Density-dependent natural selection mediates harvest-induced trait changes

Alix Bouffet-Halle¹, Jacques Mériguet^{2,3}, David Carmignac⁴, Simon Agostini⁵, Alexis Millot⁶, Samuel Perret^{7,8}, Eric Motard⁹, Beatriz Decenciere¹⁰, Eric Edeline^{11,12}*

- 1: Sorbonne Université, Université Paris Diderot, UPEC, CNRS, INRA, IRD, Institut d'Ecologie et des
- 6 Sciences de l'Environnement de Paris (iEES-Paris), F-75252 Paris, France. alix.halle@gmail.com
 - 2: Ecole Normale Supérieure, CEREEP Ecotron Île-de-France, UMS CNRS ENS 3194, 78 rue du
- 8 Château, 77140 Saint Pierre-lès-Nemours, France. jacques.meriguet@ens.fr
 - 3: Institut de Biologie de l'Ecole Normale Supérieure, CNRS, INSERM, PSL Research University,
- 10 Paris, France.

2

- 4: Sorbonne Université, Université Paris Diderot, UPEC, CNRS, INRA, IRD, Institut d'Ecologie et des
- 12 Sciences de l'Environnement de Paris (iEES-Paris), F-75252 Paris, France. david.carmignac@upmc.fr
 - 5: CNRS, CEREEP Ecotron Île-de-France, UMS CNRS ENS 3194, 78 rue du Château, 77140 Saint
- 14 Pierre-lès-Nemours, France. simon.agostini@ens.fr
 - 6: Ecole Normale Supérieure, CEREEP Ecotron Île-de-France, UMS CNRS ENS 3194, 78 rue du
- 16 Château, 77140 Saint Pierre-lès-Nemours, France. alexis.millot@ens.fr
 - 7: CNRS, CEREEP Ecotron Île-de-France, UMS CNRS ENS 3194, 78 rue du Château, 77140 Saint
- 18 Pierre-lès-Nemours, France.
 - 8: Centre d'Ecologie Fonctionnelle et Evolutive CEFE, UMR 5175, Campus CNRS, Université de
- 20 Montpellier, Université Paul-Valéry Montpellier, Montpellier Cedex 5, France. samuel.perret@cefe.cnrs.fr
- 9: Sorbonne Université, Université Paris Diderot, UPEC, CNRS, INRA, IRD, Institut d'Ecologie et des Sciences de l'Environnement de Paris (iEES-Paris), F-75252 Paris, France. eric.motard@upmc.fr
- 10: CNRS, CEREEP Ecotron Île-de-France, UMS CNRS ENS 3194, 78 rue du Château, 77140 Saint Pierre-lès-Nemours, France. <u>beatriz.decenciere@ens.fr</u>
- 26 11: Sorbonne Université, Université Paris Diderot, UPEC, CNRS, INRA, IRD, Institut d'Ecologie et des Sciences de l'Environnement de Paris (iEES-Paris), F-75252 Paris, France.
- 28 12: ESE, Ecology and Ecosystem Health, INRAE, Agrocampus Ouest, Rennes, France. * Corresponding author: Eric.Edeline@inrae.fr

- 30 **Statement of authorship:** ABH performed the laboratory F₁ experiment, contributed to data analysis, wrote the first draft of the manuscript and contributed to subsequent versions. EE designed the study,
- 32 contributed to the pond experiment, performed data analysis, and led manuscript writing from the second version. JM, DC, SA, AM, SP, EM and BD contributed to both the pond and laboratory
- 34 experiments.
 - **Data accessibility statement:** Should the manuscript be accepted, the data supporting the results will
- 36 be archived in an appropriate public repository and the data DOI will be included at the end of the article.
- 38 **Key words:** Adaptive landscapes, Asymmetric competition, Body size, Cannibalism, Eco-evolutionary feedback loops, Energy allocation rules, Fisheries, Global change, Harvesting yields, Life-history
- 40 change.

Type of article: Letter.

Manuscript length: 144 words in the abstract, 4996 words in the main text, 59 references, 0 table, 6 figures, 0 text boxes.

44 ABSTRACT

46

48

50

52

54

Rapid life-history changes caused by size-selective harvesting are often interpreted as a response to direct harvest selection against a large body size. However, similar trait changes may result from a harvest-induced relaxation of natural selection for a large body size via density-dependent selection. Here, we show evidence of such density-dependent selection favouring large-bodied individuals at high population densities, in replicated pond populations of medaka fish. Harvesting, in contrast, selected medaka directly against large-bodied medaka and, in parallel, decreased medaka population densities. Five years of harvesting were enough for harvested and unharvested medaka populations to inherit the classically-predicted trait differences, whereby harvested medaka grew slower and matured earlier than unharvested medaka. We demonstrate that this life-history divergence was not driven by direct harvest selection for a smaller body size in harvested populations, but by density-dependent natural selection for a larger body size in unharvested populations.

INTRODUCTION

56

58

60

62

64

66

68

70

72

74

Phenotypic changes caused by harvesting can be exceptionally rapid (Darimont *et al.* 2009) and may have cascading effects on harvesting yields and ecosystem function (Conover & Munch 2002; Dunlop *et al.* 2015). However, the underlying mechanisms that control harvest-induced trait changes are potentially complex and often remain cryptic in empirical studies. The most immediate effect of harvesting is to reduce population density and increase food resources in survivors (Fig. 1, Removal \rightarrow Density-dependent plasticity pathway), which results in higher rates of somatic growth and reproduction (Verhulst 1838; Hilborn & Walters 1992). However, in parallel, harvesting is often

directly size-selective (Fig. 1, Direct harvest selection pathway) and generates complex selective pressures on body size and size-related traits (Matsumura et al. 2011 and references therein).

The form and strength of direct harvest selection depends on the specific pattern of selectivity of the fishing gear combined with fishing pressure (Kuparinen *et al.* 2009). For instance, gears targeting large-bodied individuals directly select against fast-growing genotypes (Conover & Munch 2002; Edeline *et al.* 2007, 2009; Swain *et al.* 2007) and, in parallel, select against late-maturing genotypes through reducing life expectancy (Ernande *et al.* 2004; Dunlop *et al.* 2009; Heino *et al.* 2015). Accordingly, a number of empirical and experimental studies have associated harvesting with change towards earlier maturation at a smaller body size and/or towards slower somatic growth (see reviews by Diaz Pauli & Heino 2014; Heino et al. 2015; Kuparinen & Festa-Bianchet 2017). Note, however, that selection for an earlier maturation may also result in evolution of faster somatic growth, allowing for earlier maturation (Dunlop *et al.* 2009; Eikeset *et al.* 2016; Diaz Pauli *et al.* 2017).

Recently, eco-genetic models have further revealed that the presence and strength of density-dependent plasticity in somatic growth can alter the amount and direction of direct harvest selection (Gobin et al. 2018, Modulation pathway in Fig. 1). Such interactions occur because density-dependent plastic changes in somatic growth can shift the timing of maturation in the same direction as harvest selection, thus reducing the strength of direct harvest selection (Lester et al. 2014; Eikeset et al. 2016, but see Arlinghaus et al. 2009). Ultimately, this effect may shift harvesting of large-bodied individuals from removing both immature and mature individuals to removing only mature individuals, in which case selection changes from favouring an early maturation to favouring a late maturation (Ernande *et al.* 2004; Heino *et al.* 2015).

Density-dependent plasticity is not the only pathway through which population density may affect body sizes. The possibility for intraspecific interactions to induce density-dependent, natural-selection on body sizes also exists (Fig. 1, Natural selection arrow). Pioneering studies in *Drosophila* have demonstrated that juvenile (larval) competitive ability in laboratory populations can rapidly evolve in response to crowding, in particular through changes in foraging rates, food conversion efficiency and development time, all of which are traits that affect body size (Mueller 1988, 1997; Sgrò & Partridge 2000; Sarangi *et al.* 2016). More recently, it was shown that density-dependent selection may also be a major driver of trait change in natural populations. For instance, predators relax the strength of density-dependent regulation in wild guppy populations (*Poecilia reticulata*), and degrade the evolved ability of guppies to cope with increasing population density (Bassar *et al.* 2013).

86

88

90

92

94

96

98

100

102

104

In particular, increased interference competition at high population densities often favours larger-bodied individuals (Post *et al.* 1999). An example of this is the brown anole (Anolis sagrei), where increased interference competition at high population density yields strong natural selection for large body size (Calsbeek & Smith 2007; Calsbeek & Cox 2010). Cannibalism is an extreme form of interference present in a number of taxa, and that also typically selects for large body size (see Claessen et al. 2004 and references therein). In Windermere pike (*Esox lucius*), natural selection is thought to select for larger body sizes through cannibalism (Carlson *et al.* 2007), resulting in a positive relationship between pike density and body size (Edeline *et al.* 2007). Therefore, by lowering population density, harvesting may relax interference competition and cannibalism, in turn decreasing the strength of density-dependent natural selection for large body size. Our aim in this study is to explore this density-mediated, Removal → Natural selection pathway (Fig. 1).

The density-mediated, Removal → Natural selection pathway for harvest selection operates simultaneously with the Direct harvest selection pathway, and both favour smaller body sizes, such that detecting the signature of density-mediated harvest selection is challenging. One way to tackle this challenge is to use an organism in which body size evolves in one direction only. For instance, if body size has already evolved to some lowest possibly physiological limit (Silva *et al.* 2013; Marty *et al.* 2014; Dunlop *et al.* 2015), it can only evolve towards larger body sizes. Therefore, in such an organism any body-size difference between harvested and non-harvested populations would indicate response to density-dependent selection for a larger body size in unharvested populations.

Another potential way to disentangle the effects of natural selection and direct harvest selection is through examining genotype-by-environment interactions on phenotypes. Falconer (1990) showed that,

116 when a trait is selected in environment A, the phenotypic response to selection is expressed in both environment A and other environments (say B), but the amplitude of the phenotypic response to 118 selection is often less in B than in A due to genotype-by-environment interactions. A key environmental effect of harvesting is to increase the levels of available food resources for survivors. Therefore, response to direct harvest selection for a smaller body size should be larger in a high-food than in a 120 low-food environment (Figs. 2a & 2c) while, in turn, response to natural, density-dependent selection 122 for a larger body size should be larger in a low-food than in a high-food environment (Figs. 2b & 2d). Here, we used both approaches to separate the effects of density-mediated and direct harvest selection 124 on body sizes. Specifically, we measured harvest-by-food interactions on somatic growth rate, maturation and energy acquisition rate in the medaka fish (Oryzias latipes), an organism that was 126 shown in the laboratory to have an asymmetric body-size evolvability.

MATERIALS AND METHODS

Pond medaka populations

128

Origin and maintenance

- The medaka is an oviparous fish belonging to the group of Beloniformes, a sister group of Cyprinodontiformes which includes killifishes (Kinoshita *et al.* 2009). The medaka, which naturally inhabits slow-moving fresh- and brackish-water habitats of south-east Asia, has a wide thermal range, and can both overwinter under the ice and have a 2-3 months generation time at 27°C in the laboratory. Hence, the medaka is ideal to perform parallel experiments in both the laboratory and in outdoor ponds
 - under temperate latitudes.

136 Our starting medaka populations descended from 100 parents wild-caught in Kiyosu (Toyohashi, Aichi Prefecture, Japan) in June 2011. Medaka from this populations were shown to respond to selection for a 138 larger body size by evolving faster somatic growth rates and delayed maturation, but were unable to respond to selection for a smaller body size (Le Rouzic *et al.* in press; Renneville *et al.* in press). About 160 progeny were used to seed each of 12 circular outdoor ponds (10 m², 1.2 m deep). Further details 140 on how populations were formed and tanks were installed may be found in the SI Appendix I. No food 142 was added to the ponds so that natural density-dependent processes could take place. To measure the effects of medaka harvesting on their food, we sampled zooplankton (11 dates in 2012) and filamentous algae (7 dates from 2012 to 2013), which are the two major food sources for medaka in ponds (SI 144 Appendix I). Food was manipulated only subsequently during the common garden experiment in the 146 laboratory, so as to test for harvest-by-food interactions on somatic growth, maturation and food intake (see below).

Medaka harvesting and phenotyping in ponds

148

150

152

154

From 2012 to 2016, each of the 12 pond populations was sampled in March before medaka reproduction (pre-recruitment) and in November after medaka reproduction (post-recruitment). Fish were concentrated using a seine net and then fished using handnets (catchability = $98 \pm 0.6\%$ SD estimated using removal sampling). All sampled fish were individually weighed to the nearest mg (no anaesthesia required) and estimated for standard body length (from the tip of the snout to the base of the caudal fin) using a body mass-length relationship obtained in March 2013, when fish were also measured individually using the imageJ software ($R^2 = 0.98$ on a log-log scale, n = 2722).

Each year in March in the 6 harvested populations, we removed all the fish that were too large to pass through a screen made from 2 mm-spaced parallel bars (i.e., selection on body girth), while in unharvested populations all fish were released after phenotyping. The fishery removed on average 79% of individuals (i.e., 98% catchability × 81% removal rate). Such a high exploitation rate is comparable to those imposed by some industrial marine fisheries (Hutchings 2000; Myers & Worm 2003). In November, all fish from both harvested and unharvested populations were released after phenotyping. The fishing operation and manipulations resulted in an incidental 0.8% mortality rate, which was independent of harvest treatments but that decreased with increasing medaka body size.

Larvae counts

164

166

168

170

172

174

In fish, negative density dependence of population dynamics is generally considered to reflect juvenile mortality due to interference competition and cannibalism from large-bodied adults (Ricker 1954; Claessen *et al.* 2004). However, in theory overcompensating recruitment may also operate through decreased adult fecundity. To discriminate between the two mechanisms, we counted the number of newly-hatched larvae hiding in two pairs of floating plastic brushes (summed for analyses) at irregular intervals during the 2014, 2015 and 2016 spawning periods (April to September). In total, for each pond we performed 167 counts spread over 118 dates with one to three counts per date.

Common garden in the laboratory

Common garden experiments alleviate plastic effects of the environment on the phenotype, and can thus reveal an heritable response to harvesting. We examined harvest-by-food interactions under

common garden conditions in the F_1 medaka generation born from parents sampled in each of the 12 populations. Maternal effects are not alleviated in the F_1 generation, but they are unlikely to have had large effects in our experiment (see the Discussion).

178 Parental fish

176

180

182

184

186

188

190

192

194

In November 2016, between 6 and 10 medaka (mean 9.6) were randomly sampled from each of the 12 pond populations to serve as parents for a F_1 generation in the laboratory. These 12 random parental samples, which represented from 3 to 29% of the catch (mean 9%), were maintained in a greenhouse at air temperature in 150L tanks with live food. In January 2017, parental fish were weighed to the nearest mg, measured for standard body length with ImageJ software, and grouped to form 3 breeding pairs per population (36 pairs in total), except for one parental sample which had only one mature female still alive in January 2017 (this female and its mate became parents of all progeny produced from this pond). In parents, we found no significant effect of harvesting on body size (random pond intercept ANOVA, Chisq = 0.353, df = 1, p > 0.552) or body condition (ANCOVA, Chisq = 0.456, df = 1, p > 0.499). Each of the 36 pairs was transferred to the laboratory in a 3.5L aquarium and induced to spawn by progressively raising temperature from 20.0 to 27.0 \pm 0.3°C (mean \pm SD) and setting a 15-h light: 9-h dark photoperiod. Water conductivity was 375 \pm 43 μ S/cm.

Dry food (Skretting Gemma Micro, 300 µm pellets) was provided twice per day and live *nauplii* of *Artemia salina* once per day. After initiation of spawning by all breeding pairs, eggs from each breeding pair were collected daily during a 4-day period, enumerated and incubated in separate jars so as to keep track of individual parental identity (but not spawning day). We found no significant effect

of harvesting on either probability of a non-zero clutch (zero-inflated negative binomial, random pond intercept, logit scale, estimate of harvest effect = -0.128 (\pm 0.200 SE), p = 0.522) or on the size of a non-zero clutch (log scale, estimate of harvest effect = 0.071 (\pm 0.090 SE), p = 0.431).

Progeny birth and feeding treatments

196

198

200

202

204

206

208

210

We collected larvae hatched from the 7th to the 10th day after the weighted average date of spawning. Larvae born from the same breeding pair on the same day were transferred to 1.5 L aquariums by groups of three larvae (1-4 groups of larvae, mean 2.9, per breeding pair), and were maintained under the same temperature and light regime as their parents. From 15 days post hatch (dph) onwards, we varied resource levels by applying to F₁ progeny three food environments. In the low-food environment, medaka were fed once every second day with *nauplii* of *Artemia salina*, alternated with dry food. In the high-food environment, medaka were fed twice daily, once with *nauplii* and once with dry food. In the medium-food environment, medaka were fed once daily alternating *nauplii* and dry food. Compared to low-food, the medium- and high-food environments multiplied food supply by 2 and 4, respectively, which loosely corresponds to the *relative* effect of harvesting on zooplankton availability in ponds (see results). Further details on the feeding protocol may be found in the SI Appendix I.

Progeny phenotyping

At 15 dph, all F₁ individuals were weighed and measured as described above and only one individual per aquarium was randomly kept for subsequent phenotyping, making it possible to track individual developmental trajectories.

Individual phenotyping was repeated at 30 dph, 40 dph and then once per week until 90 dph (11

individual measurements for a total of 104 individuals). From 40 dph onwards, phenotyping further

included detection of the maturity status from the presence of secondary sexual characters (Yamamoto

1975). Specifically, the maturity criteria were appearance of a round-shaped anal papilla in females,

and of the papillar process on the anal fin in males. On average, females matured at 58.4 dph and 15.1

mm, while males matured at 56.5 dph and 14.7 mm. We found no significant difference in somatic

growth rate between males and females (random pair intercept ANCOVA, chisq = 0.102, df = 1, p =

222 0.749).

216

218

220

224

226

228

230

232

We measured individual feeding rate three times during a period ranging from 46 to 66 dph. We fasted

fish overnight and acclimatized them for five minutes in a 80 mL container under the same temperature

and light conditions as during rearing. Fish were then presented with 20 prey (nauplii of Artemia

salina), and we counted the number of prey eaten during 5 minutes.

Statistical analyses

A full description is given in the SI Appendix I. Briefly, each year in both harvested and unharvested

pond medaka populations, the number of age classes was inferred using model-based clustering. The

number of fish in each age class at each sampling event was estimated by fitting a mixture of Gaussian

distributions to individual body lengths (n = 17908). These estimated numbers allowed us to visualize

the strength of negative density-dependence by plotting Ricker stock-recruitment relationships fitted to

mean point estimates. We estimated the relationship between individual standard body length and

probability to survive through the fishery in March (n = 3970 individuals) using a mixed effects

Bernoulli GLM with a logit link function. Finally, medaka larvae and zooplankton counts were

modelled using mixed-effects zero-inflated negative binomial models, while % of pond surface covered by filamentous algae was modelled using a negative binomial model.

In the common garden experiment, we estimated harvest-by-food interactions on somatic growth rates 238 and growth trajectories of the F₁ medaka progeny using a second order polynomial regression of 240 standard body length on age. We modelled medaka maturation using two complementary approaches: probabilistic maturation reaction norms (PMRNs "direct estimation" method, Heino & Dieckmann 2008) and maturation rates (Van Dooren et al. 2005; Harney et al. 2013). We modelled PMRNs using a 242 Bernoulli GLM that accounted for the effects of both age (days post hatch) and body size (mm) on 244 maturation probability, but that did not include any harvest-by-food interaction so as to gain statistical power in estimating the effect of harvesting on the PMRNs. By doing so, we assumed that the plastic 246 effects of the food environment on the PMRN was fully mediated by somatic growth rates. We modelled harvest-by-food interactions on medaka maturation rates, measured in logit maturation 248 probability day⁻¹, using a GLM approximation of a maturation rate model, as described by Harney et al. (2013). Specifically, we fitted to the maturation data a Bernoulli GLM including the time interval 250 between two observations as an offset term. Finally, in the feeding trials, counts of the number of nauplii larvae eaten by individual medaka were modelled using a mixed-effects zero-inflated negative 252 binomial model.

RESULTS

254 Age structure and population dynamics in ponds

256

258

262

264

266

268

270

Both harvested and unharvested pond medaka populations included two age groups corresponding to

age 0+ and age 1+ fish (Table S1, Fig. 3a). The experimental fishery targeted medaka larger than 15

mm in standard body length (Fig. 3a, length at 50% removal probability = 18.7 mm) and removed

about 50 % of age-0+ recruits and 100 % of 1+ individuals (Fig. 3a), thus reproducing a typical direct

harvest selection pattern, which is predicted to favour slow-growing and early-maturing genotypes.

260 In parallel with imposing this Direct harvest selection pathway on medaka body size (Fig. 1), our

experimental fishery removed negative density-dependence on juvenile recruitment in medaka

population dynamics. Pond medaka populations followed "overcompensating" stock-recruitment

dynamics (Fig. 3b), which is typical of many other fish populations (Ricker 1954; Hilborn & Walters

1992). Specifically, fishing consistently decreased the stock of spawners (population size in March)

below circa 50 individuals (28 on average, red squares in Fig. 2b), a density region in which increasing

stock size had a positive effect on the absolute number of age-0+ recruits (black curves, Fig. 3b),

indicating demographic "undercompensation" due to density-independence of vital rates (Bellows

1981). In contrast, unharvested medaka populations had stock sizes above circa 50 individuals (137 on

average, blue triangles in Fig. 3b), a density region where increasing stock size had a negative effect on

the absolute number of age-0+ recruits, indicating demographic "overcompensation" due to negative

density-dependence of vital rates (black curves, Fig. 3b).

Newborn medaka larvae were on average more numerous in unharvested than in harvested populations (P-value = 0.011, Fig. 4, Table S2), i.e., opposite to what one might expect if overcompensating medaka recruitment in unharvested populations occurred via reduced adult fecundity. Hence, as is typical for fish (Ricker 1954), overcompensating recruitment in medaka was instead mediated by increased post-larval (juvenile) mortality, indicating that large-bodied adults dominated smaller-bodied juveniles in pond populations (see also SI Appendix II).

Food resources in ponds

278

280

282

Fishing for medaka strongly increased abundances of large-bodied zooplankton and filamentous algae, the two major food sources for medaka in ponds. Specifically, medaka fishing multiplied mean abundances of *Asplanchna* sp., Calanoids and Cladocerans by 1.6, 1.9 and 2.9, respectively, and multiplied % pond surface covered by filamentous algae by 14.0 (Table S3).

Life history in the laboratory common garden

Under all three food environments, F₁ progeny from harvested populations had significantly lower somatic growth rates than progeny born from unharvested populations (Figs. 5a, MCMC P-values = 0.008 at low food, < 0.001 at medium food and = 0.002 at high food), resulting in a similar effect of harvesting on somatic growth trajectories (Fig. 6a). Accordingly, a deviance analysis shows that there was no significant harvest-by-food interaction on medaka somatic growth rates (P-value of Age × Harvesting × Food interaction = 0.265, Table S4), indicating that the harvest-induced decrease in somatic growth was food-independent. This absence of any harvest-by-food interaction does not permit

selective drivers to be disentangled, and therefore can not support any of our two predictions (Figs. 2a, 2b).

In a low-food environment (MCMC P-value = 0.015), but not in a high-food or medium-food environments (MCMC P-values = 0.881 and 0.506, respectively), maturation rates were higher in F_1 294 progeny from harvested populations than from unharvested populations (Fig. 5b). The deviance 296 analysis confirmed the significant Age × Harvesting × Food interaction (P-value < 0.001 in Table S4). This harvest-by-food interaction on maturation rates resulted in harvesting having no influence on 298 the height of medaka PMRN, but on its slope, which shifted from positive to negative (Fig. 6b). Specifically, in a high-food environment medaka reached a 50% maturation probability at around 50 300 days post hatch (dph) in both harvested and unharvested populations (Fig. 6b). In contrast, in a lowfood environment the 50% maturation probability was reached at around 65 dph by harvested medaka, 302 but was reached at about 80 days by unharvested medaka (Fig. 6b). Taken together, these results support the prediction that medaka maturation rates responded to selection in a low-food environment, 304 i.e., responded to density-dependent natural selection (Fig. 2d), but not to direct harvest selection (Fig. 2c). Note that we did not find any strong population pattern on somatic growth rates or maturation 306 probabilities (Fig. S2).

Feeding trials in the laboratory

292

F₁ progeny from harvested and unharvested medaka populations ate a similar number of prey in a low-food and in a high-food environments (Fig. 6c, MCMC P-values = 0.673 and 0.405, respectively), indicating no difference in food acquisition rate. In contrast, in a medium-food environment F₁ progeny

from harvested populations ate significantly less prey than progeny from unharvested populations (Fig. 6c, MCMC p-value = 0.015).

312

DISCUSSION

Our results show that, as expected, harvested medaka grew slower and matured earlier than unharvested medaka. Convergent elements suggest that these differences were inconsistent with an adaptive response of medaka owing entirely to direct harvest selection for a smaller body size, but that were driven in part by natural selection for larger body size (Fig. 1).

In ponds, medaka populations included two age classes which, in March of year *t*, were age 0+ fish born in summer of year *t*-1 and age 1+ fish born in summer of year *t*-2. In harvested populations, however, almost all 1+ fish were removed by the fishery in March (Fig. 3a) and only 0+ individuals were effectively able to reproduce, similar to the wild in Japan where medaka die during their first reproductive season (Edeline et al. 2016 and references therein). Therefore, in ponds fishing mortality replaced the natural mortality regime that prevails in the wild, and harvesting is thus unlikely to have driven any large response in medaka life histories. In contrast, in unharvested populations large-bodied individuals were favoured at high population densities (Fig. 3b) and enjoyed a second reproductive season (full discussion in SI Appendix II), probably causing the observed changes towards faster somatic growth and delayed maturation.

328 This conclusion is supported by harvest-by-food interactions on maturation. Both probabilistic maturation reaction norms and maturation rates consistently showed that harvested medaka matured

330 significantly earlier than unharvested medaka in a low-food environment only. This result suggests that medaka maturation responded to selection in a low-food environment (Falconer 1990) and, hence, 332 responded to selection under high population densities in unharvested populations (Fig. 2). This result fits with results in the Trinidadian killifish (*Rivulus hartii*), in which predation-induced change to earlier maturation was more pronounced in a high-food environment, presumably because predators 334 decrease the population density of *Rivulus* and thus favour genes conferring an early maturation in a high-food environment (Walsh & Reznick 2008, see also Walsh & Reznick 2010, 2011).

336

338

340

342

344

346

348

Finally, the primacy of natural selection as a driver of life-history divergence among harvested and unharvested pond medaka populations is also supported by a previous selection experiment in the laboratory. In this experiment, medaka were either randomly size-selected (control line), selected against a large body size or selected against a small body size, with populations sizes (about 200 individuals) and intensity of selection on body size (80%) similar as those in the ponds. In the laboratory, response to size-selection was asymmetric: medaka were unable to respond to selection for a smaller body size, but were able to evolve delayed maturation and faster somatic growth in response to selection for a larger body size (Le Rouzic et al. in press; Renneville et al. in press). This is an important result suggesting that, in ponds also, medaka did not respond to direct harvest selection for a smaller body size but rather to density-dependent natural selection for a larger body size (Fig. 1).

The absence of any harvest-by-food interaction on medaka somatic growth rates did not support either harvest or natural selection exclusive as a primary driver of trait change. This negative result may have two mutually-exclusive causes: (1) harvest selection at high food and natural selection at low food were

identical in strengths and, at the same time, induced similar correlated responses in a low- and highfood environments, respectively, or (2) somatic growth rates are not subject to gene-by-food
interactions in medaka. Identical strengths of natural and harvest selection are improbable because
harvest selection was shown to be stronger than natural selection in multiple systems (Darimont *et al.*2009), and similar correlated responses in multiple environments are uncommon (Falconer 1990). In
contrast, somatic growth rate was reported not to show any gene-by-food interaction in several
experiments in mice and rats (Falconer 1990). Hence, we conclude that medaka somatic growth rates
were most likely not subject to any gene-by-food interaction, such that selection on somatic growth in
one food environment resulted in a similar response in any food environment.

Rates of food acquisition

350

352

354

356

358

In the medaka, it was recently shown in the laboratory that selection for a larger body size favoured higher rates of food acquisition in females and lower boldness in males (Diaz Pauli *et al.* 2019), supporting the prediction that size-dependent selection may alter metabolic rates (Claireaux et al. 2018 and references therein). Combining our results on rates of food acquisition, somatic growth and maturation allows us to propose qualitative hypotheses for how different energy pathways responded to size-dependent selection in different food environments.

In a low-food environment, unharvested and harvested medaka had similar food-acquisition rates (Fig. 6c), but unharvested medaka had higher somatic growth rates (Figs. 5a, 6a) and delayed maturation (Figs. 5b, 6b), suggesting energy re-allocation from reproduction to somatic growth. In a medium-food environment, unharvested medaka had higher food acquisition rates than harvested medaka, which may

explain also why unharvested medaka grew faster (Figs. 5a, 6a), but the higher feeding rate of unharvested medaka did not induce any earlier maturation (Figs. 5b, 6b), again suggesting energy reallocation to somatic growth. Finally, in a high-food environment unharvested and harvested medaka had similar rates of food intake (Fig. 6c) and maturation (Figs. 5b, 6b), but unharvested medaka grew faster (Figs. 5a, 6a), suggesting that they had a higher energy assimilation efficiency. Note that our conclusions hold true only if energy pathways other than somatic growth and maturation show negligible response to harvesting.

It is also important to keep in mind that we obtained our results in a simplified experimental system, where medaka could virtually not avoid interacting with the fishing gear (98% catchability) and, hence, where the fishery selected directly on body size only (Fig. 1). In the wild, fish can avoid interacting with the fishing gear through adopting different escape behaviours or habitat choices, and fisheries thus select directly on both behaviour and body size in parallel (Diaz Pauli & Heino 2014; Claireaux et al. 2018). Compared to strict size-dependent selection, this added complexity might potentially lead to different rearrangements in the rates of energy acquisition, assimilation and/or allocation.

Maternal effects

378

380

382

384

386

388

By examining harvest-by-food interactions in F₁ medaka progeny, our design avoided the possible selective effects of the captive environment on medaka phenotypes (i.e., domestication), but did not remove possible maternal effects, defined as traits or genes expressed in the mother that influence offspring phenotypes (Lynch & Walsh 2018). Parental body condition was not found to differ between harvested and unharvested populations in either sex, which makes maternal or paternal effects unlikely

to have occurred. Additionally, in fish environmental maternal effects often occur when high-quality maternal environments result in larger-sized eggs and in faster early somatic growth in the offspring, an effect that generally vanishes as individuals develop and approach maturity (Einum & Fleming 1999; Heath *et al.* 1999; Lindholm *et al.* 2006). We did not measure egg size in our experiment. However, F₁ progeny born from females experiencing a high-food environment (i.e., from harvested population) had a similar size at hatch and grew slower, not faster, than did F₁ progeny born from females experiencing a low-food environment (i.e., from unharvested populations). This result is opposite to what we would expect from an environmental maternal effect on early somatic growth rates. Therefore, we conclude that environmental maternal effects were small and did not strongly influence our results.

Conclusions

400 For the first time, we provide experimental evidence suggesting that life-history divergence between harvested and unharvested populations may result from natural, density-dependent selection for a larger 402 body size in unharvested populations. This result strengthens the mounting evidence showing that density-dependent selection may be a primary driver of trait dynamics (Mueller 1988, 1997; Moorcroft 404 et al. 1996; Reznick et al. 2002; Calsbeek & Smith 2007; Edeline et al. 2007; Calsbeek & Cox 2010; Sarangi et al. 2016). The effects of density-dependent natural selection, which often favours larger body sizes at higher population densities, come in conflict with the effects of density-dependent 406 plasticity, which favours smaller body sizes at higher densities due to food limitation or social stress 408 (Edeline et al. 2010). This conflict blurs the phenotypic effects of density-dependent selection, which are thus likely to remain unnoticed (Wolf 2003; Hadfield et al. 2011; Kinnison et al. 2015) and, hence, 410 might be more common than the literature suggests and might help to explain why the relationship

between biomass productivity and population density is often tenuous (Vert-pre et al. 2013). Density-

dependent natural selection has important ramifications to the management of harvested populations

(Engen et al. 2014) and may drive eco-evolutionary feedback loops, which are critical for the

maintenance of biodiversity in the context of global changes (Dieckmann & Ferrière 2004; Edeline &

Loeuille 2020). In particular, introduction of invasive species, habitat destruction, climate warming and

harvesting are all strong disruptors of both population density and size structure and, hence, are

potentially strong drivers of density-dependent eco-evolutionary feedbacks on body size.

418 **Acknowledgements.** We are indebted to Prof. Kiyoshi Naruse (NIBB, Okazaki, Japan) for his support

in obtaining and maintaining wild medaka from Kiyosu. Tony Masson, Solène Boursault, Angevine

Masson, and Nicolas Florès contributed to count medaka larvae in ponds. Iago Bonnici, Stéphane

Loisel, Antoine Vallier and Paul Hübner contributed to counting zooplankton, and David Rozen, Gérard

422 Lacroix and Stéphane Loisel contributed to collecting data on filamentous algae. We thank Prof. L.

Asbjørn Vøllestad (Universitetet i Oslo) and Nicolas Loeuille (Sorbonne Université) for helpful

comments on an early version of the manuscript. Jenilee Gobin (Trent University) and four anonymous

reviewers further provided highly relevant comments that helped us to strongly improve the quality of

426 later versions.

412

414

416

420

424

428

Funding. This work has benefited from technical and human resources provided by CEREEP-Ecotron

IleDeFrance (CNRS/ENS UMS 3194) as well as financial support from the Regional Council of Ile-de-

France under the DIM Program R2DS bearing the references I-05-098/R and 2015-1657. It has

430 received a support under the program "Investissements d'Avenir" launched by the French government

and implemented by ANR with the references ANR-10-EQPX-13-01 Planagua and ANR-11-INBS-

432 0001 AnaEE France, and from Pépinière interdisciplinaire CNRS de site PSL (Paris-Sciences et

Lettres) "Eco-Evo-Devo". EE also acknowledges support from the Research Council of Norway

434 (projects EvoSize RCN 251307/F20 and REEF RCN 255601/E40), from IDEX SUPER (project

Convergences MADREPOP J14U257) and from Rennes Métropole (AIS program – project number

436 18C0356).

438

440

Author contributions. ABH performed the laboratory F₁ experiment, contributed to data analysis,

wrote the first draft of the manuscript and contributed to subsequent versions. EE designed the study,

contributed to the pond experiment, performed data analysis, and led manuscript writing from the

second version. JM, DC, SA, AM, SP, EM and BD performed the pond and laboratory experiments.

Competing interests. The authors declare no competing interests.

Data archiving statement. All data used in this paper will be archived.

Ethical statement. The protocols used in this study were designed to minimize discomfort, distress and

pain of animals, and were approved by the Darwin Ethical committee (case file #Ce5/2010/041).

Supplementary Information

446 SI Appendix I: Supplementary Methods.

SI Appendix II: Natural selection on body size in medaka.

448 SI Appendix III: Supplementary Results:

• Table S1: Inference of number of age classes in pond medaka populations.

- Table S2: Structure and MCMC parameter estimates for models 4-6 and 8.
 - Table S3: Effect of medaka fishing on medaka food in ponds.
- Table S4: Analysis of deviance.
 - Fig. S1: Experimental design.
- Fig. S2: Random effects of breeding pairs on their progeny's somatic growth rate and maturation probability.

456 References

- Arlinghaus, R., Matsumura, S. & Dieckmann, U. (2009). Quantifying selection differentials caused by recreational fishing: development of modeling framework and application to reproductive investment in pike (*Esox lucius*). *Evol. Appl.*, 2, 335–355.
- Bassar, R.D., Lopez-Sepulcre, A., Reznick, D.N. & Travis, J. (2013). Experimental evidence for density-dependent regulation and selection on Trinidadian guppy life histories. *Am. Nat.*, 181, 25–38.
- Bellows, T.S. (1981). The descriptive properties of some models for density dependence. *J. Anim. Ecol.*, 50, 139–156.
- Calsbeek, R. & Cox, R.M. (2010). Experimentally assessing the relative importance of predation and competition as agents of selection. *Nature*, 465, 613.
- Calsbeek, R. & Smith, T.B. (2007). Probing the adaptive landscape using experimental islands: density-dependent natural selection on lizard body size. *Evol. Int. J. Org. Evol.*, 61, 1052–1061.
- Carlson, S.M., Edeline, E., Vøllestad, L.A., Haugen, Thrond.O., Winfield, I.J., Fletcher, J.M., *et al.* (2007). Four decades of opposing natural and human-induced artificial selection acting on Windermere pike (*Esox lucius*). *Ecol. Lett.*, 10, 512–521.

- Claessen, D., de Roos, A.M. & Persson, L. (2004). Population dynamic theory of size–dependent cannibalism. *Proc. R. Soc. Lond. B Biol. Sci.*, 271, 333.
- Claireaux, M., Jørgensen, C. & Enberg, K. (2018). Evolutionary effects of fishing gear on foraging behavior and life-history traits. *Ecol. Evol.*, 8, 10711–10721.
- Conover, D.O. & Munch, S.B. (2002). Sustaining fisheries yields over evolutionary time scales. *Science*, 297, 94–96.
- Darimont, C.T., Carlson, S.M., Kinnison, M.T., Paquet, P.C., Reimchen, T.E. & Wilmers, C.C. (2009).

 Human predators outpace other agents of trait change in the wild. *Proc. Natl. Acad. Sci. U. S. A.*, 106, 952–954.
- Diaz Pauli, B., Garric, S., Evangelista, C., Vøllestad, L.A. & Edeline, E. (2019). Selection for small body size favours contrasting sex-specific life histories, boldness and feeding in medaka, *Oryzias latipes. BMC Evol. Biol.*, 19, 127.
- Diaz Pauli, B. & Heino, M. (2014). What can selection experiments teach us about fisheries-induced evolution? *Biol. J. Linn. Soc.*, 111, 485–503.
- Diaz Pauli, B., Kolding, J., Jeyakanth, G. & Heino, M. (2017). Effects of ambient oxygen and size-selective mortality on growth and maturation in guppies. *Conserv. Physiol.*, 5, cox010–cox010.
- Dieckmann, U. & Ferrière, R. (2004). Adaptive dynamics and evolving biodiversity. In: *Evolutionary conservation biology* (eds. Ferrière, R., Dieckmann, U. & Couvet, D.). Cambridge University Press, Cambridge, pp. 188–224.
- Dunlop, E.S., Eikeset, A.M. & Stenseth, N.C. (2015). From genes to populations: how fisheries-induced evolution alters stock productivity. *Ecol. Appl.*, 25, 1860–1868.

- Dunlop, E.S., Heino, M. & Dieckmann, U. (2009). Eco-genetic modeling of contemporary life-history evolution. *Ecol. Appl.*, 19, 1815–1834.
- Edeline, E., Carlson, S.M., Stige, L.C., Winfield, I.J., Fletcher, J.M., James, J.B., *et al.* (2007). Trait changes in a harvested population are driven by a dynamic tug-of-war between natural and harvest selection. *Proc. Natl. Acad. Sci. U. S. A.*, 104, 15799–15804.
- Edeline, E., Haugen, T.O., Weltzien, F.-A., Claessen, D., Winfield, I.J., Stenseth, N.C., *et al.* (2010). Body downsizing caused by non-consumptive social stress severely depresses population growth rate. *Proc. Biol. Sci.*, 277, 843–851.
- Edeline, E., Le Rouzic, A., Winfield, I.J., Fletcher, J.M., James, J.B., Stenseth, N.Chr., *et al.* (2009). Harvest-induced disruptive selection increases variance in fitness-related traits. *Proc. R. Soc. Lond. B Biol. Sci.*, 276, 4163–4171.
- Edeline, E. & Loeuille, N. (2020). Size-dependent eco-evolutionary feedback loops in exploited ecosystems. *bioRxiv*, 2020.04.03.022905.
- Edeline, E., Terao, O. & Naruse, K. (2016). Empirical evidence for competition-driven semelparity in wild medaka. *Popul. Ecol.*, 58, 371–383.
- Eikeset, A.M., Dunlop, E.S., Heino, M., Storvik, G., Stenseth, N.C. & Dieckmann, U. (2016). Roles of density-dependent growth and life history evolution in accounting for fisheries-induced trait changes. *Proc. Natl. Acad. Sci. U. S. A.*, 113, 15030–15035.
- Einum, Sigurd. & Fleming, Ian.A. (1999). Maternal effects of egg size in brown trout (*Salmo trutta*): norms of reaction to environmental quality. *Proc. R. Soc. Lond. B Biol. Sci.*, 266, 2095–2100.

- Engen, S., Lande, R., Sæther, B.-E., Associate Editor: Jürgen Groeneveld & Editor: Troy Day. (2014). Evolutionary consequences of nonselective harvesting in density-dependent populations. *Am. Nat.*, 184, 714–726.
- Ernande, B., Dieckmann, U. & Heino, M. (2004). Adaptive changes in harvested populations: plasticity and evolution of age and size at maturation. *Proc. Biol. Sci.*, 271, 415–423.
- Falconer, D.S. (1990). Selection in different environments: effects on environmental sensitivity (reaction norm) and on mean performance. *Genet. Res.*, 56, 57–70.
- Gobin, J., Lester, N.P., Fox, M.G. & Dunlop, E.S. (2018). Ecological change alters the evolutionary response to harvest in a freshwater fish. *Ecol. Appl.*, 28, 2175–2186.
- Hadfield, J.D., Wilson, A.J. & Kruuk, L.E.B. (2011). Cryptic evolution: does environmental deterioration have a genetic basis? *Genetics*, 187, 1099–1113.
- Harney, E., Van Dooren, T.J.M., Paterson, S. & Plaistow, S.J. (2013). How to measure maturation: a comparison of probabilistic methods used to test for genotypic variation and plasticity in the decision to mature. *Evolution*, 67, 525–538.
- Heath, D.D., Fox, C.W. & Heath, J.W. (1999). Maternal effects on offspring size: variation through early development of chinook salmon. *Evolution*, 53, 1605–1611.
- Heino, M., Díaz Pauli, B. & Dieckmann, U. (2015). Fisheries-induced evolution. *Annu. Rev. Ecol. Evol. Syst.*, 46, 461–480.
- Heino, M. & Dieckmann, U. (2008). Detecting fisheries-induced life-history evolution: an overview of the reaction-norm approach. *Bull. Mar. Sci.*, 83, 69–93.
- Hilborn, R. & Walters, C. (1992). *Quantitative fisheries stock assessment: choice, dynamics and uncertainty*. 1st edn. Springer US.

- Hutchings, J.A. (2000). Collapse and recovery of marine fishes. *Nature*, 406, 882–885.
- Kinnison, M.T., Hairston, N.G. & Hendry, A.P. (2015). Cryptic eco-evolutionary dynamics. *Ann. N. Y. Acad. Sci.*, 1360, 120–144.
- Kinoshita, M., Murata, K., Naruse, K. & Tanaka, M. (2009). *Medaka. Biology, management and experimental protocols*. 1st edn. Wiley, Ames (USA).
- Kuparinen, A. & Festa-Bianchet, M. (2017). Harvest-induced evolution: insights from aquatic and terrestrial systems. *Philos. Trans. R. Soc. B Biol. Sci.*, 372.
- Kuparinen, A., Kuikka, S. & Merilä, J. (2009). Estimating fisheries-induced selection: traditional gear selectivity research meets fisheries-induced evolution. *Evol. Appl.*, 2, 234–243.
- Le Rouzic, A., Renneville, C., Millot, A., Agostini, S., Carmignac, D. & Edeline, E. (in press).

 Unidirectional response to bidirectional selection on body size. II Quantitative genetics. *Ecol. Evol.*
- Lester, N.P., Shuter, B.J., Venturelli, P. & Nadeau, D. (2014). Life-history plasticity and sustainable exploitation: a theory of growth compensation applied to walleye management. *Ecol. Appl.*, 24, 38–54.
- Lindholm, A.K., Hunt, J. & Brooks, R. (2006). Where do all the maternal effects go? Variation in offspring body size through ontogeny in the live-bearing fish *Poecilia parae*. *Biol. Lett.*, 2, 586–589.
- Lynch, M. & Walsh, B. (2018). *Evolution and selection of quantitative traits*. 1st edn. Oxford University Press, New York.
- Marty, L., Rochet, M.-J. & Ernande, B. (2014). Temporal trends in age and size at maturation of four North Sea gadid species: cod, haddock, whiting and Norway pout. *Mar. Ecol. Prog. Ser.*, 497.

- Matsumura, S., Arlinghaus, R. & Dieckmann, U. (2011). Assessing evolutionary consequences of size-selective recreational fishing on multiple life-history traits, with an application to northern pike (*Esox lucius*). *Evol. Ecol.*, 25, 711–735.
- Moorcroft, P.R., Albon, S.D., Pemberton, J.M., Stevenson, I.R. & Clutton-Brock, T.H. (1996). Density-dependent selection in a fluctuating ungulate population. *Proc. Biol. Sci.*, 263, 31–38.
- Mueller, L.D. (1988). Evolution of competitive ability in *Drosophila* by density-dependent natural selection. *Proc. Natl. Acad. Sci. U. S. A.*, 85, 4383–4386.
- Mueller, L.D. (1997). Theoretical and empirical examination of density-dependent selection. *Annu. Rev. Ecol. Syst.*, 28, 269–288.
- Myers, R.A. & Worm, B. (2003). Rapid worldwide depletion of predatory fish communities. *Nature*, 423, 280–283.
- Post, J.R., Parkinson, E.A. & Johnston, N.T. (1999). Density-dependent processes in structured fish populations: interaction strengths in whole-lake experiments. *Ecol. Monogr.*, 69, 155–175.
- Renneville, C., Millot, A., Agostini, S., Carmignac, D., Maugars, G., Dufour, S., *et al.* (in press). Unidirectional response to bidirectional selection on body size. I. Phenotypic, life history and endocrine response. *Ecol. Evol.*
- Reznick, D., Bryant, M.J. & Bashey, F. (2002). r- and K-selection revisited: the role of population regulation in life-history evolution. *Ecology*, 83, 1509–1520.
- Ricker, W.E. (1954). Stock and recruitment. J. Fish. Res. Board Can., 11, 559–623.
- Sarangi, M., Nagarajan, A., Dey, S., Bose, J. & Joshi, A. (2016). Evolution of increased larval competitive ability in *Drosophila melanogaster* without increased larval feeding rate. *J. Genet.*, 1–13.

- Sgrò, C.M. & Partridge, L. (2000). Evolutionary responses of the life history of wild caught *Drosophila melanogaster* to two standard methods of laboratory culture. *Am. Nat.*, 156, 341–353.
- Silva, A., Faria, S. & Nunes, C. (2013). Long-term changes in maturation of sardine, *Sardina pilchardus*, in Portuguese waters. *Sci. Mar.*, 77, 429–438.
- Swain, D.P., Sinclair, A.F. & Mark Hanson, J. (2007). Evolutionary response to size-selective mortality in an exploited fish population. *Proc. R. Soc. B Biol. Sci.*, 274, 1015.
- Van Dooren, T.J.M., Tully, T. & Ferrière, R. (2005). The analysis of reaction norms for age and size at maturity using maturation rate models. *Evolution*, 59, 500–506.
- Verhulst, P.F. (1838). Notice sur la loi que la population suit dans son accroissement. *Corresp. Mathématique Phys.*, 10, 113–121.
- Vert-pre, K.A., Amoroso, R.O., Jensen, O.P. & Hilborn, R. (2013). Frequency and intensity of productivity regime shifts in marine fish stocks. *Proc. Natl. Acad. Sci.*, 110, 1779.
- Walsh, M.R. & Reznick, D.N. (2008). Interactions between the direct and indirect effects of predators determine life history evolution in a killifish. *Proc. Natl. Acad. Sci.*, 105, 594.
- Walsh, M.R. & Reznick, D.N. (2010). Influence of the indirect effects of guppies on life-history evolution in *Rivulus hartii*. *Evolution*, 64, 1583–1593.
- Walsh, M.R. & Reznick, D.N. (2011). Experimentally induced life-history evolution in a killifish in response to the introduction of guppies. *Evolution*, 65, 1021–1036.
- Wolf, J.B. (2003). Genetic architecture and evolutionary constraint when the environment contains genes. *Proc. Natl. Acad. Sci. U. S. A.*, 100, 4655–4660.
- Yamamoto, T. (1975). *Medaka (killifish): biology and strains*. 1st edn. Keigaku Pub. Co, Tokyo.

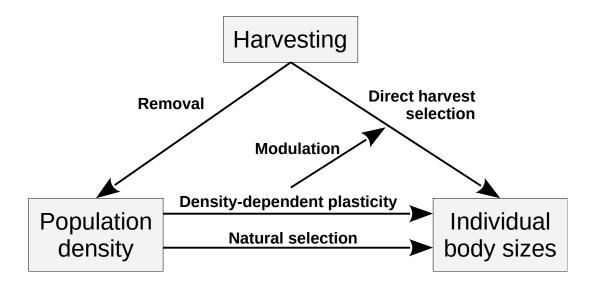


Fig. 1. Interaction pathways between size-selective harvesting, population density and individual body sizes. Harvesting simultaneously selects directly against large-bodied individuals (Direct harvest selection pathway) and lowers population density (Removal pathway). The later translates into increased food availability and faster somatic growth in survivors (Density-dependent plasticity pathway). Faster somatic growth rates may shift size-dependent harvesting from targeting both immature and mature fish, to targeting mature only fish (Modulation pathway). Lower population density further relaxes density-dependent natural selection acting on body sizes (Natural selection pathway).

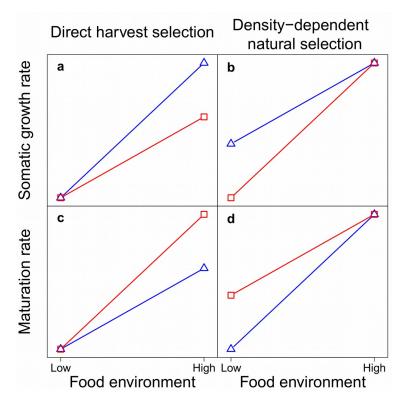


Fig. 2. Predicted harvest-by-food interaction patterns on somatic growth rates (e.g., mm day⁻¹) and maturation rates (e.g., probability day⁻¹) under direct harvest selection vs. density-dependent natural selection on body size. Red squares: mean trait value in harvested populations, blue triangles: mean trait value in unharvested populations. Genotype-by-environment interactions often result in response to selection being larger in the environment in which selection was performed than in other environments (Falconer 1990). Harvesting selects for a small body size and simultaneously increases food availability through decreased population density. Hence, body-size response to direct harvest selection is expectedly larger in a high-food than in a low-food environment. In contrast, natural selection for a large body size occurs at high population density and low food. Hence, body-size response to density-dependent natural selection is expectedly larger in a low-food than in a high-food environment.

466

468

470

472

474

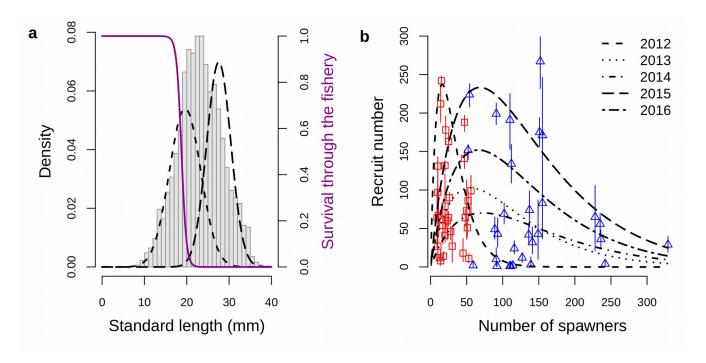


Fig. 3. Direct harvest selection and density-dependent natural selection on medaka in ponds. a: Direct harvest selection on age and size-at-age. Grey bars represent raw standard length data in harvested populations. Superimposed Gaussians represent mean MCMC estimates for the density of 0+ juveniles (short-dashed curve) and 1+ and older adults (long-dashed curve) individuals. The purple logistic curve is the mean relationship between individual probability to survive through the fishery and standard body length, showing that direct harvest selection was both selecting against an old age and against large-bodied 0+ individuals. b: Harvesting relaxes negative density-dependence on juvenile medaka recruitment. Points show mean MCMC estimate for the absolute number of age-0+ recruits with 95% credible intervals from the Gaussian mixture model (see Methods), for harvested (red squares) and unharvested (blue triangles) populations. Each point represents one pond in a given year. Black curves show year-specific Ricker stock-recruitment functions fitted to mean estimates using maximum likelihood.

476

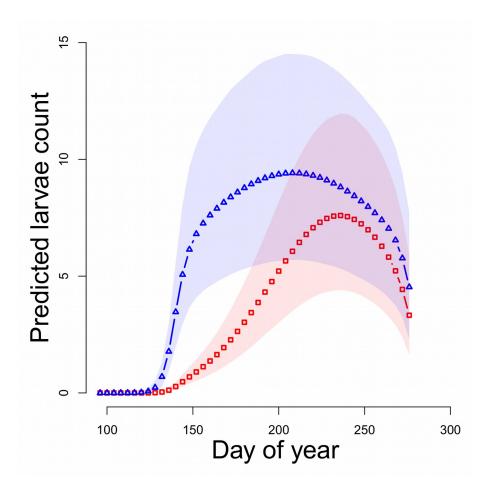
478

480

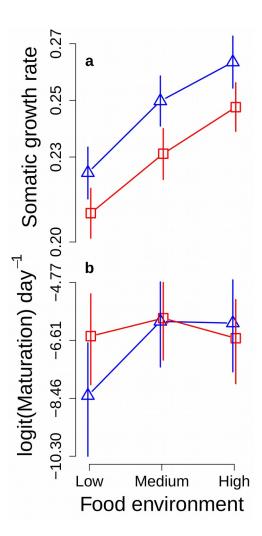
482

484

486



488 Fig. 4. Larvae count seasonal dynamics in ponds during three years. Dots represent mean MCMC estimates for daily counts of newly-hatched larvae for unharvested (blue triangles) and harvested (red squares) populations. Shaded areas show 95% credible intervals around mean MCMC estimates for unharvested (blue) and harvested (red) populations. Raw data are not shown due to a large range (from 0 to 95 larvae, mean 12.4 in unharvested populations; from 0 to 120 larvae, mean 7.4 in harvested populations).



494 Fig. 5. Harvest-by-food interaction patterns on somatic growth rates and maturation rates in individually-raised F₁ progeny in the laboratory. Points represent mean MCMC estimates with 95%
 496 credible intervals for medaka originating from unharvested (blue triangles) and harvested (red squares) populations, and maintained in a low-food, medium-food or high-food environment. Under low food
 498 conditions, offspring from harvested populations grew slower and matured at a faster rate than unharvested populations. Somatic growth rates are in mm day-1, maturation rates are in logit of
 500 maturation probability per day.

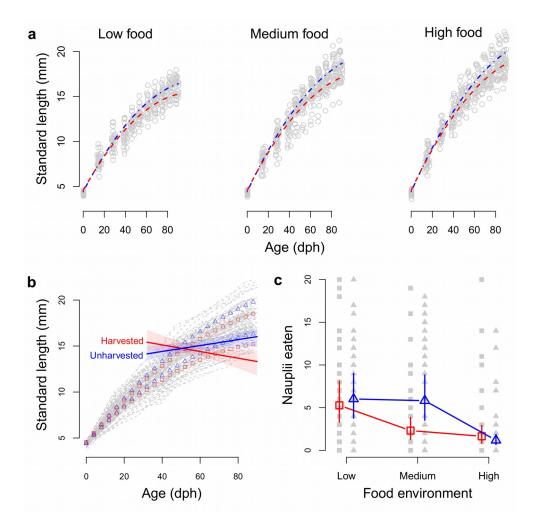


Fig. 6. Life-history of individually-raised F₁ **progeny in the laboratory. a: Somatic growth trajectories.** Somatic growth curves from mean MCMC parameter estimates for individuals originating from unharvested (dot-dashed blue curves) and harvested (dashed red curves) populations in a low-, medium- or high-food environments. Grey dots show the raw data. Dph: days post hatch. **b: Probabilistic maturation reaction norms (PMRNs).** Solid lines show MCMC mean estimates for 50% maturation probability with 95% credible intervals (not to be confounded with the PMRN width). In the background, thin grey curves show raw somatic growth trajectories for medaka originating from

502

504

506

unharvested (dot-dashed) and harvested (dashed) populations, and dots show mean somatic growth trajectories from panel (a) for unharvested (blue triangles) and harvested (red squares) medaka in a low-food and high-food environments (medium-food environment ommitted for clarity). **c: Feeding behaviour.** Coloured, open points symbols show mean MCMC estimates with 95% credible intervals for the number of prey eaten by medaka originating from unharvested control (blue triangles) and harvested (red squares) populations and maintained in a low-, medium- or high- food environment. Grey, filled symbols show the raw data.