Population productivity of shovelnose rays: inferring the potential for recovery

Short title: Population productivity of shovelnose rays

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Abstract

Recent evidence of widespread and rapid declines of shovelnose ray populations (Order Rhinopristiformes), driven by a high demand for their fins in Asian markets and the quality of their flesh, raises concern about their risk of over-exploitation and extinction. Using life history theory and incorporating uncertainty into a modified Euler-Lotka model, maximum intrinsic rates of population increase (r_{max}) were estimated for nine species from the four families of rhinopristiforms. Estimates of median r_{max} varied from -0.04 to 0.60 year^{-1} among the nine species, but generally increased with increasing maximum size. In comparison to 115 other species of chondrichthyans for which r_{max} values were available, the families Rhinidae and Glaucostegidae are relatively productive, while most species from Rhinobatidae and Trygonorrhinidae had relatively low r_{max} values. If the demand for their high value products can be addressed, then population recovery for this species is likely possible but will vary depending on the species.

Keywords: Convention on International Trade in Endangered Species of Wild Fauna and Flora, wedgefish, giant guitarfish, data-poor fisheries, natural mortality

Introduction

Understanding the ability of species to recover from declines following implementation of management measures is important to rebuilding depleted populations. This can be approximated through measuring the species' population productivity using various demographic techniques such as rebound potential models [1-3], age or stage structured life history tables and matrix models [4, 5], and demographic invariant methods [6, 7]. These demographic techniques utilise the known relationships between life history traits and demography, known as the Beverton-Holt dimensionless ratios [8] that can be used to infer a species' ability to rebound following population declines [9-11]. One commonly used parameter is the maximum intrinsic rate of population increase r_{max}, which

reflects the theoretical productivity of depleted populations in the absence of density dependent regulation [12]. This method can be used to provide insights into demography and fisheries sustainability [13], particularly for poorly monitored species with limited available information, like many chondrichthyan (sharks, rays, and chimeras, Class Chondrichthyes) populations [14, 15].

An estimated 25% of chondrichthyans populations have an elevated risk of extinction [16], raising significant ecological and conservation concerns [17-19]. Chondrichthyans, generally, have low biological productivity (slow growth, late maturity, few offspring, and long generational times), which limits their recovery from population declines [20, 21]. Declines of chondrichthyan populations are typically the result of the rapid expansion of fisheries [22-24] and the globalisation of trade [25, 26], and can be exacerbated by habitat degredation [27]. Among species, larger elasmobranchs (sharks and rays, Subclass Elasmobranchii) have some of the lowest intrinsic rates of population increase [28, 29], and as a result are unlikely to sustain high levels of fishing pressure before population collapse [30-33].

The order Rhinopristiformes, commonly referred to as shovelnose rays, is considered one of the most threatened orders of chondrichthyans [16]. All five species of sawfishes (Pristidae), five out of six species of giant guitarfishes (Glaucostegidae), seven out of 10 species of wedgefishes (Rhinidae), six of 31 species of guitarfishes (Rhinobatidae) and two of eight species of banjo rays (Trygonorrhinidae) are listed in a Threatened category (i.e. Vulnerable, Endangered, or Critically Endangered) on the International Union for Conservation of Nature's (IUCN) Red List of Threatened Species [34]. These large shovelnose rays are strongly associated with soft-bottom habitats in shallow (<100 m) tropical and temperate coastal waters [35-37], resulting in high exposure to intensive and expanding fisheries, which can also result in habitat degradation [38, 39]. These groups are extremely sensitive to overexploitation as a result of their large body size, slow life history [16] and use of inshore habitat in some of the world's most heavily fished coastal regions [40-42]. There is increasing evidence of extensive and rapid declines in populations of wedgefishes and giant guitarfishes throughout most of their ranges, including Indonesia [43], South Africa [44], Madagascar

[45], Mozambique [46], Tanzania [47] Arabian Seas and surrounding region [34, 48], India [49] and Brazil [50].

The widespread declines of wedgefishes and giant guitarfishes are linked to the international shark fin trade, as their fins are considered amongst the most lucrative shark and ray products [22, 38, 43, 51]. The fins of wedgefishes and giant guitarfishes are considered the highest grade fins and are prevalent in shark fin trading hubs such as Hong Kong [52] and Singapore [53, 54]. Given high demand of the shark fin trade and increasing evidence of population declines, there is considerable concern that wedgefishes and giant guitarfishes are following a similar pattern of global decline as the sawfishes [34]. All five species of sawfish declined rapidly over 30 years throughout their range, driven by unregulated fisheries, the interational fin trade, and delayed scientific attention [55-58]. Yet despite a global conservation strategy [38], restriction of international trade (i.e. listing on CITES Appendix I), and evidence that some species of sawfish have the ability to recover from fishing pressure in the foreseeable future [59], the recovery of the populations is projected to take at least several decades. To avoid the same fate of sawfishes, precautionary management and conservation for shovelnose rays will be vital to maintain populations.

Currently, shovelnose rays fisheries and trade are not regulated through species-specific fishing or trade restrictions. The magnitude of the declines and the subsequent conservation issues have attracted the focus of major international management conventions and agencies, with *Rhynchobatus australiae* and *Rhinobatos rhinobatos* listed on the Appendix II of Convention on the Conservation of Migratory Species of Wild Animals (CMS) in 2017 [60]. In addition, *R. australiae, Rhynchobatus djiddensis, Rhynchobatus laevis,* and *R. rhinobatos* were listed on Annex 1 of the CMS Memorandum of Understanding (MOU) on the Conservation of Migratory Sharks in 2018 [61], and the families Rhinidae and Glaucostegidae have been proposed for listing on the Appendix II of the Convention on International Trade in Endangered Species (CITES) [62]. Given the global concerns for this group of species, and the importance of trade in their high value fins, the use of trade regulations through CITES listings may help achieve positive conservation outcomes [63]. However,

management and conservation efforts can be hampered by the lack of life history (e.g. age, growth and maturity) and demographic information. Despite the high exposure to fisheries, there is limited information available globally on biology (age, growth, maturity, demography), and abundance and distribution of shovelnose rays.

Estimating r_{max} values can provide the demographic basis for evaluating the sustainability of wedgefish and guitarfish fisheries and international trade, and their ability to recover from population declines, while accounting for uncertainty about their life histories [28]. While the maximum intrinsic population rate of population increase has previously been estimated for *Pseudobatos horkelii* and *Pseudobatos productus* as a part of multi-species comparison [15, 64], there has not been a comprehensive analysis on the population productivity for rhinopristiforms.The focal shovelnose ray families studied in this paper were Rhinidae (wedgefishes), Glaucostegidae (giant guitarfishes), Rhinobatidae (guitarfishes) and Trygonorrhinidae (banjo rays), while the family Pristidae (sawfishes) were excluded as they have been previously assessed in detail by. Dulvy *et al.* (58). The aim of this paper was to use life history data and theory, and comparative analysis, to estimate the population productivity for shovelnose rays, particularly within the context of other sharks and rays.

Materials and Methods

Life history data collection

A literature search was conducted for all species from four families of Rhinopristiformes: Rhinidae, Glaucostegidae, Rhinobatidae and Trygonorrhinidae, to provide data for estimation of population productivity. Life history information required for analyses consisted of age at maturity (α_{mat}, range of years), maximum age (α_{max}, in years), range of litter size (in number of female pups), sex ratio, breeding intervals (*i,* years), and von Bertalanffy growth coefficient (k, year^{-1}). Out of the

127 four families, with a total of 57 species, only nine species had enough published life history

128 information to estimate r_{max} (Table 1).

129

130 **Table 1.** The nine species of shovelnose rays included in this study, their threat status according

131 the International Union of Conservation of Nature (IUCN)'s Red List of Threatened Species, and

132 whether the species are listed on the appendixes of Convention of International Trade of

133 Endangered Species (CITES) and/or Convention on the Conservation of Migratory Species of Wild

134 Animals (CMS) and the CMS Memorandum of Understanding (MOU). IUCN categories are CR,

135 Critically Endangered; EN, Endangered; VU, Vulnerable; LC, Least Concern; DD, Data Deficient.

Family	Species	IUCN	Year	CITES	CMS	Year
Rhinidae	*Rhynchobatus australiae*	VU	2003	Proposed (2018)	Appendix II	2017
					MOU Annex 1	2018
Glaucostegidae	*Glaucostegus cemiculus*	EN	2007	Proposed (2018)	-	-
	Glaucostegus typus	VU	2006	-	-	-
Rhinobatidae	*Acroteriobatus annulatus*	LC	2006	-	-	-
	Pseudobatos horkelii	CR	2007	-	-	-
	Pseudobatos productus	NT	2014	-	-	-
	Rhinobatos rhinobatos	EN	2007	-	Appendix II	2017
					MOU Annex 1	2018
Trygonorrhinidae	*Zapteryx brevirostris*	VU	2006	-	-	-
	Zapteryx exasperata	DD	2015	-	-	-

136

137 Growth coefficient data for *R. australiae* was reported as *Rhynchobatus spp.* by White *et al.* (65)

138 as results from the species complex including *R. australiae, Rhynchobatus palpebratus* and

139 *Rhynchobatus laevis* along the eastern coast of Australia. Recent taxonomic revision has resolved

140 this species complex in this area, with *R. laevis* primarily found in the Indian Ocean and Indo-West

141 Pacific Ocean [66], and further examination of data associated with specimen examined by White *et*

142 *al.* (65) have demonstrated they were primarily *R. australiae*.

143 For *R. australiae*, *G. typus* and *Z. brevirostris* the age at maturity was back-calculated using:

144 $$Age_x = \frac{\left(ln(L_\infty - TL_x) - ln(L_\infty) - (k * t_0)\right)}{-k}$$

where Age_x is age at time x, TL_x is total length (cm TL) at time x, L_∞ is the asymptotic length (cm TL), t_0 is the length at time zero, and k is the von Bertalanffy growth coefficient. For *R. australiae,* the age at maturity was back-calculated using the von Bertalanffy parameters reported for *Rhynchobatus spp.* by White *et al.* (65) and the size at maturity of 150cm TL from *Rhynchobatus djiddensis* [66]. The age at maturity for *Glaucostegus typus* was estimated using the estimated size at maturity in Last *et al.* (66) and growth coefficient in White *et al.* (65).There is no reported litter size for *G. typus*, thus we assumed to have the same litter size and breeding interval as *Glaucostegus cemiculus* to calculate annual reproductive output. For *R. australiae*, *Acroteriobatus annulatus, Zapteryx exasperata* and *Z. brevirostris,* the breeding interval was assumed to be one year, as there was no information available (Table 2).

Table 2. Life history values and sources used to estimate (r_{max}) for the nine species studied, including the maximum size (Max. size in centimetres total length, cm TL), lower, upper and mean (standard deviation, S.D.) values of the age at maturity (α_{mat}, years), lower and upper values for litter size, breeding interval (i, years), lower and upper annual reproductive output of females (b), lower and upper values for von Bertalanffy growth coefficient (k, year^{-1}), the observed, and lower (T_{lower}) and upper (T_{upper}) and mean (S.D.) values of theoretical maximum age (α_{max}, years).

		Max. size	α_{mat}				Litter size		i	*b*		*k*		α_{max}					
Family	**Species**	(cm TL)	lower	upper	mean	± S.D.	lower	upper	(years)	lower	upper	lower	upper	Observed	T_{lower}	T_{upper}	mean	± S.D.	**References**
Rhinidae	*Rhynchobatus australiae*	300	3.00	6.00	4.50	0.450	7	19	1	3.5	9.5	0.400	0.400	12.0	11.3	11.3	11.51	0.346	[65, 66]
Glaucostegidae	*Glaucostegus cemiculus*	290	2.89	6.50	4.70	0.680	5	24	1	2.5	12	0.200	0.275	14.0	13.9	16.1	14.67	0.50	[66-70]
	Glaucostegus typus	270	6.50	8.00	7.25	0.245	5	24	1	2.5	12	0.150	0.150	19.0	18.1	18.1	18.42	0.16	[65, 66, 71]
Rhinobatidae	*Acroteriobatus annulatus*	140	2.30	2.80	2.55	0.080	2	10	1	1.0	5.0	0.240	0.240	7.00	14.8	14.8	12.23	1.30	[66, 72]
	Pseudobatos horkelii	140	7.00	9.00	8.00	0.300	4	12	1	2.0	6.0	0.194	0.194	28.0	16.3	16.3	22.17	1.86	[66, 73]
	Pseudobatos productus	170	7.00	8.40	7.70	0.200	1	10	1	0.5	5.0	0.016	0.240	33.8	14.8	33.8	33.80	3.50	[66, 74-76]
	Rhinobatos rhinobatos	185	2.20	4.10	3.15	0.350	1	14	1	0.5	7.0	0.134	0.310	18.9	13.1	18.9	18.92	1.00	[66, 77-82]
Trygonorrhinidae	*Zapteryx brevirostris*	66.0	7.71	11.5	9.61	0.700	1	8	1	0.5	4.0	0.110	0.130	10.0	19.1	20.3	16.48	1.55	[66, 83-86]
	Zapteryx exasperata	103	5.41	9.65	7.53	0.800	2	13	1	1.0	6.5	0.144	0.174	22.6	17.1	18.4	19.85	0.80	[66, 87-90]

Estimation of maximum intrinsic population growth rate, r_{max}

Maximum intrinsic rate of population increase was estimated using an unstructured derivation of the Euler-Lotka model. This model accounts for juvenile survivorship that depends on age at maturity and species-specific natural mortality, and incorporates uncertainty within the parameters through Monte Carlo simulation [64, 91]. Requirements of this model are estimates of three biological parameters: annual reproductive output, age at maturity, and natural mortality. This model is founded on the principle that a breeding female only has to produce one mature female in her lifetime to ensure a stable population [92-95]:

$$l_{\alpha_{mat}} b = e^{r_{max}\alpha_{mat}} - e^{-M}(e^{r_{max}})^{\alpha_{mat}-1}$$

where $l_{\alpha_{mat}}$ is survival to maturity in the absence of fishing and is calculated as $l_{\alpha_{mat}} = (e^{-M})^{\alpha_{mat}}$, b is the annual rate of production of females, α_{mat} is the age of maturity and M is instantaneous natural mortality. The annual reproductive output of females was calculated as $b = 0.5l/i$, where l is litter size (in number of males and females) and i is breeding interval (in years). Annual reproductive output estimates were derived from uniform distributions constrained by the minimum and maximum litter sizes published in the literature (Table 2). If the litter sex ratio was unknown, it was assumed to be 1:1. Age at maturity estimates were derived from normal distributions with means and standard deviations (S.D.) calculated from the available ages at maturity published in the literature for each species (Table 2). Normal distributions were truncated to be positive and within "reasonable bounds". The von Bertalanffy growth coefficients (*k)* for each species were derived from uniform distributions ranging between the minimum and maximum published values (Table 2). As the observed maximum age may not reflect the longevity of the species [96], the theoretical maximum age (T_{max}) was calculating using minimum and maximum *k* reported for each species in the literature, using the following the formula from Mollet *et al.* (97):

$$T_{max} = 7 \times ln(2/k)$$

Maximum age (α_{max}) estimates were derived from a normal distribution using the mean and S.D., calculated from the observed maximum age reported in the literature, minimum theoretical

maximum age (T_{lower}) and maximum theoretical age (T_{upper}). As there was no current consensus on the best indirect method to estimate the instantaneous natural mortality, it was estimated using four common methods, Jensen's First mortality estimate [98], modified Hewitt and Hoeing estimator [99], Frisk's estimator [9], and reciprocal of the lifespan [10] (Table 3).

Table 3. Natural mortality (*M*) methods used to estimate (r_{max}) in this study, where α_{mat} is age at maturity in years, α_{max} is maximum age in years, and *k* is the von Bertalanffy growth coefficient in year^{-1}.

Method	Equation	References
1. Jensen's First Estimator	$M = 1.65/\alpha_{mat}$	Jensen (98)
2. Modified Hewitt & Hoeing Estimator	$M = 4.22/\alpha_{max}$	Hewitt and Hoenig (99)
3. Frisk's Estimator	$M = 0.4/k$	Frisk *et al.* (9)
4. Reciprocal of lifespan	$M = 1/(\alpha_{mat} + \alpha_{max}/2)$	Pardo *et al.* (64)

The annual reproductive output and age at maturity were highly uncertain parameters, while the natural mortality was estimated indirectly, which can result in additional uncertainty [28]. Monte Carlo simulation was used to account for uncertainty of input parameters. Model parameters were drawn from their respective distributions iteratively 20,000 times [14]. To incorporate uncertainty into *M,* for each iteration the values for age at maturity, maximum age and von Bertalanffy growth coefficients were drawn from their respective distributions and used to estimate natural mortality for the four natural mortality estimators, which in turn is required to estimate r_{max} [14]. Scenarios were investigated where uncertainty was only incorporated into a single parameter. Values of one parameter were drawn from its distribution, while the remaining parameters were set as deterministic by using the median values of their respective distributions. This was done for the age at maturity, annual reproductive output and natural mortality. The *M* value was set as a deterministic in the other scenarios, even when the parameters used to estimate *M* were being drawn from distributions. In each iteration, the r_{max} equation was solved using the *nlminb* optimisation function by minimising the sum of squared differences. This range of r_{max} values was

generated to encompass the widest range of plausible life histories and should therefore include the true parameter values. This method was applied to each of the nine species for which data were available and from this range, median and mean r_{max} values and standard deviation were calculated.

Comparison of shovelnose ray r_{max} estimates among chondrichthyans

Median r_{max} of the nine shovelnose ray species were compared to all available estimates using recalculated values by Pardo *et al.* (64) to incorporate survival to maturity. The reciprocal of the lifespan natural mortality method was chosen to estimate the natural morality in order to compare to values generated by Pardo *et al.* (64) as that was the method used in their study. The age at maturity (years), maximum age (years), growth rate (k, years^{-2}) and maximum size in centimetres (cm) were plotted against the r_{max} estimates for 115 chondrichthyan species, including the nine species of shovelnose rays. Maximum sizes were total length (TL) for all species except for Myliobatiformes, where the disc width (DW) were used [15, 28]. All models and figures were built in the R version 3.4.1 version [100].

Results

Estimates of maximum intrinsic rate of population increase for the nine species of shovelnose ray varied considerably among species, between families and by the method of estimating natural mortality. There was a high level of uncertainty in the annual reproductive output and age at maturity across all species (Fig 1). Uncertainty in the natural mortality values were low (Fig 1), but it resulted in high uncertainty in the r_{max} estimates, which was highly influenced by the natural mortality estimator (Fig 2; Table 4).

Table 4. Maximum intrinsic rates of population increase estimates (r_{max}, year^{-1}) for nine species of shovelnose ray, using four estimators of natural mortality. The minimum (Min.), median, mean (± S.E.), and maximum (Max.) r_{max} values are reported for each species and natural mortality estimator.

		Jensen's First estimator					**Hewitt & Hoeing's estimator**					**Frisk's estimator**					**Reciprocal of lifespan**				
Family	Species	Min.	Median	Mean	±S.E.	Max.	Min.	Median	Mean	±S.E.	Max.	Min.	Median	Mean	±S.E.	Max.	Min.	Median	Mean	±S.E.	Max.
Rhinidae	*Rhynchobatus australiae*	0.12	0.23	0.22	0.0004	0.35	0.02	0.22	0.22	0.0005	0.59	0.24	0.43	0.43	0.0005	0.79	0.29	0.47	0.47	0.0005	0.81
Glaucostegidae	*Glaucostegus cemiculus*	0.02	0.24	0.23	0.0005	0.48	0.04	0.30	0.30	0.0005	0.86	0.23	0.49	0.49	0.0007	1.00	0.24	0.48	0.49	0.0007	1.00
	Glaucostegus typus	0.08	0.19	0.18	0.0003	0.26	0.06	0.19	0.18	0.0003	0.28	0.23	0.36	0.35	0.0003	0.46	0.21	0.34	0.33	0.0003	0.43
Rhinobatidae	*Acroteriobatus annulatus*	-0.25	0.05	0.03	0.0008	0.20	-0.14	0.30	0.28	0.0008	0.54	0.30	0.59	0.57	0.0008	0.80	0.22	0.54	0.52	0.0008	0.75
	Pseudobatos horkelii	0.06	0.13	0.12	0.0002	0.17	0.00	0.14	0.14	0.0002	0.25	0.17	0.25	0.25	0.0002	0.34	0.18	0.26	0.26	0.0002	0.34
	Pseudobatos productus	-0.06	0.09	0.08	0.0004	0.15	-0.06	0.13	0.12	0.0004	0.23	0.09	0.26	0.25	0.0004	0.37	0.09	0.24	0.23	0.0004	0.33
	Rhinobatos rhinobatos	-0.31	0.13	0.10	0.0010	0.32	-0.04	0.38	0.35	0.0011	0.93	0.12	0.55	0.53	0.0011	1.00	0.13	0.54	0.51	0.0011	1.00
Trygonorrhinidae	*Zapteryx brevirostris*	-0.05	0.07	0.06	0.0003	0.13	-0.08	-0.04	0.04	0.0003	0.10	0.08	0.19	0.19	0.0003	0.31	0.04	0.16	0.16	0.0003	0.27
	Zapteryx exasperata	-0.05	0.12	0.11	0.0004	0.21	-0.03	0.12	0.12	0.0004	0.34	0.12	0.27	0.27	0.0004	0.49	0.12	0.27	0.26	0.0004	0.47

235 **Figure 1.** Predicted values of maximum intrinsic rate of population increase ($r_{max,}$ year^{-1}) for nine
236 shovelnose ray species when including uncertainty in the reciprocal of the lifespan natural mortality
237 estimator (*M*, top/grey boxplot), annual reproductive output (*b,* middle/blue boxplot), and age at
238 maturity (α_{mat}, bottom/orange boxplot). Specie are (A) *Rhynchobatus australiae,* (B), *Glaucostegus*
239 *cemiculus,* (C) *Glaucostegus typus,* (D) *Acroteriobatus annulatus,* (E) *Pseudobatos horkelii*, (F)
240 *Pseudobatos productus,* (G) *Rhinobatos rhinobatos,* (H) *Zapteryx brevirostris*, and (I) *Zapteryx*
241 *exasperata*. The natural mortality estimator used for this graph is the. Boxes indicate median, 25 and
242 75% quantiles, whereas the lines encompass 95% of the values (2.5 and 97.5% quantiles). For plots
243 incorporating uncertainty with other three natural mortality methods, see Supplementary Material.
244
245 The ranges of r_{max} for each species were relatively large as a result of the high uncertainty in the
246 life history parameters and method of estimating natural mortality (Fig 2). *Acroteriobatus annulatus*
247 and *R. rhinobatos* had the largest range of r_{max}, regardless of the natural mortality estimation
248 method used (Fig 2; Table 4). *Pseudobatos horkelii* and *P. productus* had the smallest range of r_{max}
249 (Fig 2; Table 4). For all species, the highest r_{max} values were estimated using the maximum age
250 natural mortality method, excluding *P. productus* and *Z. brevirostris,* which had highest r_{max} values
251 estimated from Frisk's Estimator (Table 4). Frisk's estimator, Maximum Age and Lifespan methods
252 produced similar r_{max} estimates for each species, with 7% or less difference between median values
253 (Table 4, Fig 2).
254
255 **Figure 2**. The range of maximum intrinsic rate of population increase (r_{max}, year^{-1}) for nine species of
256 shovelnose rays, obtained with five different methods of estimating the instantaneous natural
257 mortality: Jensen's First Estimator (red), modified Hoeing & Hewitt's Estimator (yellow), Frisk's
258 Estimator (green), and Reciprocal of lifespan (blue). Means (triangle) and standard deviation (black
259 line) are presented for each method. Species are (A) *Rhynchobatus australiae,* (B), *Glaucostegus*
260 *cemiculus,* (C) *Glaucostegus typus,* (D) *Acroteriobatus annulatus,* (E) *Pseudobatos horkelii*, (F)

261 *Pseudobatos productus,* (G) *Rhinobatos rhinobatos,* (H) *Zapteryx brevirostris*, and (I) *Zapteryx*
262 *exasperata*. Values below the black dashed line indicate implausible r_{max} estimates.
263
264 *Acroteriobatus annulatus* and *R. rhinobatos* had the highest estimates of r_{max}, followed by *G.*
265 *cemiculus* and *R. australiae*. *Pseudobatos horkelii, P. productus* and *Z. exasperata* had lower r_{max}
266 estimates with similar ranges across all natural mortality estimators (Table 4). The lowest r_{max} values
267 from every species were generated using the Jensen's First estimator and modified Hewitt and
268 Hoeing's methods. These methods estimated negative minimum r_{max} values for *A. annulatus, P.*
269 *productus, R. rhinobatus, Z. brevirostris* and *Z. exasperata, a*nd a negative median r_{max} for *Z.*
270 *brevirostris* (Table 4; Fig 2). *Zapteryx brevirostris*, the smallest species in the study, had one of the
271 lowest estimates of r_{max}, across of natural mortality methods (Table 4).
272 As the age at maturity decreased, the estimates of maximum intrinsic rate of population
273 increased for the nine species of shovelnose rays (Fig 3A). The species with the highest median
274 estimates of r_{max}, *R. australiae*, *G. cemiculus*, *R. rhinobatos* and *A. annulatus* had the youngest age at
275 maturity, while *Z. brevirostris* had the oldest age at maturity and lowest median estimate for r_{max} (Fig
276 3A). The estimates of r_{max} increased with the increasing the number of female offspring produced
277 annually (Fig 3B). *Rhynchobatus australiae* and *G. cemiculus* had the highest annual reproductive
278 output and r_{max}, while *G. typus* had lower r_{max} estimates but the same annual reproductive output as
279 the two species (Fig 3B). *Rhinobatos rhinobatos, P. horkelii* and *Z. exasperata* had similar estimates
280 of annual reproduction, yet *R. rhinobatos* had a higher estimates of r_{max} than *P. horkelii* and *Z.*
281 *exasperata* (Fig 3B). *Z.* brevirostris had the lowest annual reproductive output and r_{max} estimate (Fig
282 3B). Maximum rate of population growth increased with maximum size of the species (Fig 4A). The
283 largest species (i.e. *R. australiae, G. cemiculus* and *G. typus*) were estimated to have a higher
284 maximum rate of population increase than the smaller species in the order, such as *P. horkelii* and *Z.*
285 *brevirostris* (Table 4; Fig 4A). The largest species, *R. australiae, G. cemiculus* and *G. typus*, had the
286 highest mean annual reproductive output and the largest mean size at birth in relation to the

287 maximum size, compared to the six smaller species (Fig 4B, C). The smallest species, *Z. exasperata*
288 and *Z. brevirostris* had the smallest annual reproductive output and size at birth, in relation to their
289 maximum size (Fig 4B, C).
290
291 **Figure 3.** Maximum intrinsic rate of population increase (r_{max}) for the nine species of shovelnose rays
292 in relation to the (A) mean ± S.E. age at maturity (a_{mat}, years) and (B) mean ± S.E. annual
293 reproductive output (*b*). The shapes represent the four families, black circle represents the Family
294 Glaucostegidae (giant guitarfishes), black triangle signifies the Rhinidae (wedgefishes), black squares
295 represents species from Rhinobatidae (guitarfishes) and black crosses are the species from
296 Trygonorrhinidae (banjo rays).
297
298 **Figure 4.** Maximum size (cm TL) for the nine species of shovelnose rays in relation to the (A) median
299 maximum intrinsic rate of population increase ($r_{max,}$ year^{-1}) using the reciprocal of lifespan to
300 estimate natural mortality, (B) mean (± S.E.) annual reproduction rate of females (*b*), and (C) mean
301 (± S.E.) size at birth (cm TL). The shapes represent the four families, black circle represents the
302 Family Glaucostegidae (giant guitarfishes), black triangle signifies the Rhinidae (wedgefishes), black
303 squares represents species from Rhinobatidae (guitarfishes) and black crosses are the species from
304 Trygonorrhinidae (banjo rays).
305
306 For the 115 chondrichthyan species, maximum intrinsic rate of population increase ranged from
307 0.035 year^{-1} for the gulper shark *Centrophorus granulosus* (Family Centrophoridae) to 1.395 year^{-1}
308 for the brown skate *Raja miraletus* (Family Rajidae) (Fig 5). The maximum intrinsic rate of population
309 increase of shovelnose rays was similar to that of the four sawfish species (Family Pristidae), *Pristis*
310 *pristis* (0.172 year^{-1}), *Pristis pectinate* (0.179 year^{-1}), *Pristis clavata* (0.236 year^{-1}) and *Pristis zijsron*
311 (0.272 year^{-1}). Estimates of r_{max} for *Acroteriobatus annulatus, R. rhinobatos, G. cemiculus* and *R.*
312 *australiae* were among the highest among chondrichthyans (median = 0.522, 0.539, 0.486 and 0.467

313 year^{-1}, respectively). *Rhynchobatus australiae*, *G. cemiculus* and *G. typus* had relatively high r_{max}

314 estimates, compared to species with similar maximum sizes (Fig 6A). *Pseudobatos horkelii, P.*

315 *productus* and *Z. exasperata* had mid-range estimates of r_{max} compared to species of a similar size

316 (Fig 6A). Compared to other species of the same maximum size, *A. annulatus* and *R. rhinobatos* had

317 relatively high r_{max}, while *Z. brevirostris* had a lower r_{max} compared to similar sized species (Fig 6A).

318 The majority of the largest chondrichthyan species for which r_{max} are all listed on CITES and CMS,

319 however they are not the least productive species (Fig 6A).

320 Chondrichthyans species with the oldest age at maturity tend to have the lowest r_{max} estimates

321 (Fig 6B). *Acroteriobatus annulatus*, *G.* cemiculus and *R. australiae* mature at the youngest ages and

322 had higher estimates of r_{max}, compared to the other rhinopristiformes and chondrichthyans (Fig 6B).

323 As maximum age increased, r_{max} estimates decreased for the 115 chondrichthyans species (Fig 6C).

324 *Acroteriobatus annulatus*, *R. rhinobatos, G.* cemiculus and *R. australiae* are among the

325 chondrichthyans species with the lowest maximum age estimates, and hence high r_{max} (Fig 6C).

326 *Glaucostegus typus, Z. exasperata, P. horkelii* and *P. productus* have mid-range maximum age

327 compared to other species, while *Z. brevirostris* had a lower r_{max} estimate compared to other species

328 with a similar maximum age (Fig 6C).

329 Across chondrichthyan species, as the number of female offspring produced annually increases,

330 r_{max} also increases (Fig 6D). However, *A. annulatus*, *R. rhinobatos, G.* cemiculus and *R. australiae*

331 have a relatively higher r_{max} estimates compared to species with similar annual reproductive output.

332 *Zapteryx exasperata, P. horkelii* and *P. productus* are estimated to have a mid-range annual

333 reproductive estimate, compared to species with similar r_{max} (Fig 6D). *Glaucostegus typus* has a

334 relatively high r_{max} estimate compared to species with similar annual reproductive output, while *Z.*

335 *brevirostris* has a low r_{max} estimate compared to species with similar annual reproductive output (Fig

336 6D). The majority of currently CITES listed species have a low annual reproductive output and r_{max}

337 (Fig 6D).

There is a positive correlation between growth coefficient and maximum intrinsic rate of population increase (Fig 6E). *Acroteriobatus annulatus*, *R. rhinobatos, G.* cemiculus and *R. australiae* have fast somatic growth and a high r_{max}, in comparison to the other chondrichthyan species (Fig 6E). *Glaucostegus typus, Z. exasperata* and *P. horkelii* have a mid-range r_{max} estimate compared to species with similar growth rates, while *P. productus* and *Z. brevirostris* have a lower r_{max} estimate compared to other species with similar growth rates (Fig 6E). The majority of the CITES and CMS listed species have a relatively low growth coefficients and r_{max} estimates (Fig 6E).

Figure 5. Median maximum intrinsic rate of population increase (r_{max}) for 115 chondrichthyans, including the nine shovelnose ray species, which are displayed on the figure and grouped by their r_{max} values. Blank line denote the mean (r_{max} = 0.30) and blue line represents the median (r_{max} = 0.23). Species illustrations are from [66]

Figure 6. Maximum intrinsic rate of population increase (r_{max}) estimates for 115 chondrichthyans, including the nine shovelnose ray species compared with (A) maximum size (cm TL/DW/FL), (B), age at maturity (α_{mat} years), (C) maximum age (α_{max}, years), (D) annual reproductive output *b*, (E) the von Bertanlaffy growth coefficient (*k*, year^{-1}). All axes are on a logarithmic scale. The nine shovelnose ray species are labelled on the graph, and the median r_{max} value is reported, using the reciprocal of the lifespan method to estimate natural mortality. Species that are listed on CITES appendix I or II represented in blue, non-CITES species in grey, and species listed on CMS appendix I or II represented as triangle, with non-CMS species are closed circle.

Discussion

Our analysis show that the larger species of shovelnose rays, such as the wedgefishes and giant guitarfishes appear to have a greater intrinsic ability to recover from population depletion than

363 many other large chondrichthyans including those currently list on multilateral environmental
364 agreements like CITES and CMS, such as manta and devil rays [28, 29]. Maximum rate of intrinsic
365 population growth increased with increasing body size for nine species of rhinopristiformes. The
366 largest species, *R. australiae*, *G. cemiculus* and *G. typus* had among the highest estimates of
367 population increase for shovelnose rays, while the smaller banjo rays (family Trygonorrhinidae) such
368 as the *Z. brevirostris* and *Z. exasperata* had the lowest estimates. The two species of guitarfish
369 (family Rhinobatidae), *A. annulatus* and *R. rhinobatos*, did not fall within the positive relationship
370 between maximum sizes and r_{max} seen in this group. *Acroteriobatus annulatus* and *R. rhinobatos* only
371 attaining 140 cm TL and 100 cm TL, respectively [66], and had the highest estimates of r_{max}. The high
372 r_{max} estimates for *A. annulatus* and *R. rhinobatos* could be explained by the young age at maturity
373 (1.9 – 4.5 years), fast somatic growth (0.134 – 0.289 *k* years^{-1}), and high annual reproductive output
374 (2.5–12 females pups per year). The high maximum intrinsic rate of population growth for larger
375 shovelnose rays, such as wedgefishes and giant guitarfishes, which both reach maximum sizes close
376 to 300cm TL [66], is the result of the higher annual reproductive outputs, combined with early at age
377 maturity, compared to other species of a similar maximum size. While the banjo rays, *Z. brevirostris*
378 and *Z. exasperata* which reach a maximum size of 66 cm TL and 97cm TL, respectively, had slower
379 growth, later maturity compared to other similar sized chondrichthyans species. However, this trend
380 is highly unusual among elasmobranchs [101]. These findings contrasts other multi-species
381 comparative studies, such as Dulvy *et al.* (28), where the maximum intrinsic rate tends to decrease
382 with increasing maximum size, however this study had more sharks than large tropical rays. The
383 productivity of shovelnose rays was more similar to four sawfish species: largetooth sawfish *Pristis*
384 *pristis* (0.172 year^{-1}), smalltooth sawfish *Pristis pectinata* (0.179 year^{-1}), dwarf sawfish *Pristis clavata*
385 (0.236 year^{-1}) and green sawfish *Pristis zijsron* (0.272 year^{-1}). Despite their large size (ranging from
386 318 – 700 cm TL), some species of sawfish have been estimated to have a relatively high productivity
387 for elasmobranchs [59]. Typically large bodied marine animals are associated with factors of
388 vulnerability such as lower intrinsic rate of population growth, late maturity, and dependence on

389 vulnerable habitat, while smaller bodied species are linked to factors providing resilience including
390 faster population growth and early maturity [14, 27, 76]. While body size has been used to predict
391 extinction risk in elasmobranchs, with the larger species predicted to be most at risk of extinction
392 [16], some studies have found little [9, 15] to no correlation [20] between body size and rate of
393 population increase. The relationship between body size and rate of population growth has been
394 hypothesised to be the result of correlations between body size and other more influential life
395 history traits such as age at maturity and litter size [102, 103], which may be the case for the
396 shovelnose rays.

397 Many elasmobranch species have a low reproductive rate [20]. Three species had annual
398 reproductive outputs less than 5 pups per year, *P. productus*, *Z. brevirostris* and *A. annulatus*.
399 Species with reproductive output lower than five pups per year have been demonstrated to have the
400 highest demographic uncertainty when estimating r_{max}, and are likely to have low r_{max} estimates
401 [14], which was observed for *P. productus* and *Z. brevirostris*. However, while *A. annulatus* had very
402 low annual reproductive output (b = 1-5), this species also had one of the highest estimates of r_{max}.
403 The high estimate of r_{max} may be attributed instead to the very young age at maturity (α_{mat} = 2-3
404 years), as well as the fast growth (k = 0.240 year^{-1}) and long lifespan (α_{max} = 15 years). The annual
405 reproductive output was higher in the larger species, such as *R. australiae and G. cemiculus*, which
406 have the largest annual reproductive output parameters (b = 3.5 – 9.5 and b = 2.5 – 12, respectively).
407 This reproductive output is likely to be driving the high r_{max} estimates for these species. While the
408 similar sized *G. typus* had the same annual reproductive output as *G. cemiculus*, it has lower
409 estimate of r_{max}, likely the result of an older age at maturity (α_{mat} = 6.5 – 8 years) than *R. australiae*
410 (α_{mat} = 3 – 6 years) and *G. cemiculus* (α_{mat} = 2.9 – 6.5 years). As r_{max} estimates are sensitive to
411 increasing variation in age at maturity [14], the early maturity of many shovelnose rays, particularly
412 compared to species of similar sizes, may help to explain the relatively high r_{max} estimates for this
413 group. The larger body size of wedgefish and giant guitarfish species (max. size estimated between
414 270-300 cm TL) allows these species to produce numerous and large offspring in relation to their

415 maximum size. Whereas the guitarfishes and banjo rays have smaller birth size and smaller litters
416 relative to their maximum size. Larger offspring will likely have a greater survival probability than the
417 smaller offspring of species with a similar r_{max} [14], and for long lived species, juvenile survival is a
418 key contributor to the population growth rate [9]. While the model used in this study incorporates
419 juvenile survival, it also assumes that juvenile mortality is equal to adult mortality [64]. As juveniles
420 tend to have higher mortality rates than adults [104], which then can vary with local differences in
421 habitat [105], this assumption of equal mortality is likely to result in conservative estimates of *M*
422 [64]. The differential juvenile mortality among species was not accounted for in this model but
423 should be the focus of further study [14].

424 Natural mortality, referring to the death of individuals in the population from natural causes
425 such as predation, disease and old age [94], is one of the most important parameters in fisheries and
426 conservation modelling, yet it is one of the hardest to estimate [10, 106, 107]. While in some models
427 uncertainty in the natural mortality parameter has had little influence on r_{max} [14], the different
428 estimators can have substantial effects on r_{max} estimates [107]. There is considerable debate as to
429 which empirical model should be used to estimate adult natural mortality, as there are numerous
430 and diverse approaches using life history information to estimate this parameter [106, 108]. Jensen's
431 First Estimator [98] and the modified Hewitt and Hoeing method [99] systematically resulted in
432 negative, and thus unrealistic estimates, of r_{max} for five out of the nine species of shark-like ray
433 species, which was also demonstrated in [64], and were considered to be less appropriate for this
434 group. The implausible estimates are likely the consequence of overestimating natural mortality (e.g.
435 > 0.1 year^{-1}) for these species, particularly when the annual reproductive output is low (e.g. b < 5)
436 and age at maturity is high [14, 64]. The other two natural mortality estimators, the Frisk's estimator
437 and Reciprocal of lifespan, provided similar estimates of maximum rate of population increase for *G.*
438 *cemiculus, G. typus, R. rhinobatus, P. horkelii, Z. exasperata* and *Z. brevirostris*. The plausible r_{max}
439 estimates produced by these two natural mortalities suggest these natural mortality estimators may
440 be the more appropriate methods for elasmobranchs. Frisk's estimator and Reciprocal of life span

441 are more suited for elasmobranchs, given they have a relatively high juvenile survival [9, 64], while
442 the Jensen's First Estimator and the modified Hewitt and Hoeing method were explicitly designed for
443 adult mortality. Most teleost species have different life history strategies to chondrichthyan species
444 (i.e. higher fecundity, younger age at maturity, faster growth, and shorter lifespans) [109, 110]. Thus,
445 it is likely that Jensen's First Estimator and the modified Hewitt and Hoeing are less appropriate
446 methods of estimating natural mortality for chondrichthyans. Unfortunately estimating natural
447 mortality with useful accuracy is remarkably difficult [108]. Identifying, or improving the best
448 indirect estimator would require data-intensive methods, such as catch data to analyse catch curves,
449 mark re-capture experiments, virtual population analysis (VPA), or fully integrated stock assessments
450 [108]. These methods all require extensive prior knowledge of the species biology which is lacking
451 for many chondrichthyan species. Presenting the results from multiple natural mortality estimators
452 provides a better understanding of the uncertainty associated with the maximum intrinsic rate of
453 population increase.

454 The greatest obstacle to accurately estimate r_{max} and natural mortality is the accuracy of the
455 biological information used [91]. The use of inaccurate surrogate information can reduce the
456 accuracy of the demographic models [91, 111, 112]. Of the 56 species across the four families of
457 shovelnose rays, only nine species had sufficient information to estimate their maximum intrinsic
458 rate of population increase, and with relatively high levels of uncertainty associated with the life
459 history parameters. For example, there were only age and growth two studies for wedgefishes and
460 giant guitarfishes, one from the eastern coast of Australia for *R. australiae* and *G. typus* [65], and one
461 from Central Mediterranean Sea for *G. cemiculus* [70]. Neither study estimated age at maturity, nor
462 aged individuals at the maximum sizes. Given that the age at maturity is a pivotal parameter when
463 estimating r_{max}, yet highly uncertain for all shovelnose rays examined, these estimates must be taken
464 with caution. Furthermore, numerous reviews have reported sampling biases and failures in ageing
465 protocols, including lack of validation [113, 114] that often result in overestimation or underestimate
466 of age and growth parameters [115]. As there has been no validation studies in the ages of

467 shovelnose rays, the maximum ages for these species are likely to be underestimated, while the age
468 at maturity estimates could also be inaccurate. This can lead to inaccurate estimates of natural
469 mortality and r_{max} [91, 116]. The information on the reproductive biology for rhinopristiforms is
470 generally limited, but is more available for species in the Rhinobatidae and Trygonorrhinidae
471 families. For example, there is evidence that species such as *P. productus, P. horkelii*, and Z.
472 *exasperata* experience embryonic diapause or delayed development [87, 117], potentially as a result
473 of unfavourable environmental conditions [118] or sex segregation [119]. In addition, Simpfendorfer
474 (120) hypothesised that diapause allowed other elasmobranch species (*Rhizoprionodon taylori*) to
475 have larger litter sizes than other similar sized species in the same family (Carcharhinidae). As
476 possibility of diapause was not able to be taken into account during this study, the breeding interval
477 and annual reproductive output may be inaccurate for *Z. exasperata*, thus resulting in an
478 inappropriate maximum intrinsic rate of population growth. Directing research efforts to obtain data
479 from more species, as well as improving the accuracy of life history parameters for data poor
480 species, such as age at maturity and annual reproductive output, would be the most pragmatic
481 option to improve the accuracy of r_{max} for shovelnose rays.
482 Measuring the population productivity of a species allows for a greater understanding of the
483 species' ability to recover from declines, and has direct practical importance in conservation and
484 fisheries [91, 121]. Compared to other chondrichthyans, some families of shovelnose rays, such as
485 wedgefishes, giant guitarfishes and guitarfishes appear to have a high maximum intrinsic rate of
486 population increase, therefore possessing biological traits that can potentially allow them to recover
487 from population declines more rapidly than other threatened species. While other families, such as
488 banjo rays, are less productive and may have a slower ability to rebound from population depletions.
489 However, the current and unregulated fishing pressure that most shovelnose ray species currently
490 experience is likely unsustainable, with reported severe declines and localised extinctions of
491 wedgefishes and guitarfishes throughout their range [34, 48]. Therefore, it appears that these
492 groups are facing a widespread conservation crisis [34]. Yet globally, wedgefishes, guitarfishes, and

banjo rays are currently not managed through international trade or fishing restrictions, and there is limited market and trade information available [122]. Given the lack of national level management for these species, and the role of international trade, especially in fins, the use of trade controls through CITES listings may be a useful way encouraging better management of their shovelnose ray species. Currently, the families Rhinidae and Glaucostegidae are being proposed for CITES Appendix II listing, which are for species that not necessarily currently threatened with extinction but may become so unless trade is controlled [123]. However, trade restrictions alone will not reduce fishing mortality since these species are rarely taken in directed fisheries (although there are exceptions e.g. in Indonesia [43]). Regional and national fisheries management strategies will be required to address the overfishing of stocks, which will require reductions in fishing effort. Lastly, increasing the accuracy of information and research on the biology and life history of these species will be essential for their management and conservation.

Conclusion

Using current life history data, incorporating uncertainty in parameters, and taking into account juvenile mortality, this study provides the first analysis into the population productivity for nine species from four families of rhinopristiforms. The larger wedgefishes and giant guitarfishes were found to be relative productive families, while the smaller guitarfishes and banjo rays were less productive, compared to other chondrichthyans. While the maximum intrinsic rate of population growth varied with the different natural mortality estimator, r_{max} appears to increase with increasing maximum size for the four families, which is counter to most studies of shark populations. There was considerable uncertainty in the age at maturity and annual reproductive output for all species. There is a need for accurate life history information for these data poor species, as well directing efforts to obtain data for more species as there was only nine of out 56 species with sufficient life history information. We recommend presenting the results from multiple natural mortality estimators to

provide a greater understanding of the uncertainty for the maximum intrinsic rate of population increase. It appears that wedgefishes and giant guitarfishes could recover from population depletion faster than guitarfishes and banjo rays, if fishing mortality is kept low. Extensive regional, national and international management strategies will be required to address the overfishing of stock and reduction of fishing mortality, and the use of international trade regulations such as CITES listing, may help to achieve positive conservation outcomes. The results of this study emphasise the urgency for effective management for the conservation of these rays.

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Supporting Information

S1 Figure. Predicted values of maximum intrinsic rate of population increase (r_{max}) for nine shovelnose ray species when including uncertainty in the Jensen's natural mortality estimator (*M*, top/grey boxplot), annual reproductive output (*b,* middle/blue boxplot), and age at maturity (α_{mat}, bottom/orange boxplot). Species are (A) *Rhynchobatus australiae,* (B), *Glaucostegus cemiculus,* (C) *Glaucostegus typus,* (D) *Acroteriobatus annulatus,* (E) *Pseudobatos horkelii*, (F) *Pseudobatos productus,* (G) *Rhinobatos rhinobatos,* (H) *Zapteryx brevirostris*, and (I) *Zapteryx exasperata*. Boxes indicate median, 25 and 75% quantiles, whereas the lines encompass 95% of the values (2.5 and 97.5% quantiles).

S2 Figure. Predicted values of maximum intrinsic rate of population increase (r_{max}) for nine shovelnose ray species when including uncertainty in the modified Howitt & Hewitt's natural mortality estimator (*M*, top/grey boxplot), annual reproductive output (*b,* middle/blue boxplot), and age at maturity (α_{mat}, bottom/orange boxplot). Species are (A) *Rhynchobatus australiae,* (B), *Glaucostegus cemiculus,* (C) *Glaucostegus typus,* (D) *Acroteriobatus annulatus,* (E) *Pseudobatos horkelii*, (F) *Pseudobatos productus,* (G) *Rhinobatos rhinobatos,* (H) *Zapteryx brevirostris*, and (I) *Zapteryx exasperata*. Boxes indicate median, 25 and 75% quantiles, whereas the lines encompass 95% of the values (2.5 and 97.5% quantiles).

S3 Figure. Predicted values of maximum intrinsic rate of population increase (r_{max}) for nine shovelnose ray species when including uncertainty in the Frisk's natural mortality estimator (*M*, top/grey boxplot), annual reproductive output (*b,* middle/blue boxplot), and age at maturity (α_{mat},

bottom/orange boxplot). Species are (A) *Rhynchobatus australiae,* (B), *Glaucostegus cemiculus,* (C) *Glaucostegus typus,* (D) *Acroteriobatus annulatus,* (E) *Pseudobatos horkelii*, (F) *Pseudobatos productus,* (G) *Rhinobatos rhinobatos,* (H) *Zapteryx brevirostris*, and (I) *Zapteryx exasperata*. Boxes indicate median, 25 and 75% quantiles, whereas the lines encompass 95% of the values (2.5 and 97.5% quantiles).

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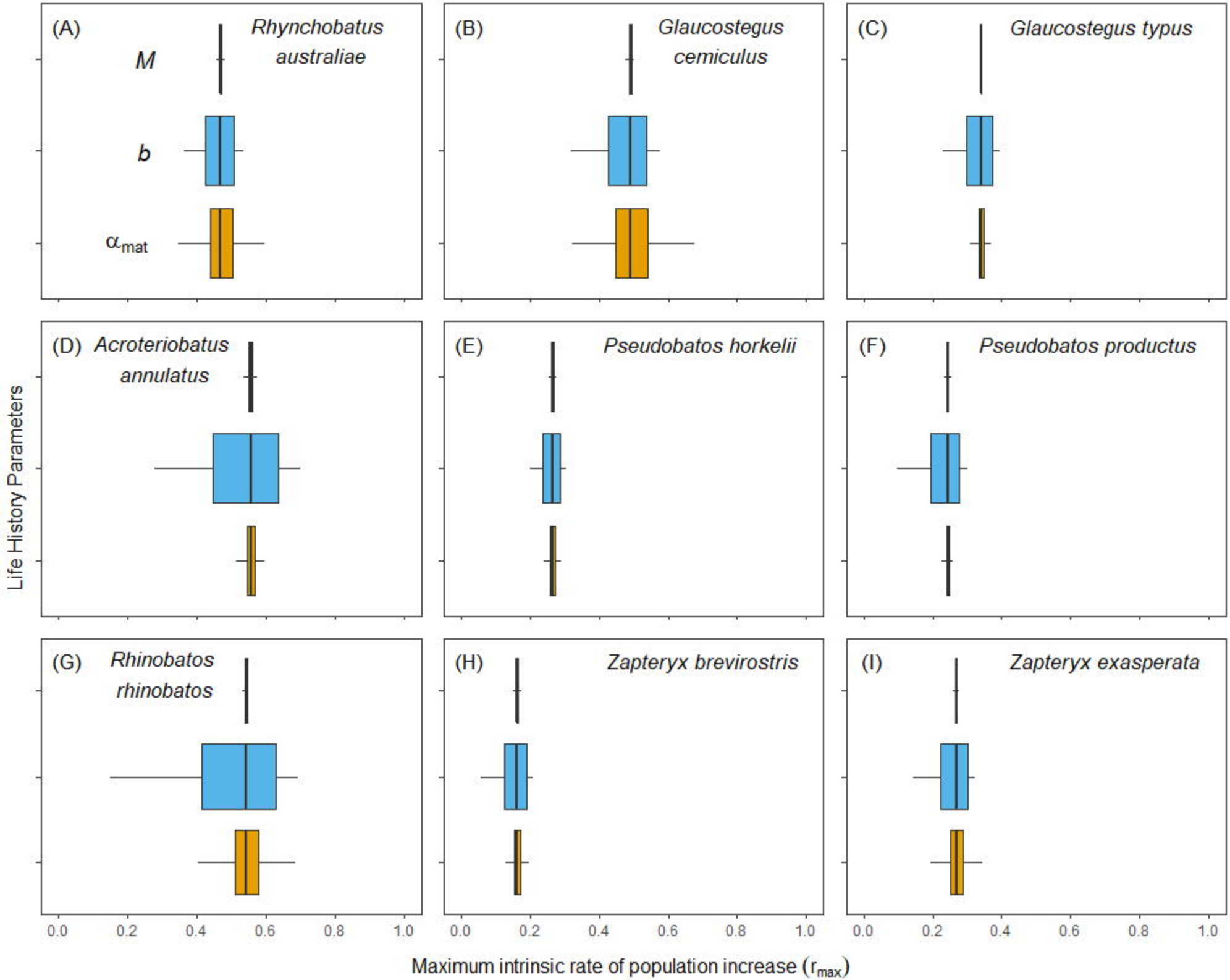
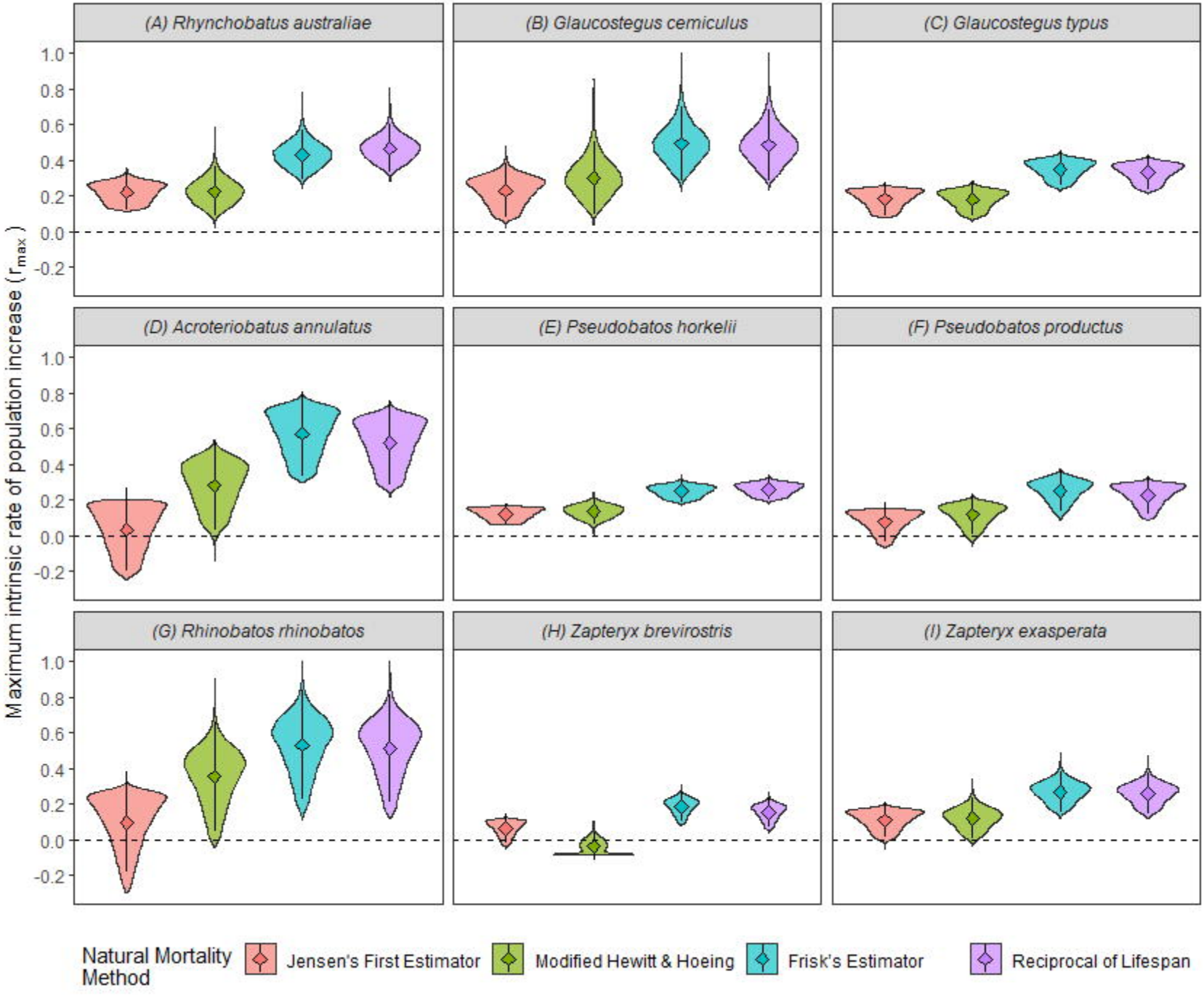
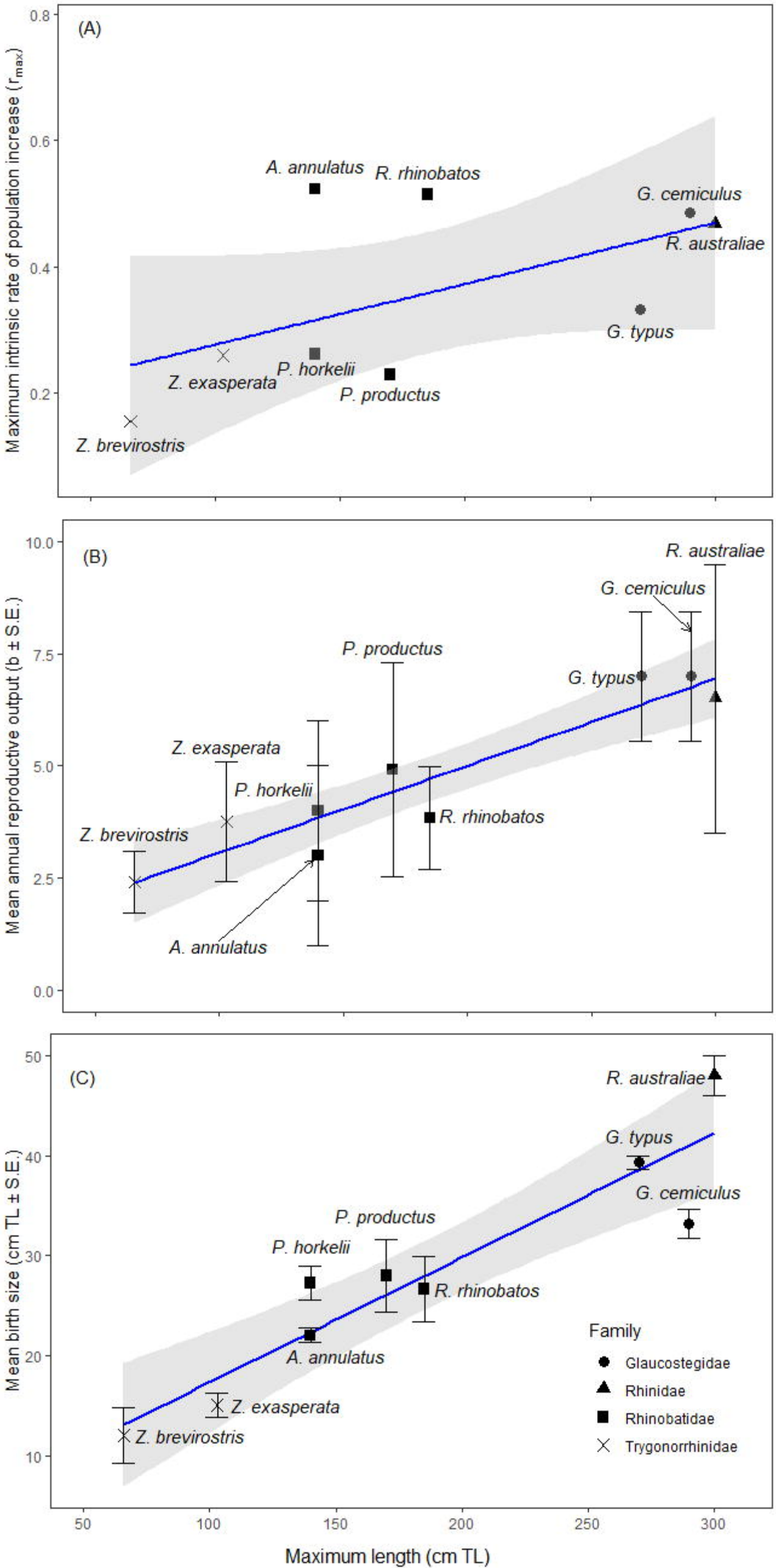
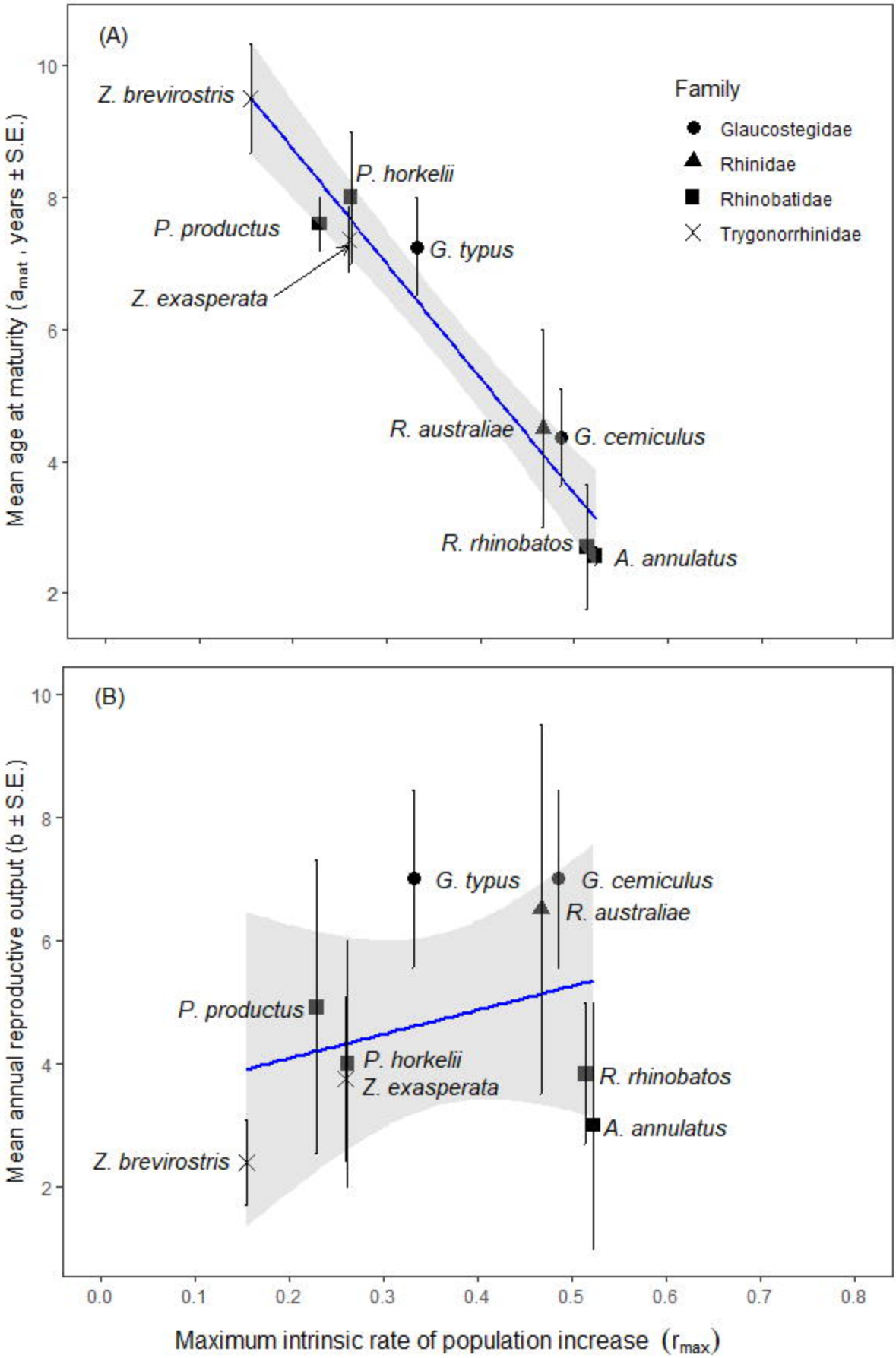
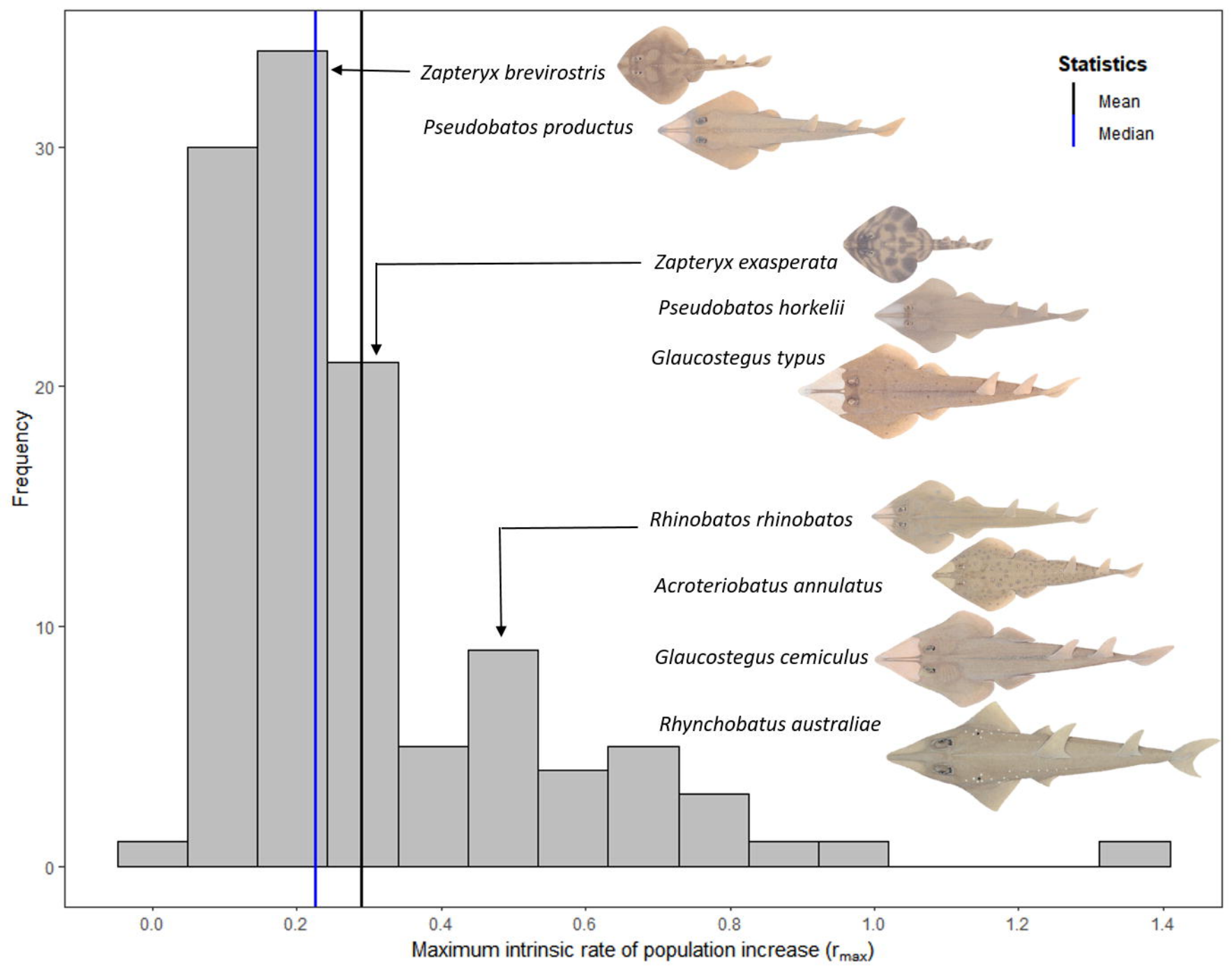
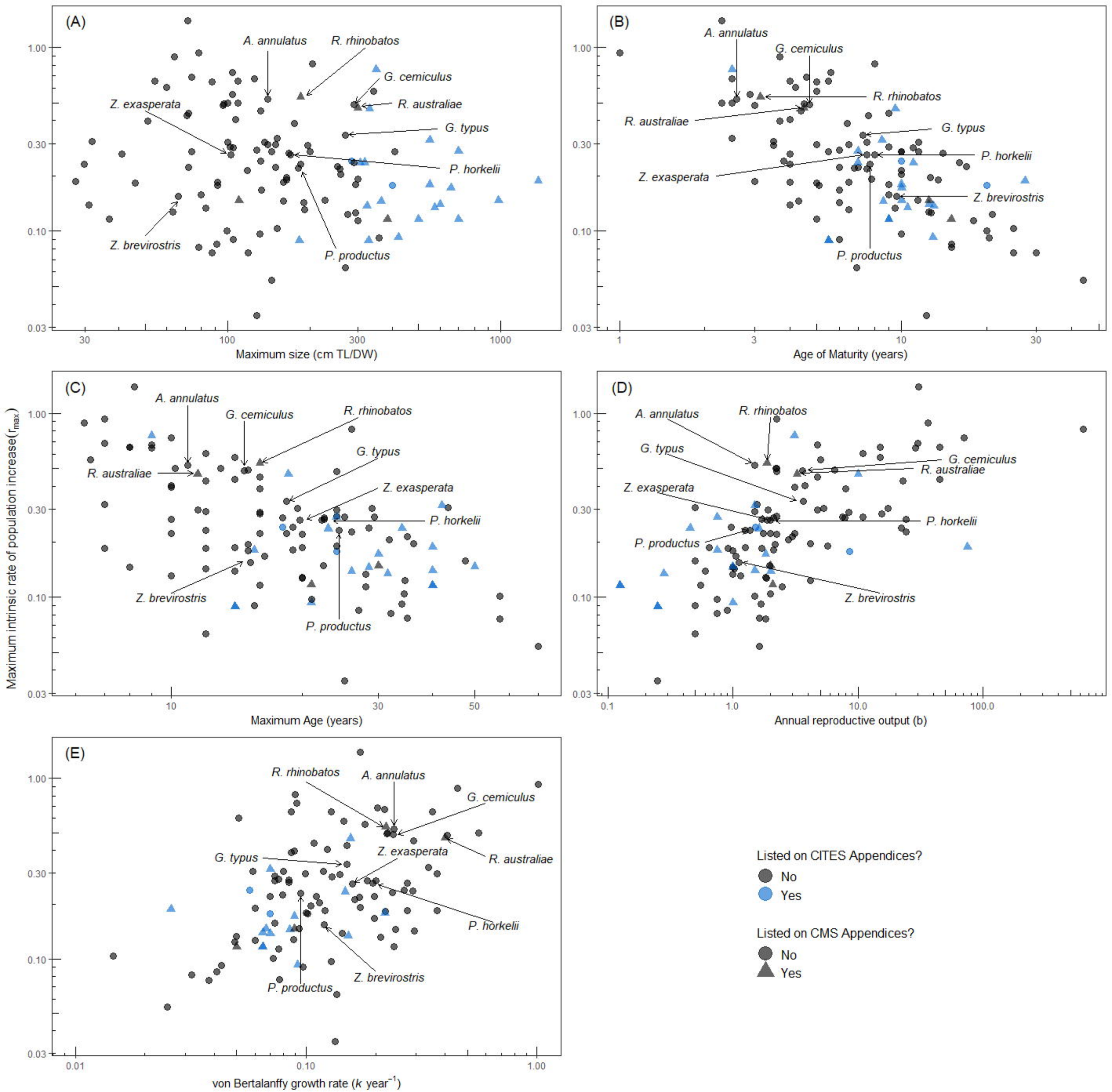

(A)
Rhynchobatus australiae
M
b
α_{mat}
(B)
Glaucostegus cemiculus
(C)
Glaucostegus typus
(D)
Acroteriobatus annulatus
(E)
Pseudobatos horkelii
(F)
Pseudobatos productus
(G)
Rhinobatos rhinobatos
(H)
Zapteryx brevirostris
(I)
Zapteryx exasperata
Life History Parameters
0.0
0.2
0.4
0.6
0.8
1.0
Maximum intrinsic rate of population increase (r_{max})


(A) Rhynchobatus australiae
(B) Glaucostegus cemiculus
(C) Glaucostegus typus
(D) Acroteriobatus annulatus
(E) Pseudobatos horkelii
(F) Pseudobatos productus
(G) Rhinobatos rhinobatos
(H) Zapteryx brevirostris
(I) Zapteryx exasperata
Maximum intrinsic rate of population increase (r_{max})
1.0
0.8
0.6
0.4
0.2
0.0
-0.2
Natural Mortality Method
Jensen's First Estimator
Modified Hewitt & Hoeing
Frisk's Estimator
Reciprocal of Lifespan


(A)
Maximum intrinsic rate of population increase (r_{max})
0.8
0.6
0.4
0.2
A. annulatus
R. rhinobatos
G. cemiculus
R. australiae
G. typus
P. horkelii
Z. exasperata
P. productus
Z. brevirostris
(B)
Mean annual reproductive output (b ± S.E.)
10.0
7.5
5.0
2.5
0.0
R. australiae
G. cemiculus
P. productus
G. typus
Z. exasperata
P. horkelii
R. rhinobatos
Z. brevirostris
A. annulatus
(C)
Mean birth size (cm TL ± S.E.)
50
40
30
20
10
R. australiae
G. typus
G. cemiculus
P. productus
P. horkelii
R. rhinobatos
A. annulatus
Z. exasperata
Z. brevirostris
Family
Glaucostegidae
Rhinidae
Rhinobatidae
Trygonorrhinidae
50
100
150
200
250
300
Maximum length (cm TL)


(A)
Mean age at maturity (a_{mat}, years ± S.E.)
Z. brevirostris
P. horkelii
P. productus
G. typus
Z. exasperata
R. australiae
G. cemiculus
R. rhinobatos
A. annulatus
Family
Glaucostegidae
Rhinidae
Rhinobatidae
Trygonorrhinidae
(B)
Mean annual reproductive output (b ± S.E.)
G. typus
G. cemiculus
R. australiae
P. productus
P. horkelii
Z. exasperata
R. rhinobatos
A. annulatus
Z. brevirostris
Maximum intrinsic rate of population increase (r_{max})


Zapteryx brevirostris
Pseudobatos productus
Zapteryx exasperata
Pseudobatos horkelii
Glaucostegus typus
Rhinobatos rhinobatos
Acroteriobatus annulatus
Glaucostegus cemiculus
Rhynchobatus australiae
Statistics
Mean
Median
Frequency
0
10
20
30
0.0
0.2
0.4
0.6
0.8
1.0
1.2
1.4
Maximum intrinsic rate of population increase (r_{max})


(A)
A. annulatus
R. rhinobatos
G. cemiculus
R. australiae
Z. exasperata
G. typus
P. horkelii
Z. brevirostris
P. productus
Maximum size (cm TL/DW)
(B)
A. annulatus
G. cemiculus
R. rhinobatos
R. australiae
G. typus
P. horkelii
Z. exasperata
Z. brevirostris
P. productus
Age of Maturity (years)
(C)
A. annulatus
G. cemiculus
R. rhinobatos
G. typus
R. australiae
Z. exasperata
P. horkelii
Z. brevirostris
P. productus
Maximum Age (years)
(D)
A. annulatus
R. rhinobatos
G. typus
G. cemiculus
R. australiae
Z. exasperata
P. horkelii
P. productus
Z. brevirostris
Annual reproductive output (b)
(E)
R. rhinobatos
A. annulatus
G. cemiculus
G. typus
Z. exasperata
R. australiae
P. horkelii
Z. brevirostris
P. productus
von Bertalanffy growth rate (k year^{-1})
Maximum intrinsic rate of population increase(r_{max})
Listed on CITES Appendices?
No
Yes
Listed on CMS Appendices?
No
Yes