

1 **Title:**

2 Do El Niño events beget generations of reproductively deficient adults?

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17 **Abstract**

18 1. Warm ocean waters during El Niño events deplete primary productivity, with cascading
19 effects through the food chain that profoundly affect many marine and terrestrial species,
20 commonly increasing adult mortality and offspring starvation. With global warming, events
21 will double and increasingly threaten the depletion or extinction of some animal
22 populations.

23 2. Because adverse environments experienced during infancy generally induce reproductive
24 deficits in adulthood, El Niño events are also expected to affect animals born during them,
25 engendering generations of adults with reduced reproductive potential and exacerbating
26 demographic impacts.

27 3. We made the first test of this idea, using the blue-footed booby, a piscivorous apex
28 predator of the eastern tropical Pacific Ocean. Surprisingly, detailed monitoring of 18
29 generations over a 27-year period detected no deficits in the annual breeding success,
30 offspring viability, lifespan or lifetime reproductive success of generations of adults born
31 during El Niño years.

32 4. These results testify to remarkable developmental resilience extending across the
33 lifespan. However, there was evidence that this resilience was supported by two
34 mechanisms of quality control of adult generations from El Niño years.

35 5. First, viability selection on nestlings and fledglings was more severe for El Niño birth
36 cohorts than ordinary cohorts.

37 6. Second, in El Niño years, adult boobies self-selected for breeding. There was no
38 increase in the proportional representation of either high quality breeders or breeders in
39 their peak years (middle-age), but there was an increase in old-young adult pairings, which
40 in this population produce the most viable fledglings.

41 7. The blue-footed booby appears to be immune to the expected developmental impact of
42 El Niño on the reproductive value of adult generations. The susceptibilities and resilience
43 of other species need to be explored, to better predict the demographic impacts of this
44 accelerating climatic oscillation.

45 **Keywords:**

46 Adverse environments, cohort effects, resilience, lifetime reproductive success, El Niño
47 Southern Oscillation, viability selection, self-selection

48

49

50 **Introduction**

51 With global warming, our planet's most dramatic short-term climate oscillation, El Niño
52 Southern Oscillation (ENSO), is predicted to increase in intensity, with extreme events
53 doubling in frequency in the near future (Cai et al., 2014; Ham, 2018; Wang et al., 2017).
54 This raises urgent questions about the ability of animal populations of the eastern Pacific
55 Ocean to withstand its accelerating demographic impacts. During El Niño events, separated
56 by 2-7 year intervals (Zhang & Zhao, 2015), warm ocean waters deplete primary
57 productivity, with broadly deleterious consequences throughout the food chain. Impacts on
58 marine vertebrates in the reproductive season, especially seabirds, include physiological
59 suppression and delay (Guerra, Fitzpatrick, Aguilar, & Venables, 1988), reduced adult
60 survival and reproduction, nest abandonment, reduced offspring growth and survival,
61 population decline and changes in distribution (Hays, 1986; Hodder & Graybill, 1985; Le
62 Boeuf & Reiter, 1991; Tovar, Guillen, & Cabrera, 1987; Valle, Cruz, Cruz, Merlen, &
63 Coulter, 1987); with increased frequency of warm water events limiting population
64 recovery and increasing risk of extinction (Vargas, Harrison, Rea, & Macdonald, 2006;
65 Vargas et al., 2007).

66 Another likely consequence, with demographic implications, is diminished reproductive
67 value of adults born in El Niño years. No study has addressed the possibility of El Niño
68 cohort effects, but numerous experiments have shown that early-life environmental
69 conditions usually have developmental effects on physiological, behavioural,
70 morphological and life history traits of vertebrates during adulthood (Grace, Martin-
71 Gousset, & Angelier, 2017; Haywood & Perrins, 1992; Mumme, Bowman, Pruett, &
72 Fitzpatrick, 2015; Nettle et al., 2017; Plard et al., 2015). Furthermore, in mammals, natural

73 early-life adversity has been linked to harmful consequences on fitness components and
74 senescence rates (e.g. Nussey, Kruuk, Morris, & Clutton-brock, 2007; Pigeon, Festa-
75 Bianchet, & Pelletier, 2017; Plard et al., 2015), and in 5 of 9 bird species studied in nature,
76 adults that experienced poor weather, habitats, prey availability, parental care or late
77 fledging in the natal year showed reduced lifetime reproductive success (review in
78 Drummond & Ancona, 2015). There is also experimental and descriptive evidence for
79 intergenerational effects: negative impacts of nutritional and social stress in infancy on the
80 quality and viability of the infant's eventual offspring (Burton & Metcalfe, 2014;
81 Drummond & Rodríguez, 2013; Naguib & Gil, 2005).

82 These findings raise the expectation that in many species, generations of adults from El
83 Niño birth cohorts will show survival or fitness deficits or produce offspring of low quality,
84 at some stage of their lifetime. However, in wild animals such effects are sometimes
85 mitigated in one or both sexes by simple developmental resilience or by making life-history
86 adjustments that seemingly neutralize fitness penalties (e.g. Ancona & Drummond, 2013;
87 Cartwright, Nicoll, Jones, Tatayah, & Norris, 2014; Dmitriew & Rowe, 2007; Douhard et
88 al., 2016). Thus, in some species, adults exposed to warm ocean waters in their natal year
89 might sidestep impacts on their lifetime reproductive success, although complete evasion of
90 fitness impacts of early-life environmental stress has not been documented for any long-
91 lived animal.

92 Our case study of a population of blue-footed boobies (*Sula nebouxii*) tested for effects of
93 ENSO conditions (mean water temperature in the 8-month breeding season) in the natal
94 year on the quality of adult generations by comparing the lifetime reproductive
95 performance of recruits from 18 birth cohorts. This long-lived (up to 25 years), socially

96 monogamous apex predator of the eastern tropical Pacific Ocean lays clutches of 1-3 eggs,
97 but often raises fewer chicks when underfeeding triggers siblicidal brood reduction. When
98 an El Niño event depletes populations of its fish prey, circulating corticosterone increases
99 in females (Wingfield, Ramos-Fernandez, Nuñez-de la Mora, & Drummond, 1999),
100 mortality of young adults increases, fewer boobies breed, breeding is delayed, clutch and
101 brood sizes are smaller, hatching fledging and breeding success decline, and nestlings grow
102 more slowly (Ancona, Sánchez-Colón, Rodríguez, & Drummond, 2011; Kiere &
103 Drummond, 2016; Oro, Torres, Rodríguez, & Drummond, 2010). Nonetheless, previous
104 analysis of the first ten years of life suggested that male and female recruits from El Niño
105 birth cohorts largely escape developmental impacts: they recruited at an earlier age, bred
106 less frequently (skipped more years) and matched the survivorship and accumulated
107 breeding success (production of fledglings) of recruits born under favourable cold water
108 conditions (Ancona & Drummond, 2013). However, these observations do not demonstrate
109 full equivalence to recruits from cold-water cohorts because fitness penalties could
110 manifest after the first 10 years or in other fitness components such as rate of reproductive
111 senescence, adult lifespan or offspring quality. To determine whether El Niño events
112 impact the reproductive value of adult generations, fitness accounting across the whole
113 lifespan of recruits is required. This was possible for the focal population because, with rare
114 exceptions, fledglings nest for the first time (mostly at ages 2-5 years) close to where they
115 hatched, then nest close to their first nest site during the rest of their lives (Kim, Torres,
116 Domínguez, & Drummond, 2007).

117 We also tested for two largely unstudied filtering processes which theoretically could
118 buttress populations against demographic impacts of adverse natal environments and mask

119 their developmental impacts. The first is self-selection for breeding in adverse conditions
120 by high quality breeders able to confer their own quality on offspring. In support of this
121 idea, there is some evidence that in stressful circumstances animals that are in poor
122 condition, very young or very old may skip breeding, trading reproduction for survival in
123 the service of lifetime reproductive success (Cubaynes, Doherty, Schreiber, Gimenez, & Jr,
124 2010; Goutte, Kriloff, Weimerskirch, & Chastel, 2011; Skjaeraasen et al., 2012), and
125 resulting in an increase in the proportion of high quality individuals among breeders. The
126 second is increased viability selection in adverse conditions: differential survival and
127 recruitment into the breeding population of high quality offspring (cf. Garratt et al., 2015;
128 Mojica & Kelly, 2010).

129

130

131 **Materials and methods**

132 *Study population and data collection*

133 Demographic data were collected between 1989 and 2018 during lifelong annual
134 monitoring of individual blue-footed boobies (*Sula nebouxii*) on Isla Isabel, Nayarit
135 (21°53'N, 105°54'W), off the Pacific coast of Mexico (Drummond, Torres, & Krishnan,
136 2003; Kim et al., 2007).

137 Breeders were individually identified by bands fitted at fledging. Throughout each
138 reproductive season (roughly February through July) all pairs in two study areas were
139 monitored by recording their nest contents every 3-6 days through fledging (details in
140 Drummond et al., 2003). Nests in the study area and within 20 m of its borders, which
141 correspond with natural borders of the colony, are too conspicuous on the beach and forest
142 floor to escape detection by the team of two monitors (Drummond & Rodríguez, 2015).
143 This monitoring protocol allowed detection of nearly all recruitment and breeding of the
144 focal fledglings through 2018.

145 For all analyses of breeders, samples included all members of the 18 birth cohorts (1989-
146 2009) that recruited by age 6 years (the age by which 87% of fledglings recruit), thereby
147 excluding individuals whose early nesting may have been missed. We analysed annual
148 breeding success (number of 70-d old fledglings produced, regardless of brood size),
149 fledgling viability (proportion of fledglings that recruited), adult lifespan (age at the last
150 observed breeding), lifetime reproductive success (total recruits produced), fledging
151 success (proportion of nestlings that managed to fledge), and recruitment (whether an
152 individual was resighted as a breeder; 1-0 variable), separately for females and males

153 because pair mates are not statistically independent. Because nestlings and fledglings are
154 not sexed until they return to the colony as breeders, the analyses of fledging success and
155 recruitment were not separated by sex.

156 The sample for analysis of annual breeding success included all breeders from the 18
157 cohorts (2143 male and 1893 female boobies); the sample for analysis of fledgling viability
158 included the offspring produced by all breeders from those cohorts that fledged at least one
159 chick at any age (1646 male and 1329 female breeders), and excluded fledglings born after
160 2013 because their recruitment could not be adequately scored (they were monitored for
161 less than 6 years). The sample for analyses of adult lifespan and lifetime reproductive
162 success included only the 1319 male and 1105 female breeders that died before 2013,
163 allowing each of their fledglings at least 6 years to recruit during annual monitoring. Death
164 was assumed when a breeder was absent during three consecutive breeding seasons; in the
165 cohorts of 1989-1995, only 5-6% of breeders were ever resighted after three consecutive
166 absences. For males and females, respectively, ages of first reproduction (mean \pm SD) were
167 4.14 ± 0.95 and 3.54 ± 0.98 years, and adult lifespans were 8.05 ± 3.49 and 8.02 ± 3.99
168 years (range 1 to 22 years for both sexes).

169 Self-selection for breeding in warm water conditions was tested in the 4036 males and
170 females of the 18 cohorts that recruited by looking for increase in the proportional
171 representation of individuals of above average quality, favourable ages (10-14 years) or
172 from cool water (La Niña) natal years (lest this confer any long-term benefit). Individual
173 quality was expressed as the mean of all z-normal standardized annual breeding success
174 values (fledglings produced) over each individual's lifespan.

175 Increased viability selection on warm water cohorts was tested by analysing the proportion
176 of nestlings that managed to fledge each year and the recruitment probability of each year's
177 fledglings. For fledging success, we used data from one-, two-, and three- chick broods
178 (sample sizes: 7630, 6975, and 1326, respectively) reared between 1989 and 2018; for
179 recruitment, we included all 10604 fledglings produced between 1989 and 2013 for which
180 body measures were available and recruitment could be adequately scored. The ratio
181 between weight (grams) and ulna length (millimetres) at age 70 days was used as an
182 indicator of body condition at fledging. This ratio accounts for the sexual size dimorphism
183 of the boobies; at 79 days old, females weigh approximately 26% more than males, and
184 their ulnas are 10% longer (Torres & Drummond, 1999). To remove year effects, individual
185 body condition was calculated as the z-normal standardized annual body condition at
186 fledging of the colony. Boobies weighed (mean \pm SD) 1593.51 ± 256.2 grams at fledging
187 with an ulna length of 205.57 ± 12.98 millimetres.

188 ENSO conditions during the reproductive season in which a focal breeder was born were
189 expressed as the average of the 8 monthly (December to July) sea surface temperature
190 anomalies (SSTA) at the closest oceanographic station, 55 km from the colony, obtained
191 from the IRI Data Library
192 http://iridl.ldeo.columbia.edu/SOURCES/.NOAA/.NCEP/.EMC/.CMB/.GLOBAL/.Reyn_SmithOIv2/.monthly/.sst/. Niño and La Niña years were characterized by mean SSTA of at
194 least $+0.5^{\circ}\text{C}$ or -0.5°C , respectively, using the criteria of NOAA, obtained from the El Niño
195 and La Niña Alert System <https://www.climate.gov/news-features/understanding-climate/el-ni%C3%B1o-and-la-ni%C3%B1a-alert-system>. The Southern Oscillation Index
196 (SOI) was not used because SSTAs are more closely related to the breeding parameters and
197

198 age at first reproduction of this population (Ancona & Drummond, 2013; Ancona et al.,
199 2011).

200 *Statistical analysis*

201 All analyses were performed in R statistical environment (R Development Core Team,
202 2018). All variables were standardized prior to model fitting to minimize collinearity of
203 linear and quadratic terms and to facilitate the interpretation of the relative strength of
204 parameter estimates (Cade, 2015; Grueber, Nakagawa, Laws, & Jamieson, 2011). Because
205 high or even moderate collinearity can inflate standard errors of coefficient estimates and
206 make genuine effects harder to detect (Zuur, Ieno, & Elphick, 2010), the variance inflation
207 factor (VIF) was calculated for each variable in every hypothesis-based model, so that
208 variables with VIF values higher than 2 could be excluded from the final analyses (none
209 were). All VIF values are reported in the supplementary materials (Supporting
210 information). Models were fitted with a Gaussian error distribution and an identity link
211 function. Variable standardization was carried out using the *rescale* function in the R
212 package *arm* (Gelman et al., 2016). We used the *lmer* function in the package *lme4* (Bates,
213 Maechler Martin, & Walker, 2016) to build the LMMs; LMs were built using the built-in
214 function *lm* in R. Model selection was carried out with *model.sel* function in the MuMIn
215 package (Bartón, 2016)³⁵, *mod.avg* function was used to average models within a 95%
216 confidence interval (CI; Burnham, Anderson, & Huyvaert, 2011), and *confint* function to
217 calculate the CI of each parameter.

218 To assess the effects of recruits' early-life conditions on their annual breeding success and
219 their fledglings' viability, we first constructed a base model (a LMM accounting for

220 previously demonstrated effects in the species; *sensu* Panagakis, Hamel, & Côté, 2017) for
221 each analysis; this contained the linear and quadratic expression of breeder age and laying
222 date, with breeder identity as random effect. For annual breeding success, SSTA in each
223 reproductive year was also included in base models. The linear expression of year of birth
224 was included in every model, irrespective of its statistical significance (cf. Bouwhuis,
225 Vedder, & Becker, 2015), to reflect the fact that the earlier an individual was born, the
226 higher its lifespan could be, biasing our dataset. From these base models, we constructed a
227 series of competing models which included SSTA in the natal year (the early-life-condition
228 variable), whether it was the individual's terminal reproductive event (1-0 variable), and the
229 linear and quadratic expressions of age of first reproduction along with two-way
230 interactions: SSTA in the natal year \times age², age of first reproduction \times age² and SSTA in the
231 natal year \times age of last reproduction. SSTA in the reproductive year was added to
232 competing models for the analysis of offspring viability. Age of first reproduction was
233 included to evaluate the indirect effects of early adversity, as this life history trait has been
234 shown to vary with natal ENSO conditions (Ancona & Drummond, 2013) and is a major
235 determinant of fitness (Pärt, 1995); its quadratic expression was added to detect an optimal
236 onset of reproduction (e.g. Mourocq et al., 2016). The interactions were included to detect
237 differences in senescence patterns due to natal conditions or age of first reproduction.

238 To test whether adult lifespan and lifetime reproductive success were influenced by early-
239 life conditions, we used linear models (LMs). For both analyses, year of birth was also
240 included in all competing models as a fixed variable. Variables included in the models were
241 SSTA in the natal year and the linear and quadratic expression of age of first reproduction.

242 For lifetime reproductive success adult lifespan was added to the competing models, along
243 with its interaction with age of first reproduction.

244 To test for increased representation in warm water years of breeders of above average
245 quality or peak reproductive ages (10-14 years, middle aged), or from cool water natal
246 years (< -0.5 SSTA), as well as diminished representation of both young (1-9 years) and old
247 (15-23 years) breeders, we built linear models including sex as a factor and using the linear
248 expression of year of reproduction to account for progressive increase in the proportions of
249 middle-aged and old birds among ringed birds with additional years of monitoring and
250 ringing.

251 We compared the proportions of nestlings that fledge under different ENSO conditions by
252 building a LM for each brood size. To compare the probabilities of recruiting after fledging
253 in different ENSO conditions, and test for self-selection of above average breeders for
254 reproduction in warm water conditions (self-selection), we built LMMs. For all three
255 analyses, year of birth and SSTA in the natal year were included as fixed variables. For the
256 recruitment analysis, body condition at fledging and chick rank (1st, 2nd or 3rd hatched) were
257 included in the competing models, along with three two-way interactions: SSTA in the
258 natal year x body condition, SSTA in the natal year x chick rank, and chick rank x body
259 condition. Nest identity was added as a random effect because, in some cases, siblings were
260 included in the analysis. Body condition and rank were included to test whether low values
261 affect recruitment. The interactions tested whether poor natal environments increase the
262 cost of poor body condition or low rank, and evaluated whether body condition at fledging
263 varies with rank.

264 On the basis of AICc, Δ AICc, and AICc weights (w_i) of the competing models, we selected
265 the best-supported model, or averaged a subset of the best-supported models with a
266 cumulative Akaike weight of ~ 0.95 . For each resulting model, we calculated the estimates
267 (all reported effect sizes were standardized to two SDs) and 95% confidence intervals (CI).
268 Model selection tables and summaries of best-supported and averaged models are provided
269 in the supplementary materials.
270

271 **Results**

272 *Fitness effects of natal ENSO conditions*

273 Annual breeding success (number of fledglings produced) of 1893 female recruits from the
274 18 birth cohorts was unaffected by water temperature in the natal year (Table S2; Figure
275 1a). Natal water temperature affected the annual breeding success of 2143 male recruits,
276 but not in the expected direction. Males that experienced warm natal waters produced
277 slightly more fledglings annually than those born under more favourable conditions ($\beta =$
278 0.009 [95% Confidence Interval: 0.008 to 0.048]; Figure 1), with males fledged under a
279 +0.8 °C SSTA (sea surface temperature anomaly) predicted to produce roughly 6% more
280 fledglings than males fledged under a -0.8 °C anomaly (the extreme anomalies observed).
281 Consistent with previous reports (Ancona et al., 2011; Drummond et al., 2003), boobies of
282 both sexes increased their annual breeding success with earliness of laying and current
283 coolness of the sea, and showed senescence after a peak in reproductive success at roughly
284 ages 10-14 years (Beamonte-Barrientos, Velando, Drummond, & Torres, 2010; Alberto
285 Velando, Drummond, & Torres, 2006). Importantly, natal SSTA did not interact with age
286 (Table S3), showing that warm natal waters had no effect on either the rate of increase in
287 breeding success during early adult years or the rate of decline during late adult years
288 (reproductive senescence).

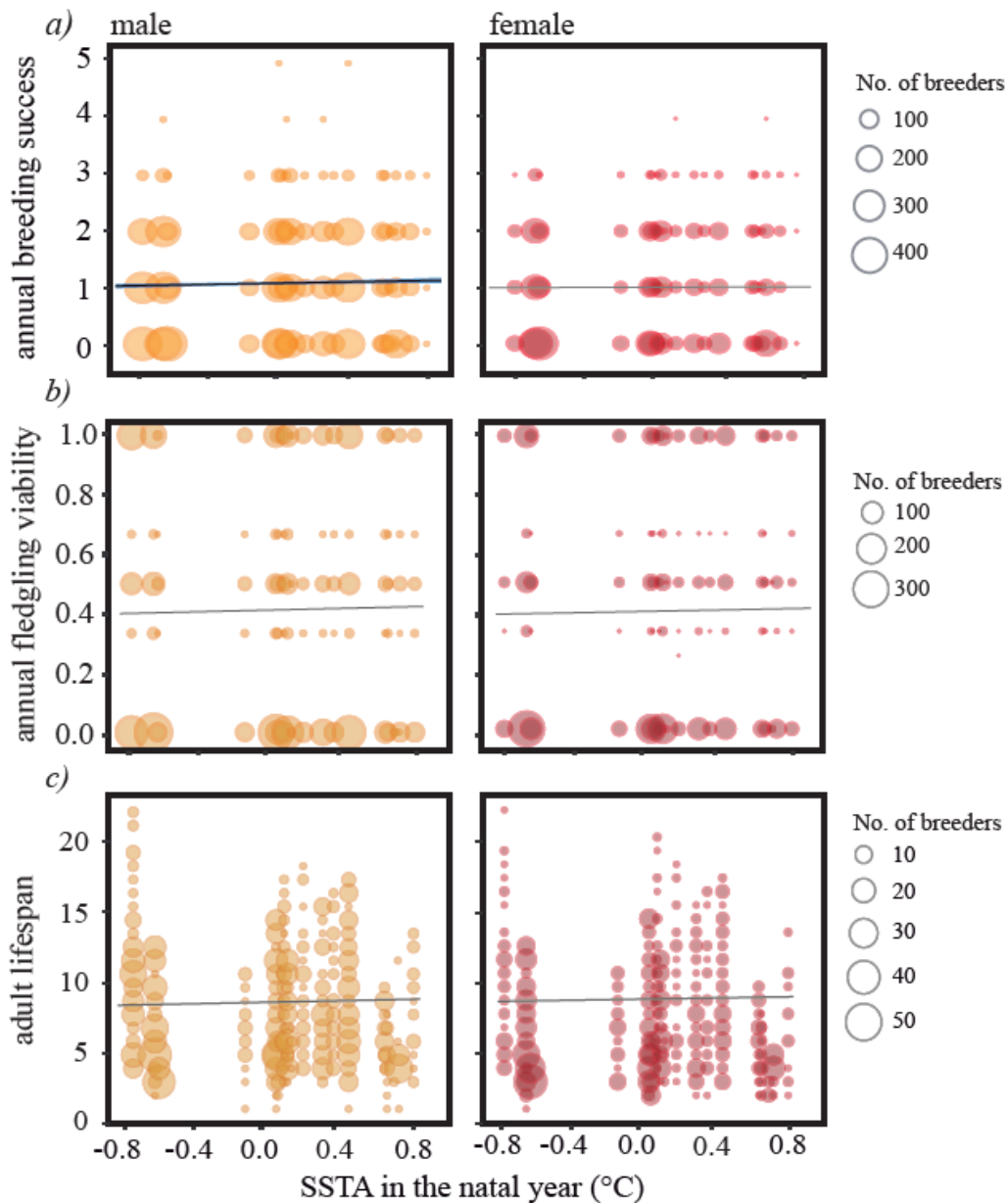
289 Viability (probability of recruitment) of the fledglings produced by 1646 male and 1329
290 female recruits at all ages over the lifespan was unaffected by those recruits' natal water
291 temperature (Table S5; Figure 1b), although other variables were influential. Fledglings
292 produced by both sexes in warm water years were less likely to recruit (Males: $\beta = -0.194$

293 [95% CI: -0.222 to -0.166]; Females: $\beta = -0.195$ [95% CI: -0.231 to -0.159]) and, at all
294 ages, earliness of laying increased the viability of fledglings produced (Males: $\beta = -0.128$
295 [95% CI: -0.156 to -0.101]; Females: $\beta = -0.112$ [95% CI: -0.146 to -0.077]). A recruit's
296 age had a quadratic effect on the probability its offspring would recruit (Table S6), with
297 peaks at roughly 10 years in males and 11 years in females.

298 The lifespan of recruits was unaffected by natal water temperature (Table S8; Figure 1c),
299 but increased with late recruitment in both sexes (Males: 0.139 [95% CI: 0.090 to 0.0.189];
300 Females: 0.213 [95% CI: 0.160 to 0.266]), although this effect might arise simply because
301 late recruits cannot die early.

302

303

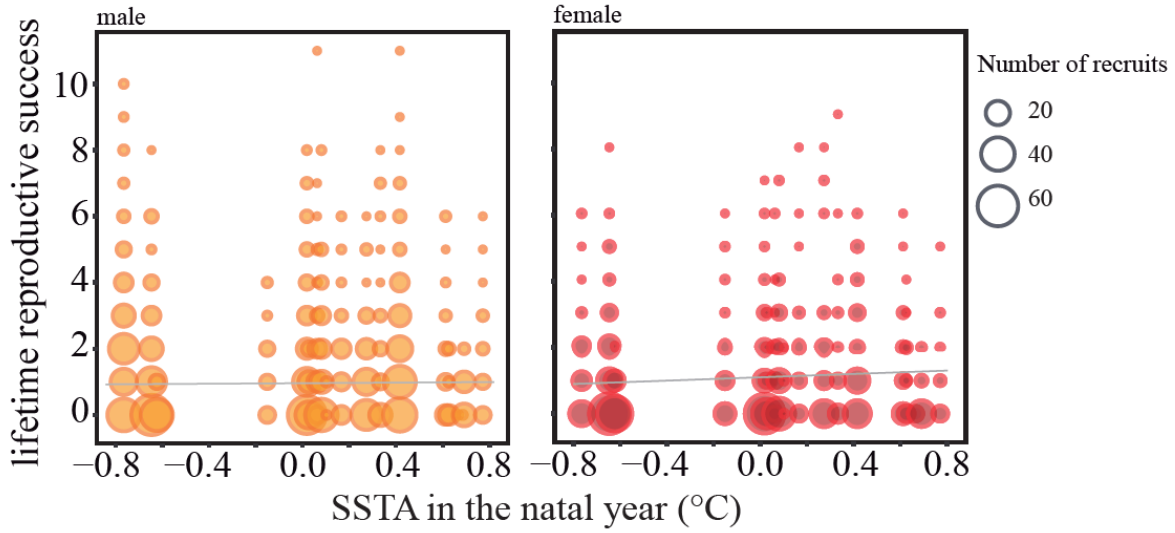


304 **Fig. 1. Effect of SSTA in natal year on annual breeding success, fledgling viability**
305 **and lifespan. a)** A male booby's breeding success (fledglings produced) increased slightly
306 with its natal SSTA (Sea Surface Temperature Anomaly), whereas a female's breeding
307 success was unaffected. Points represent 2143 male and 1893 female breeders observed
308 over their lifespans. **b)** For 1646 male and 1329 female breeders, the proportion of their

309 annually produced fledglings that recruited was unaffected by their natal SSTA. c) Adult
310 lifespan of 1319 male and 1105 female breeders was not affected by natal SSTA. Solid line
311 represents model predictions, with 95% confidence intervals in blue; grey lines show the
312 statistically unsupported tendencies for the other five analyses.

313 Finally, lifetime reproductive success (total recruits produced) of both sexes was unaffected
314 by natal water temperature (Table S11; Figure 2), although in both sexes age of recruitment
315 and adult lifespan were influential. Boobies with younger ages of first reproduction
316 produced more recruits over their lifetime than those that started later in life (Males: $\beta = -$
317 0.086 [95% CI: -0.126 to -0.046]; Females: $\beta = -0.062$ [95% CI: -0.119 to -0.013]). Greater
318 adult lifespan, itself associated with delayed age of first reproduction, also increased
319 lifetime production of recruits in both sexes (Males: $\beta = 0.715$ [95% CI: 0.670 to 0.761];
320 Females: $\beta = 0.619$ [95% CI: 0.561 to 0.677]). Note that despite achieving roughly 6%
321 higher annual breeding success than cool natal water males over the lifespan, warm natal
322 water males showed no corresponding increase in lifetime reproductive success, probably
323 because they skipped more breeding seasons than cool water males (Ancona & Drummond,
324 2013). Warm natal water females showed no such increase in annual breeding success, but
325 females are nearly one third heavier than males and more susceptible to food deprivation
326 (Torres & Drummond, 1997; Velando, 2002).

327



328

329 **Fig. 2.** Lifetime production of recruits was unaffected by natal water temperature. Data for

330 1319 male and 1105 female breeders. Grey lines show the statistically unsupported

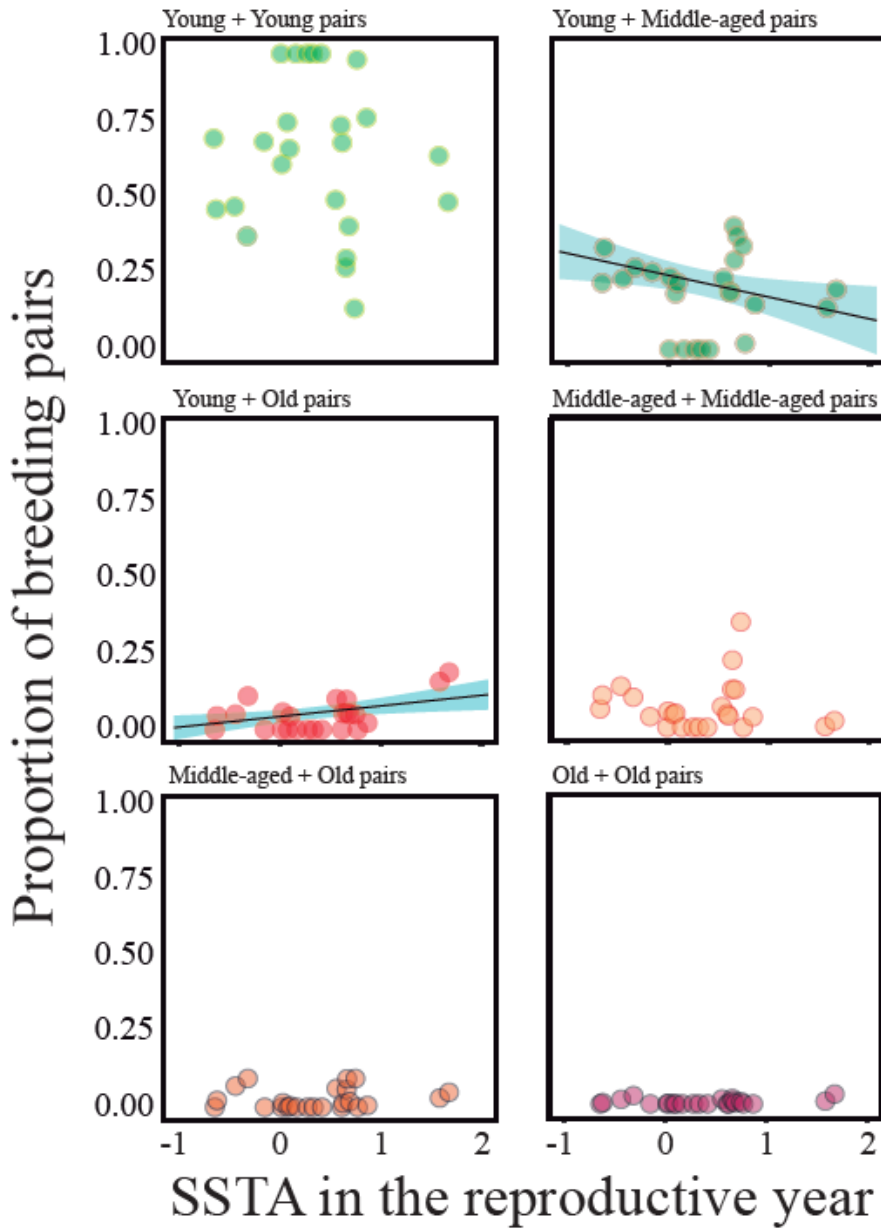
331 tendency for production to increase in both sexes with warm natal water.

332

333 *Self-selection for breeding in El Niño conditions*

334 In warm water years neither boobies of above average quality nor boobies born in cool
335 water (La Niña) years increased their proportional representation among breeders ($n =$
336 4036; Tables S13-S15) but, contrary to expectation, middle aged boobies substantially
337 diminished theirs ($\beta = -0.317$ [95% CI: -0.470 to -0.164]) and the proportions of both
338 young (1-9 years) and old (>14 years) boobies among breeders increased (Young: $\beta = 0.193$
339 [95% CI: 0.068 to 0.318]; Old: $\beta = 0.175$ [95% CI: 0.020 to 0.331]).

340 In this booby, despite young and old individuals showing inferior breeding success (see
341 fitness effects), young-with-old parental pairings produce the most viable (likely to recruit)
342 fledglings (Drummond & Rodríguez, 2015), so we hypothesized that overrepresentation of
343 young and old breeders in warm years could increase the proportion of young + old
344 pairings and hence the average quality of fledglings and recruits. We tested for this effect
345 on pairings by calculating the proportional representation of all six possible pairings of
346 young, middle aged and old boobies in every year between 1993 and 2017 (Table S16).
347 Sex, as a factor, was not included in the LMs because greater viability of offspring of
348 young-old pairs is independent of which sex is the older pair mate (Boyko, 2008;
349 Drummond & Rodríguez, 2015). These analyses showed that warm water is indeed
350 associated with an increase in the proportion of young + old pairings, and also with a
351 decrease in the proportion of young + middle aged pairings (Young + old: $\beta = 0.280$ [95%
352 CI: 0.072 to 0.487]; Young + Middle aged: $\beta = -0.340$ [95% CI: -0.614 to -0.067]; Figure
353 3).



354

355 **Fig. 3.** In warm water years the proportion of young + old pairs increased, and the
356 proportion of young + middle aged pairs decreased.

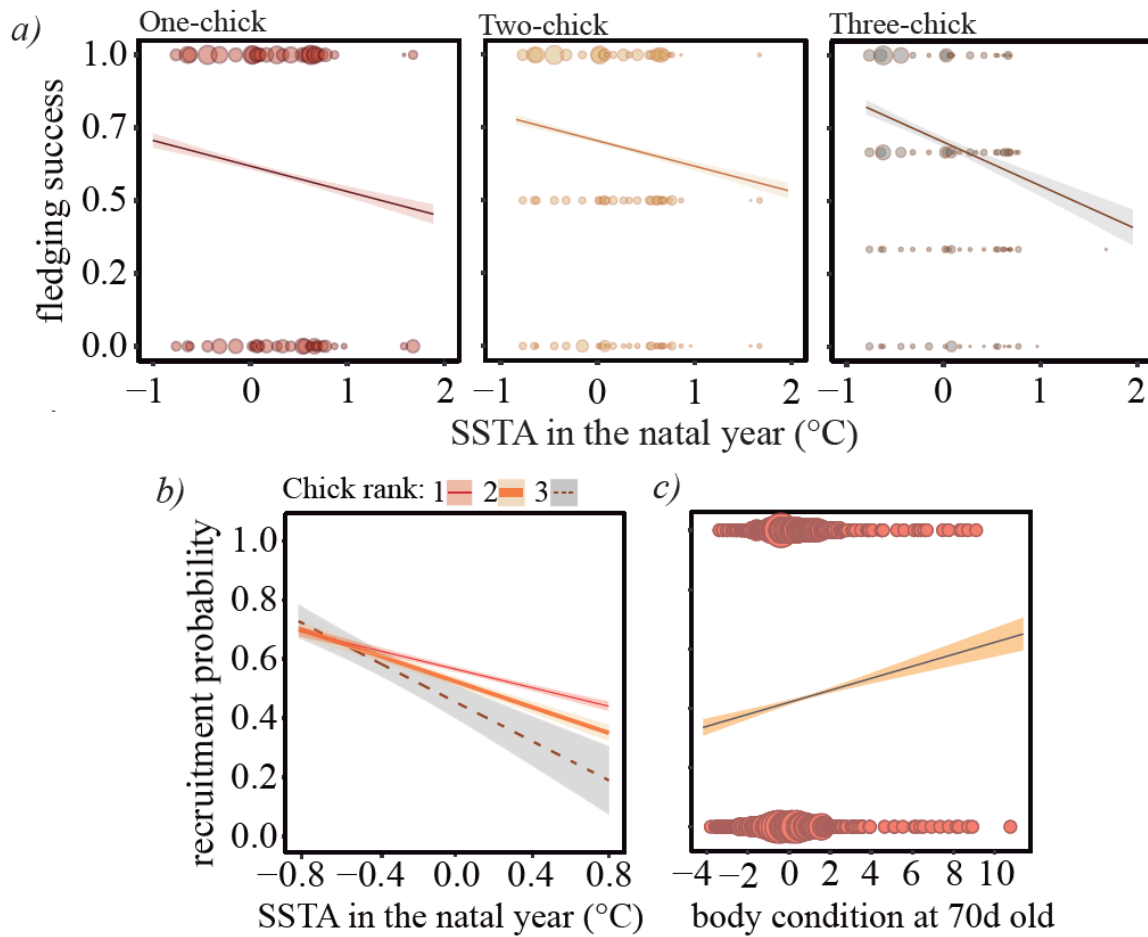
357

358 *Viability selection in El Niño conditions*

359 Both nestling and juvenile mortality increased with the warmth of natal waters, potentially
360 increasing the influence of any recruitment bias in favour of higher quality cohort members.
361 Nestlings from all three brood sizes were less likely to fledge in the warmest natal
362 condition than in the coolest (Table S17; Fig 4a), with a greater differential in three-chick
363 broods ($\beta = -0.27$ [95% CI: -0.32 to -0.22]) than in one- and two-chick broods ($\beta = -0.10$
364 [95% CI: -0.12 to -0.08] and $\beta = -0.14$ [95% CI: -0.16 to -0.12], respectively), and with 2nd
365 and 3rd hatched chicks especially liable to fail, independent of natal water temperature
366 (Table S19; Figure. 4b). Fledgling body condition was key: in all cohorts, poor condition
367 reduced the probability of recruiting ($\beta = 0.07$ [95% CI: 0.05 to 0.08]; Figure. 4c).
368 Importantly, compared to cohorts from the coolest water, 23% fewer chicks from cohorts of
369 the warmest water fledged, and 26% fewer fledglings from cohorts of the warmest water
370 went on to recruit.

371

372



373

374 **Fig. 4.** a) The proportion of chicks from 15621 broods that managed to fledge decreased
375 with the warmth of natal waters. The probability of recruiting for 10604 unsexed fledglings
376 b) decreased with the warmth of natal waters, especially for second and third chicks, and c)
377 in every year was higher for offspring that fledged with good body condition.

378

379 **Discussion**

380 Given the developmental susceptibility of birds to experimental stresses in infancy,
381 including food scarcity, compensatory growth, elevated corticosterone and parasitic
382 infection, as well as poor weather, habitats, prey availability, parental care and late fledging
383 (Drummond & Ancona, 2015), it is remarkable that the annual breeding success, offspring
384 viability, adult lifespan and lifetime reproductive success of adult booby generations are not
385 affected by exposure to El Niño in their natal year. However, it remains to be shown that
386 the viable offspring of recruits from warm water cohorts pay no reproductive costs after
387 recruitment (an intergenerational effect; Naguib & Gil, 2005; cf. Bouwhuis et al., 2015).
388 We conclude that, even though warm waters can severely prejudice breeding participation
389 of the focal population and impact not only clutch sizes but also hatching, fledging and
390 recruitment success (Ancona, Zúñiga-Vega, Rodríguez, & Drummond, 2018), as well as
391 adult survival (Oro et al., 2010), the recruits derived from warm water cohorts can
392 contribute normally to the population's growth rate.

393 We identified two mechanisms that probably contribute to warm water recruits performing
394 as well as cool water recruits. Self-selection for breeding in warm water years probably
395 makes a small contribution. High-quality boobies, boobies raised in cool water conditions,
396 and boobies of prime reproductive ages (middle aged) do not increase their representation
397 among breeders in warm years. However, young + old pairs increase their representation to
398 7% of breeding pairs in El Niño years, so their progeny could boost the average quality of
399 recruits from El Niño years provided the high quality conferred on fledglings by those
400 pairings persists into adult life. Considering the extremes, a fledgling produced by a pair
401 1yr + 21yr old is roughly 80% more likely to recruit than that of a pair 1yr + 1yr old

402 (Drummond & Rodríguez, 2015). Nonetheless if, as hinted in Figure 3, the relationship
403 between water temperature and proportion of young + old pairings is a step function rather
404 than a linear function, then that representation may ascend to roughly 20% of breeding
405 pairs in El Niño years, implying a greater potential boost to average recruit quality. Note
406 that increases in the proportions of young and old breeders in El Niño years could serve
407 functions unrelated to the benefit of pairing with opposite-age partners. For example, red-
408 footed booby (*S. sula*) pre-breeders show preference for recruiting in El Niño years and are
409 thought to do so because competition may be weaker (Cubaynes et al., 2010), and after age
410 14 years blue-footed boobies may be penalized for skipping reproduction because their
411 annual survival is in decline and increasingly variable (Drummond, Rodriguez, & Oro,
412 2011).

413 The second mechanism, viability selection on offspring, is more prevalent and probably
414 more influential than filtering of breeders. The increase in mortality among booby offspring
415 in warm water years should raise the average reproductive value of recruits to the extent
416 that mortality falls disproportionately on low quality individuals. In the nestling period,
417 dominance ranks in the brood derive from 4-day hatch intervals and selective mortality in
418 broods of two and three is dictated by hatch order rather than individual quality
419 (Drummond & Garcia Chavelas, 1989; Drummond, González, & Osorno, 1986), so
420 viability selection may be weak. By contrast, in the pre-breeding period, when juvenile
421 boobies transition to independent feeding and compete for reproductive opportunity in the
422 colony, their survival and recruitment, associated with high body condition at fledging, may
423 depend largely on individual quality. If so, the increase in fledgling mortality from roughly
424 25% in cohorts from the coolest years to 60% in cohorts from the warmest years marks an

425 increase in viability selection that could substantially raise the average phenotypic quality
426 of recruits from El Niño cohorts.

427 The explanation for reproductive equivalence of generations of recruits from El Niño
428 cohorts lies in a combination of developmental resilience of blue-footed booby offspring,
429 increased viability selection among them, especially during the pre-breeding period, and
430 increased participation of young + old breeding pairs. Developmental resilience probably
431 depends on increased investment by parents caring for late, slow-growing offspring during
432 food scarcity, and also on physiological adaptations and life history adjustments of those
433 offspring, which tend to recruit at an early age and skip more breeding years (Ancona &
434 Drummond, 2013). We found that early recruitment generally increases lifetime
435 reproductive success, and warm water cohorts obtain this benefit while avoiding both
436 increase in mortality at any adult age (Ancona et al., 2018) and the longevity cost (Table
437 S12) of reproductive investment early in life, apparently by skipping some breeding
438 seasons. Blue-footed boobies may not be alone in this ability to parry impacts of a stressful
439 natal environment by making life history adjustments. Females of the reintroduced
440 Mauritius kestrel (*Falco punctatus*) mitigate fitness penalties of an anthropogenically
441 altered natal environment by modifying their schedule of reproductive effort (Cartwright et
442 al., 2014), although it is unclear whether their lifespan is affected.

443 The prevalence among animal species of El Niño cohort effects on the reproductive value
444 of recruits is unknown, and we may find that many species have evolved ways of coping.
445 Exposure of species of the southern and eastern Pacific ocean to ENSO-like oscillations
446 since at least the mid-Holocene (Corrège et al., 2000) has probably honed the evolution of
447 developmental, behavioural and life history traits that contribute to the reproductive

448 equivalence of adult generations. Nonetheless, given the well-documented developmental
449 susceptibility of infant vertebrates to adverse natal environments and the increasing severity
450 of El Niño events, it is likely that natal El Niño conditions compromise the reproductive
451 potential of affected generations of some species, and that such effects will increase in the
452 near future. We should examine other species, to document natal El Niño impacts on
453 reproductive potential, to identify filtering effects and other means by which species
454 sidestep or resist impacts, and to explore their demographic consequences.

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466 **Author's contributions**

467 SO and HD conceived the study, HD and CR collected the data, SO analysed the data, SO
468 and HD led the writing of the manuscript. All authors contributed critically to the drafts and
469 gave final approval for publication.

470 **Data accessibility**

471 Data files are available from Dryad upon manuscript acceptance.

472 **Supporting information**

473 Addition Supporting Information may be found in the online version of this article.

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