Stoichiometric constraints modulate the effects of temperature and nutrients on biomass distribution and community stability

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Author contributions

A.S., B.H., and J.M.M. conceived the study. B.H. and A.S. developed and analysed the models. A.S. wrote the first draft of the manuscript. All authors contributed substantially to revisions.

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Legal and ethical requirements

The authors declare that they have no conflict of interest.

Abstract

Temperature and nutrients are two of the most important drivers of global change. Both can modify the elemental composition (i.e. stoichiometry) of primary producers and consumers. Yet their combined effect on the stoichiometry, dynamics, and stability of ecological communities remains largely unexplored. To fill this gap, we extended the Rosenzweig-MacArthur consumer-resource model by including thermal dependencies, nutrient dynamics, and stoichiometric constraints on both the primary producer and the consumer. We found that stoichiometric constraints dampen the paradox of enrichment and increased persistence at high nutrient levels. Nevertheless, they also reduced consumer persistence at extreme temperatures. Finally, we also found that stoichiometric constraints can strongly influence biomass distribution across trophic levels by modulating consumer assimilation efficiency and resource growth rates along the environmental gradients. Our findings highlight the importance of accounting for stoichiometric constraints as they can mediate the temperature and nutrient impact on the dynamics and functioning of ecological communities.

Introduction

Temperature and nutrients regulate many biological processes, including species geographical distribution, primary production, species interactions, and energy and material fluxes (Falkowski *et al.* 1998; Enquist *et al.* 1999; Elser *et al.* 2007; Thomas *et al.* 2017). They are at the core of several ecological theories. While temperature is a fundamental component of metabolic scaling theory (Brown *et al.* 2004), nutrients are at the core of resource competition theory (Tilman 1982) and ecological stoichiometry theory (Sterner & Elser 2002). Nevertheless, less is known about how temperature and stoichiometry (i.e. the element composition of organisms) interact to influence ecological processes across level of organisation (but see Cross et al. 2015, Woods et al. 2003). Cross et al. (2015) reviewed the evidence and suggested that a better understanding of the interactions between temperature and nutrients is crucial for developing realistic predictions about ecological responses to multiple drivers of global change, including climate warming and elevated nutrient supply. While they provided a road map on how to investigate the combined effects of temperature and nutrient on ecological processes, we still lack an integrative theory to understand better how the links between stoichiometry, nutrient enrichment, and temperature influence the dynamics and stability of multispecies communities. Such a theory will allow us to understand how and when stoichiometric variation modulates the consequences of single and combined components of global change on trophic interactions, community dynamics, and ecosystem functioning.

Nutrients can modulate the effects of warming on communities directly by altering primary production, and/or indirectly by changing the elemental composition of primary producers. Conversely, thermal effects on trophic interaction strengths (i.e. the *per capita* effect of predators on prey population densities) and on consumer energetic efficiencies (i.e. ingestion relative to metabolic demand) depend on both the quantity and quality of their resources. Thus, to understand

the combined effects of temperature and nutrients on ecosystems we need a theory that includes

both thermal and stoichiometric dependencies that regulate species interactions and community

dynamics.

Predicting the effects of global warming and nutrient changes on ecosystems is challenging as

species are embedded within communities of multiple interacting species (Petchey *et al.* 1999;

Tylianakis *et al.* 2008; Montoya & Raffaelli 2010; Gilbert *et al.* 2014). Increased resource

availability (hereafter: enrichment) and warming can jointly affect food-web stability and structure

by modifying the strength of trophic interactions (O'Connor *et al.* 2009; Binzer *et al.* 2012; Kratina

et al. 2012; Sentis *et al.* 2014; Binzer *et al.* 2016). Enrichment typically increases energy flux from

resources to higher trophic levels which often leads to the well-known paradox of enrichment

where the amplitude of population fluctuations increase with nutrients, leading to extinctions at

high nutrient concentrations (Rosenzweig 1971; Rip & McCann 2011; Gilbert *et al.* 2014).

Nevertheless, most consumer species become less efficient at processing matter and energy at

warmer temperatures as their metabolic rates often increase faster with temperature than their

feeding rates (Vucic-Pestic *et al.* 2011; Fussmann *et al.* 2014; Iles 2014). This reduction of

energetic efficiency lessens energy flow between trophic levels and hence stabilizes food-web

dynamics by reducing population fluctuations (Rip & McCann 2011; Binzer *et al.* 2012; Gilbert *et*

al. 2014). As a result, mild warming may alleviate the paradox of enrichment by decreasing

consumer energetic efficiency (Binzer *et al.* 2012; Sentis *et al.* 2017).

The theoretical expectations and results described above have already improved our ability to

understand and predict the effects of temperature and enrichment on food webs (Boit *et al.* 2012;

Tabi *et al.* 2019). However, most previous studies using metabolic scaling theory assumed that

nutrient enrichment lead to an increase in resource carrying capacity without influencing resource

elemental composition (Vasseur & McCann 2005; Binzer *et al.* 2012; Gilbert *et al.* 2014; Binzer

et al. 2016; Sentis *et al.* 2017). Yet nutrient enrichment effects are more complex. The elemental

composition of primary producers is likely to be altered, in response to the supplies of energy and

materials relative to their growth and nutrient intake rates (Rastetter *et al.* 1997; Robert W. Sterner

et al. 1997; Finkel *et al.* 2009) Sterner et al. 1997). This, in turn, can affect the dynamics of the

producer population and the herbivores feeding on it. For instance, previous modelling studies

showed that introducing stoichiometric heterogeneity in predator-prey population dynamic models

can dampen the negative effect of nutrient enrichment on system persistence by reducing

population biomass fluctuations (Andersen 1997; Loladze *et al.* 2000; Andersen *et al.* 2004; Elser

et al. 2012). More generally, the stoichiometric flexibility of primary producers, in particular the

flexibility in carbon to nutrient ratios (e.g. C:N or C:P), has important implications for animal

feeding behaviour (White 1993), consumer population stability (White 1993; Sterner & Hessen

1994; Hessen *et al.* 2002), community structure (Andersen 1997), and ecosystem processes such

as biogeochemical cycling (Andersen 1997; Hessen *et al.* 2004).

Previous theoretical and empirical studies reported that stoichiometric variations can have a strong

influence on the stability of consumer-resource interactions (Andersen 1997; Andersen *et al.* 2004;

Diehl *et al.* 2005; Elser *et al.* 2012). For instance, populations of crustacean *Daphnia* feeding on

low quality (i.e. low nutrient: carbon ratio) algae cannot persist even when resource quantity is not

a limiting factor (Elser *et al.* 2007). Consumer extinction is explained by the fact that the consumer

assimilation efficiency is, for most organisms, a function of resource quality (Elser *et al.* 2000).

When resource quality is low, the consumers assimilate only few nutrients relative to the biomass

they ingest, which limits their growth and reproduction (Elser *et al.* 2000; Elser *et al.* 2012).

Temporal variations in resource quality can stabilize the system by weakening interaction strength

and dampening population fluctuations (Andersen *et al.* 2004; Diehl *et al.* 2005)but see(Loladze *et*

al. 2000; Elser *et al.* 2012). However, it remains unclear whether and how temporal variations in

the elemental composition of primary producers and consumers can modulate the effects of

temperature and nutrients on important community features such as stability and biomass

distribution across trophic levels. Previous studies indicated that the spatial and temporal

intraspecific variations in the elemental composition of primary producers are expected to increase

in response to global change drivers such as temperature, CO_2, and nutrient availability (Bezemer

& Jones 1998; Woods *et al.* 2003; Finkel *et al.* 2009). This increased variation can be of importance

for both primary producer and consumer populations as the growth rate of primary producers is

well known to depend on their elemental composition (Droop 1974) as is the assimilation efficiency

of the consumers (Sterner & Elser 2002).

Altogether, previous studies indicated that both temperature and stoichiometric variations can have

important effects on species interactions and community dynamics (Andersen *et al.* 2004; Diehl *et*

al. 2005; Fussmann *et al.* 2014; Binzer *et al.* 2016; Sentis *et al.* 2017). However, the effects of

temperature and nutrient stoichiometry on food web dynamics and stability have only been studied

in isolation. Here we focus on the combined effects of temperature and nutrients on the

stoichiometry of primary producers and consumers and how this affects community stability and

biomass distribution across trophic levels in a consumer-resource system. Understanding the

determinants of stability and biomass distribution has been at the core of ecology for a long time

(Elton (1927), Lindeman (1942)). Recent theory aims at explaining empirical observations of

trophic pyramids (i.e. population biomass decreases with trophic levels), inverted trophic pyramids

(i.e. population biomass increases with trophic levels), trophic cascades and the link between

biomass distribution and stability (McCauley *et al.* 2018; Barbier & Loreau 2019). However, we

lack theory on temperature and stoichiometry as drivers of biomass and stability patterns.

Here, we used the Rosenzweig-MacArthur model as a baseline nonstoichiometric model because

this model is one of the most studied models used to investigate the effects of temperature and

nutrient enrichment on community dynamics (Vasseur & McCann 2005; Binzer *et al.* 2012;

Fussmann *et al.* 2014; Sentis *et al.* 2017). Inspired by previous temperature-independent

stoichiometric consumer-resource models (Andersen 1997; Andersen *et al.* 2004; Diehl *et al.*

2005), we then extended the Rosenzweig-MacArthur model to account for nutrient dynamics as

well as for the simultaneous dependence of community dynamics on temperature and flexible

resource stoichiometry. Our objective here was not to develop a complex and very realistic

stoichiometric model that would include additional important abiotic and biotic features such as

light intensity (Diehl 2007) or compensatory feeding (Cruz-Rivera & Hay 2000). Instead, we aimed

at introducing two fundamental stoichiometric features (i.e. stoichiometric flexibility and

stoichiometric imbalance) and investigating how these stoichiometric considerations can change

predictions of the Rosenzweig-MacArthur model. We thus used our extended Rosenzweig-

MacArthur model to predict the effects of warming and nutrient enrichment on population

dynamics and biomass distribution across trophic levels and compared these predictions with the

predictions of the nonstoichiometric Rosenzweig-MacArthur model. We particularly addressed

two questions: (i) How do stoichiometric constraints modulate the effects of enrichment and

warming on community stability and persistence? and (ii) How do stoichiometric constraints

modulate the effects of enrichment and warming on biomass distribution across multiple trophic

levels?

Methods: Population dynamic models

The Rosenzweig-MacArthur (RM) model.

Rates of change of the consumer and resource biomass densities $\dot{C}$ and $\dot{R}$ depend on their respective

biomass densities C and R (g.m^{-3}):

$\dot{R} = r(1 - \frac{R}{K})R - \frac{aR}{1+ahR}C$ (1)

$\dot{C} = (e\frac{aR}{1+ahR} - m)C$ (2)

The population growth rate of the resource is given by the logistic equation where r is the resource

maximum growth rate and K is the resource carrying capacity. The population growth rate of the

consumer is equal to its feeding rate multiplied by its assimilation efficiency e (i.e. the fraction of

resource biomass converted into consumer biomass) minus a loss term associated to metabolic

losses m. The feeding rate of the consumer C depends on the density of its resource R and follows

a Holling type II functional response, with consumer-resource attack rate a and handling time h.

Our choice for a type II functional response is motivated by a meta-analysis reporting that most

consumers feed on their prey following a saturating type II functional response rather than a linear

type I or a sigmoidal type III functional response (Jeschke *et al.* 2004).

In the RM model, consumer and resource population growth rates are only limited by nutrient or

resource density. Nutrient enrichment is assumed to increase resource carrying capacity, which

often leads to the well-known paradox of enrichment where populations fluctuates up to extinctions

(Rosenzweig 1971). Nevertheless, this model considers neither nutrient dynamics nor temporal

variations of resource stoichiometry and their consequences on population dynamics. To

circumvent these limitations of the RM model, we extended it to better consider nutrient dynamics,

resource stoichiometry and the way they can affect resource and consumer population dynamics.

The Stoichiometric Rosenzweig-MacArthur (SRM) model.

We derived a stoichiometric extension of the Rosenzweig-MacArthur consumer–resource model

with additional stoichiometric and temperature dependencies of several biological rates. We

considered two stoichiometric constraints: one on the resource population growth rate, and the

other on the consumer assimilation efficiency (see below for more details). These stoichiometric

constraints have been observed for several consumer-resource pairs suggesting that they are core

components of species growth and interactions (Sterner & Elser 2002).

Stoichiometric constraint on the resource population growth rate

Inspired by previous stoichiometric models (Andersen 1997; Loladze *et al.* 2000; Andersen *et al.*

2004; Diehl *et al.* 2005), we extended the RM model by considering explicit nutrient dynamics and

nutrient effects on resource population growth rate. The system is assumed to be closed for

nutrients. Thus, nutrient supply originates exclusively from biomass excretion and

remineralisation. The total amount of nutrients in the system (N_{tot}) is then a measure of nutrient

enrichment. In contrast to the very high plasticity in C:N or C:P exhibited by autotrophs,

heterotrophs regulate elemental composition within narrower bounds, even when consuming food

with large variation in elemental composition (Andersen & Hessen 1991; Sterner & Hessen 1994;

Andersen 1997; Elser *et al.* 2000). In other words, the elemental homeostasis is much stronger for

consumers compared to primary producers. We thus assumed the nutrient quota (i.e. the nutrient

to carbon ratio) of the consumer Q_C to be conserved whereas the one of the resource Q_R is flexible

over time with the only constraint that $Q_R > 0$. As in the RM model, rates of change of the consumer

and resource biomass densities $\dot{C}$ and $\dot{R}$ depend on their respective carbon biomass densities C and

R (gC.m^{-3}), except that the resource population growth rate follows the Droop equation (Droop

1974) given by $r(1-Q_{min}/Q_R)R$ and is now limited by Q_R relative to the minimum nutrient quota

Q_{min}:

$\quad N_{tot} = Q_R R + Q_C C$ $\hfill$ (3)

$\quad \dot{R} = r(1 - \frac{Qmin}{Q_R})R - \frac{aR}{1+ahR}C$ $\hfill$ (4)

$\quad \dot{C} = (e\frac{aR}{1+ahR} - m)C$ $\hfill$ (5)

$\quad$ From the nutrient conservation equation (eqn 3) we obtain $Q_R = \frac{N_{tot} - Q_C C}{R}$. The intuitive

$\quad$ interpretation is that the resource nutrient quota Q_R changes instantaneously with the density of the

$\quad$ resource population R and with the density of the nutrient stored in the consumer biomass $Q_C C$, to

$\quad$ maintain nutrient balance (see Text S1 for details).

$\quad$ *Stoichiometric constraint on the consumer population growth rate*

$\quad$ In the RM model, the growth rate of the consumer population only depends on resource density.

$\quad$ We relaxed this assumption by making the population growth rate of the consumer dependent on

$\quad$ both resource quality (i.e. nutrient quota) and quantity (i.e. biomass density). In the SRM model,

$\quad$ consumer production is also limited by resource quality as the consumer assimilation efficiency e

$\quad$ is a saturating function of resource nutrient quota Q_R:

$\quad e(Q_R) = e_{max}\frac{Q_R}{Q_R+Q_C}$ $\hfill$ (6)

$\quad$ The intuitive interpretation of eqn. 6 is that resource quality is not a limiting factor for consumer

$\quad$ growth as long as the nutrient content of the resource is superior to the nutrient content of the

$\quad$ consumer (i.e. $Q_R > Q_C$). In other words, $e(Q_R)$ is at its maximum for $Q_R > Q_C$ and proportional to

$\quad Q_R$ for $Q_R < Q_C$. By replacing e by $e(Q_R)$ in eqn. 5, we obtain the SRM model.

$\quad$ **Temperature dependence of model parameters**

$\quad$ To investigate the effect of temperature and stoichiometric constraints on consumer-resource

$\quad$ dynamics, we next extended the RM and SRM models described above by adding thermal

$\quad$ dependencies of the parameters. Following Uszko et al. (2017), we assumed that the total amount

of nutrient N_{tot}, the maximum food conversion efficiency e_{max}, and fixed stoichiometric traits (Q_C)

are independent of temperature, as there is no evidence of systematic temperature dependence for

any of them (Peters 1983; Ahlgren 1987; Borer *et al.* 2013; Yvon-Durocher *et al.* 2015). Rate of

maintenance respiration and natural background mortality m typically increases exponentially with

temperature (Fig. S1a and b). We thus used the Arrhenius equation to describe the effect of

temperature T (in Kelvin) on m:

$$m(T) = m_0 e^{\frac{-E_m}{kT}} \tag{7}$$

where m_0 is a parameter-specific constant calculated at temperature of 0°C (= 273.15 K). The

temperature dependence is characterized by the respective activation energy E_m (eV) and the

Boltzmann constant $k=8.62\times10^{-5}$ eVK^{-1}. As the temperature dependencies of resource intrinsic

growth rate r and functional response parameters (a, $1/h$) are often unimodal rather than

exponential (Englund *et al.* 2011; Rall *et al.* 2012; Sentis *et al.* 2012; Thomas *et al.* 2012), we used

Gaussian functions for r and a and an inverted Gaussian function for h:

$$g(T) = g_0 e^{\pm\frac{(T-Topt)^2}{2s^2}} \tag{8}$$

where T_{opt} is the temperature at which the rate g reaches its minimum or maximum, s is the function

width and g_0 is a parameter-specific constant calculated at T_{opt}. The minus-sign corresponds to

Gaussian functions and the plus-sign to inverted Gaussian functions.

Model parameterisation and simulations

To parameterise the models we assumed the resource and consumer species to be a unicellular

freshwater algae and a *Daphnia* grazer, respectively. The choice for this system was motivated by

the good characterization of both the stoichiometric parameters and thermal dependencies for this

system (Andersen 1997; Uszko *et al.* 2017). Uszko *et al.* (2017) recently estimated the thermal

dependencies for biological rates of the green algae *Monoraphidium minutum* and the grazer

Daphnia hyalina. We thus used their estimates of stoichiometric parameters and thermal

dependencies (See Table S1 and Fig. S1 for further details).

To investigate the individual and combined effects of enrichment, warming, and stoichiometric

constraints, we varied temperature (401 values ranging from 0 to 40°C by 0.1°C) and total amount

of nutrients (parameter N_{tot} in eqn. 10; 60 values ranging from 0.001 to 0.06 gP.m^{-3} by 0.001 gP.m^{-}

^{3}, overlapping with reported mean phosphorus concentration in European peri-alpine lakes

(Anneville *et al.* 2005)). For the RM model, we used the minimum nutrient quota to convert

nutrients into resource (i.e. $K = N_{tot}/Q_{min}$). This implies that carrying capacity is independent of

temperature which is expected for closed, nutrient-limited systems (Uszko *et al.* 2017) although

more experimental evidence are needed to verify this assumption. We then simulated the consumer-

resource dynamics for 1000 days to enable the system to reach an attractor (either an equilibrium

point or a limit cycle) before we assessed the final state. Therefore, for each model, we simulated

24060 combinations of environmental conditions (401 temperatures by 60 nutrient concentrations).

Initial biomass density of each species was set to 0.98 times its equilibrium density in the two-

species system (calculated by solving for the two-species equilibrium, using either eqns 1-2 for

model RM or eqns 3-5 for model SRM). The value of 0.98 was chosen to be (1) close enough to

equilibria to avoid extinctions caused solely by transient dynamics and (2) not exactly the

equilibrium value to probe the stability of the equilibrium. Any population falling below the

extinction threshold of 10^{-9} g.m^{-3} during the simulation was deemed extinct and its biomass set to

zero to exclude ecologically unrealistic limit cycles. For each model, we calculated system

persistence as the percentage of simulations with the two species remaining extant at the end of the

simulations.

Results

Stability: population fluctuations and persistence

Stoichiometric constraints dampened the paradox of enrichment, reducing fluctuations at high

nutrient levels and hence increasing persistence. However, stoichiometric constraints also reduced

the persistence of the consumer at low and high temperatures. As a result, the overall effect of

stoichiometric constraints on stability depends on its relative influence on population fluctuations

versus consumer persistence. In the two following paragraphs, we explain in more detail these

results and highlight key differences between the outcomes from RM and SRM models.

The RM model predicts that increasing nutrient concentration is strongly destabilizing: the system

shifts from a stable equilibrium point to limit cycles (i.e. the system crosses a Hopf bifurcation).

This agrees with the paradox of enrichment. As population biomass fluctuations (i.e. cycle

amplitude) increase with nutrient concentration, minimal population densities are very low at high

nutrient concentrations leading to the extinction of both the consumer and resource once the

extinction threshold is crossed (Fig. 1). In the range of temperatures where the consumer persists,

warming does not have a strong influence on the nutrient concentration at which the system shifts

from the stable equilibrium point to limit cycles, although this qualitative shift is absent at very

high temperatures (i.e. 32°C) when the consumer is close to extinction. Warming decreases

fluctuation amplitude and thus dampens extinctions driven by the paradox of enrichment, which

results in warming enhancing the persistence of the consumer-resource system at high nutrient

concentrations. However, very warm and cold temperatures cause the extinction of the consumer

(see below for the mechanisms underlying extinctions), releasing resources from top-down control.

Overall, we found that, without considering the extinction threshold of 10^{-9} g.m^{-3} (see Model

parametrisation and simulations), both the consumer and the resource can persist in 74% of the

temperature-nutrient concentration scenarios (i.e. black + orange areas in Fig 1C). Nevertheless,

when considering the extinction threshold, they persist in only 21% of the temperature-nutrient

scenarios (i.e. black area in Fig. 1c) indicating that extinctions driven by population fluctuations

are highly prevalent in the RM model.

In contrast, the SRM model shows that increasing nutrient concentrations causes fewer fluctuations

than those observed for the RM model (Fig. 1). This is because: (1) more nutrients are needed to

shift the system from a stable equilibrium point to limit cycles—the system can indeed persist

without fluctuations up to 0.02 $gP.m^{-3}$ whereas it was only up to 0.0005 $gP.m^{-3}$ in the RM model—

and (2) when the system fluctuates, the amplitude of the fluctuations is smaller in the SRM than in

the RM model. As a result, stoichiometric constraints dampen the amplitude of population

fluctuations (i.e. the paradox of enrichment) and hence increase system persistence at high nutrient

levels. While the qualitative effect of temperature is similar to that observed in the RM model, the

thermal thresholds for consumer persistence are reduced at low and high temperatures in the SRM

predictions. Moreover, thermal thresholds remain almost constant along the nutrient gradient in the

RM model, whereas in the SRM model they depend on nutrient concentration, with a smaller

thermal range at low nutrient levels compared to high nutrient levels (Fig. 1). The consumer is thus

more likely to go extinct at low nutrient concentrations and extreme temperatures in the SRM

model than in the RM model. Overall, system persistence for the SRM model was 44% without

considering the extinction threshold and 37% when considering it. In other words, without

considering extinctions driven by very low biomass densities, the SRM model predicts lower

persistence of the consumer compared to RM model but it is the opposite pattern when considering

extinctions driven by very low biomass densities. We thus conclude that the RM model predicts

larger population fluctuations leading to high probabilities of populations extinctions in

comparison to the SRM model.

Biomass distribution

We next compared the predictions of both models for consumer-resource biomass ratios along the

temperature and nutrient gradients (Fig. 2). We found that the RM model systematically predicts

inverted biomass pyramids. In contrast, the SRM model predicts both biomass pyramids and

inverted pyramids depending on temperature and nutrient levels. The RM model systemically

predicts that, as soon as the consumer can persist, its population biomass density systematically

exceeds the resource population biomass density (Fig. 2). With the SRM model, the biomass ratios

are below one at low nutrient levels (Fig. 2). However, at medium and high nutrient levels, the

ratios are above one as soon as the consumer can persist. We found qualitatively similar results

when considering unstable equilibrium points (Fig. S2).

Mechanisms underlying stability and biomass distribution patterns

Here, we detail the mechanisms underlying the stability and biomass distribution patterns to better

understand how and when stoichiometric constraints modulate the effects of temperature and

nutrients on consumer-resource dynamics. Stoichiometric constraints induce two effects. The first

one is a static shift in the parameter values of consumer assimilation efficiency and of the resource

carrying capacity. We call this static effect. The second one corresponds to a population dynamical

feedback. We call this dynamic effect. Both static and dynamic effects operate through consumer

efficiency and resource population growth rate.

Consumer energetic efficiency of the consumer and assimilation efficiency

The persistence of the consumer at low and high temperatures is driven by the energetic efficiency

EE of the consumer (i.e. its feeding rate relative to metabolic losses) calculated as follows:

$$EE = \frac{ef(R^*)}{m} \qquad (9)$$

Where $f(R^*)$ is the functional response of the consumer at resource density R^* (i.e. the resource

equilibrium density in absence of the consumer). We recall that the assimilation efficiency e is a

function of resource quality Q_R in the SRM model whereas it is constant in the RM model. The

intuitive interpretation of eqn. 9 is that *EE* should be above one for the consumer population to

growth and persist.

To better understand the influence of stoichiometric constraints on consumer persistence, we thus

investigated differences in the RM and SRM model predictions regarding the consumer energetic

efficiency *EE* along the temperature gradient at two nutrient concentrations (Fig. 3). For both

models, energetic efficiency at equilibrium (i.e. static effect) has a hump-shaped relationship with

temperature with maximal efficiency values at medium temperatures. While this unimodal shape

is conserved across nutrient levels and models, the RM model systematically predicts higher

consumer energetic efficiency values than the SRM model because consumer assimilation

efficiency is lower in the SRM than in the RM model (Fig. S3). As a result, the temperatures at

which energetic efficiency falls below one and drives consumers extinct are more extreme in the

RM model compared to the SRM model (Fig. 3). In other words, energetic efficiency is above one

for a narrower thermal range in the SRM model.

Dynamical feedbacks due to the stoichiometric constraints

The second mechanism by which stoichiometric constraints influence consumer-resource stability

and biomass distribution is the dynamical feedbacks due to stoichiometric constraints on the

resource population growth rate and on the consumer energetic efficiency. In the SRM model, the

growth rate of the resource population depends on both the total nutrient load and the consumer

population density as $Q_R = (N_{tot} - Q_C C)/R$. In other words, when consumer population increases,

this decreases resource population growth leading to a negative feedback on consumer population

growth rate. In addition to this first dynamical feedback, there is a second dynamical feedback as

the consumer population growth rate also depends on Q_R and thus on its own biomass density.

Thus, also this second negative feedback loop limits the consumer population growth rate when its

density increases. Altogether, dynamical feedbacks reduce strongly the amplitude of population

fluctuations, which in turn increases resource and consumer persistence.

To separate the static and dynamic effects of the stoichiometric constraints, we calculated the

values of assimilation efficiencies and carrying capacities predicted by the SRM model for each

temperature-nutrient scenario (Fig. S3) and used these effective parameter values to replace the

values of parameters e and K in the RM model for each temperature-nutrient scenario. In other

words, we calculated average values of e and K in the dynamic SRM model and used them as

constant input parameters in the RM model. We then simulated population dynamics along the

temperature-nutrient gradient using the RM model with these effective parameters; referred

hereafter as effective RM model (Fig. 4). Comparing predictions from the RM, effective RM, and

SRM models allowed to disentangle the static stoichiometric effects (when going from the RM to

the effective RM predictions; Fig. 4, panels a to b) from the dynamical stoichiometric effects (when

going from the effective RM to the SRM predictions; Fig. 4, panels b to c). We found that, at low

nutrient concentrations, population fluctuations and consumer persistence predicted by the

effective RM model agreed with predictions of the SRM model. However, the system shifted from

a stable equilibrium point to a limit cycle at lower nutrient concentrations for the effective RM

model than for the SRM model. This suggests that more nutrients are needed to destabilize the

system with the SRM model. Moreover, the effective RM model predicts ampler population

fluctuations than the SRM model. As a result, the effective RM predicts high extinction rates at

high nutrient concentrations compared to the SRM model. Overall, we found that the effective RM

model cannot fully reproduce the dynamics predicted by the SRM, which indicates that including

stoichiometric constraints in the RM model involves more than only changing parameter values.

Discussion

Temperature and nutrient enrichment are two of the most important drivers of global change

(Nelson 2005). However, most research on the effects of temperature and nutrients on community

dynamics assumes that the elemental composition of primary producers and consumers are constant

and independent of changes on energy and material fluxes (Binzer *et al.* 2012; Boit *et al.* 2012;

Amarasekare & Coutinho 2014; Gilbert *et al.* 2014; Amarasekare 2015; Binzer *et al.* 2016;

Gilarranz *et al.* 2016). Yet, the elemental composition of primary producers is known to be flexible,

which can have important consequences for community dynamics and ecosystem processes (Elser

et al. 2000). We have shown how stoichiometric constraints that account for flexible stoichiometry

can affect predictions on how temperature and nutrients influence community stability and biomass

distribution across trophic levels. We thus argue that considering stoichiometric constraints is an

important step toward a better understanding of the effects of global change on ecosystems.

Stoichiometric constraints and temperature can dampen the paradox of enrichment

We showed that both stoichiometric constraints and temperature dampen the negative effect of

nutrient enrichment on consumer-resource fluctuations and increase system persistence at high

nutrient levels. Temperature effects are driven by physiological mechanisms. In agreement with

previous empirical studies, our model parametrization reflects the observation ~~we found~~ that

metabolic loss rates increase faster with warming than consumer feeding rates (Vucic-Pestic *et al.*

2011; Sentis *et al.* 2012; Fussmann *et al.* 2014; Iles 2014). Consumers are thereby less energetically

efficient at higher temperatures which stabilizes food-web dynamics by reducing energy flow

between trophic levels (Binzer *et al.* 2012; Kratina *et al.* 2012; Fussmann *et al.* 2014; Sentis *et al.*

2017). In contrast, the effect of stoichiometric constraints is mainly linked to two mechanisms: a

static shift in the position of the Hopf bifurcation and negative dynamical feedbacks of the

consumer and resource on their population growth rates. Both resources and consumers are

composed of the same essential elements (N, P, and C), which implies that, when consumer or

resource population biomass increases, it reduces the pool of free nutrients available for the growth

of the resource population. Therefore, more nutrients are needed to shift the system from a stable

equilibrium to population cycles. In other words, the paradox of enrichment is displaced to higher

nutrient concentrations (i.e., the position of the Hopf bifurcation is shifted to higher nutrient levels).

In addition to the static effect above, we found two dynamic effects that correspond to negative

dynamical feedbacks of the consumer and the resource on themselves. When consumer population

increases, it decreases the population growth rate of the resource by limiting nutrient availability,

diminishing resource biomass, which, in turn, decreases the consumer population growth rate.

Conversely, when the resource biomass increases, this decreases the nutrient content of the

resource, which, in turn, limits the growth rates of both the resource and consumer populations.

These stoichiometric negative feedback loops strongly decrease the amplitude of population

fluctuations and thus dampen the paradox of enrichment. Interestingly, our comparisons of the RM,

effective RM and SRM model predictions indicate that the dynamical effects contribute more to

the reduction of fluctuations than the static effects: population fluctuations are large in the effective

RM model accounting for the static effect only, whereas they are much smaller in SRM model

accounting for both static and dynamical effects (Fig. 4). This implies that the impact of

stoichiometric constraints on community dynamics goes beyond a simple modification of

parameter values and encompass more complex population feedbacks between the consumer and

the resource.

Overall, these results demonstrate that both flexible stoichiometry and temperature can

synergistically dampen the paradox of enrichment by two different mechanisms: population

dynamic feedbacks and physiological constraints. Our consumer-resource model is simplified

compared to natural communities composed of numerous species. Yet, the predictions of the

stoichiometric model fit with empirical observations. In eutrophic lakes and experimental

mesocosms, populations can persist at relatively high nutrient concentrations even if fertilisation

enhance population fluctuations (O'Connor *et al.* 2009; Boit *et al.* 2012; Kratina *et al.* 2012), as

our stoichiometric model predicts. In contrast, the Rosenzweig-MacArthur model tends to produce

very large population fluctuations and extinctions at low nutrient concentrations which can explain

why these predictions are not well supported by empirical observations (McAllister *et al.* 1972;

Jensen & Ginzburg 2005).

Effects of stoichiometric constraints on system persistence across environmental gradients

While stoichiometric constraints dampen the paradox of enrichment and thus increase persistence

at high nutrient levels, they also reduce the persistence of the consumer at low and high

temperatures. Stoichiometric constraints affect the thermal thresholds for consumer extinctions.

Consumers can only persist over a narrower range of intermediate temperatures when they are

constrained by stoichiometry. This is due to the reduced assimilation efficiency of the consumer at

low and high temperatures that, in turn, decreases its energetic efficiency and thus fastens consumer

extinction. The finding that assimilation efficiency is temperature dependent contrasts with

previous theoretical studies that used the RM model and assumed that the assimilation efficiency

is temperature independent (Binzer *et al.* 2012; Gilbert *et al.* 2014; Sentis *et al.* 2017; Uszko *et al.*

2017). In the SRM model, the thermal dependency of the consumer assimilation efficiency is fully

driven by the change in the resource stoichiometry induced by temperature. The SRM model thus

predicts an additional mechanism by which temperature can influence trophic interactions:

temperature changes resource stoichiometry which, in turn, impacts the consumer assimilation

efficiency and its population growth rate. To sum up, the overall effect of stoichiometric constraints

on system persistence thus depends on the temperature range considered and on their relative

influence on population fluctuations versus consumer persistence.

Effects of stoichiometric constraints on biomass distribution

We found that stoichiometric constraints can modulate the effects of temperature and nutrients on

biomass distribution across trophic levels. Without stoichiometric constraints (i.e. with the

Rosenzweig-MacArthur model), biomass ratios are above one for almost all temperatures or

nutrient levels as the biomass produced by the resource is efficiently transferred to the consumer

level consistently along the environmental gradients. This finding agrees with theoretical studies

reporting that Lotka-Volterra and RM models predict biomass ratios above one and fail to

reproduce biomass pyramids for a substantial region of parameter values (Jonsson 2017; Barbier

& Loreau 2019). However, in nature, consumer-resource biomass ratios are often below one

(McCauley & Kalff 1981; Del Giorgio & Gasol 1995; McCauley *et al.* 1999; Irigoien *et al.* 2004)

suggesting that additional mechanisms should be included to better understand and predict biomass

distribution patterns in natural food webs. Our stoichiometric model agrees with experimental

observations. It predicts that, at low nutrient concentrations (i.e. < 0.01 gP.m^{-3}), the biomass ratio

never exceeds one along the entire temperature gradient. This is observed in oligotrophic aquatic

systems where primary production is too low to sustain high consumer populations (O'Connor *et*

al. 2009). In addition, we also found that increasing nutrient levels decreased the temperature

ranges within which biomass ratio is below one. This corresponds to results from manipulated

nutrient concentrations and temperature in aquatic mesocosms, where zooplankton to

phytoplankton biomass ratio only exceeds one in the enriched mesocosms at medium or warm

temperatures (i.e. 27°C) (O'Connor *et al.* 2009). This suggests that the models with stoichiometric

constraints better reproduce the biomass patterns observed in experimental and natural systems.

Nevertheless, further experiments investigating the links between stoichiometric flexibility and

consumer-resource dynamics are needed to determine if these stoichiometric mechanisms are

underlying patterns of biomass distribution in nature.

Implications of our findings for global change

Temperature and nutrients do not act in isolation from each other. Climate warming, for example,

causes stronger water stratification which, in turn, can limit nutrient cycling (Sarmiento *et al.* 2004;

Tranvik *et al.* 2009). Environmental policies such as the European water framework directive (i.e.

Directive 2000/60/EC of the European Parliament and of the Council establishing a framework for

the Community action in the field of water policy) effectively reduces input of nutrients in aquatic

ecosystems (Anneville *et al.* 2005) while the climate keeps warming. With these two phenomena,

water will often be warmer and contain fewer nutrients in aquatic systems. Our models consistently

predict that warmer temperatures should stabilise consumer-resource dynamics but, if temperature

further increases, the consumer goes extinct as energetic efficiency decreases with warming.

Moreover, we found that stoichiometric constraints can reduce this thermal extinction threshold

(i.e. the consumer persists in a narrower thermal range), especially at low nutrient levels. Our

stoichiometric model thus suggests that decreasing nutrient concentrations alongside warmer

temperatures should fasten the extinction of consumer populations. This prediction matches

empirical observations of consumer extinctions at warm temperatures in oligotrophic aquatic

systems (Petchey *et al.* 1999; O'Connor *et al.* 2009). Altogether, these results indicate that

considering stoichiometric constraints can be of importance for the management of nutrient inputs

and the conservation of natural populations and communities under climate change.

Conclusion

Knowledge of how temperature and nutrient simultaneously influence the elemental composition

of primary producers and consumers is crucial to better understand and predict the effects of global

change on species interactions, community dynamics and fluxes of energy and material within and

among ecosystems. Here we showed that stoichiometric constraints dampen the negative effect of

enrichment on stability by reducing population fluctuations through population dynamics

feedbacks. However, stoichiometric constraints also decrease consumer energetic efficiency,

which increases consumer extinction risk at extreme temperatures and low nutrient concentrations.

Finally, stoichiometric constraints can reverse biomass distribution across trophic level by

modulating consumer efficiency and resource population growth rate along the temperature and

nutrient gradients. Our study provides a first step in the exploration of the consequences of

stoichiometric constraints and temperature on ecological communities. It suggests that accounting

for stoichiometric constraints can strongly influence our understanding of how global change

drivers impact important features of ecological communities such as stability and biomass

distribution patterns.

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References

Ahlgren, G. (1987). Temperature functions in biology and their application to algal growth

constants. *Oikos*, 49, 177-190.

Amarasekare, P. (2015). Effects of temperature on consumer-resource interactions. *Journal of*

Animal Ecology, 84, 665-679.

Amarasekare, P. & Coutinho, R.M. (2014). Effects of temperature on intraspecific competition in

ectotherms. *The American Naturalist*, 184, E50-E65.

Andersen, T. (1997). *Pelagic nutrient cycles: herbivores as sources and sinks*. Springer-Verlag,

Berlin, Germany.

Andersen, T., Elser, J.J. & Hessen, D.O. (2004). Stoichiometry and population dynamics.

Ecology Letters, 7, 884-900.

Andersen, T. & Hessen, D.O. (1991). Carbon, nitrogen, and phosphorus content of freshwater

zooplankton. *Limnology and Oceanography*, 36, 807-814.

Anneville, O., Gammeter, S. & Straile, D. (2005). Phosphorus decrease and climate variability:

mediators of synchrony in phytoplankton changes among European peri-alpine lakes.

Freshwater Biology, 50, 1731-1746.

Barbier, M. & Loreau, M. (2019). Pyramids and cascades: a synthesis of food chain functioning

and stability. *Ecology Letters*, 22, 405-419.

Bezemer, T.M. & Jones, T.H. (1998). Plant-insect herbivore interactions in elevated atmospheric

C02: quantitative analyses and guild effects. *Oikos*, 82, 212-222.

Binzer, A., Guill, C., Brose, U. & Rall, B.C. (2012). The dynamics of food chains under climate

change and nutrient enrichment. *Philosophical Transactions of the Royal Society B:*

Biological Sciences, 367, 2935-2944.

Binzer, A., Guill, C., Rall, B.C. & Brose, U. (2016). Interactive effects of warming,

eutrophication and size structure: impacts on biodiversity and food-web structure. *Global*

Change Biology, 22, 220-227.

Boit, A., Martinez, N.D., Williams, R.J. & Gaedke, U. (2012). Mechanistic theory and modelling

of complex food-web dynamics in Lake Constance. *Ecology Letters*, 15, 594-602.

Borer, E.T., Bracken, M.E., Seabloom, E.W., Smith, J.E., Cebrian, J., Cleland, E.E. *et al.* (2013).

Global biogeography of autotroph chemistry: is insolation a driving force? *Oikos*, 122,

1121-1130.

Brown, J.H., Gillooly, J.F., Allen, A.P., Savage, V.M. & West, G.B. (2004). Toward a metabolic

theory of ecology. *Ecology*, 85, 1771-1789.

Cruz-Rivera, E. & Hay, M.E. (2000). Can quantity replace quality? Food choice, compensatory

feeding, and fitness of marine mesograzers. *Ecology*, 81, 201-219.

Del Giorgio, P.A. & Gasol, J.M. (1995). Biomass Distribution in Freshwater Plankton

Communities. *The American Naturalist*, 146, 135-152.

Diehl, S. (2007). Paradoxes of enrichment: effects of increased light versus nutrient supply on

pelagic producer-grazer systems. *The American Naturalist*, 169, E173-E191.

Diehl, S., Berger, S. & Wöhrl, R. (2005). Flexible nutrient stoichiometry mediates environmental

influences on phytoplankton and its resources. *Ecology*, 86, 2931-2945.

Droop, M. (1974). The nutrient status of algal cells in continuous culture. *Journal of the Marine*

Biological Association of the United Kingdom, 54, 825-855.

Elser, J., Sterner, R., Gorokhova, E.a., Fagan, W., Markow, T., Cotner, J. *et al.* (2000). Biological

stoichiometry from genes to ecosystems. *Ecology Letters*, 3, 540-550.

Elser, J.J., Bracken, M.E., Cleland, E.E., Gruner, D.S., Harpole, W.S., Hillebrand, H. *et al.*

(2007). Global analysis of nitrogen and phosphorus limitation of primary producers in

freshwater, marine and terrestrial ecosystems. *Ecology Letters*, 10, 1135-1142.

Elser, J.J., Loladze, I., Peace, A.L. & Kuang, Y. (2012). Lotka re-loaded: modeling trophic

interactions under stoichiometric constraints. *Ecological Modelling*, 245, 3-11.

Elton, C. (1927). *Animal ecology*. Sidgwick & Jackson, LTD, London.

Englund, G., Ohlund, G., Hein, C.L. & Diehl, S. (2011). Temperature dependence of the

functional response. *Ecology Letters*, 14, 914-921.

Enquist, B.J., West, G.B., Charnov, E.L. & Brown, J.H. (1999). Allometric scaling of production

and life-history variation in vascular plants. *Nature*, 401, 907-911.

Falkowski, P.G., Barber, R.T. & Smetacek, V. (1998). Biogeochemical controls and feedbacks on

ocean primary production. *Science*, 281, 200-206.

Finkel, Z.V., Beardall, J., Flynn, K.J., Quigg, A., Rees, T.A.V. & Raven, J.A. (2009).

Phytoplankton in a changing world: cell size and elemental stoichiometry. *Journal of*

Plankton Research, 32, 119-137.

Fussmann, K.E., Schwarzmüller, F., Brose, U., Jousset, A. & Rall, B.C. (2014). Ecological

stability in response to warming. *Nature Climate Change*, 4, 206-210.

Gilarranz, L.J., Mora, C. & Bascompte, J. (2016). Anthropogenic effects are associated with a

lower persistence of marine food webs. *Nature communications*, 7, 10737.

Gilbert, B., Tunney, T.D., McCann, K.S., DeLong, J.P., Vasseur, D.A., Savage, V. *et al.* (2014).

A bioenergetic framework for the temperature dependence of trophic interactions.

Ecology Letters, 17, 902-9014.

Hessen, D.O., Ågren, G.I., Anderson, T.R., Elser, J.J. & De Ruiter, P.C. (2004). Carbon

sequestration in ecosystems: the role of stoichiometry. *Ecology*, 85, 1179-1192.

627 Hessen, D.O., Færøvig, P.J. & Andersen, T. (2002). Light, nutrients, and P:C ratios in algae:

628   grazer performance related to food quality and quantity. *Ecology*, 83, 1886-1898.

629 Iles, A.C. (2014). Towards predicting community level effects of climate: relative temperature

630   scaling of metabolic and ingestion rates. *Ecology*, 95, 2657–2668.

631 Irigoien, X., Huisman, J. & Harris, R.P. (2004). Global biodiversity patterns of marine

632   phytoplankton and zooplankton. *Nature*, 429, 863-867.

633 Jensen, C.X. & Ginzburg, L.R. (2005). Paradoxes or theoretical failures? The jury is still out.

634   *Ecological Modelling*, 188, 3-14.

635 Jeschke, J.M., Kopp, M. & Tollrian, R. (2004). Consumer-food systems: why Type I functional

636   responses are exclusive to filter feeders. *Biological Reviews*, 79, 337-349.

637 Jonsson, T. (2017). Conditions for eltonian pyramids in Lotka-Volterra food chains. *Scientific*

638   *Reports*, 7, 10912.

639 Kratina, P., Greig, H.S., Thompson, P.L., Carvalho-Pereira, T.S.A. & Shurin, J.B. (2012).

640   Warming modifies trophic cascades and eutrophication in experimental freshwater

641   communities. *Ecology*, 93, 1421-1430.

642 Lindeman, R.L. (1942). The trophic-dynamic aspect of ecology. *Ecology*, 23, 399-417.

643 Loladze, I., Kuang, Y. & Elser, J.J. (2000). Stoichiometry in producer-grazer systems: linking

644   energy flow with element cycling. *Bulletin of Mathematical Biology*, 62, 1137-1162.

645 McAllister, C., LeBrasseur, R., Parsons, T. & Rosenzweig, M. (1972). Stability of enriched

646   aquatic ecosystems. *Science*, 175, 562-565.

647 McCauley, D.J., Gellner, G., Martinez, N.D., Williams, R.J., Sandin, S.A., Micheli, F. *et al.*

648   (2018). On the prevalence and dynamics of inverted trophic pyramids and otherwise top-

649   heavy communities. *Ecology Letters*, 21, 439-454.

McCauley, E. & Kalff, J. (1981). Empirical relationships between phytoplankton and

zooplankton biomass in lakes. *Canadian Journal of Fisheries and Aquatic Sciences*, 38,

458-463.

McCauley, E., Nisbet, R.M., Murdoch, W.W., de Roos, A.M. & Gurney, W.S.C. (1999). Large-

amplitude cycles of *Daphnia* and its algal prey in enriched environments. *Nature*, 402,

653-656.

Montoya, J.M. & Raffaelli, D. (2010). Climate change, biotic interactions and ecosystem

services. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365,

2013-2018.

Nelson, G.C. (2005). *Millennium ecosystem assessment: drivers of ecosystem change: summary*

chapter. World Resources Institute, Washington, DC.

O'Connor, M.I., Piehler, M.F., Leech, D.M., Anton, A. & Bruno, J.F. (2009). Warming and

resource availability shift food web structure and metabolism. *PLoS Biology*, 7,

e1000178.

Petchey, O.L., McPhearson, P.T., Casey, T.M. & Morin, P.J. (1999). Environmental warming

alters food-web structure and ecosystem function. *Nature*, 402, 69-72.

Peters, R.H. (1983). *The ecological implications of body size*. Cambridge University Press,

Cambridge.

Rall, B.C., Brose, U., Hartvig, M., Kalinkat, G., Schwarzmüller, F., Vucic-Pestic, O. *et al.*

(2012). Universal temperature and body-mass scaling of feeding rates. *Philosophical*

Transactions of the Royal Society B: Biological Sciences, 367, 2923-2934.

Rastetter, E.B., Ågren, G.I. & Shaver, G.R. (1997). Responses of N-limited ecosystems to

increased CO_2: a balanced-nutrition, coupled-element-cycles model. *Ecological*

Applications, 7, 444-460.

Rip, J.M.K. & McCann, K.S. (2011). Cross-ecosystem differences in stability and the principle of

energy flux. *Ecology Letters*, 14, 733-740.

Robert W. Sterner, James J. Elser, Everett J. Fee, Stephanie J. Guildford &

Thomas H. Chrzanowski (1997). The Light: Nutrient Ratio in Lakes: The Balance of

Energy and Materials Affects Ecosystem Structure and Process. *The American Naturalist*,

150, 663-684.

Rosenzweig, M.L. (1971). Paradox of enrichment: destabilization of exploitation ecosystems in

ecological time. *Science*, 171, 385-387.

Sarmiento, J.L., Slater, R., Barber, R., Bopp, L., Doney, S.C., Hirst, A. *et al.* (2004). Response of

ocean ecosystems to climate warming. *Global Biogeochemical Cycles*, 18, 1-23.

Sentis, A., Binzer, A. & Boukal, D.S. (2017). Temperature-size responses alter food chain

persistence across environmental gradients. *Ecology Letters*, 20, 852-862.

Sentis, A., Hemptinne, J.L. & Brodeur, J. (2012). Using functional response modeling to

investigate the effect of temperature on predator feeding rate and energetic efficiency.

Oecologia, 169, 1117-1125.

Sentis, A., Hemptinne, J.L. & Brodeur, J. (2014). Towards a mechanistic understanding of

temperature and enrichment effects on species interaction strength, omnivory and food-

web structure. *Ecology Letters*, 17, 785-793.

Sterner, R.W. & Elser, J.J. (2002). *Ecological stoichiometry: the biology of elements from

molecules to the biosphere*. Princeton University Press.

Sterner, R.W. & Hessen, D.O. (1994). Algal nutrient limitation and the nutrition of aquatic

herbivores. *Annual Review of Ecology and Systematics*, 25, 1-29.

Tabi, A., Petchey, O.L. & Pennekamp, F. (2019). Warming reduces the effects of enrichment on

stability and functioning across levels of organisation in an aquatic microbial ecosystem.

Ecology Letters, 22, 1061-1071.

Thomas, M.K., Aranguren-Gassis, M., Kremer, C.T., Gould, M.R., Anderson, K., Klausmeier,

C.A. *et al.* (2017). Temperature–nutrient interactions exacerbate sensitivity to warming in

phytoplankton. *Global Change Biology*, 23, 3269-3280.

Thomas, M.K., Kremer, C.T., Klausmeier, C.A. & Litchman, E. (2012). A global pattern of

thermal adaptation in marine phytoplankton. *Science*, 338, 1085-1088.

Tilman, D. (1982). *Resource competition and community structure*. Princeton university press.

Tranvik, L.J., Downing, J.A., Cotner, J.B., Loiselle, S.A., Striegl, R.G., Ballatore, T.J. *et al.*

(2009). Lakes and reservoirs as regulators of carbon cycling and climate. *Limnology and*

Oceanography, 54, 2298-2314.

Tylianakis, J.M., Didham, R.K., Bascompte, J. & Wardle, D.A. (2008). Global change and

species interactions in terrestrial ecosystems. *Ecology Letters*, 11, 1351-1363.

Uszko, W., Diehl, S., Englund, G. & Amarasekare, P. (2017). Effects of warming on predator–

prey interactions – a resource-based approach and a theoretical synthesis. *Ecology Letters*,

20, 513-523.

Vasseur, D.A. & McCann, K.S. (2005). A mechanistic approach for modeling temperature-

dependent consumer–resource dynamics. *American Naturalist*, 166, 184-198.

Vucic-Pestic, O., Ehnes, R.B., Rall, B.C. & Brose, U. (2011). Warming up the system: higher

predator feeding rates but lower energetic efficiencies. *Global Change Biology*, 17, 1301-

1310.

White, T. (1993). The inadequate environment. Nitrogen and the abundance of animals. Springer

Verlag, Berlin.

Woods, H.A., Makino, W., Cotner, J.B., Hobbie, S.E., Harrison, J.F., Acharya, K. *et al.* (2003).

Temperature and the chemical composition of poikilothermic organisms. *Functional*

Ecology, 17, 237-245.

Yvon-Durocher, G., Dossena, M., Trimmer, M., Woodward, G. & Allen, A.P. (2015).

Temperature and the biogeography of algal stoichiometry. *Global Ecology and*

Biogeography, 24, 562-570.

Figure legends

Fig. 1. Population fluctuations (consumer biomass coefficient of variation; panels *a* and *b*) and

species persistence (number of species; panels *c* and *d*) across the temperature (*y* axis) and nutrient

(*x* axis) gradients as predicted by the Rosenzweig-MacArthur (RM; panels *a* and *c*) and by the

Stoichiometric Rosenzweig-MacArthur (SRM; panels *b* and *d*) models. In panels *a* and *b*, the white

colour corresponds to the temperature-nutrient scenario for which the consumer has gone extinct

whereas the orange to red to dark red represent population fluctuations of increasing amplitude. In

panels *c* and *d*, in black: both consumer and resource persist; in red: only the resource persists; in

orange: none persists. Resource biomass CV is not shown; it is qualitatively similar to the consumer

biomass CV as resource and consumer biomass fluctuation are strongly coupled.

Fig. 2. Consumer-resource biomass ratio along the temperature gradient for the Rosenzweig-

MacArthur (RM, green lines) and the Stoichiometric Rosenzweig-MacArthur (SRM, black lines)

models at three nutrient concentrations (0.008, 0.02, and 0.032 gP.m^{-3}). In each panel, the dotted

line represents biomass ratio of one; i.e. the biomass densities of the resource and the consumer are

equal. Biomass values shown at equilibrium points. For unstable equilibrium points (i.e. limit

cycles), see Fig. S2.

Fig. 3. Consumer energetic efficiency along the temperature gradient for the Rosenzweig-

MacArthur (RM, in green) and the Stoichiometric Rosenzweig-MacArthur (SRM, in black) models

at two nutrient concentrations (0.008 and 0.02 gP/m^3). In each panel, the dotted line represents

energetic efficiency equal to one.

Fig. 4. Population fluctuations (consumer biomass coefficient of variation) across the temperature

(y axis) and nutrient (x axis) gradients as predicted by the Rosenzweig-MacArthur (RM; panel a),

the RM with effective parameters (panel b), and the Stoichiometric Rosenzweig-MacArthur (SRM;

panel c) models.

Figure 1

Figure 2

Figure 3

769 **Figure 4**

776   **Stoichiometric constraints modulate the effects of temperature and nutrients on biomass**
777   **distribution and community stability**
778
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780

Text S1. Derivation of the Stoichiometric Rosenzweig-MacArthur (SRM) model

The model studied in the main text is very similar to previous stoichiometric consumer-resource

models (Andersen 1997; Loladze *et al.* 2000; Andersen *et al.* 2004; Diehl *et al.* 2005). To make

our paper self-contained, we here present the model assumptions and derive the model equations

(eqs. 3-6 in main text). Our objective was not to develop a complex and very realistic stoichiometric

model that would include additional important abiotic and biotic features such as light intensity

(Diehl 2007) or compensatory feeding (Cruz-Rivera & Hay 2000). Instead, we aimed at introducing

two fundamental stoichiometric features (i.e. stoichiometric flexibility and stoichiometric

imbalance) and investigate how these stoichiometric considerations can change predictions of the

Rosenzweig-MacArthur model. We assumed that resource and consumer production are limited by

energy and a single mineral nutrient. Moreover, we assume the system is closed for nutrients. Thus,

nutrient supply originates exclusively from excretion and remineralization of biomass. The total

amount of nutrients in the system (N_{tot}) is then a measure of nutrient enrichment. As elemental

homeostasis is much stronger for consumers compared to primary producers (Andersen 1997), we

assumed the nutrient quota of the consumer Q_C to be constant whereas the nutrient quota of the

resource Q_R is flexible. Four differential equations determine the dynamics of four state variables,

that is, the concentrations of resource (R) and consumer (C) carbon biomasses and of dissolved

mineral nutrients (N), and the nutrient quota of the resource (Q_R):

$$\dot{C} = (e\frac{aR}{1+ahR} - m)C \tag{S1}$$

$$\dot{R} = r(1 - \frac{Qmin}{Q_R})R - \frac{aR}{1+ahR}C \tag{S2}$$

$$\dot{Q}_R = h(N) - r(Q_R - Qmin) \tag{S3}$$

$\dot{N} = (Q_R - Q_C e)\frac{aR}{1+ahR} C + Q_C mC - h(N)R$ (S4)

As in the RM model, rates of change of the consumer and resource biomass densities $\dot{C}$ and $\dot{R}$

depend on their respective carbon biomass densities C and R (gC.m^{-3}), except that the resource

population growth rate follows the Droop equation (Droop 1974) and is now limited by its nutrient

quota Q_R relative to the minimum nutrient quota Q_{min}. Rate of change of Q_R depends on the nutrient

uptake rate by the resource species $h(N)$ and the amount of nutrient invested in growth (eqn S3).

$h(N)$ is the specific resource nutrient uptake rate and can be represented by a Michaelis-Menten

model where the amount of nutrient uptake saturates at high nutrient concentrations.

With the mass-balance equation, we get that the total amount of nutrient is the sum of the free

nutrient plus the nutrient fixed in the resource biomass plus the nutrient fixed in the consumer

biomass: $N_{tot} = N + Q_R R + Q_C C$. As Eqns S1-S4 conserve total biomass (the system is closed), the

time derivative of N_{tot} is zero. We can thus replace one of the four differential equations S1-S4 with

the algebraic equation $N_{tot} = N + Q_R R + Q_C C$:

$\dot{C} = (e\frac{aR}{1+ahR} - m)C$

(S5)

$\dot{R} = r(1 - \frac{Qmin}{Q_R})R - \frac{aR}{1+ahR}C$

(S6)

$\dot{N} = (Q_R - Q_C e)\frac{aR}{1+ahR} C + Q_C mC - h(N)R$

(S7)

$N_{tot} = N + Q_R R + Q_C C$ (S8)

It is possible to derive a simpler model by reducing the number of dimensions in the above model

from three to two. This model reduction is based on the assumption that free nutrients are taken up

very quickly relative to the dynamics of the consumer and resource biomasses. This corresponds

to taking $h(N)$ large, $h(N) = \frac{1}{\varepsilon}\tilde{h}(N)$ for small ε. The fast dynamics (on the timescale t ~ ε) are

$\dot{N} = -\frac{1}{\varepsilon}\tilde{h}(N)R + (slower\ contributions)$

(S9)

Which converge to N → 0, and $Q_R \rightarrow \frac{N_{tot}-Q_C C}{R}$ with N_{tot} the total nutrient in the system. In other

words, N in dead and excreted matter is immediately recycled and acquired by the resource species.

When substituting the quasi-steady-state in eqns. (S5, S6), we get the resulting dynamics (on the

timescale t ~1):

$\dot{C} = (e\frac{aR}{1+ahR} - m)C$ (S10)

$\dot{R} = r(1 - \frac{Qmin}{Q_R})R - \frac{aR}{1+ahR}C$ (S11)

$N_{tot} = Q_R R + Q_C C$ (S12)

From the nutrient conservation equation (eqn. S12) we obtain that $Q_R = \frac{N_{tot}-Q_C C}{R}$. The intuitive

interpretation is that the resource nutrient quota Q_R decreases with the density of the resource

population and with the density of nutrient stored in the consumer biomass. In contrast to eqns S5-

S8, the reduced model has only two differential equations and one algebraic equation. It can be

equivalently written as a set of three differential equations with $\dot{C}$ and $\dot{R}$ similar as equations S10

and S11 and with $\dot{Q}_R = \frac{d\frac{N_{tot}-Q_C C}{R}}{dt} = (Q_R - Q_C e)\frac{aR}{1+ahR}C + \frac{Q_C mC}{R} - r(Q_R - Qmin)$.

In the RM model, the growth rate of the consumer population is assumed to depend only on

resource density. We relaxed this assumption by making the population growth rate of the

consumer dependent on both the resource quality (i.e. nutrient quota) and quantity (i.e. density). In

the SRM model, consumer production is also limited by resource quality as the consumer

assimilation efficiency e is a saturating function of resource nutrient quota Q_R:

$\quad e(Q_R) = e_{max}\frac{Q_R}{Q_R+Q_C}$ $\hspace{10cm}$ (S13)

$\quad$ The intuitive interpretation of eqn. S13 is that resource quality is not a limiting factor for consumer

$\quad$ growth as long as the nutrient content of the resource is superior to the nutrient content of the

$\quad$ consumer (i.e. $Q_R > Q_C$). In other words, when $Q_R >> Q_C$, $e(Q_R) \rightarrow e_{max}$ and when $Q_R << Q_C$, $e(Q_R)$

$\quad \rightarrow 0$. By replacing e by $e(Q_R)$ in eqn. S10, we obtain the SRM model.

Table S1. Definitions and units of model parameters, from Uszko *et al.* (2017). For temperature-

dependent parameters, we list the value of the scaling constant Q_0 (in units of the parameter) and

the values of either the activation energy E_Q (eV, when temperature dependence is monotonous,

eqn. 7) or of the temperature T_{opt} (Kelvin) at which the parameter value reaches a

maximum/minimum and the width s (Kelvin) of this bell-/U-shaped function (when temperature

dependence is non-monotonous, eqn. 8). Biomass and nutrients are expressed in units of carbon

(C) and phosphorus (P), respectively

Parameter	Temperature independent value	Unit	Definition	Reference
	Thermal parameters			
r	$r_0 = 2.2$; $T_{opt} = 298.15$; $s = 12.0$	1/d	Intrinsic rate of resource net production (gross production – biosynthesis costs)	Uszko et al. 2017
h	$h_0 = 0.17$; $T_{opt} = 294.1$; $s = 6.4$	d	Handling time	Uszko et al. 2017
a	$a_0 = 8.9$; $T_{opt} = 296.0$; $s = 9.4$	$m^3/(gC\ d)$	Attack rate	Uszko et al. 2017
m	$m_0 = 4.4 \times 10^8$; $E_m = 0.55$	1/d	Consumer mortality plus maintenance rate	
e_{max}	0.385	-	Maximum assimilation efficiency	Yodzis
Q_C	0.042	g P/g C	Consumer P:C ratio	Diehl 2005
Q_{min}	0.009	g P/g C	Minimum nutrient quota	Diehl 2005
Q_R	Variable	g P/g C	Resource P:C ratio	
N_{tot}	Variable	$g\ P/m^3$	Total nutrients in the system	
T	Variable	K	Temperature	

866   **Fig. S1.** Thermal functions used to parametrize the model (adapted from Uszko et al. 2017)

Fig. S2. Consumer-resource biomass ratio (log scale) along the temperature gradient for the

Rosenzweig-MacArthur (RM, green lines) and the Stoichiometric Rosenzweig-MacArthur (SRM,

black lines) models at three nutrient concentrations (0.008, 0.02, and 0.032 gP.m^{-3}). In each panel,

the dotted lines represent unstable solutions whereas full lines represent stable solutions. The thin

horizontal dotted line represents biomass ratio of one; i.e. the biomass densities of the resource and

the consumer are equal.

884   **Fig. S3**. Effective assimilation efficiency e_{ef} and carrying capacity K_{ef} from the

885   Stoichiometric Rosenzweig-MacArthur (SRM) model along the temperature gradient at

886   three nutrient levels (0.004, 0.015, and 0.032 gP.m^{-3}) with $Q_C = 0.042$. Full lines represent

887   temperature and nutrient scenarios for which both the resource and consumer persist

888   whereas dotted lines represent scenarios for which only the resource persists. Effective

889   assimilation efficiency was calculated as $e_{ef} = e_{max}Q_R/(Q_R+Q_C)$, with Q_R the equilibrium

890   solution of the SRM model and the effective carrying capacity as $K_{ef} = Q_R R/Q_{min} = (N_{tot}-$

891              $Q_C C)/Q_{min}$, with Q_R, R and C the equilibrium solutions of the SRM model.

References

Andersen, T. (1997). *Pelagic nutrient cycles: herbivores as sources and sinks*. Springer-Verlag, Berlin, Germany.

Andersen, T., Elser, J.J. & Hessen, D.O. (2004). Stoichiometry and population dynamics. *Ecology Letters*, 7, 884-900.

Cruz-Rivera, E. & Hay, M.E. (2000). Can quantity replace quality? Food choice, compensatory feeding, and fitness of marine mesograzers. *Ecology*, 81, 201-219.

Diehl, S. (2007). Paradoxes of enrichment: effects of increased light versus nutrient supply on pelagic producer-grazer systems. *The American Naturalist*, 169, E173-E191.

Diehl, S., Berger, S. & Wöhrl, R. (2005). Flexible nutrient stoichiometry mediates environmental influences on phytoplankton and its resources. *Ecology*, 86, 2931-2945.

Droop, M. (1974). The nutrient status of algal cells in continuous culture. *Journal of the Marine Biological Association of the United Kingdom*, 54, 825-855.

Loladze, I., Kuang, Y. & Elser, J.J. (2000). Stoichiometry in producer-grazer systems: linking energy flow with element cycling. *Bulletin of Mathematical Biology*, 62, 1137-1162.

Uszko, W., Diehl, S., Englund, G. & Amarasekare, P. (2017). Effects of warming on predator–prey interactions – a resource-based approach and a theoretical synthesis. *Ecology Letters*, 20, 513-523.