr, Fe), and rocky and shallow,

with poor moisture retention (Proctor and Woodell 1975, Oze et al. 2004, 2008, McNaughton

1968, Harrison and Viers 2007, Eviner 2016).

Soil properties determine a variety of factors likely to influence the occurrence and

outcomes of plant infections, including soil moisture levels and plant chemistry and physiology

(Cunningham et al. 1999, McElrone et al. 2005, Smith 2007). Spores of different fungal species

vary in both their ability to survive in dry and moist environments and their tolerance of heavy

metals (Coley-Smith and Cooke 1971, Iram et al. 2009). Moreover, plants of the same species

grown in drier and lower-nutrient soils can have thicker epidermises and higher concentrations of

phenolic compounds, many of which have antifungal properties (Osbourne 1996, Cunningham et

al. 1999). Finally, soil chemistry may influence pathogen infection either through

hyperaccumulation of heavy metals, which can bolster chemical defenses against natural

enemies, or by limiting common mechanisms of plant defense (Springer et al. 2006, Rascio and

Navari-Izzo 2011).

In this study, we ask: (1) how does disease burden differ on plants growing in different

soil types? (2) how does the species composition of pathogen communities differ on plants

growing in different soil types? and (3) by what mechanisms—including host community

composition and plant tissue chemistry—does soil type affect disease burden and pathogen

composition? We hypothesize that both disease burden and pathogen species composition differ

on *S. pulchra* individuals growing in serpentine versus nonserpentine soils, either because of

differences in the composition of the surrounding plant community, differences in plant

physiological responses, or both. We further hypothesize that the soil type *S. pulchra* grows in

alters foliar tissue chemistry in ways that may affect disease burden, with higher heavy metal

content and/or lower C, N, and P content in serpentine plants. If so, these foliar chemistry

differences may translate into either (a) higher disease in serpentine plants relative to

nonserpentine plants, facilitated by negative impacts of limited plant resource availability on

plant health (as in Springer 2006), or (b) lower disease in serpentine plants due to heightened

chemical and/or physiological plant defenses in response to serpentine soil properties. The results

of this study will aid in assessing disease risk for plant populations in natural and agricultural

contexts by providing information about the spatial and temporal scales over which soil type and

other environmental factors can affect the spread of outbreaks.

Methods

All surveys and sample collection took place at Jasper Ridge Biological Preserve (JRBP),

a 485-hectare site in San Mateo County, CA managed by Stanford University (37.4° N, 122.2°

W). JRBP has a Mediterranean climate with cool winters (mean 9.2°C), warm summers (mean

20.1°C), and annually variable precipitation that predominantly occurs in the winter and spring

(averaging 622.5 mm per year) (Ackerly et al. 2002). The growing season occurs during the

rainy season and plants senesce in the early summer. This study was conducted during the

growing seasons from mid-April to mid-May in 2015, 2016, and 2017. There was highly variable

annual precipitation across the three years (491 mm, 604 mm, and 985mm in 2015, 2016, and

2017, respectively; Weather Underground).

Substantial populations of *S. pulchra* occur in both serpentine and nonserpentine

grasslands at JRBP (McNaughton 1968). The serpentine plant communities at JRBP are more

diverse than the nearby nonserpentine plant communities and include many native grasses and

forbs, in contrast to the invasive-dominated nonserpentine grasslands (McNaughton 1968). Plant

species commonly found in serpentine grasslands at JRBP include *Stipa pulchra*, *Elymus*

glaucus, *Elymus multisetus, Eschscholzia californica*, and the invasive grass *Bromus hordeaceus*

(McNaughton 1968, Hobbs and Mooney 1985). The nonnative annual species that dominate the

nonserpentine grasslands but are absent from serpentine grasslands include *Avena barbata,*

Avena fatua, *Bromus diandrus,* and *Erodium botrys* (McNaughton 1968, Hobbs and Mooney

1985).

Quantification of disease

To assess plant disease burden, we quantified the percentage of living leaf area exhibiting

symptoms of fungal disease in serpentine and nonserpentine populations of *S. pulchra*. We

selected three serpentine grassland sites and three nonserpentine grassland sites such that each

serpentine site was paired with a nearby (~170m away) nonserpentine site. Individual sites

within each soil type were >300m apart. Each year, we randomly placed four 5-meter transects at

each site and assessed the infection status of the *S. pulchra* individual nearest each meter mark,

resulting in five individual plants per transect surveyed each year (N = 60 plants per soil type per

year). We visually estimated the percentage of living foliar tissue damaged by pathogens for six

arbitrarily-selected leaves per plant. All survey sites were located along the ridgetop at JRBP,

where serpentine and nonserpentine (greenstone) soils are present in discrete adjacent bands

(Figure 1a, Oze et al. 2004). The relative flatness of the ridgetop ensured that all survey sites

were at roughly the same elevation and had similar slopes, aspects, and water availability (Oze et

al. 2008).

We compared the mean percentage of diseased leaf area and proportion of surveyed

leaves with disease in serpentine and nonserpentine *S. pulchra* using Welch's two-tailed t-tests

for the data over all years. We used analysis of variance (ANOVA) to test for the effects of year

on disease levels in serpentine and nonserpentine *S. pulchra*. When ANOVA results were

statistically significant, we used pairwise t-tests with Bonferroni adjustments of p-values to

account for multiple comparisons to identify the interactions contributing to the significant

result.

Isolation and molecular identification of foliar fungal pathogens

To identify the fungal foliar pathogens infecting serpentine and nonserpentine *S. pulchra,*

we harvested one segment of symptomatic leaf tissue for culturing and identification of the

causative fungal pathogen(s) from each surveyed individual with disease. We excised, surface

sterilized, and plated a $2mm^2$ square of tissue from each sample at the leading edge of infection.

We observed the plates for growth for eight weeks and isolated morphologically distinct hyphae

into axenic culture. We extracted and sequenced genomic DNA from the ITS region for each

isolate. Fungal isolation and sequencing methods followed Spear and Mordecai (2018). We

estimated taxonomic placement of fungal isolates by grouping sequences into operational

taxonomic units with 97% sequence similarity (OTUs), a proxy for species delineation based on

the range of intraspecific ITS divergence (O'Brien et al. 2005), and then comparing assembled

DNA sequences to the entries in the UNITE fungal database (see Supplemental Information for

additional details).

Analyses of fungal community composition

To characterize pathogen community composition in serpentine and nonserpentine *S.*

pulchra populations, we compared the diversity and assessed the similarity of communities of

fungi cultured from the diseased tissue of surveyed plants. To describe fungal community

diversity, we (1) calculated observed taxa richness and Fisher's alpha, a widely used measure of

species richness that is robust to uneven sample sizes (Fisher et al. 1943, Magurran 1988, Hansen

and Coleman 1998); (2) generated taxa accumulation curves to understand sampling efficacy;

and (3) counted the number of fungal isolates in each genus, family, order, and class from each

grassland type. We compared fungal community composition across grassland types and across

years visually, using non-metric multidimensional scaling (NMDS), and statistically, using

permutational multivariate analyses of variance (PERMANOVA) with the Adonis and

pairwise.perm.manova functions (Anderson 2001, Oksanen et al. 2016, Hervé 2017). To account

for low sample size in some transects, we considered the combined isolates from two transects at

each site in each year to be a distinct community (for a total of two communities per site per

year) for these analyses. For the NMDS and PERMANOVA analyses, we used the function

vegdist with the abundance-based Chao method, accounting for unsampled species, to create a

matrix of pairwise community dissimilarities (Chao et al. 2016, Oksanen et al. 2016). Finally, we

used the Morisita-Horn index, which is independent of sample size and diversity, to make

pairwise comparisons of similarity within and between the serpentine and nonserpentine fungal

communities from each year (N = 265 isolates, 200 bootstrap replicates) (Wolda 1981, Jost et al.

2011).

Chemical analyses of leaf tissue

We analyzed C, N, Ca, Cr, Mg, Ni, and P content of the foliar tissue of *S. pulchra* grown

in serpentine and nonserpentine soils *in situ* and in pots. In 2016, we collected the youngest fully

expanded and entirely asymptomatic leaf of each individual surveyed in the field (N = 58

nonserpentine, 60 serpentine). We dried the samples at room temperature for three months, then

at 60°C for 48 hours, ground them to a powder, and then dried them again at 60°C for 24 hours.

We measured C and N content of these samples with a Carlo Erba NA-1500 elemental analyzer

using atropine standards.

To compare Ca, Cr, Mg, Ni, and P content, which requires more tissue biomass than

could be collected from plants in the field, we used a set of plants grown in pots in the

greenhouse and then moved to JRBP for a separate experiment (see Supplement for additional

information). In total, we grew 480 plants in serpentine soil sourced from our survey sites at

JRBP and 480 plants in nonserpentine greenhouse soil (PROMIX HP Mycorrhizae TM High

Porosity Professional Growing Medium, Premier Horticulture Inc.). To ensure that serpentine-

and nonserpentine-grown plants encountered similar soil biota, we triple autoclave-sterilized

serpentine soil (40-minute cycles at 120°C) and then immediately inoculated it with greenhouse

soil in a ratio of 93:7 by weight (serpentine soil treatment; Heinze et al. 2015) and mixed

sterilized serpentine soil into greenhouse soil in the same ratio (nonserpentine soil treatment, to

account for possibility of microorganisms surviving autoclaving). Autoclave sterilization of soil

has been shown to alter physical and chemical properties and thus has the potential to affect our

results. However, the documented instances of autoclaving-induced changes to extractable

levels of the elements we measured, including increased extractable Ca:Mg and decreased

extractable Ni, suggest that this procedure is more likely to increase than decrease the similarity

of nutrient availability in serpentine and nonserpentine soils (Tadros 1957, Wolf et al. 1989,

Abou-Shanab et al. 2003). Therefore, any significant differences observed between the foliar

chemistry of *S. pulchra* grown in autoclaved serpentine and unsterilized nonserpentine soils that

indicate plant tissue chemistry matching soil chemistry should not be due to autoclave

sterilization effects.

Elemental content of these plants was measured via inductively coupled plasma optical

emission spectrometry (ICP-OES). Because the minimum amount of leaf tissue for which some

heavy metals can be detected via ICP-OES is 0.5g and most of our *S. pulchra* had a total

aboveground biomass of less than 0.5g, each measurement represents the combined elemental

content of equal weights of oven-dried foliar tissue from two plants. Plants were randomly

selected for analysis from among all individuals with an aboveground biomass of at least 0.29g.

All asymptomatic leaves were removed from each plant's base and processed as described

above. A 0.5g portion of each pooled sample was digested in 10ml of 70% trace metal quality

nitric acid with microwave assistance. Digested samples were diluted to 16.7% acid

concentration with DI water and the concentrations of Ca, Cr, Mg, Ni, and P were measured in

parts per billion by a Thermo Scientific iCAP 6300 spectrometer (Thermo Scientific, Waltham,

MA, USA) (Cantarelli et al. 2010).

To assess the potential for soil chemistry to mediate plant-pathogen interactions, we

related soil type to leaf tissue chemistry of *in situ* and potted *S. pulchra*, and chemistry of *in situ*

plants to disease burden. We compared the mean mol/kg dried weight of leaf tissue for Ca, Cr,

Mg, and Ni for potted serpentine and nonserpentine plants with Wilcoxon rank sum tests. We

also compared mean C:N and Ca:Mg ratios that are indicative of plant health (Ciompi et al.

1996, Brady et al. 2005). We used Pearson's product-moment correlation to test for significant

monotonic relationships between the amount of diseased tissue observed on individual plants in

the field and their foliar C and N content.

All statistical analyses were done in R version 3.4.1. We used the packages ggplot2

(Wickham 2016), vegan (Oksanen et al. 2016), SpadeR (Chao et al. 2016), bipartite (Dormann et

al. 2008), fossil (Vavrek 2011), BiodiversityR (Kindt and Coe 2005), rich (Rossi 2011), and

RVAideMemoire (Hervé 2017).

Results

Pathogen damage

Pathogen damage was similarly ubiquitous and low-burden in serpentine and

nonserpentine soils. Every plant surveyed exhibited evidence of foliar fungal-caused disease.

The mean percentage of diseased leaf area observed across all years was 1.66 for nonserpentine

and 1.63 for serpentine *S. pulchra,* respectively (t = 0.11, df = 356.2, p-value = 0.91; Figure 1b),

and the mean proportion of pathogen-damaged leaves was 0.83 for both soil types. Although the

percentage of diseased tissue on serpentine plants increased slightly each year from 2015 to 2017

(F-value = 4.49, p-value = 0.035), the Bonferroni-adjusted p-values for t-tests for each pair of

years were not statistically significant (p-values = 0.16, 0.45, 1.00). There was no statistically

significant effect of year on nonserpentine pathogen damage (F-value = 1.35, p-value=0.246).

Fungal pathogen community

We isolated 267 unique fungal isolates from 258 symptomatic tissue pieces with fungal

growth (144 nonserpentine, 114 serpentine from 36 nonserpentine and 34 serpentine transects,

out of 360 total tissue pieces). Of the 267 unique fungal isolates, we successfully sequenced 256

isolates. The sequenced isolates clustered into 30 operational taxonomic units (OTUs) based on

97% sequence similarity, representing 23 genera, 15 families, 8 orders, and 5 fungal classes. The

largest OTU contained 48 isolates and the smallest OTUs consisted of one isolate each. OTUs

are hereafter referred to as "species" (O'Brien et al. 2005).

The serpentine fungal community was more diverse than the nonserpentine community at

every taxonomic level (Tables S-1 to S-6). Twenty-two fungal species were isolated from

nonserpentine (Fisher's alpha = 7.261, 95% CI = 3.597, 10.925) and 24 from serpentine *S.*

pulchra (Fisher's alpha = 9.324, 95% CI = 4.664, 13.983); 16 of these species (53%) were shared

(Figure 2, Tables 1 and S-1). Novel genera were isolated for both community types in every year

of surveying (Table S-6). Species accumulation curves did not approach horizontal asymptotes

for either soil type, and the serpentine curve had a steeper slope than the nonserpentine curve

(Figure S-1). These curves suggest that neither fungal community was fully described, and that

there were more unsampled fungal species associated with the plants growing in serpentine soil

than those in nonserpentine.

Both fungal communities were dominated by a few abundant species (isolated >10

times), but the numerically dominant species differed between nonserpentine and serpentine

soils: *Alternaria* sp. 2 and *Ramularia* sp., respectively (Figure S-2, Table S-1). The

nonserpentine community included five abundant species, representing 62% of the isolates,

while the serpentine community included two abundant species, representing 51% of serpentine

isolates (Figure S-2, Table S-1). The serpentine community had a higher proportion of rare

species (<3 isolates) than the nonserpentine community, 63% versus 36%, respectively. One

Drechslera species was abundant in the nonserpentine community and rare in the serpentine

(Table S-1).

Soil type significantly affected pathogen community composition, based on pairwise

PERMANOVA analysis of two-transect fungal communities and NMDS visualization of

community similarity (F = 4.585, R^2 = 0.07567, p = 0.001; Figure 3). Serpentine fungal

communities were generally more similar to one another than they were to nonserpentine fungal

communities, and vice versa (Figure 3, Table 1).

Serpentine fungal community composition was more consistent across years than

nonserpentine fungal community composition (Table 1). While nonserpentine fungal community

similarity was unpredictable between years, with some very dissimilar communities (Morisita-

Horn overlap = 0.295) and some very similar ones (Morisita-Horn overlap = 0.882), serpentine

fungal communities were always relatively similar between years, with Morisita-Horn overlap

ranging from 0.707 to 0.786. Morisita-Horn overlap for nonserpentine and serpentine fungal

communities in each year was also variable, ranging from 0.535 to 0.725.

Leaf tissue chemistry and disease burden

As expected from soil chemistry differences, leaf chemistry in plants grown on serpentine

versus nonserpentine soil differed in mean C, N, Ca, Mg, Ni, and P content, and in Ca:Mg ratio.

Serpentine plants had significantly lower mean Ca, C, N, and P content and Ca:Mg ratio than

nonserpentine plants and significantly higher Ni and Mg (Table S-7, Figure 4). Cr content and

C:N ratio were similar for both soil types. Nitrogen content in asymptomatic foliar tissue from

plants surveyed in the field was positively related to the percentage of diseased tissue observed

for nonserpentine, but not serpentine plants, based on Pearson's product-moment correlation (for

N: nonserpentine cor. = 0.328, p-value = 0.012, serpentine cor. = -0.19, p-value = 0.14; for C:

nonserpentine cor. = 0.08, p-value = 0.53, serpentine cor. = -0.04, p-value = 0.76) (Figure 5).

Discussion

Here we show that a diverse suite of pathogens caused ubiquitous, low-grade damage on

host individuals across growing seasons and environmental conditions, and that plant populations

and communities appeared to shape the diversity and composition of fungal pathogen

communities in this system. The consistent presence of low-grade foliar disease in *S. pulchra* and

evidence from recent work suggesting that these symptoms inflict only minimal plant fitness

costs (Spear and Mordecai 2018, Uricchio et al. 2018) imply a coevolutionary relationship

between *S. pulchra* and its foliar pathogens that has led to plant-fungal interactions nearer the

commensal than the parasitic end of the symbiotic spectrum. This finding contrasts with

prevalent hypotheses that disease plays a major role in structuring plant communities by

regulating host population dynamics through mechanisms including enemy escape effects,

density- or distance-dependent Janzen-Connell effects, and epidemic-driven community shifts

(Janzen 1970, Connell J.H. 1971, Day and Monk 1974, Keane and Crawley 2002, Mitchell and

Power 2003, Petermann et al. 2008, Mordecai 2011). While evidence from tropical climates

supports these hypotheses (Mangan et al. 2010, Bagchi et al. 2014, Spear et al. 2015), this study

adds to a small body of literature from Mediterranean and semi-arid grasslands that indicates

little to no role for such mechanisms (Mordecai 2013, Spear and Mordecai 2018, Uricchio et al.

2018). However, we did not address pathogens with potential for more direct impacts on plant

survival and fecundity, such as flower, seed, and seedling pathogens (Gilbert 2002).

As predicted, the chemical composition of *S. pulchra* reflected the chemistry of the soil in

which it grew (except for Cr levels) (Oze et al. 2008) (Figure 4, Table S-7). Contrary to our

hypothesis, differences in soil chemistry that extended to plant chemistry did not affect disease

burden (Figures 1 and 4, Table S-6). Elevated heavy metal content in serpentine-grown plants

was far below "hyperaccumulation" levels that have been shown to inhibit pathogenic infection

in plants (e.g., 1.92×10^{-2} mol Cr/kg required for hyperaccumulation versus 3.30×10^{-4} mol

Cr/kg observed in serpentine *S. pulchra*) (Jaffre et al. 1976, Brooks et al. 1977, Boyd et al. 1994,

Martens and Boyd 1994, Table S-7), but the consistently low disease burden observed in both

soil types suggests that *S. pulchra* is able to effectively resist or suppress damage by foliar

pathogens such that heightened defenses via opportunistic hyperaccumulation would provide

little to no fitness benefit.

This study provides some of the first evidence that local growing conditions influence

structure of the pathogen communities of a single host species even over relatively small spatial

scales of a few hundred meters (Figure 1a, Figure 3). The close proximity of the sampled

nonserpentine and serpentine plants suggests that environmental filtering, rather than dispersal

limitation, plays a key role in fungal community assembly. We hypothesize that either tissue

chemistry, plant community composition, or both influence the composition of fungal

communities and their variation across environments.

Differing heavy metal content in the leaves of nonserpentine and serpentine *S. pulchra*

may influence fungal community assembly through a variety of mechanisms. Soil Ca content is

known to influence the outcomes of plant-pathogen interactions through its roles in cellular and

biochemical pathways in both the plant immune system (Lamb et al. 1989, Levine et al. 1996,

Blumwald et al. 1998, Romeis et al. 2001) and fungal pathogen virulence, growth, and

reproduction (Magalhães et al. 1991, Warwar and Dickman 1996, Lee and Lee 1998, Sebghati et

al. 2000, Shaw and Hoch 2000, Uhm et al. 2003, Brandhorst et al. 2005). Fungal pathogen

interspecific variation in reliance on Ca for successful infection and/or fungal response to Ca-

dependent plant immune defenses could therefore explain differences in nonserpentine and

serpentine fungal community composition. Previous studies showed that *Hesperolinon* sp.

infection frequency and severity by the fungal rust *Melampsora lini* in California grasslands

depended on plant tissue Ca levels linked to serpentine soil chemistry (Springer et al. 2006,

Springer 2009). Divergent tolerances of heavy metals other than Ca among fungi may act as

additional filters on the species present in nonserpentine and serpentine communities (Iram et al.

2009).

Differences in nonserpentine and serpentine fungal communities may also result from

differences in the plant community. Serpentine *S. pulchra* pathogen communities were more

diverse than nonserpentine communities at the species, genus, family, order, and class levels

(Tables S-1 through S-6). Results from previous studies that demonstrate a positive relationship

between host diversity and pathogen diversity indicate that higher plant species diversity in

serpentine grasslands might contribute to higher fungal diversity if native forbs and bunchgrasses

that specialize on serpentine soils share foliar fungal pathogens with *S. pulchra* (Hechinger and

Lafferty 2005, Kamiya et al. 2014, Spear 2017). Research on five common invasive and native

grasses (including *S. pulchra*) at JRBP shows that multi-host foliar fungal pathogens dominate

this system, supporting this hypothesis (Spear and Mordecai 2018).

Alternatively, serpentine and nonserpentine fungal communities could differ due to

genetic divergence between the *S. pulchra* populations found in these soils. The gene-for-gene

model of the evolution of virulence hypothesizes that a given pathogen can infect a given host

only if the pathogen does not possess an avirulence gene recognized by a matching resistance

gene in the host (Flor 1956, Parker and Gilbert 2004). To date, there have been few studies of

genetic differences between serpentine and nonserpentine populations of the same plant species.

However, genetic and chemical analyses of serpentine and nonserpentine populations of the

native forb *Lasthenia californica* at JRBP show that these populations represent two

phylogenetically distinct species and that each species both accumulates ions in significantly

different concentrations and has a unique flavonoid profile (Desrochers and Bohm 1993,

Rajakaruna and Bohm 1999, Chan et al. 2002, Rajakaruna et al. 2003). Since many flavonoids

have antifungal properties, serpentine and nonserpentine *S. pulchra* could also contain different

genotypes with different potential for resistance to any given fungus due to divergent resistance

gene and/or flavonoid profiles (Parker and Gilbert 2004, Treutter 2006).

Species turnover between years contributed substantially to fungal diversity (Table 1,

Table S-6, Figure 3), and species accumulation curves suggest that many more fungal species

remain to be discovered in this system (Figure S-1). Between-year variation in these

communities may be due to highly variable annual precipitation (2015 rainfall = 491mm; 2016 =

604mm, 2017 = 985mm) (Weather Underground). Timing and amount of precipitation might

alter infection dynamics by washing away fungal spores, transmitting spores to new hosts, and/or

influencing plant community composition (Hobbs and Mooney 1991, Madden 1997).The year-

to-year dissimilarity of fungal pathogen communities hosted by *S. pulchra* in both serpentine and

nonserpentine soils highlights the necessity of long-term data collection to accurately

characterize plant-disease interaction networks. This temporal variability also points to the

potential importance of the storage effect, in which temporally variable environmental conditions

act as ecological niches that facilitate species diversity, in maintaining diverse fungal

communities in California grasslands (Warner and Chesson 1985).

The variation in fungal community composition we observed on a single host species

with soil chemistry and over time suggests possible mechanisms preventing foliar disease from

regulating plant host populations in California grasslands. Transitions between distinct soil types

over spatial scales within the range of fungal spore dispersal may prevent disease outbreaks by

hampering the spread of pathogens that are more or less virulent depending on leaf tissue

chemistry. Variable climatic conditions favoring successful infection by different fungal species

each growing season (Table S-6), combined with plant senescence during the dry season each

year, suggest temporal barriers to pathogens evolving high virulence on both nonserpentine and

serpentine plants. The processes that limit disease burden and impact in *S. pulchra* suggest that

climate and growth form may explain differences with previous studies that reported larger

pathogen impacts. For example, seasonal senescence and local environmental barriers to

transmission in semi-arid and Mediterranean systems may limit the severity of individual

pathogen outbreaks, in contrast to year-round suitable climates and potentially more homogenous

biotic and abiotic environments in tropical systems. Life histories that promote the accumulation

of large disease burdens, along with relatively high and constant humidity, may be important

determinants of the potential for fungal pathogens to structure plant communities.

Clarifying the specific mechanisms that drive differences in serpentine and nonserpentine

fungal communities will require inoculation experiments to test how plant susceptibility to

infection by different fungal species changes with plant tissue chemistry and plant genetics.

Additionally, characterization of comprehensive plant-pathogen networks including all plant

species co-occurring with *S. pulchra* in serpentine and nonserpentine grasslands will improve our

understanding of the role of the surrounding plant community in structuring fungal pathogen

communities.

Conclusions

Fungi associated with foliar disease in the native bunchgrass *S. pulchra* are relatively benign and are unlikely to maintain the diversity of grassland plants, but contribute substantially to the overall biodiversity of California grasslands. The high taxonomic diversity and rarity of serpentine fungi demonstrates that this community is potentially at risk of species loss via host-parasite coextinction as annual grasses continue to expand their ranges at the cost of native plants, and human activities including development and increased C and N deposition contribute to the loss of serpentine grasslands that already make up <2% of the California's surface area (Huenneke et al. 1990, Vallano et al. 2012). This work demonstrates that, in addition to >200 endemic plant species, California's serpentine grasslands support unique communities of symbiotic fungi, providing additional motivation for active conservation of native grassland communities based on the far-reaching loss of biodiversity across trophic levels associated with their disappearance (Sprent 1992, Dobson et al. 2008, Dunn et al. 2009, Lafferty 2012).

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References

Abbate, J. L., and J. Antonovics. 2014. Elevational disease distribution in a natural plant–
pathogen system: insights from changes across host populations and climate. Oikos
123:1126–1136.

Abou-Shanab, R. A., J. S. Angle, T. A. Delorme, R. L. Chaney, P. Van Berkum, H. Moawad, K.
Ghanem, and H. A. Ghozlan. 2003. Rhizobacterial effects on nickel extraction from soil and
uptake by Alyssum murale. New Phytologist 158:219–224.

Ackerly, D. D., C. A. Knight, S. B. Weiss, K. Barton, and K. P. Starmer. 2002. Leaf size,
specific leaf area and microhabitat distribution of chaparral woody plants: Contrasting
patterns in species level and community level analyses. Oecologia 130:449–457.

Anderson, M. J. 2001. A new method for non-parametric multivariate analysis of variance.
Austral Ecology 26:32–46.

Bagchi, R., R. E. Gallery, S. Gripenberg, S. J. Gurr, L. Narayan, C. E. Addis, R. P. Freckleton,
and O. T. Lewis. 2014. Pathogens and insect herbivores drive rainforest plant diversity and
composition. Nature 506:85–88.

Blumwald, E., G. S. Aharon, and B. C-H. Lam. 1998. Early signal transduction pathways in
plant-pathogen interactions. Trends in Plant Science 3:342–346.

Borer, E. T., A. Laine, and E. W. Seabloom. 2016. A multiscale approach to plant disease using
the metacommunity concept. Annual Review of Phytopathology 54:397–418.

Boyd, R. S., J. J. Shaw, and S. N. Martens. 1994. Nickel hyperaccumulation defends
Streptanthus polygaloides (Brassicaceae) against pathogens. American Journal of Botany

81:294–300.

Brady, K. U., A. R. Kruckeberg, and H. D. Bradshaw Jr. 2005. Evolutionary ecology of plant

adaptation to serpentine soils. Annual Review of Ecology, Evolution, and Systematics

36:243–266.

Brandhorst, T. T., G. M. Gauthier, R. A. Stein, and B. S. Klein. 2005. Calcium binding by the

essential virulence factor BAD-1 of Blastomyces dermatitidis. Journal of Biological

Chemistry 280:42156–42163.

Brooks, R. R., J. Lee, R. D. Reeves, and T. Jaffre. 1977. Detection of nickeliferous rocks by

analysis of herbarium specimens of indicator plants. Journal of Geochemical Exploration

7:49–57.

Cantarelli, M. A., R. G. Pellerano, L. A. Del Vitto, E. J. Marchevsky, and J. M. Camiña. 2010.

Characterisation of two South American food and medicinal plants by chemometric

methods based on their multielemental composition. Phytochemical Analysis 21:550–555.

Chan, R., B. G. Baldwin, and R. Ornduff. 2002. Cryptic goldfields: A molecular phylogenetic

reinvestigation of Lasthenia californica sensu lato and close relatives (Compositae:

Heliantheae sensu lato). American Journal of Botany 89:1103–1112.

Chao, A., K. H. Ma, T. C. Hsieh, and C.-H. Chiu. 2016. SpadeR: Species prediction and

diversity estimation with R.

Ciompi, S., E. Gentili, L. Guidi, and G. F. Soldatini. 1996. The effect of nitrogen deficiency on

leaf gas exchange and chlorophyll fluorescence parameters in sunflower. Plant Science

118:177–184.

Coley-Smith, J. R., and R. C. Cooke. 1971. Survival and Germination of Fungal Sclerotia.

Annual Review of Phytopathology:65–92.

Connell J.H. 1971. On the role of natural enemies in preventing competitive exclusion in some

marine animals and in rain forest trees. Dynamics of numbers in populations. Proceedings

of the Advanced Study Institute. Centre for Agricultural Publication and Documentation,

Wageningen, The Netherlands:298–312.

Cunningham, S. A., B. Summerhayes, and M. Westoby. 1999. Evolutionary divergences in leaf

structure and chemistry, comparing rainfall and soil nutrient gradients. Ecology 69:569–

588.

Day, F. P., and C. D. Monk. 1974. Vegetation patterns on a southern Appalachian watershed.

Ecology 55:1064–1074.

Desrochers, A. M., and B. A. Bohm. 1993. Flavonoid variation in Lasthenia california

(Asteraceae). Biochemical Systematics and Ecology 21:449–453.

Dobson, A., K. D. Lafferty, A. M. Kuris, R. F. Hechinger, and W. Jetz. 2008. Homage to

Linnaeus: how many parasites? how many hosts? Proceedings of the National Academy of

Sciences 105:11482–11489.

Dormann, C. F., B. Gruber, and J. Fruend. 2008. Introducing the bipartite package: analysing

ecological networks. R news 8:8–11.

Dunn, R. R., N. C. Harris, R. K. Colwell, L. P. Koh, and N. S. Sodhi. 2009. The sixth mass

coextinction: are most endangered species parasites and mutualists? Proceedings of the

Royal Society B: Biological Sciences 276:3037–3045.

Eviner, V. T. 2016. Grasslands. Pages 449–477 Ecosystems of California. University of

California Press.

Fisher, R. A., A. S. Corbet, and C. B. Williams. 1943. The relation between the number of

species and the number of individuals in a random sample of an animal population. Source

Journal of Animal Ecology 12:42–58.

Flor, H. H. 1956. The complementary genic systems in flax and flax rust. Advances in Genetics

8:29–54.

Gilbert, G. S. 2002. Evolutionary ecology of plant diseases in natural ecosystems. Annual

Review of Phytopathology 40:13–43.

Hansen, R. A., and D. C. Coleman. 1998. Litter complexity and composition are determinants of

the diversity and species composition of oribatid mites (Acari: Oribatida) in litterbags.

Applied Soil Ecology 9:17–23.

Harrison, S. P., and J. H. Viers. 2007. Serpentine Grasslands. Pages 145–155 California

Grasslands: Ecology and Management. University of California Press.

Hechinger, R. F., and K. D. Lafferty. 2005. Host diversity begets parasite diversity: bird final

hosts and trematodes in snail intermediate hosts. Proceedings of the Royal Society B:

Biological Sciences 272:1059–1066.

Heinze, J., J. Bergmann, M. C. Rillig, and J. Joshi. 2015. Negative biotic soil-effects enhance

biodiversity by restricting potentially dominant plant species in grasslands. Perspectives in

Plant Ecology, Evolution and Systematics 17:227–235.

Hervé, M. "RVAideMemoire: testing and plotting procedures for biostatistics. . R. package

version 0. 9–6. (2017). 2017. RVAideMemoire: testing and plotting procedures for

biostatistics.

Hobbs, R. J., and H. A. Mooney. 1985. Community and population dynamics of serpentine

annual grassland in relation to gopher disturbances. Oecologia 67:342–351.

Hobbs, R. J., and H. A. Mooney. 1991. Effects of rainfall variability and gopher disturbance on

serpentine annual grassland dynamics. Ecology 72:59–68.

Huenneke, L. F., S. P. Hamburg, R. Koide, H. A. Mooney, P. M. Vitousek, L. F. Huenneke, S. P.

Hamburg, R. Koide, H. A. Mooney, and P. M. Vitousek. 1990. Effects of soil resources on

plant invasion and community structure in Californian serpentine grassland. Ecology

71:478–491.

Iram, S., I. Ahmad, B. Javed, S. Yaqoob, K. Akhtar, M. R. Kazmi, and Badar-uz-Zaman. 2009.

Fungal tolerance to heavy metals. Pakistan Journal of Botany 41:2583–2594.

Jaffre, T., R. R. Brooks, J. Lee, and R. D. Reeves. 1976. Sebertia acuminata: a hyperaccumulator

of nickel from New Caledonia. Science 193:579–580.

Janzen, D. H. 1970. Herbivores and the number of tree species in tropical forests. The American

Naturalist 104:501–528.

Jost, L., A. Chao, and R. L. Chazdon. 2011. Compositional similarity and beta diversity. Page

345 *in* A. E. Magurran and B. J. McGill, editors. Biological diversity: frontiers in

measurement and assessment. Oxford University Press, Oxford, England.

Kamiya, T., K. O'Dwyer, S. Nakagawa, and R. Poulin. 2014. Host diversity drives parasite

diversity: meta-analytical insights into patterns and causal mechanisms. Ecography 37:689–

697.

Keane, R. M., and M. J. Crawley. 2002. Exotic plant invasions and the enemy release hypothesis.

Trends in Ecology and Evolution 17:164–170.

Kindt, R., and R. Coe. 2005. BiodiversityR. Tree diversity analysis. A manual and software for

common statistical methods for ecological and biodiversity studies. World Agroforestry

Centre (ICRAF), Nairobi.

Lafferty, K. D. 2012. Biodiversity loss decreases parasite diversity: Theory and patterns.

Philosophical Transactions of the Royal Society B: Biological Sciences 367:2814–2827.

Lamb, C. J., M. A. Lawton, M. Dron, and R. A. Dixon. 1989. Signals and transduction

mechanisms for activation of plant defenses against microbial attack. Cell 56:215–224.

Lee, S. C., and Y. H. Lee. 1998. Calcium/calmodulin-dependent signaling for appressorium

formation in the plant pathogenic fungus Magnaporthe grisea. Molecules & Cells 8:698–

704.

Levine, A., R. I. Pennell, M. E. Alvarez, R. Palmer, and C. Lamb. 1996. Calcium-mediated

apoptosis in a plant hypersensitive disease resistance response. Current Biology 6:427–437.

Madden, L. V. 1997. No effects of rain on splash dispersal of fungal pathogens. Canadian

Journal of Plant Pathology 19:225–230.

Magalhães, B. P., R. Wayne, R. A. Humber, E. J. Shields, and D. W. Roberts. 1991. Calcium-

regulated appressorium formation of the entomopathogenic fungus Zoophthora radicans.

Protoplasma 160:77–88.

Magurran, A. E. 1988. Ecological diversity and its measurement. Chapman & Hall.

Mangan, S. A., S. A. Schnitzer, E. A. Herre, K. M. L. MacK, M. C. Valencia, E. I. Sanchez, and

J. D. Bever. 2010. Negative plant-soil feedback predicts tree-species relative abundance in a

tropical forest. Nature 466:752–755.

Martens, S. N., and R. S. Boyd. 1994. The ecological significance of nickel hyperaccumulation:

a plant chemical defense. Oecologia 98:379–384.

McElrone, A. J., C. D. Reid, K. A. Hoye, E. Hart, and R. B. Jackson. 2005. Elevated CO2

reduces disease incidence and severity of a red maple fungal pathogen via changes in host

physiology and leaf chemistry. Global Change Biology 11:1828–1836.

McNaughton, S. J. 1968. Structure and function in California grasslands. Ecology 49:962–972.

Mitchell, C. E., and A. O. Power. 2003. Release of invasive plants from fungal and viral

pathogens. Nature 421:625–627.

Mordecai, E. A. 2011. Pathogen impacts on plant communities: unifying theory, concepts, and

empirical work. Ecological Monographs 81:429–441.

Mordecai, E. A. 2013. Consequences of pathogen spillover for cheatgrass-invaded grasslands:

coexistence, competitive exclusion, or priority effects. The American Naturalist 181:737–

47.

O'Brien, H. E., J. L. Parrent, J. A. Jackson, J. Moncalvo, and R. Vilgalys. 2005. Fungal

community analysis by large-scale sequencing of environmental samples. Applied and

Environmental Microbiology 71:5544–5550.

Oksanen, J., R. Kindt, P. Legendre, and B. O'Hara. 2016. Vegan: community ecology package.

Osbourne, A. E. 1996. Preformed antimicrobial compounds and plant defense against fungal

attack. The Plant Cell 8:1821–1831.

Oze, C., S. Fendorf, D. K. Bird, and R. G. Coleman. 2004. Chromium geochemistry in

serpentinized ultramafic rocks and serpentine soils from the Franciscan complex of

California. American Journal of Science 304:67–101.

Oze, C., C. Skinner, A. W. Schroth, and R. G. Coleman. 2008. Growing up green on serpentine

soils: biogeochemistry of serpentine vegetation in the Central Coast Range of California.

Applied Geochemistry 23:3391–3403.

Parker, I. M., and G. S. Gilbert. 2004. The evolutionary ecology of novel plant-pathogen

interactions. Annual Review of Ecology, Evolution, and Systematics 35:675–700.

Parker, I. M., M. Saunders, M. Bontrager, A. P. Weitz, R. Hendricks, R. Magarey, K. Suiter, and

G. S. Gilbert. 2015. Phylogenetic structure and host abundance drive disease pressure in

communities. Nature 520:542–544.

Petermann, J. S., A. J. F. Fergus, L. A. Turnbull, and B. Schmid. 2008. Janzen-Connell effects

are widespread and strong enough to maintain diversity in grasslands. Ecology 89:2399–

2406.

Proctor, J., and S. R. J. Woodell. 1975. The ecology of serpentine soils. Advances in Ecological

Research 9:255–366.

Rajakaruna, N., B. G. Baldwin, R. Chan, A. M. Desrochers, B. A. Bohm, and J. Whitton. 2003.

Edaphic races and phylogenetic taxa in the Lasthenia californica complex (Asteraceae:

Heliantheae): an hypothesis of parallel evolution. Molecular Ecology 12:1675–1679.

Rajakaruna, N., and B. A. Bohm. 1999. The edaphic factor and patterns of variation in Lasthenia

californica (Asteraceae). American Journal of Botany 86:1576–1596.

Rascio, N., and F. Navari-Izzo. 2011. Heavy metal hyperaccumulating plants: how and why do

they do it? and what makes them so interesting? Plant Science 180:169–181.

Romeis, T., A. A. Ludwig, R. Martin, and J. D. G. Jones. 2001. Calcium-dependent protein

kinases play an essential role in a plant defence response. EMBO Journal 20:5556–5567.

Rossi, J.-P. 2011. rich: an R package to analyse species richness. Diversity 3:112–120.

Seabloom, E. W., E. T. Borer, K. Gross, A. E. Kendig, C. Lacroix, C. E. Mitchell, E. A.

Mordecai, and A. G. Power. 2015. The community ecology of pathogens: coinfection,

coexistence and community composition. Ecology Letters 18:401–415.

Sebghati, T. S., J. T. Engle, and W. E. Goldman. 2000. Intracellular parasitism by Histoplasma

capsulatum: fungal virulence and calcium dependence. Science 290:1368–1372.

Shaw, B. D., and H. C. Hoch. 2000. Ca2+regulation of Phyllosticta ampelicida pycnidiospore

germination and appressorium formation. Fungal Genetics and Biology 31:43–53.

Smith, V. 2007. Host resource supplies influence the dynamics and outcome of infectious

disease. Integrative and Comparative Biology 47:310–316.

Spear, E. R. 2017. Phylogenetic relationships and spatial distributions of putative fungal

pathogens of seedlings across a rainfall gradient in Panama. Fungal Ecology 26:65–73.

Spear, E. R., P. D. Coley, and T. A. Kursar. 2015. Do pathogens limit the distributions of tropical

trees across a rainfall gradient? Journal of Ecology 103:165–174.

Spear, E. R., and E. A. Mordecai. (n.d.). Foliar pathogens do not stabilize coexistence of

competing species in a California grassland. Ecology.

Spear, E. R., and E. A. Mordecai. 2018. Foliar pathogens do not stabilize coexistence of

competing species in a California grassland. Ecology 0:1–10.

Sprent, J. F. A. 1992. Parasites lost? International Journal for Parasitology 22:139–151.

Springer, Y. P. 2009. Do extreme environments provide a refuge from pathogens? A

phylogenetic test using serpentine flax. American Journal of Botany 96:2010–2021.

Springer, Y. P., B. A. Hardcastle, and G. S. Gilbert. 2006. Soil calcium and plant disease in

serpentine ecosystems: A test of the pathogen refuge hypothesis. Oecologia 151:10–21.

Tadros, T. M. 1957. Evidence of the presence of an edapho-biotic factor in the problem of

serpentine tolerance. Ecology 38:14–23.

Telfer, S., X. Lambin, R. Birtles, P. Beldomenico, S. Burthe, S. Paterson, and M. Begon. 2010.

Species interactions in a parasite community drive infection risk in a wildlife population.

Science 330:243–247.

Treutter, D. 2006. Significance of flavonoids in plant resistance: a review. Environmental

Chemistry Letters 4:147–157.

Uhm, K.-H., I.-P. Ahn, S. Kim, and Y.-H. Lee. 2003. Calcium/calmodulin-dependent signaling

for prepenetration development in Colletotrichum gloeosporioides. Phytopathology 93:82–

87.

Uricchio, L. H., S. C. Daws, E. R. Spear, and E. A. Mordecai. 2018. Exotic species dominate due

to niche overlap in a complex grassland community. in review:1–35.

Vallano, D. M., P. C. Selmants, and E. S. Zavaleta. 2012. Simulated nitrogen deposition

enhances the performance of an exotic grass relative to native serpentine grassland

competitors. Plant Ecology 213:1015–1026.

Vavrek, M. J. 2011. Fossil: palaeoecological and palaeogeographical analysis tools.

Palaeontologia Electronica 14.

Warner, R. R., and P. L. Chesson. 1985. Coexistence mediated by recruitment fluctuations: a

field guide to the storage effect. The American Naturalist 125:769–787.

Warwar, V., and M. B. Dickman. 1996. Effects of calcium and calmodulin on spore germination

and appressorium development in Colletotrichum trifolii. Applied and Environmental

639  Microbiology 62:74–79.

640 Weather Underground. (n.d.). Jasper Ridge Biological Preserve KCAWOODS19.

641  https://www.wunderground.com/personal-weather-station/dashboard?ID=KCAWOODS19.

642 Wickham, H. 2016. Ggplot2: elegant graphics for data analysis. Springer-Verlag, New York.

643 Wolda, H. 1981. Similarity indices, sample size and diversity. Oecologia 50:296–302.

644 Wolf, D. C., T. H. Dao, H. D. Scott, and T. L. Lavy. 1989. Influence of sterilization methods on

645  selected soil microbiological, physical, and chemical properties. Journal of Environment

646  Quality 18:39.

651 **Figure legends**

652 **Figure 1. Pathogen damage was consistently low across three years and both grassland**

653 **types.** Panel (a) maps the locations of surveyed nonserpentine sites (blue) and serpentine sites

654 (green) along the ridgetop at JRBP. Panel (b) shows mean percentage (bars) and standard error

655 (error bars) of diseased leaf area for surveyed serpentine and nonserpentine *S. pulchra,* by year.

656 **Figure 2. Pathogens isolated from *Stipa pulchra* growing in nonserpentine and serpentine**

657 **grasslands.** Bipartite network showing species interactions between *S. pulchra* plants (left)

658 growing in serpentine grasslands (green) and nonserpentine grasslands (blue) and fungal species

659 cultured from symptomatic leaf tissue (white bars on the right). The length of each white bar

correlates with the number of isolates in the operational species unit (based on 97% sequence

similarity). The thickness of lines connecting left and right bars represents the number of times a

particular fungal species was isolated from *S. pulchra* in a particular grassland type. The

nonserpentine fungi came from 15 genera, 8 families, 4 orders, and 3 classes; the serpentine

fungi came from 20 genera, 14 families, 7 orders, and 5 classes (Tables S-1 to S-6).

Figure 3. Distinct fungal communities isolated from *Stipa pulchra* growing in nearby

serpentine and nonserpentine sites across years. Non-metric multidimensional scaling

visualization of serpentine and nonserpentine fungal community similarity. Each point

corresponds to the combined fungal community of two transects at the same site in the same

year. Serpentine communities are shown in green and nonserpentine communities are shown in

blue. Circles, triangles, and squares represent fungal communities sampled in 2015, 2016, and

2017, respectively. Ellipses enclose 95% confidence intervals for ordination of serpentine and

nonserpentine communities. Community dissimilarity increases with the distance between

points. PERMANOVA analysis showed a significant effect of soil type on fungal community

composition (F = 3.623, R^2 = 0.0618, p = 0.001).

Figure 4. Plants grown in nonserpentine soil had higher C, N, P, Ca, and Ca:Mg and lower

Mg and Ni than those grown in serpentine soil. Box plots comparing elemental content of

dried foliar tissue from nonserpentine- and serpentine-grown *S. pulchra*. In each plot, the thick

black bar represents the median, the height of the box represents the interquartile range (IQR),

the whiskers represent the maximum and minimum points within 1.5*IQR of the 25th and 75th

percentiles, and the circles represent points outside this range. An asterisk above a panel

indicates a statistically significant difference between serpentine- and nonserpentine-grown

plants (p < 0.05) indicated by a Wilcoxon ranked-sum test. Panel (a) shows percent C of dried

foliar tissue (W = 2713, p-value = 1.65×10^{-7}); panel (b), percent N (W = 2346, p-value =

0.0011); panel (c), C:N ratio (W = 1398, p-value = 0.066); panel (d), moles P per kilogram dried

leaf tissue (W = 390, p-value = 2.83 x 10^{-9}); panel (e), moles Ca per kilogram dried leaf tissue

(W = 400, p-value = 1.45 x 10^{-11}); panel (f), Ca:Mg ratio (W = 400, p-value = 1.451 x 10^{-11});

panel (g), moles Cr per kilogram dried leaf tissue (W = 192.5, p-value = 0.8498); panel (h),

moles Mg per kilogram dried leaf tissue (W = 1, p-value = 2.90 x 10^{-11}); and panel (i), moles Ni

per kilogram dried leaf tissue (W = 0, p-value = 6.79 x 10^{-8}).

Figure 5. C, N, and C:N predict the mean foliar pathogen damage with opposite trends in

serpentine versus nonserpentine grasslands. Foliar percent C (Panels (a) and (b)), percent N

(Panels (c) and (d)) and C:N ratio (Panels (e) and (f)) versus mean percentage of diseased leaf

area for serpentine and nonserpentine *S. pulchra* plants, with linear regression (lines) and 95%

confidence intervals (grey bands). An asterisk following a p-value indicates a statistically

significant result (p<0.05).

Figures

(a)

(b)

699   **Figure 1.**

701   **Figure 2.**

Figure 3.

707 **Figure 4.**

Figure 5.

Tables

Table 1. Fungal community diversity was higher and communities were distinct in *S. pulchra* plants growing in serpentine versus nonserpentine grasslands, across all years.
Fungal community similarity by soil type and year. The leftmost column lists the fungal communities considered, with the number of species in each community in parentheses. The second column lists Fisher's alpha for each community. The third column shows the number of shared species and, in parentheses, the percentage of total observed species that were shared for each pair of communities. The rightmost column lists the estimated Morisita-Horn community overlap value, based on absolute species abundances, for each pair of communities. The Morisita-Horn index ranges from 0 to 1, with 1 indicating complete overlap.

Fungal community pair	Fisher's alpha	Shared species	Morisita-Horn Similarity Index
All years serpentine (24)	9.324	16	0.675
All years nonserpentine (22)	7.261	(53.0%)	
2015 serpentine (11)	7.955	5	0.535
2015 nonserpentine (12)	5.205	(27.8%)	
2016 serpentine (16)	7.897	11	0.766
2016 nonserpentine (13)	5.861	(61.1%)	
2017 serpentine (12)	6.784	8	0.725
2017 nonserpentine (12)	5.136	(50.0%)	
2015 serpentine (11)	7.955	7	0.707
2016 serpentine (16)	7.897	(35.0%)	
2015 serpentine (11)	7.955	5	0.721
2017 serpentine (12)	6.784	(27.8%)	
2016 serpentine (16)	7.897	7	0.786
2017 serpentine (12)	6.784	(33.3%)	
2015 nonserpentine (12)	5.205	6	0.668
2016 nonserpentine (13)	5.861	(31.5%)	
2015 nonserpentine (12)	5.205	5	0.295
2017 nonserpentine (12)	5.136	(26.3%)	
2016 nonserpentine (13)	5.861	7	0.882
2017 nonserpentine (12)	5.136	(38.9%)	