Effect of time-series length and resolution on abundance- and trait-based early warning signals of

population declines

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Abstract

Natural populations are increasingly threatened with collapse at the hands of anthropogenic effects.

Predicting population collapse with the help of generic early warning signals (EWS) may provide a

prospective tool for identifying species or populations at highest risk. However, pattern-to-process

methods such as EWS have a multitude of challenges to overcome to be useful, including the low

signal to noise ratio of ecological systems and the need for high quality time-series data. The inclusion

of trait dynamics with EWS has been proposed as a more robust tool to predict population collapse.

However, the length and resolution of available time series are highly variable from one system to

another, especially when generation time is considered. As yet it remains unknown how this variability

with regards to generation time will alter the efficacy of EWS. Here we take both a simulation- and

experimental-based approach to assess the impacts of relative time-series length and resolution on the

forecasting ability of EWS. We show that EWS' performance decreased with decreasing length and

resolution. Our simulations suggest a relative time-series length between ten and five generations and a

resolution of half a generation are the minimum requirements for accurate forecasting by abundance-

based EWS. However, when trait information was included alongside abundance-based EWS, we

found positive signals at lengths and resolutions half of what was required without them. We suggest

that, in systems where specific traits are known to affect demography, trait data should be monitored

and included alongside abundance data to improve forecasting reliability.

Keywords: early warning signals, population collapse, sampling, time-series length, trait-based EWS

Introduction

Anthropogenic pressures have long been known to reduce the resilience of ecological systems, leaving

them vulnerable to transitioning into undesirable states where the systems' ability to provide valuable

ecosystem services is diminished. Such undesirable transitions have occurred in multiple systems such

as in the global whale stock collapse of the 20th century due to overfishing (Hilborn et al. 2003,

Clements et al. 2017); lake eutrophication through heavy nutrient input (Smith and Schindler 2009); or

coral bleaching as a result of increased ocean warming (Hughes et al. 2017). In many cases recovery

from such a perturbed state can be difficult as complex systems such as those seen in ecology often

show hysteresis (Folke et al. 2004, Scheffer et al. 2009), thus driving a need to minimize impacts on

biological systems, as well as a developing effective methods to monitor them (Costanza et al. 1997).

Early warning signals (EWS) have been shown to predict population collapses (Wissel 1984, Drake and

Griffen 2010, Dai et al. 2012, Clements and Ozgul 2018) and shifts in ecosystem states (Scheffer et al.

2009, Carpenter et al. 2011). These indicators provide the possibility to intervene and reverse these

undesirable events (Biggs et al. 2009, Pace et al. 2017). Classical EWS are statistical signatures which

arise as a result of a phenomenon known as critical slowing down (CSD) that occurs prior to an

ecosystem transition (Dakos et al. 2008, Scheffer et al. 2009, Clements and Ozgul 2018). CSD occurs

as a system loses stability in the face of increasing external stress and takes longer to return to its

original equilibrium state (Wissel 1984). Directly measuring CSD requires monitoring the return rate of

the system which is challenging to do in natural systems. Alternatively, whether a system is

experiencing CSD can be inferred through other statistical metrics measured over state-based data, for

example abundance time series. Some proposed statistical signatures related to the return rate of a

dynamical system are variance and autocorrelation, although various other metrics have also been

developed (Dakos et al. 2012a). Increases in variance and autocorrelation in abundance time series,

known as classical abundance-based EWS (Drake and Griffen 2010, Dai et al. 2012), are shown both

theoretically and experimentally to occur in systems approaching transition (Drake and Griffen 2010,

Carpenter et al. 2011, Dakos et al. 2012b, Clements and Ozgul 2016b).

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Increases in variance and autocorrelation, along with other classical abundance-based EWS, provide an

ideal generic method that responds to the dynamics of the system independent of any system-specific

data such as intrinsic demographic processes and extrinsic environmental factors. However, the

performance of these classical EWS has been questioned in numerous simulations (Hastings and

Wysham 2010, Boerlijst et al. 2013, Clements et al. 2015, Burthe et al. 2016, Clements and Ozgul

2016a, Dutta et al. 2018) as well as in experimental and field data (Wilkinson et al. 2017, Pace et al.

2017). Recent studies on data quality have shown that these signals might require high-resolution time-

series data to produce reliable forecasting (Clements et al. 2015). Given that EWS are statistics derived

from abundance time series, the quality of data available is critical to obtain a strong forecast by EWS

and temporal limitations might have many consequences.

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In monitoring programs from natural ecological systems high resolution data can be hard to achieve. In

addition, data can be spatially and temporally limited due to constraints on resources. Thus, data from

ecological systems can often present with short time-series lengths, low sampling resolutions, or both

(Clements et al. 2015). Further, abundance time-series data collected in field or experimental

populations can vary greatly in temporal quality. For instance, in a laboratory experiment, populations

of *Didinium nasutum* were sampled roughly once every two generations for 45 days before a collapse

of the population occurred (Clements and Ozgul 2016a) while in a field experiment data on a lake

system was collected once a day during the summer season for three years (Carpenter et al. 2011).

Similarly, in wild populations demographic data is generally collected monthly (van Benthem et al.

2017) or annually (Walle et al. 2018). Due to these differences in sampling effort the temporal quality

of the time series relative to the process rate of the system (for example the generation time of the

species sampled) will vary (Clements and Ozgul 2018). Previous work suggested that the length and

resolution of data being analyzed in relation to the process rate, that is: the relative length and

resolution, could alter the rate at which a tipping point occurs (Spanbauer et al. 2016). Following this, it

has been shown that the speed at which a tipping point occurs can affect the detectability of abundance-

based EWS (Clements and Ozgul 2016b). It is thus unknown how the temporal quality of the time

series relative to the generation time of the organisms being monitored will affect the detectability of

abundance-based EWS. Given that wild populations are monitored using varied sampling efforts, it is

important to further our understanding of the relative length and resolution of time series needed to

derive reliable EWS. In turn, this would help lay the foundation for generalizable guidelines for the

monitoring of populations with the aim of reliably predicting population declines, or whether time

series that are currently available will be suitable for detecting abundance-based EWS.

Composite EWS have been proposed as a more reliable method whereby multiple leading indicators

are combined to increase overall forecasting ability (Drake and Griffen 2010). Recent work has used

this composite approach to drive the inclusion of fitness related phenotypic trait data, specifically body

size, alongside abundance-based methods to create trait-based EWS (Clements and Ozgul 2016a,

Clements et al. 2017). The motivation behind the use of fitness-related trait data combined with leading

indicators comes from a body of work providing evidence that individual traits affected by changes in

the external environment are linked with concurrent demographic changes (Ozgul et al. 2009, Pigeon et

al. 2017, Baruah et al. 2018b). In particular, individual plasticity in body size has been shown to buffer

environmental change to higher trophic levels for example in the face of reduced prey availability,

climate change, and increased pollution (Brown et al. 2004, Cheung et al. 2013), For example, previous

study has shown that shifts in body size can accompany a transition in diatom communities (Spanbauer

et al. 2016). In addition, trait-based signals derived from body size data have been shown to be more

robust than traditional abundance-based leading indicators (Clements and Ozgul 2016a, Clements et al.

2017, Baruah et al. 2018a). There remains a need to fully assess whether the inclusion of body size data

leads to any considerable improvement in forecasting population collapses in the face of common data

quality issues, such as shortened time-series lengths.

In this paper, we use model-simulations and data from microcosm populations of *Didnium nasutum*

(Clements and Ozgul 2016a) to test and compare the effects of sampling length and resolution on the

detectability of population collapse by both classical abundance-based EWS as well as trait-based

EWS. We first investigated the strength and reliability of abundance-based EWS across a range of

sampling lengths and resolutions. Subsequently, we tested whether the inclusion of trait dynamics

(body size) can increase forecasting ability even when data is sparse.

Methods

2.1 Simulations: transcritical and fold bifurcation model

We first modeled the logistic growth of a population that moves from an underexploited state to critically

exploited state through a non-catastrophic transcritical bifurcation. The population is forced through the

transcritical bifurcation via a linear harvesting regime. The dynamics of this population are given by:

$$dN/dt = rN(1 - N/K) + \sigma N dW - c_t N \quad (1)$$

where, r is the growth rate of the population (0.5 individuals/day), K is the carrying capacity (100

individuals), c_t is the harvesting rate, and $\sigma N dW$ is the Gaussian distributed white noise process with

mean 0 and standard deviation σ (1.5). Time step dt used was 0.3 for each of the stochastic simulations

and was implemented using the Euler approximation. The simulation was run for 100 total time steps.

Next, to simulate the dynamics of population collapse via a fold catastrophe, we used a second model

where harvesting of individuals of the population followed a non-linear function (May 1977, Scheffer

2009). The parameters in this model are identical to those of the transcritical catastrophe model except

for the addition of $h,$ the half-saturation constant:

$$dN/dt = rN(1 - N/K) + \sigma N dW - c_t\, N^2/(h^2 + N^2) \quad (2)$$

2.2. Population collapse and abundance-based EWS:

We simulated population collapse for the two models by increasing the value of the harvest parameter

c_t linearly with time. We used three different levels of forcing (Clements and Ozgul 2016b) : 1) fast

forcing: where the forcing parameter c_t increased linearly at the rate of 0.07 and 0.004 in fold and

transcritical bifurcation models respectively; 2) moderate forcing: where the forcing parameter c_t

increased linearly at the rate of 0.045 and 0.0025 in fold and transcritical models respectively; 3) slow

forcing: where the rate of forcing increased linearly from 0.03 and 0.0015 in fold and transcritical

models respectively. For each simulated population's time-series data, we estimated the bifurcation

time point by fitting GAMS (Generalized Additive Modelling) to $1/N\,(dN/dt)$ over time $t.$ The time

point at which $1/N\,(dN/dt) < 0$ is then our estimated bifurcation time point. For abundance-based

EWS' analyses we discarded abundance time-series data after the estimated bifurcation point. We then

applied generic EWS of population collapse by using the *earlywarnings* package (Dakos et al. 2012a)

in R version 3.5.2 (R Core Team 2018). Specifically, we used two early warning indicators:

autocorrelation at first-lag (*ar(1)*) and standard deviation (*sd*). Other indicators such as return rate, first-

order autoregressive coefficient, coefficient of variation can theoretically be derived from these two

main indicators. We used Gaussian detrending to remove any trends in the abundance time-series data.

To quantify the strength of population collapse, we calculated Kendall's Tau correlation coefficients of

the statistical indicators over time. Strong positive Kendall's Tau correlation of the statistical indicators

(sd, ar(1)) with time would indicate an approaching population collapse (Dakos et al. 2012b). We

further quantified the rate of false negatives as the number of times Kendall's Tau value calculated to

be less than or equal to zero within the set of replicate time series. While previous work has suggested

that a strong trend is indicated by a Kendall's tau correlation approaching one (Dakos et al. 2012b), we

use this false negative metric as an alternative visualization to presenting raw Kendall's tau values.

2.3 Effect of sampling in relation to simulated population dynamics on EWS

To study the effect of varying resolutions on the detectability of collapse, we subset the data into four

datasets that varied in their resolution: as the generation time for the populations in the simulation models

is t=1, sampling of the abundance time series was done every quarter of *t*, every half of *t,* every *t,* and

every two *t*. Interpolation between points was not performed. With these four datasets, we explored the

effect of different sampling resolutions on the efficacy of EWS forecasting. Specifically, we assessed the

resolution required to detect positive EWS. We assessed the effect of these sampling regimes on EWS

for the three different levels of environmental forcing as mentioned in section 2.2. Next, we calculated

Kendall's tau correlation coefficients as a measure of the strength of EWS and rates of false negatives as

a measure of the reliability of EWS for these sampling regimes and for each level of environmental

forcing.

2.4 Effect of varying length of simulation abundance time series on EWS

To quantify the effect of varying lengths of abundance time series on EWS we used the quarter

generation sampling resolution dataset acquired following section 2.3. We used this resolution as it

allowed us to explore the largest range of time-series lengths. Next, we quantified the strength and

reliability of EWS on the entire length of the abundance time series and then re-quantified it using the

same time series but with the earliest data point removed (or the furthest data point from the bifurcation

point). We repeated this process by sequentially removing the earliest data point and re-performing

EWS analysis until the time series was too short for any meaningful analysis. We performed this length

reduction analysis on time series from the three forcing scenarios mentioned in section 2.2. Finally, we

calculated Kendall's tau correlation coefficients as a measure of the strength of EWS and rates of false

negatives as a measure of the reliability of EWS for these sampling regimes and for each level of

environmental forcing.

2.5. Experimental data

In addition to the model simulations, we also analyzed an experimental microcosm dataset. In this

experiment, microcosm populations were forced to collapse by varying the rate of decline in prey

availability over time in four different scenarios: 1) fast decline in food availability, 2) moderate decline

in prey availability, 3) slow decline in prey availability, 4) constant prey availability as the control

treatment. The microcosm populations consisted of protozoan ciliate *Didinum nasutum* feeding on

Paramecium caudatum. This particular experiment used a total of 60 replicate populations, where 15

replicates were used per treatment. In our study, we used the microcosm data only from the three different

deteriorating environments (fast, moderate and slow decline in prey availability). For details of the

experimental design refer to Clements and Ozgul 2016a.

2.5.1 Effect of varying length of experimental time series on EWS

For each of the three experimental treatments (fast, moderate, and slow), we estimated the effect of

different lengths of time series on the efficacy of EWS as was done with simulation data in section 2.4

to directly compare our simulation results with the experimental data.

2.5.2 Effect of sampling in relation to microcosm population dynamics on EWS

We subset the experimental data into two sampling regimes. Since *Didinum nasutum* has a generation

time of roughly 2 days (Beers 1926), sampling of the abundance time-series data was done: 1) every

half a generation (everyday), 2) every generation (every 2 days). Subsampling of this kind was done for

each of the three experimental treatments of the microcosm population collapse. Next, as was done

with simulation data, we quantified Kendall's tau correlation coefficients as a measure of the strength

of EWS and rates of false negatives as a measure of the reliability of EWS for these sampling regimes

and for each level of environmental forcing.

2.6 Inclusion of body size data: trait-based EWS

We wanted to assess whether including trait dynamic information (body size) alongside abundance-

based EWS would improve the predictability of population collapse for the scenarios of time-series

length and for the sampling resolutions in the experimental data for the three forcing experimental

treatments. To evaluate the utility of trait-based EWS for the different sampling resolutions and lengths

of abundance time series, data from mean body size was incorporated with the leading indicators in an

additive manner (Clements and Ozgul 2016a). We z-standardized abundance-based EWS (standard

deviation (sd), autocorrelation at-lag-1 (ar(1)), and body size time-series data. The length of body-size

time series that was used was the same as the corresponding replicate abundance time series. Before a

population collapse, sd and ar(1) are anticipated to increase linearly over time, while body-size is

expected to decline as food availability in the experimental treatment decreases. As a consequence, z-

standardized body-size time series were multiplied by -1 so that they could be included alongside

standardized abundance-based EWS. Next, standardized abundance-based EWS were then added to

standardized mean body-size time series to create trait-based EWS. Next, we evaluated two trait-based

EWS metrics namely *ar(1)+mean body size, sd+mean body size*. Finally, we compared these two trait-

based EWS in terms of strength and reliability of forecasting a population collapse with the abundance-

based statistical EWS for both the scenarios of varying time-series length and varying sampling

resolution.

Results

3.1 Effect of sampling resolution on abundance-based EWS (simulated data)

Reducing the resolution of the time series used for analysis with abundance-based EWS in simulated

populations led to reductions in the strength of abundance-based indicators of collapse. Decreases in

the resolution of the time series generally led to small decreases in the median strength of the signal of

population collapse (Fig. 1) and these were mostly seen only in *sd* especially when forcing was strong

or moderate (Fig. 1 - yellow). While the median strength of signals dropped slightly with decreasing

resolution, the maximum predicting strength achieved (the highest Kendall's tau value) dropped more

sharply leading to less confident predictions with lower resolutions. This trend of decreased strength of

EWS was generally maintained between bifurcation types (Fig. 1). Further, as resolution was

decreased, the spread of Kendall's tau values decreased leading to a smaller number of false negative

signals. However, the proportion of these false negatives followed an upward trend with decreasing

resolution in both bifurcation types indicating a decrease in the reliability of forecasting in addition to

reduced strength (Table S1).

3.2 Effect of length of time series on abundance-based EWS (simulated data):

In simulations, decreasing the length of the sampling time series before either a fold or transcritical

bifurcation negatively affected the performance of abundance-based EWS *ar(1)* and *sd* across all three

intensities (Fig. 2). Moreover, as forcing increased, the decline of signal reliability (Fig. 2B) and signal

strength (Fig. 2A), as indicated by the rate of false negatives and the Kendall's tau value respectively,

saw a steeper decline. With strong forcing, when time series dropped below approximately ten

generations long both EWS performed poorly with Kendall's tau values near zero (Fig. 2A) and

approximately 50% false negatives (Fig. 2B). With moderate forcing, this minimum length drops closer

to 5 generation long timeseries. Finally, with slow forcing the minimum timeseries length becomes less

clear.

3.3 Effect of sampling resolution on abundance and trait-based EWS (experimental data):

Decreasing the resolution of the time series in microcosm populations led to decreases in the strength

of abundance-based EWS (Fig. 3), corroborating the results of model simulations (Fig. 1). The decrease

in strength of was most noticeable in *sd*, while *ar(1)* remained nearly constant as resolution was

decreased (Fig. 3).

Including information from mean body size data alongside abundance-based EWS to create trait-based

EWS led to significant increases in strength and reliability of predicting population collapse regardless

of forcing strength, or time-series resolutions (Fig. 3, dark colors).

3.4 Effect of length of time series on abundance and trait-based EWS (experimental data):

In agreement with the results from simulations, abundance-based EWS derived from microcosm data

saw an increase in false negatives and a decrease in signal strength as the length of the sampling time

series was decreased (Fig. 4A, 4B, dotted lines). This was most noticeable with fast forcing but also

occurred in all forcing conditions with time series of six or less generations long. However, the drop in

performance seen in microcosm data was lower than in the simulations, which is partially due to the

fact that the former performs the analysis on a smaller range of time-series lengths.

Furthermore, trait-based EWS i.e., including body size information alongside abundance EWS,

significantly improved the strength as well as reliability of population collapse across different length

in time-series data for different levels of environmental forcing. With trait-based EWS, Kendall's Tau

value always remained positive regardless of very short time series and forcing strength (Fig 4A, solid

lines). Similarly, the number of false negatives was also significantly lower than the abundance-based

EWS alone across different time-series lengths (Fig 4B, solid lines).

Discussion

Generic EWS would provide a unique tool for conservation prioritization and management of

populations facing increased stress with changes to their abiotic environment if they are detectable

prior to their collapse (Scheffer et al. 2009, Dakos et al. 2012b). The attraction of generic EWS is their

relative simplicity; they are easy to calculate and require only state data such as the abundance of a

population. Alternative approaches such as trait-based EWS have been developed with the goal of

providing more reliable predictions of population collapse but require additional data to calculate

(Clements and Ozgul 2016a, Clements et al. 2017). For both complimentary approaches how the

relative length and resolution of a time series may affect their performance has thus far remained

unknown. This question is critical for our understanding of the utility and applicability of these

methods. Here, using simulations and experimental data, we show that time-series length and resolution

relative to the process rate of the system in question significantly influence the performance of

abundance-based EWS. Further, we found that including average body size with abundance-based

EWS leads to stronger and more reliable signals even when temporal length and resolution of the time

series are low.

4.1 Effect of resolution of sampling time series on abundance-based and trait-based EWS

Reducing the resolution of time series in simulated and experimental populations subjected to different

forcing levels led to a decrease in the performance of abundance-based EWS (Fig. 1 and 3). There were

important differences in the behavior of abundance-based EWS calculated between simulated and

experimental populations. While in simulated population we found *sd* to produce the most robust signal

of population collapse (Fig. 1), in experimental populations *ar(1)* performed better across varying

resolutions (Fig. 3). Our simulated populations were likely subjected to smaller amounts of

stochasticity in population size compared to our experimental populations. Given that standard

deviation-based EWS are more sensitive to rising stochasticity in a system (Boettiger and Hastings

2012), this difference between experimental and simulated populations is expected. In addition,

abundance-based EWS showed a greater sensitivity to system stochasticity than trait-based EWS did;

the latter did not present any obvious differences in performance between metrics used (Fig. 3). This is

possibly an effect of adding phenotypic data to abundance data which stabilizes the signal through time

and reduces stochasticity. This is an important finding if abundance-based EWS are to be used to

monitor natural populations which likely have more stochasticity than laboratory populations.

As observed with abundance-based EWS, trait-based EWS also saw a decrease in their performance

with decreasing sampling resolution (Fig. 3). However, these drops were in most cases very small and

the performance of trait-based EWS was much better than abundance-based EWS alone. Trait-based

EWS calculated with the lowest resolution in either simulated or experimental populations still derived

a more reliable forecast than abundance-based EWS calculated with the highest possible resolution.

Our findings suggest that abundance-based EWS derived from autocorrelation at-first-lag would

perform better in natural populations subjected to high levels of stochasticity when compared with

standard deviation derived signals when the resolution of available abundance time series is low. On

the other hand, the inclusion of trait data with either metric led both to an increase in EWS performance

when faced with reduced sampling time-series resolutions and minimized any important differences

between the two time-series metrics used. Thus, whenever possible, trait-based EWS will likely

outperform abundance-based EWS in natural populations (but see Baruah et al. 2018a) and, when trait

data is not available, abundance-based EWS derived from autocorrelation at-first-lag are more

powerful and reliable indicators.

4.2 Effect of length of sampling time series on abundance-based and trait-based EWS

Reducing the length of the sampling time series in either simulated or experimental data negatively

affected the performance of abundance-based EWS. In our simulation study, time-series lengths of less

than 10 generations long and less than 5 generations long when forcing is fast and moderate

respectively, returned on average, incorrect forecasts indicating no population collapse (Fig. 2). The

choice of fold or transcritical bifurcation models had little influence on predictability of population

collapse with the exception of the fast forcing scenario. When forcing was fast, transcritical models

showed a shift in trend indicating the length of time series no longer had a strong effect on the

predictability of collapse (Fig. 2 – dashed lines). In the context of a ciliate population such as the one

used in the microcosms of this paper (*D. nasutum*) 10 generations corresponds to roughly 20 days.

Twenty days' worth of sampling might seem feasible to generate reliable forecasting for a ciliate

population. However, for this same reliability in a population experiencing rapid forcing, 22 to 35 years

of sampling would be required in the case of *Thynnus thynnus* (Atlantic bluefin tuna): an endangered

species of high economic concern with a generation time between 2.2 to 3.5 years (Collette et al. 2011).

For organisms with longer generation times such as *T. thynnus,* the required investment of resources

and time becomes an important consideration if there is a desire to monitor populations using

abundance-based EWS. In addition, we found that abundance time series require a minimum of 10

generations worth of data when forcing is fast and a minimum of 5 when forcing is moderate (fig. 2).

This finding suggests that in situations where forcing is more intense and, thus, populations are most at

risk the requirement for good quality data is extended. This requirement of long time series is a clear

shortcoming of abundance-based EWS for organisms with long generation times and populations

experiencing rapid forcing.

In our microcosm study, the effect of time-series length on the Kendall's tau metric was less clear than

in the simulation. However, a general decreasing trend in forecasting strength was still observed in

most forcing scenarios. In addition, abundance-based EWS rarely had the length of sampling required

to make a confident forecast of collapse. It appears that regardless of improvements brought about by

the increased length of the sampling time series, most abundance-based EWS did not provide a

confident forecast even with the maximum length of sampling time series (between 10 and 12.5

generations). In our simulations we showed that abundance time series need a minimum of 10

generations of data to provide an accurate forecast. These microcosm results show that the required

minimum length of sampling exceeds 10 generations in controlled laboratory settings. Thus, we can

expect that in natural populations subjected to higher levels of stochasticity the minimum length of

sampling might exceed even that which is required in the microcosms. However, in natural populations

there is the possibility to measure phenotypic data and derive trait-based EWS which we have

additionally assessed. The addition of phenotypic data with abundance data improved forecasting in the

microcosms. Trait-based EWS, that included body size information alongside abundance-based EWS,

offered a significant improvement over abundance-based EWS, providing a positive forecast of

collapse for all time-series lengths tested.

Our findings on the effect of sampling length on the forecasting of population collapse by EWS suggest

that the length of sampling required to calculate confident forecasts by abundance-based EWS are very

long and likely impractical from a conservation and monitoring standpoint. On the other hand, trait-

based EWS can give accurate and reliable forecasts with much less sampling data compared to their

abundance-based counterparts.

Conclusion

Our results stemming from simulations and a laboratory microcosm study make a case for the

preferential use of trait-based EWS over classical abundance-based signals based on their increased

reliability and strength when faced with varying lengths and resolutions of sampling time series. We

attempted here to evaluate these EWS with data which might more realistically resemble data gathered

from field monitoring programs. However, true field data arising from monitoring programs are likely

to be noisier than the data we have presented and analyzed here. Nonetheless, our results suggest that

abundance-based EWS perform poorly when the length or resolution of abundance time series with

regards to the process rate of the system is decreased. We found that including trait dynamics alongside

abundance-based EWS to generate trait-based EWS leads to more reliable and confident forecasts of

population collapse. We further found in our simulations that a length of ten generations and a

resolution of a half generation are the minimum for deriving confident forecasts with abundance-based

EWS. If trait-based EWS are used, this length and resolution requirement is much more relaxed. With

these results in mind, we recommend considerations be made for the length and resolution of sampling

time series required to accurately forecast populations in the design of monitoring programs. Further,

we recommend the additional monitoring of phenotypic traits in populations, which have been shown

to vary with increasing levels of environmental forcing such that trait-based EWS which have more

forecasting power and reliability can be derived.

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Figure legends

Figure 1. Boxplots of the strength of abundance-based early warning signals of population collapse

across decreasing resolutions of sub-sampling in simulated populations subjected to collapse by

harvesting. The data is split into subplots based on the bifurcation model simulated (fold, transcritical)

and forcing level (slow, medium, fast). Boxplots are further split by metric used (yellow: standard

deviation, red: autocorrelation at-first-lag). Each box represents the median Kendall's tau value (shown

on the y-axis) across replicate time series across three forcing level with lengths ranging from 1,000 to

40 time steps. The x-axis shows the resolution of sampling in numbers of generations such that a value

of 2 indicates that sampling was performed every 2 generations and 0.25 denotes sampling every

quarter generation.

Figure 2. Performance of abundance-based early warning signals of population collapse across

decreasing sampling time-series lengths in simulated populations subjected to collapse by harvesting.

The data is split into subplots based on forcing intensity (slow, moderate, fast) and further split by

metric used (yellow: standard deviation, red: autocorrelation at-lag-1) and bifurcation model simulated

(solid: fold model, dashed: transcritical model). Each point represents (A) the mean Kendall's tau

correlation coefficient or (B) rate of false negatives on the y-axis of 100 replicate simulations of

population collapse. X-axis represents length of time series analyzed.

Figure 3. Performance of abundance-based and trait-based EWS of population collapse across

decreasing sampling time-series resolutions for the experimental data of Clements & Ozgul 2016. Plots

are organized into subplots based on the forcing intensity (slow, medium, fast) and further split by

metric used (blue: standard deviation, purple: autocorrelation at-first-lag). Each box represents the

Kendall's tau value (on the y axis) for each resolution studied (on the x axis). Lighter colors represent

classical abundance-based EWS while the darker colours represent trait-based EWS.

Figure 4. Performance of abundance-based and trait-based EWS of population collapse across

decreasing sampling time-series lengths for the experimental data of Clements & Ozgul 2016. Plots are

organized into subplots based on the forcing intensity (slow, medium, fast). X-axis is the length of

time-series data analyzed and Y-axis denotes in (A) mean Kendall's Tau value and in (B) rate of false

negatives. Solid lines indicate EWS with body size information added (trait-based EWS) while dotted

lines indicate classical abundance-based EWS. Colours indicate metric used (blue: standard deviation,

purple: autocorrelation at-first-lag). False negative plots were produced using loess smoothing.

Appendices

Table S1. Rate of false negatives from abundance-based early warning signals (EWS) of population

collapse in simulated populations led to collapse through harvesting at three different rates: weak,

medium, strong. The rate of false negatives reported is the number of false negative signals returned

divided by the number of true positive signals. Results are further split by four time-series resolutions

tested (a value of 2 indicates sampling was performed every 2 generations) and by the metric used as an

EWS.

Table S1: Mean rates of false negatives (%) for varying forcing levels and timeseries resolutions

Strength of forcing	Slow				Moderate				Fast			
Resolution	2	1	50	25	2	1	50	25	2	1	50	25
Standard deviation	52.5	52.4	51.5	45.1	51.4	50.4	44.5	41.7	39.5	51	50.4	58.7
Autocorrelation at-lag-1	46.5	46.1	49.7	46.6	46.4	47.8	48	42.6	52.5	46.5	48.2	60.7

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