

36 Introduction

37 The increased rate of human-driven global change is a major threat to biodiversity [1]. Factors such as climate
38 change, habitat fragmentation, and environmental degradation are influencing the distribution and abundance of
39 species, often in ways that are impossible to predict [2]. Thus, a central theme in conservation biology is how
40 best to manage for species persistence under rapidly changing and often unpredictable conditions. When faced
41 with environmental change, species may persist by moving (or being moved) to track suitable environments.
42 Although there is sufficient evidence to suggest that species notably alter their ranges [3], facilitation of such
43 movement for larger vertebrate species (through the creation of habitat corridors, transfrontier parks or
44 translocations) often place insurmountable burdens on conservation agencies that are ultimately responsible for
45 the management of these populations. Notwithstanding, signatory countries to the Convention on Biological
46 Diversity have an obligation to manage and protect biodiversity, as also set out more recently in the Aichi
47 Biodiversity Targets .

48

49 Admixture and introgression are major threats to species conservation (these threats are dealt with specifically
50 under Aichi Target 13; see <https://www.cbd.int/sp/targets/>). The ability to accurately identify introgression is
51 critical to the management of species [4–9], and may provide unprecedented insights into evolutionary
52 processes. Although admixture, or even genetic rescue, may have beneficial outcomes through the introduction
53 of new alleles into small or isolated populations, it can lead to outbreeding depression essentially disrupting
54 locally adapted gene-complexes [10–13]. Because of the movement of animals (either natural or human-
55 facilitated), admixture and the effects thereof become increasingly more important to understand and manage.

56

57 Roan antelope (*Hippotragus equinus*) is one of Africa's most iconic large antelope species. It has a sub-Saharan
58 range, is a water-dependant species, and prefers savanna woodlands and grasslands. [14] recognised six
59 subspecies namely *H. e. equinus*, *H. e. cottoni*, *H. e. langheldi*, *H. e. bakeri*, *H. e. charicus*, and *H. e. koba* based
60 on morphological analyses. However, subsequent genetic studies by [15] and [16] provided less support for
61 these subspecies designation. Although the [15] study included relatively few specimens (only 13 animals were
62 available at the time), [16] analyzed 137 animals sampled from across the range (the only subspecies not
63 included in this study was *H. e. bakeri*) for both the mtDNA control region and eight microsatellite markers. Both
64 the mtDNA control region and microsatellite data provided strong support for a separation between the West
65 Africa population (corresponding to the *H. e. koba* subspecies) and those from East, Central and Southern Africa
66 (representing the *H. e. equinus*, *H. e. langheldi*, and *H. e. cottoni* subspecies). Although some differentiation
67 between East, Central and Southern African roan antelope was evident from the mtDNA data, the different
68 subspecies did not form monophyletic groups, with no differentiation observed for the microsatellite data. The
69 placement of the two specimens from Cameroon (corresponding to the *H. e. charicus* subspecies) were unclear,
70 and the small sample size precluded robust analyses. Based on these results, [16] argued that two evolutionary
71 significant units should be recognized for roan antelope in Africa, corresponding to a West African lineage and an
72 East, Central and Southern African lineage.

73

74 Roan antelope is listed as Least Concern, but with decreasing population sizes, notably in East and Southern
75 Africa [17]. In Southern Africa, roan antelope numbers have dramatically declined in Botswana, Namibia and
76 Zimbabwe and these animals have been eliminated from large parts of their former range including Angola and
77 Mozambique [18]. Within South Africa, roan antelope numbers in reserves and protected areas are critically low,
78 with the majority of animals residing under private ownership on game farms. Indeed, the estimated population
79 size of wild and naturally occurring roan antelope in protected areas in South Africa is less than 300 animals [19],
80 yet indications are that roan antelope is thriving on private land. Current estimates suggest that at least 3,500
81 individuals are managed on private farms [20], with numbers increasing due to these animals being considered
82 an economically important species by the South African wildlife industry. In the 1990s, a number of roan antelope
83 (approximately 40) was imported into South Africa under permit from West Africa. Subsequent to their import,
84 and based on DNA evidence [16], an embargo was placed on the trade of West African animals in South Africa.
85 Recent anecdotal evidence suggested that animals of West African descent was being traded in (based on
86 mitochondrial haplotypes; Jansen van Vuuren, pers. comm.), thereby presenting a real and significant threat to
87 the genetic integrity of roan antelope in South Africa, notwithstanding legislation prohibiting it. Furthermore,
88 animals are sometimes being exported to other Southern African countries, further endangering regional gene
89 pools.

90

91 Our aim here is to expand on the limited and non-specific suite of microsatellite markers employed by [16] to
92 specifically test the validity of these anecdotal reports of trade in West African roan. Also, we assessed the ability
93 of these markers to discriminate between non-admixed animals and hybrid offspring (F2, F3, and F4). Our results
94 will not only confirm whether suggestions of hybridization are true, but will also provide a valuable tool to ensure
95 genetic integrity in the conservation of roan antelope in Southern Africa.

96

97 **Materials and Methods**

98 *Sampling*

99 Blood, tissue or hair material was obtained from private breeders and game farm owners throughout South Africa
100 (Table 1). Reference samples were selected from the [16] study and represent animals of confirmed provenance.
101 A total of 32 West African roan antelope (populations from three farms in Limpopo Province, South Africa), and
102 98 animals representing the East, Central and Southern African ESU (populations from two farms in the Northern
103 Cape and North West provinces, South Africa) were included. In addition, eight known hybrids and 15 putative
104 hybrids were included in this study (Table 1), provided to us by game owners that legally had West African roan
105 on their farms. Ethical approval was obtained from the Animal Research Ethics Committee, University of the Free
106 State, South Africa (UFS-AED2017/0010) and the NZG Research Ethics and Scientific Committee
107 (NZG/RES/P/17/18). Samples were stored in the NZG Biobank and access for research use of the samples was
108 approved under a Section 20 permit from the Department of Agriculture, Forestry and Fisheries, South Africa
109 (S20BB1917).

110

111 **Table 1. List of roan antelope (*Hippotragus equinus*) samples.**

Population / Province	Sample size	Classification
Western roan population A, Limpopo	12	Reference Western roan
Western roan population B, Limpopo	14	Reference Western roan
Western roan population C, Limpopo	6	Reference Western roan
Rest of Africa roan population A, Northern Cape	80	Reference rest of Africa roan
Rest of Africa roan population B, North West	18	Reference rest of Africa roan
Known hybrids, Limpopo	8	Known hybrids
Putative hybrid populations, Limpopo,	15	Putative hybrids

112

113 *Microsatellite markers*

114 We selected nine cross-species microsatellite markers (HN60, HN02, HN17, HN27, HN113, HN58, HN09, HN12
115 and HN13) that were previously characterised in sable antelope (*Hippotragus niger*) by Vaz Pinto [7] and 12
116 cross-species microsatellite markers (BM3517, BM203, SPS113, BM1818, OARFCB304, CSSM19, ILST87,
117 BM719, BM757, OARCP26, OARFCB48, INRA006) that were developed for domestic livestock [21–28]. In
118 addition, species-specific microsatellite markers were developed from non-admixed East, Central and Southern
119 African roan using a Next Generation Sequencing approach. The Nextera® DNA Sample Preparation Kit
120 (Illumina, Inc., San Diego, California, USA) was used to create a paired-end library followed by sequencing on
121 the MiSeq™ sequencer (Illumina, Inc., San Diego, California, USA) using 2 x 300 bp chemistry. Library
122 construction and sequencing was carried out at the Agricultural Research Council Biotechnology Platform
123 (Onderstepoort, Gauteng, South Africa). FastQC version 0.11.4 [29] and Trimmomatic version 0.36 [30] were
124 used for quality control of the raw sequence reads. Tandem Repeat Finder version 4.09 [31] was used to search
125 the remaining reads for microsatellite motifs and Batchprimer3 software [32] was used to design primer pairs
126 flanking the repeat regions.

127

128 *Polymerase Chain Reaction (PCR) and genotyping*

129 DNA extractions were performed using the Qiagen DNeasy® Blood and Tissue Kit (Qiagen GmbH, Hilden,
130 Germany) following the manufacturer's protocols. Polymerase Chain Reaction (PCR) amplification was
131 conducted in 12.5 µl reaction volumes consisting of AmpliTaq® DNA polymerase (Roche Molecular Systems, Inc)
132 forward and reverse primers (0.5 µM each), and 50 ng genomic DNA template. The conditions for PCR
133 amplification were as follows: 5 min at 95°C denaturation, 35 cycles for 30 sec at 95°C, 30 sec at 50-62°C
134 (primer-specific annealing temperatures) and 30 sec at 72°C, followed by extension at 72°C for 10 min in a
135 T100™ Thermal Cycler (Bio-Rad Laboratories, Inc. Hercules, CA, USA). PCR products were run against a
136 Genescan™ 500 LIZ™ internal size standard on an ABI 3130 Genetic Analyzer (Applied Biosystems, Inc., Foster
137 City, CA, USA). Samples were genotyped using GeneMapper v. 4.0 software (Applied Biosystems, Inc., Foster
138 City, CA, USA).

139

140 *Genetic diversity*

141 Understanding the diversity within groups provide valuable information to identify hybrid individuals. To this end,
142 genetic diversity was evaluated for each group separately (the two different ESUs, known hybrids, and putative
143 hybrids). MICRO-CHECKER [33] was used to detect possible genotyping errors, allele dropout and null alleles.
144 The mean number of alleles per locus (A), allelic richness (AR), observed heterozygosity (H_o), unbiased
145 heterozygosity (H_z = expected heterozygosity adjusted for unequal sample sizes) [34] and number of private
146 alleles per reference group (N_p) was calculated with GenAlEx 6.5 [35,36]. Arlequin 3.5 [37,38] was used to test
147 for deviations from expected Hardy-Weinberg (HW) proportions of genotypes (Markov Chain length of 105 and
148 100,000 dememorization steps) and to evaluate loci for gametic disequilibrium (with 100 initial conditions
149 followed by ten permutations, based on the exact test described by Guo and Thompson [39]. Associated
150 probability values were corrected for multiple comparisons using Bonferroni adjustment for a significance level of
151 0.05 [40]. In addition, to determine the discriminatory power of the combined loci, the P_{ID} was calculated using
152 GenAlEx [35,36]. Finally, inbreeding (F_{IS}) and average pairwise relatedness between individuals within
153 populations was calculated using the R package Demerelate version 0.9-3 (using 1,000 bootstrap replications)
154 [41].

155

156 *Population structure and admixture analysis*

157 To estimate the degree of genetic differentiation between populations, we performed an analysis of molecular
158 variance (AMOVA) and conducted pairwise F_{ST} comparisons among populations in ARLEQUIN version 3.5
159 [37,38]. We used two approaches to assess population structure, namely a Bayesian clustering approach
160 implemented in STRUCTURE version 2.3.4 [42–44] and a Principal Component Analysis (PCA). STRUCTURE
161 was used for the identification of genetic clusters and individual assignment of non-admixed animals as well as
162 putative hybrid individuals and was run using a model that assumes admixture, correlated allele frequencies and
163 without prior population information for five replicates each with $K = 1 - 6$, with a run-length of 700,000 Markov
164 Chain Monte Carlo repetitions, following a burn-in period of 200,000 iterations. The five values for the estimated
165 $\ln(\text{Pr}(X|K))$ were averaged, from which the posterior probabilities were calculated. The K with the greatest
166 increase in posterior probability (ΔK) [45] was identified as the optimum number of sub-populations using
167 STRUCTURE HARVESTER [46]. The membership coefficient matrices (Q -matrices) of replicate runs for the
168 optimum number of sub-populations was combined using CLUMPP version 1.1.2 [47] with the FullSearch
169 algorithm and G' pairwise matrix similarity statistics. The results were visualized using DISTRUCT version 1.1
170 [48]. From the selected K value, we assessed the average proportion of membership (q_i) of the sampled
171 populations to the inferred clusters. Individuals (parental or admixed classes) were assigned to the inferred
172 clusters using an initial threshold of $q_i > 0.9$ [49]. PCA for the complete data set was achieved using the R
173 package Adegent version 2.1.1 [50].

174

175 *Maximizing the accuracy of assignments*

176 To determine which threshold Q -value (hybridization or admixture index from clustering algorithms like
177 STRUCTURE) would maximize the accuracy of assignment, simulated genotypes were created using

178 HYBRIDLAB [51]. Genotypes of non-admixed Western roan antelope, and animals from East, Central and
 179 Southern Africa (n =30) with $q_i > 0.90$ (from STRUCTURE-based analysis) were used a parental (P1)
 180 populations to create the simulated hybrid genotypes (see [9]). A dataset consisting of 180 individuals were
 181 created consisting of 30 each belonging to non-admixed Western roan antelope, non-admixed Eastern, Central
 182 and Southern roan antelope, F1 hybrids, F2 hybrids, backcrosses of F1 with Western roan (BC-Western roan)
 183 and backcrosses of F1 with rest of Africa roan antelope (BC-rest of Africa roan). The simulated dataset was
 184 analysed with STRUCTURE version 2.3.4 [42–44] using the admixed model, correlated allele frequencies and
 185 without prior population information for five replicates each with $K = 1 - 2$, a run-length of 700,000 Markov Chain
 186 Monte Carlo repetitions and a burn-in period of 200,000 iterations.

187

188 Results

189 *Species-specific microsatellite markers*

190 In this study, species specific microsatellite markers were successfully developed using DNA extracted from non-
 191 admixed roan antelope (i.e., animals of known provenance). Read lengths of 2 x 301 bp (2 x 3,306,938) were
 192 obtained and after trimming, the remaining reads ranged from 180 to 200 bp (2 x 1,596,026). A total of 14 unique
 193 loci were identified, of these only six were polymorphic and consistently amplified animals from both ESUs (Table
 194 2).

195

196 **Table 2. List of six species-specific microsatellite loci developed in *Hippotragus equinus*: F = forward**
 197 **primer; R = reverse primer; bp = base pairs. GenBank accession numbers are MN699986-MH699992.**

Marker name	Sequence (5'-3')	Repeat unit	Fluorescent dye label	Product size in bp
RAO2118F	tgccattctgtccttctca	(TG) ₁₂	FAM	120
RAO2118R	aggacatgacttatgactgaaca			
RAO4116F	agcaatcctttgcacgaaat	(AC) ₁₂	VIC	124
RAO4116R	atgccagattgggtgacat			
RAO7593F	tgcagccagattcttacca	(TG) ₁₄	NED	120
RAO7593R	caccagaggagcccatatgta			
RAO4422F	cacgagttgttgctgaatg	(AC) ₁₅	FAM	118
RAO4422R	ctcaggctaaccacaatgc			
RAO13910F	gttgagacctggcaatgat	(AC) ₁₂	PET	119
RAO13910R	actaaaggaccgctctgctc			
RAO11139F	cattgagaatcagcgtcctg	(AC) ₁₄	NED	115
RAO11139R	tttccgtacgcctcagaatc			

198

199 *Genetic differentiation and admixture analysis*

200 The final dataset included 27 microsatellite loci that yielded a total of 267 alleles, with the number of alleles
 201 ranging from 3 to 17 per locus. A total of 27 alleles were unique to the West African roan group, while 27 were

202 found exclusively in the East, Central and Southern African group (Table 3). An analysis of molecular variance
 203 (AMOVA) unequivocally retrieved the two distinct groups (corresponding to the two ESUs reported by Alpers
 204 [16]; $F_{ST} = 0.165$, $P < 0.001$), validating our two reference groups. Principle component analysis similarly
 205 revealed a clear separation between the West African versus East, Central and Southern Africa roan (Fig 1A).
 206 The two distinct genetic clusters ($K = 2$) was supported by the Bayesian assignment analysis (Fig 1B, S1 Fig).
 207 West African versus East, Central and Southern African roan antelope were assigned to two distinct clusters with
 208 individual coefficient of membership (q_i) for non-admixed Western roan $q_i > 0.881$ and for non-admixed East,
 209 Central and Southern Africa roan $q_i > 0.883$. With regards to known hybrids, six of the eight known hybrids were
 210 confirmed as hybrids, with two hybrids being identified as non-admixed Western roan ($q_i = 0.9664$ and $q_i =$
 211 0.9510 , respectively). Analysis of putative hybrids identified four out of 15 animals as hybrid (27%).

212

213 **Fig 1. Genetic differentiation analysis between populations based on (A) Principal Component Analysis**
 214 **(PCA) and (B) STRUCTURE analysis (performed with $K = 2$) of Western roan, rest of Africa roan, known**
 215 **hybrids and putative hybrids. WRA = Western roan A, WRB = Western roan B, WRC = Western roan C,**
 216 **SRA = rest of Africa A, rest of Africa B, HYB = known hybrids, PTH = putative hybrids.**

217

218 **Table 3. Private alleles in loci and allele frequency in Western and rest of Africa roan.**

Population	Locus	Allele	Frequency	Population	Locus	Allele	Frequency
Western roan	BM203	230	0.040	Rest of Africa roan	BM203	240	0.005
Western roan	Oarcp26	146	0.031	Rest of Africa roan	BM719	169	0.005
Western roan	Oarcp26	148	0.047	Rest of Africa roan	BM719	177	0.074
Western roan	OARFCB48	176	0.078	Rest of Africa roan	Oarcp26	118	0.010
Western roan	BM1818	280	0.089	Rest of Africa roan	Oarcp26	124	0.015
Western roan	BM757	180	0.031	Rest of Africa roan	BM1818	256	0.058
Western roan	ILST87	153	0.018	Rest of Africa roan	BM1818	278	0.016
Western roan	ILST87	159	0.018	Rest of Africa roan	BM1818	282	0.068
Western roan	RAO4422	115	0.017	Rest of Africa roan	BM1818	288	0.037
Western roan	RAO4422	129	0.017	Rest of Africa roan	INRA006	117	0.077
Western roan	RAO4422	137	0.050	Rest of Africa roan	INRA006	123	0.056
Western roan	RAO4422	141	0.050	Rest of Africa roan	INRA006	125	0.077
Western roan	RAO4422	149	0.017	Rest of Africa roan	INRA006	127	0.337
Western roan	RAO4422	151	0.033	Rest of Africa roan	OARFCB304	115	0.016
Western roan	RAO4422	155	0.033	Rest of Africa roan	OARFCB304	127	0.005
Western roan	RAO4422	159	0.017	Rest of Africa roan	ILST87	121	0.005
Western roan	RAO13910	115	0.031	Rest of Africa roan	ILST87	127	0.005
Western roan	RAO4116	126	0.047	Rest of Africa roan	RAO13910	141	0.040
Western roan	HN02	186	0.063	Rest of Africa roan	RAO11139	102	0.010
Western roan	HN17	202	0.109	Rest of Africa roan	RAO11139	104	0.026
Western roan	HN58	124	0.031	Rest of Africa roan	RAO11139	108	0.072
Western roan	HN58	144	0.016	Rest of Africa roan	RAO4116	112	0.086
Western roan	HN09	152	0.047	Rest of Africa roan	HN09	168	0.005
Western roan	HN09	180	0.031	Rest of Africa roan	HN09	173	0.005
Western roan	HN09	194	0.031	Rest of Africa roan	HN12	185	0.005

Population	Locus	Allele	Frequency	Population	Locus	Allele	Frequency
Western roan	HN12	171	0.032	Rest of Africa roan	HN12	195	0.005
Western roan	HN12	193	0.032	Rest of Africa roan	HN13	184	0.025

219

220 On South African farms, game owners often employ selective breeding to achieve specific outcomes. For
221 example, hybrid animals may be backcrossed with pure roan to selectively breed hybrid lineages back to pure; in
222 theory this can be achieved in $N = 4$ generations. We wanted to assess whether our markers are able to detect
223 backcrossed animals, especially in the F3 and F4 generations. In this study, we created a simulated dataset to
224 maximize the accuracy of assignment to distinguish between the two non-admixed groups (West Africa versus
225 East, Central and Southern roan antelope), F1 hybrids, F2 hybrids, F1 BC-Western roan and F1 BC-rest of Africa
226 roan. STRUCTURE analysis of simulated genotypes generated by HYBRIDLAB indicated that all (100%) of the
227 West African roan versus East, Central and Southern Africa roan, F1 and F2 genotypes were correctly assigned
228 at thresholds of $q_i > 0.80$ and $q_i > 0.85$ (Table 4). At a threshold value of $q_i > 0.90$, all F1, F2 hybrid and the East,
229 Central and Southern Africa roan were correctly assigned, however, 20% of the non-admixed Western roan
230 would be incorrectly identified as hybrid origin. At a threshold value of $q_i > 0.95$, all F1 and F2 hybrid individuals
231 would be correctly assigned, however, 40% of non-admixed Western roan and 7% of the East, Central and
232 Southern African roan would be incorrectly identified as hybrid. Our ability to distinguish non-admixed roan from
233 backcrossed individuals may be problematic in some instances with correct assignment of backcrossed Western
234 roan individuals varying from 40% at $q_i > 0.80$ to 97% at $q_i > 0.95$, and backcrossed East, Central and Southern
235 African roan individuals varying from 53% at $q_i > 0.80$ to 97% at $q_i > 0.95$. Based on the simulation results, the
236 threshold q -value of $q_i > 0.85$ was selected for analysis of the non-admixed parental populations, known hybrids
237 and putative hybrids.

238

239 **Table 4: Percentage of individuals correctly identified at different threshold values.**

Average	Western Roan 0.934 > <i>qi</i> < 0.066	Rest of Africa Roan 0.027 > <i>qi</i> < 0.973	F1 hybrid 0.507 > <i>qi</i> < 0.494	F2 Hybrid 0.4601 > <i>qi</i> < 0.539	BC-Western Roan 0.826 > <i>qi</i> < 0.174	BC-Rest of Africa roan 0.203 > <i>qi</i> < 0.797
% of Individuals correctly identified at a threshold of 0.20	100%	100%	100%	100%	40%	53%
% of Individuals correctly identified at a threshold of 0.15	100%	100%	100%	100%	56.70%	83%
% of Individuals correctly identified at a threshold of 0.10	80%	100%	100%	100%	67%	93%
% of Individuals correctly identified at a threshold of 0.05	60%	93%	100%	100%	97%	97%

240

241

242

243

244 *Genetic diversity and relatedness*

245 Deviations from HWE equilibrium were not consistent across populations, with significant deviations from HWE
246 being observed only in the East, Central and Southern African roan populations. In the East, Central and
247 Southern Africa roan population A (Northern Cape Province), 11 loci (BM3517, BM719, OARFCB48, CSSM19,
248 BM1818, BM757, SPS113, INRA006, OARFCB304, RAO4116 and HN27) deviated from HWE. In addition, two
249 loci (BM3517 and SPS113) deviated from HWE in East, Central and Southern Africa roan population B (North
250 West Province) following Bonferroni correction. These markers indicated significant heterozygote deficit in the
251 respective populations with H_o values lower than H_e values, which may be an indication of the presence of
252 possible null alleles. However, null alleles were only observed in six markers (BM3517, BM719, SPS113,
253 INRA006, RAO4116 and HN27) from the East, Central and Southern African roan group. Significant linkage
254 disequilibrium (LD) was also observed only in the East, Central and Southern African group. These departures
255 from equilibrium may be because of substructure in this group (see [16], which described three mitochondrial
256 DNA groups within this larger ESU), or because of inbreeding. To further investigate the possible causes of
257 heterozygote deficiency, we estimated the overall inbreeding coefficient per population with positive estimates
258 only being observed in the East, Central and Southern African roan group ($F = 0.102$). In addition, analysis of the
259 overall population relatedness was conducted, as mating among close relatives may cause heterozygote
260 deficiency. As shown in Fig 2, the overall population relatedness was higher in the East, Central and Southern
261 African roan group (average = 74%) compared to the West African animals (average = 39%).

262

263 **Fig 2. Mean relatedness of rest of Africa roan and Western roan. WRA = Western roan population A, WRB**
264 **= Western roan population B, WRC = Western roan population C, SRA = Rest of Africa roan population A,**
265 **rest of Africa population B, HYB = known hybrids, PTH = putative hybrids.**

266

267 Genetic diversity for each population is summarized in Table 5. Overall, the genetic diversity in the Western roan
268 populations is higher compared to populations from the East, Central and Southern African ESU, notwithstanding
269 smaller sample sizes. The mean number of alleles (A) ranged from 4.15 - 6.07 and 4.26 - 5.70, while allelic
270 richness (AR) ranged from 3.17 - 4.18 and 2.97 - 3.17 in the reference West African group, and East, Central
271 and Southern African roan groups respectively. Observed heterozygosity (H_o) in the Western roan group ranged
272 from 0.67 - 0.72 and unbiased heterozygosity (H_z) from 0.65 - 0.71 while H_o in the East, Central and Southern
273 African roan varied from 0.57 - 0.63 and H_z from 0.605 - 0.609. The P_{ID} for the 27 loci was 5.5^{20} , thus the
274 estimated probability of any two individuals by chance alone sharing the same multilocus genotype is 1.8^{19} for
275 the 27 loci combined.

276 **Table 5. Genetic diversity estimates for roan antelope (*Hippotragus equinus*).**

Samples	No. of samples	Mean no. of alleles per locus (A)	Allelic Richness (AR)	Unbiased Heterozygosity (H_z)	Observed Heterozygosity (H_o)	Inbreeding coefficient (F_{IS})
Western roan population A	12	4.926	3.418	0.652	0.673	-0.018
Western roan population B	14	6.074	4.182	0.714	0.709	-0.022
Western roan population C	6	4.148	3.165	0.667	0.719	-0.125
Rest of Africa population A	80	5.704	2.970	0.605	0.570	0.016
Rest of Africa population B	18	4.259	2.834	0.609	0.634	-0.091
Known hybrids	8	4.963	3.425	0.671	0.598	---
Putative hybrids	15	6.296	3.889	0.688	0.692	---

277

Discussion

An increasing number of species experience dramatic declining population numbers globally, with ample evidence suggesting that we are entering a mass extinction event. Although the drivers of these population declines are numerous and varied, the underlying root cause inevitably stems from anthropogenic pressures. Not surprisingly, hybridization and admixture of groups with distinct evolutionary trajectories are increasing, raising concerns about the integrity of a large number of species, especially those that experience disproportionately large human interest. For roan antelope, one of Africa's most spectacular large antelope species, this is certainly the case. Although roan antelope numbers are increasing in South Africa (largely because of protection under private ownership), real concerns exist about their genetic integrity given admixture with West African roan antelope, also for export to neighbouring countries. We discuss our results here, and provide some suggestions for roan antelope conservation in South Africa.

Evidence of hybridization

Using a suite of variable and informative microsatellite markers, we provide unequivocal evidence of hybridization and introgression between roan antelope naturally occurring in South Africa (East, Central and Southern African origin), and those of West African descent (a separate evolutionary significant unit; see [16]). More problematic, the identification of first and second generation backcrosses with q -values close to threshold values strongly suggest that hybrid individuals are viable and fertile; as also suggested from anecdotal evidence from some game farms. Although genetic diversity estimates were moderately higher in the known and putative hybrid individuals, it has previously been reported that F2 hybrids can display reduced fitness as a result of disruption of sets of co-adapted gene complexes by recombination [52,53], thereby weakening the entire gene pool of naturally occurring individuals. Our marker set was able to accurately identify F1 and F2 hybrids, as well as non-admixed individuals at thresholds of $q = 0.20$ and $q = 0.15$. However, the accurate classification of further backcrosses was less accurate at these thresholds (40% to 83%) with backcrossed individuals being incorrectly classified as non-admixed. The use of higher thresholds ($qi = 0.10$ and $qi = 0.05$) did increase the number of individuals correctly classified as backcrosses, however, this also resulted in an increase in the number of non-admixed individuals being incorrectly classified as hybrids. Thus in certain instances, backcrossed and double backcrossed individuals extend beyond the detection power of the current microsatellite marker panel.

The minimum number of markers required to accurately and consistently identify backcrosses is currently being debated. Simulation analysis in the grey wolf (*Canis lupus*) that hybridizes with domestic dogs (*C. lupus familiaris*) indicated that simply increasing the number of microsatellite markers used does not equate to an increase in the number of correctly identified admixed individuals [54]. It may be important to evaluate single nucleotide polymorphisms (SNPs) with high discriminating power to increase the ability to detect backcrossed and double backcrossed individuals, but in all likelihood thousands of SNPs may be required. Notwithstanding, the marker set described here represents the first step in assessing hybridization in roan antelope, and in the identification of hybrid individuals.

Conservation management

As signatories to the Convention on Biological Diversity, South Africa has an obligation to conserve the genetic integrity of its biological diversity. Furthermore, admixture between distinct wildlife subspecies is prohibited under national and provincial legislation. Within South Africa, wildlife can be privately owned. There has been some debate about the legal rights of an owner to act in a certain manner with its property, and whether farming with wildlife should be managed and regulated any differently than, for example, agricultural stock such as cattle. Notwithstanding, current international, national and provincial legislation is clear in prohibiting admixture, irrespective of ownership.

The private ownership of biological diversity has been advantages for a large number of species, and the high commercial value attached to many of these species has undoubtedly aided in their conservation and protection; to the point where a number of species are doing better under private ownership compared with in protected areas or national parks [55]. Roan antelope is a prime example, but others include sable antelope, white and black rhinoceros, and bontebok to name but a few. Unfortunately, many of these species are intensively managed, with selection for specific desired traits. These management practises have unintended consequences, notably a loss of genetic diversity. In our study, a number of loci showed deviations from HWE and linkage disequilibrium; all which can be ascribed to small numbers of founding individuals and genetic drift on farms [56] which may, in the long term, compromise local adaptation [57]. To fully understand the impact that farming practises, notably intensive management and selection, have on wildlife populations, comparisons need to be done with naturally occurring populations on nature reserves.

Currently, the full extent of hybridization in South Africa between roan antelope belonging to the two distinct ESUs is unknown. Laboratory screening for permitting purposes (to either sell, or translocate animals) suggest that the occurrence of widespread introgression is low, and largely confined to specific game farms.

Animals of West African decent are no longer maladapted to South African conditions and have, over the span of 20 years, adapted to local conditions. The question that needs consideration is whether South Africa should safeguard the genetic integrity and genetic variability of both roan ESUs. If historic occurrence is considered, then all West African animals should be removed from South African populations. However, the South African situation has spawned several *ex situ* breeding programmes and agreements and/or animals that could be allowed to be backcrossed to obtain some form of purity, over four or five generations. This might improve genetic variation within the national population, but may not be desirable given that the impact of hybridization on the South African roan full genome is not known. Thus, we recommend the implementation and continuation of strict genetic monitoring for hybridization in roan antelope in South Africa. With the microsatellite marker set described here, and using a threshold of $q_i = 0.15$, it is possible to detect F1 and F2 hybrids prior to translocation, thereby reducing and ultimately eliminating Western roan antelope alleles in the indigenous roan gene pools. In addition, management of roan in South Africa would benefit from a national meta-population

conservation plan to inform translocations and reintroductions and to effectively monitor genetic diversity and further hybridization events.

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Author contribution statement

AMvW, DLD, AK, JPG & BJvV wrote the main manuscript. AMvW, PSM & ASK conducted the laboratory analysis. AMvW, DLD & BJvV conducted genetic analysis of the data. All authors reviewed the manuscript.

Additional information

Conflict of interest

The authors declare that they have no conflict of interest.

Compliance with ethical standards

Ethical approval was obtained from the Animal Research Ethics Committee, University of the Free State, South Africa (UFS-AED2017/0010) and the NZG Research Ethics and Scientific Committee (NZG/RES/P/17/18). Samples were stored in the NZG Biobank and access for research use of the samples was approved under a Section 20 permit from the Department of Agriculture, Forestry and Fisheries, South Africa (S20BB1917).

Data availability

All species specific primers developed here will be submitted to GenBank following acceptance of this manuscript.

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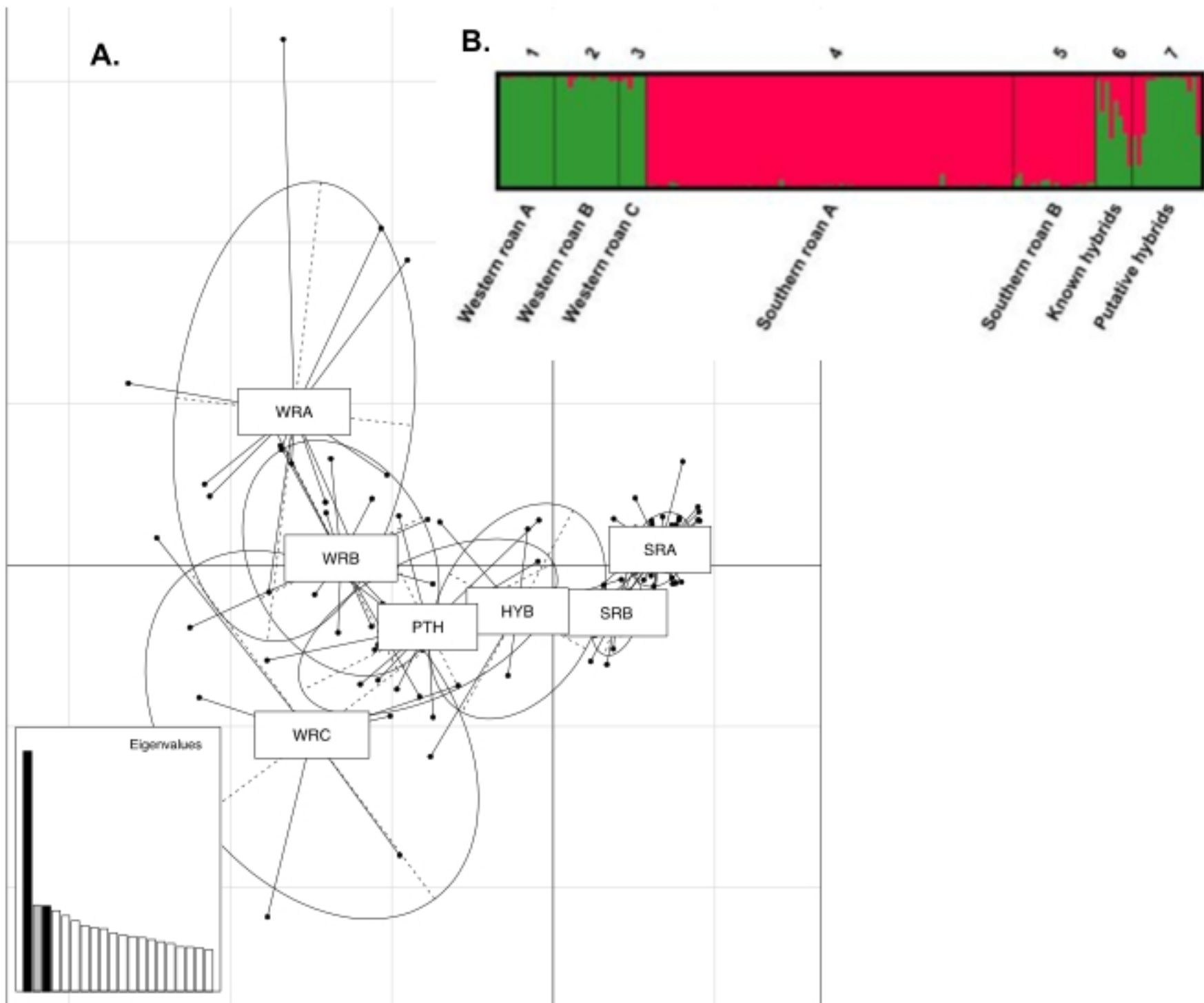
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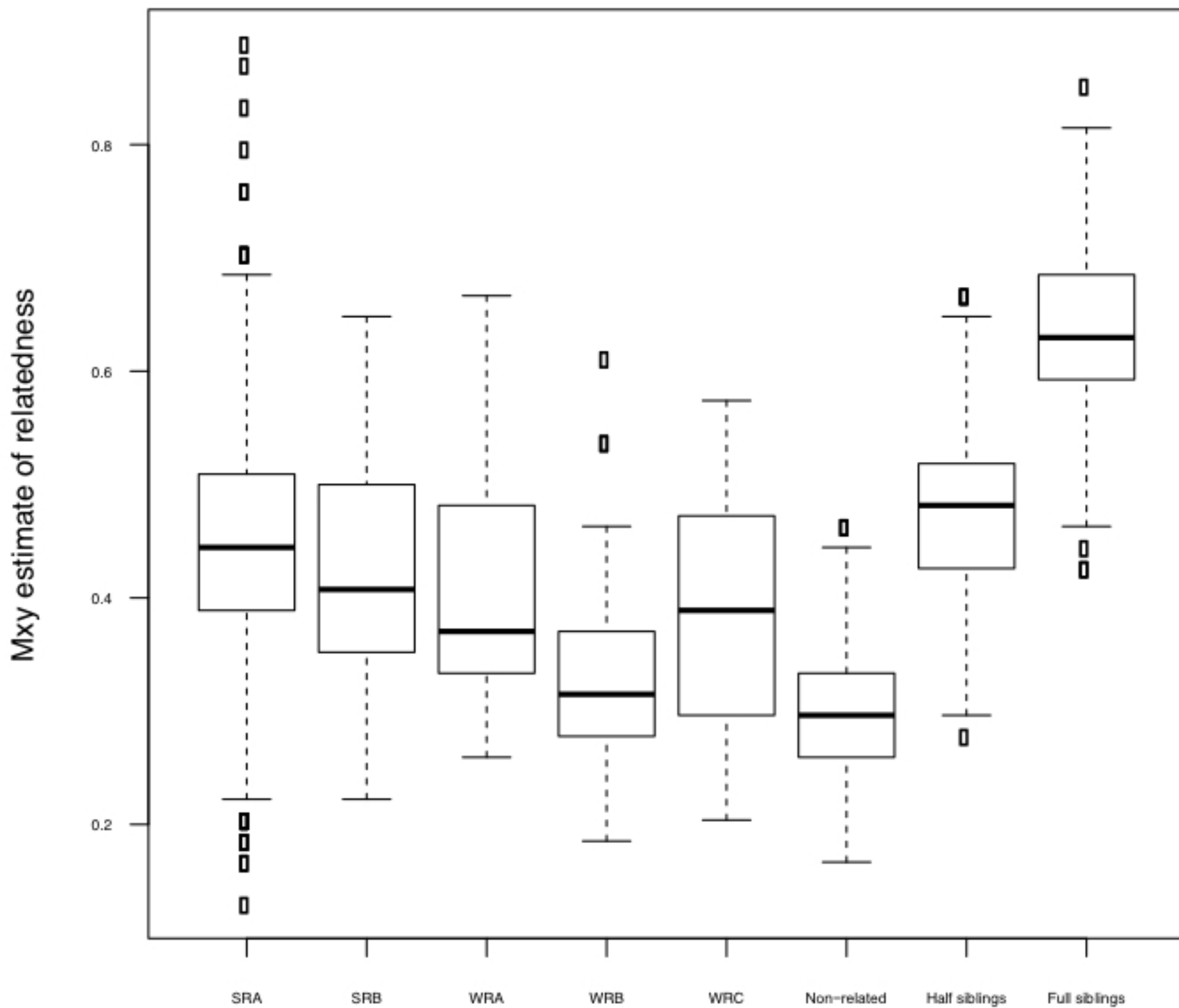
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S1 Fig. a) Probability (-LnPr) of $K = 1 - 6$ averaged over 5 runs. b) Delta K values for real population structure $K = 1 - 6$.



Figure

Mean relatedness of populations



Figure