Title page

Dual Value System and its assessment scheme for understanding and valuing ecosystem services

YANG Haile ^{a, b, 1#}, CHEN Jiakuan ^{a, b, c, *}

^a Institute of Biodiversity Science, Fudan University, Shanghai 200438, China

^b Department of Ecology and Evolutionary Biology, School of Life Sciences, Fudan University, Shanghai 200438, China

^c Centre for Watershed Ecology, Institute of Life Science, Nanchang University, Nanchang 330031, China

^{1#} Present Address: Key Laboratory of Freshwater Biodiversity Conservation, Ministry of Agriculture of China, Yangtze River Fisheries Research Institute, Chinese Academy of Fishery Sciences, 8 First Wudayuan Road, Wuhan 430223, Hubei Province, PR China

E-mail: <u>haileyang10@fudan.edu.cn</u>, <u>haileyang18@yfi.ac.cn</u> (YANG Haile); <u>jkchen@fudan.edu.cn</u> (CHEN Jiakuan)

* Corresponding author

E-mail: jkchen@fudan.edu.cn (CHEN Jiakuan)

1 Abstract

2 Valuing ecosystem services (ES) is helpful for effective ES management. However, there are 3 many limitations in traditional ES valuation approaches, including theoretical challenges and practical difficulties. To overcome these limitations, we proposed a dual value system (DVS). And 4 5 then, we presented a case study of valuing the water provision in Zhujiang River Basin (Pearl 6 River Basin) based on DVS. DVS follows the axioms that (1) human life would end if we lose any 7 of vital ES which is indispensable to human being's survival (such as oxygen, freshwater) and (2) 8 ES cannot provide any value to people without human activities. Correspondingly, DVS includes 9 two types of value: the output support value (OSV) of a vital ES refers to the total value produced 10 by human being's economic and social activities (TVPH) supported by the ES consumption; the optional capacity value (OCV) of a vital ES refers to the optional capacity of supporting TVPH 11 12 provided by total ES volume. The OCV provided by a vital ES is calculated by using the product 13 of multiplying the OSV (TVPH) by the freedom of choosing the consumption from the total 14 volume of this ES, valued in non-monetary units. Based on DVS, the OSV and OCV of water 15 provision in Zhujiang River Basin were analyzed in river basin scale and sub-basin scale, and the 16 values variation of water provision from 2006 to 2015 was analyzed in sub-basin scale. And then, based on this case study, we discussed the new insights into ES provided by DVS. Results proved 17 18 that DVS and its assessment scheme overcame the limitations on current ES valuation approaches 19 and provided an innovative quantitative framework to understand and value ES which will help to make good decisions in ES management. 20

21

22 Keywords

- 23 Ecosystem services valuation; dual value system; output support value; optional capacity value;
- 24 ecosystem services spatial subsidy; Pearl River Basin
- 25

ES: ecosystem service(s)

OSV: output support value

OCV: optional capacity value

DVS: dual value system

GDP: gross domestic product

TVPH: total value produced by human being's economic and social activities

PRWRC: Pearl River Water Resources Commission of the Ministry of Water Resources

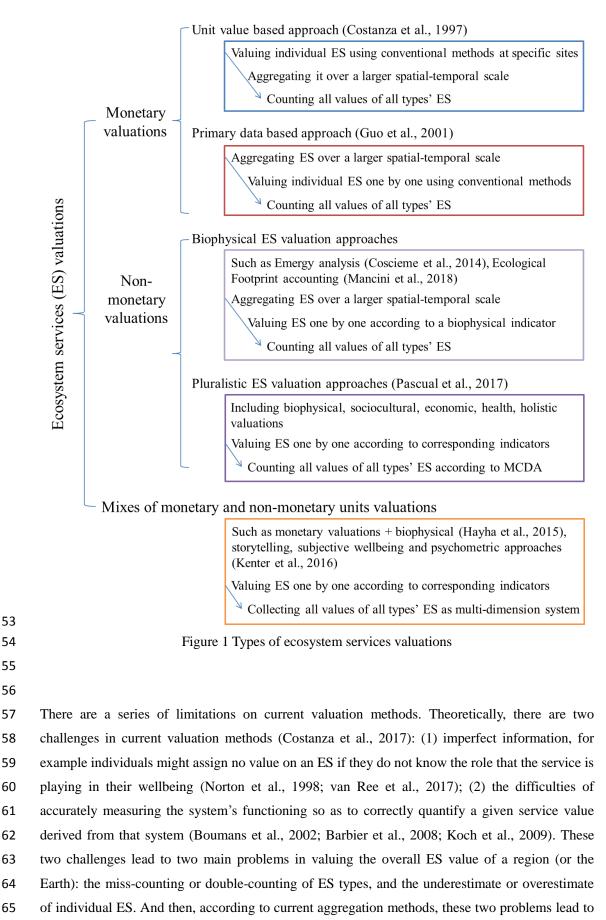
26 **1. Introduction**

Ecosystems provide a range of services that are fundamentally important for human wellbeing,
health, livelihoods, and survival (Costanza et al., 1997; Millennium Ecosystem Assessment, 2005;
TEEB Synthesis, 2010). To unravel the complex socio-ecological relationships and explicate how
human decisions would affect ecosystem services (ES), the values of ES are used to express these
changes which allow for their incorporation in public decision-making processes (Farley and
Costanza, 2010; Pascual et al., 2010; Costanza et al., 2017).

33

34 The valuation of ES could be monetary, non-monetary, or mixes of monetary and non-monetary (Figure 1). In monetary ES valuation studies, there are two approaches (Jiang, 2017): (1) unit 35 36 value based approach (i.e. benefit transfer) (Costanza et al., 1997; de Groot et al., 2012; Costanza 37 et al., 2014; Kubiszewski et al., 2017) and (2) primary data based approach (Ouyang et al., 1999; Guo et al., 2001; Dai et al., 2016) (Figure 1). In non-monetary ES valuation studies, there are 38 39 some biophysical approaches (Koellner and Geyer, 2013; Coscieme et al., 2014; Mancini et al., 40 2018), such as Emergy analysis, Ecological Footprint accounting; and some pluralistic approaches (Pascual et al., 2017; Martin and Mazzotta, 2018; Folkersen, 2018), such as the pluralistic 41 42 valuation approach provided by Intergovernmental Science-Policy Platform on Biodiversity and 43 Ecosystem Services (IPBES). IPBES considers that the ES (redefined as nature's contributions to 44 people) and a good quality of life are interdependent, and then processes biophysical, sociocultural, economic, health and holistic valuations (Pascual et al., 2017) (Figure 1). The approaches of 45 valuing ES using mixes of monetary and non-monetary units are diverse (Hayha et al., 2015; 46 47 Kenter et al., 2016) (Figure 1). In the case provided by Hayha et al. (2015), the ES are assessed in 48 biophysical and monetary units (Hayha et al., 2015). Kenter et al. (2016) integrate deliberative monetary valuation, storytelling, subjective wellbeing and psychometric approaches to 49 50 comprehensively elicit cultural ecosystem services values (Kenter et al., 2016).

51



a more serious underestimate or overestimate on the overall ES value of a region (or the Earth).

67 Practically, there are two difficulties in current valuation methods: (1) most of the current monetary ES valuation approaches are based on the conventional valuation techniques which are 68 diverse and limited (Pascual et al., 2010; Folkersen, 2018); (2) most of the current non-monetary 69 70 ES valuation approaches use multi-indicators (Pascual et al., 2017; Martin and Mazzotta, 2018), 71 which make valuation approaches complex and costly. To solve the amplified underestimate or 72 overestimate on the overall ES value of a region (or the Earth) caused by theoretical challenges 73 and current aggregation approaches, and to avoid the practical difficulties in valuing ES, we need 74 to propose a new assessment framework.

75

76 In this work, we firstly proposed a dual value system (DVS), including two new concepts: output 77 support value (OSV) and optional capacity value (OCV). And then, we presented a case study of 78 analyzing the OSV and OCV of water provision in Zhujiang River Basin (Pearl River Basin) to 79 demonstrate the application of DVS. In this case study, to explore the new insights of DVS into 80 ES and the innovative assessment scheme of DVS, we (1) analyzed the OSV and OCV of water 81 provision at two spatial scales: river basin scale and sub-basin scale, (2) analyzed the OCV of the 82 passing-by water among hydrologic units at the sub-basin scale, (3) analyzed the OSV and OCV 83 variation of the water provision from 2006 to 2015.

84

85 2. Methods and materials

86 2.1 Methods details: the DVS for valuing ES

87 Chaisson (2002) has argued that the global value of ES is infinite, as we human beings could not survive if we lost any vital ES which is indispensable to human being's survival (such as oxygen, 88 89 freshwater). The truth indicated by him shows that the total value produced by human being's 90 economic and social activities (TVPH) relies on each vital ES. Costanza et al. (2014) have 91 indicated that ecosystems could not provide any values to people without the presence of people 92 (human capital), their communities (social capital), and their built environment (built capital). In 93 other words, the value of each ES depends on the level of socioeconomic development. Here we 94 get two inferences based on these two axioms.

95

The first inference is that the consumption of vital ES supports the TVPH. To describe this inference, we proposed a new concept which is termed as "Output Support Value (OSV)". OSV of a vital ES refers to the TVPH supported by the ES consumption, which shows the benefits for human wellbeing derived from consuming a definite volume of ES, directly or indirectly.

100

101 The second inference is that the total volume of a vital ES provides the freedom of consuming 102 corresponding ES in maintaining the human being's production and survival. To describe this 103 inference, we proposed a new concept which is termed as "Optional Capacity Value (OCV)".

104 OCV of a vital ES refers to the optional capacity of supporting the OSV (i.e. TVPH) provided by

the total volume of ES, which shows the benefits for human wellbeing derived from having the

106 option of using, directly or indirectly. OCV is described as the product of multiplying the OSV (i.e.

- 107 TVPH) by the freedom of choosing the consumption from the total volume of ES.
- 108

Based on these two inferences and two new concepts, we constructed a brand new assessment
scheme for valuing ES – "dual value system (DVS)", in which OSV was valued in monetary units
and OCV was valued in non-monetary units.

112

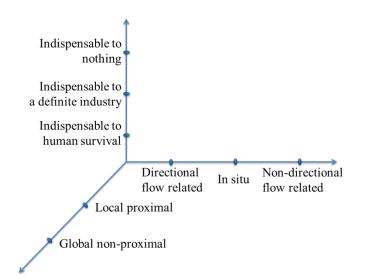
113 For valuing ES, we need first to classify ES in an appropriate way. The classification of ES mainly has two approaches (Hayha and Franzese, 2014). The first approach is the classification mainly 114 115 based on four categories: (1) provisioning, (2) regulating, (3) supporting, and (4) cultural services, 116 which is the most commonly used classification (Millennium Ecosystem Assessment, 2005; TEEB Synthesis, 2010; CICES – Haines-Young and Potschin, 2013). And the second approach is the 117 classification mainly based on spatial characteristics of ES (Costanza, 2008; Fisher et al., 2009). 118 119 To satisfy the needs of valuing ES in DVS, we proposed a classification of ES based on indispensability and spatial characteristics of ES (Table 1) (Figure 2), following the second 120 121 approach.

- 122
- 123

124 Table 1 Classification of ecosystem services based on indispensability and spatial characteristics,

125 modified from Hayha and Franzese (2014)

Classification	Example
Indispensable to human survival	Basic services for human survive; e.g. fresh water, oxygen
Indispensable to a definite industry	Basic services for an industry; e.g. honeybee
Indispensable to nothing	Services that human cannot use; e.g. hydrothermal vent ecosystem
Global non-proximal	Does not depend on proximity; e.g., carbon sequestration
Local proximal	Depends on proximity; e.g., disturbance regulation
Directional flow related	Flow from point of production to point of use; e.g., water supply
In situ	Point of use; e.g., soil formation
Non-directional flow related	Flow from point of production to adjacent area of use; e.g., climate regulation



127

Figure 2 Three-dimensional schematic diagram for the classification of ecosystem services based
 on indispensability and spatial characteristics. For example, oxygen is indispensable to human
 survival, global non-proximal, non-directional flow related; fresh water is indispensable to human

survival, local proximal, directional flow related.

132 133

131

In our classification of ES (Figure 2), indispensability means that if we lose any vital ES (indispensable to human survival, such as oxygen, freshwater), human life would end; if we lose crucial ES (indispensable to definite industries, such as honeybee), all industries based on this ES would be vanished. In DVS, we only value the currently usable ES. And an individual ES is valued only based on current socioeconomic status.

139

140 Regarding the spatial characteristics, there are ES that are not dependent on the specific location 141 (e.g., global carbon sequestration, global oxygen production), ES that are instead dependent on 142 their spatial distribution in relation to human presence (e.g., waste treatment and storm protection), 143 ES in which the direction of the flow from upstream to downstream matters (e.g., water supply 144 and sediment transport), ES in which the flow without definite direction matters (e.g., fresh air 145 supply and climate regulation), and ES in which no flow matters (e.g., soil formation).

146

Following our classification of ES (Figure 2), OSV of the ES which is indispensable to human survival could be valued using TVPH; OSV of the ES which is indispensable to definite industries could be valued using the TVPH based on these industries (Table 2). OCV of an ES is calculated as the product of multiplying its OSV by the freedom of choosing the ES consumption from the total volume of ES. The freedom is evaluated by the average uncertainty which could be described by log base 2 which indicates the uncertainty in a binary decision, and is valued in bits (Ulanowicz, 1986) (Table 2).

155

156 Table 2 The Output Support Value (OSV) and Optional Capacity Value (OCV) of an ecosystem

157 service (ES)

Ecosystem service (ES)	Output Support Value (OSV)	Optional Capacity Value (OCV)
Indispensable to human survival	OSV = TVPH	$OCV = OSV * \log_2 \frac{ES_t}{ES_c}$
Indispensable to a definite industry	<i>OSV</i> = TVPH based on corresponding industry	$OCV = OSV * \log_2 \frac{ES_t}{ES_c}$

158 In the equation, ES_t denotes the total volume of an ecosystem service in a region; ES_c denotes 159 the consumed volume of an ecosystem service in maintaining the production and survival, or in 160 supporting a definite industry; TVPH denotes the total value produced by human being's economic 161 and social activities.

- 162
- 163

Regarding the spatial characteristics, we take regions as the accounting units. In another word, 164 165 each accounting unit provides a definite ES value. As the global non-proximal ES supports the 166 global human activities, it is valued by global TVPH. As the local proximal ES only supports the 167 local human activities, it is valued by local TVPH. For the ES with (directional or non-directional) flow, both the local ES in a region and the imported ES come from other regions (accounting units) 168 support human activities in this region. Following the principle of local priority, the value of 169 170 imported ES is assigned as the value that the value of total ES (sum of the local ES and the imported ES) minus the value of local ES (Eq. 1, 2, 3, 4). For the ES in which no flow matters, its 171 172 value is the value of local ES.

- 173
- 174

175 $OSV_{l} = \begin{cases} TVPH, ES_{c} < ES_{l} \\ TVPH \times \frac{ES_{l}}{ES_{c}}, ES_{c} \ge ES_{l} \end{cases}$ (Eq. 1)

176

177
$$OCV_{l} = \begin{cases} TVPH \times \log_{2} \frac{ES_{l}}{ES_{c}}, ES_{c} < ES_{l} \\ TVPH \times \log_{2} \frac{(ES_{l} + ES_{s})}{ES_{c}} \times \frac{ES_{l}}{ES_{c}}, ES_{s} > ES_{s} \end{cases}$$
(Eq. 3)

 $OSV_s = TVPH - OSV_l$ (Eq. 2)

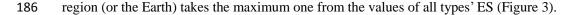
$$(IVPH \times \log_2 \frac{1}{ES_c} \times \frac{1}{ES_c}, ES_c \ge ES_l$$

$$OCV_s = TVPH \times \log_2 \frac{(ES_l + ES_s)}{ES_c} - OCV_l \quad (Eq. 4)$$

179 In the equations, OSV_l denotes the OSV of local ecosystem service (ES) in a region; OSV_s 180 denotes the OSV of ES spatial inflow in a region; OCV_l denotes the OCV of local ES in a region; 181 ES_c denotes the consumed volume of ES in a region; TVPH denotes the total value produced by 182 human being's economic and social activities.

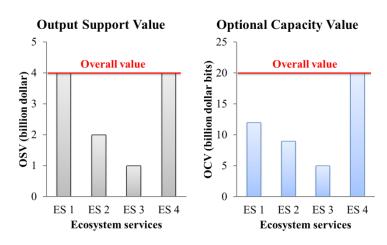
- 183
- 184

185 In the framework of DVS, all types of ES could be valued parallelly. The overall value of ES in a



187

188



189

Figure 3 Schematic representations of valuing the overall value of multi-types' ecosystem services
(ES). ES 1 represents an ecosystem service which is indispensable to human survival; and ES 4
represents another ecosystem service which is indispensable to human survival; ES 2 represents an
ecosystem service which is indispensable to a definite industry; ES 3 represents another ecosystem
service which is indispensable to another definite industry. The values in these figures are all
fictitious.

196 197

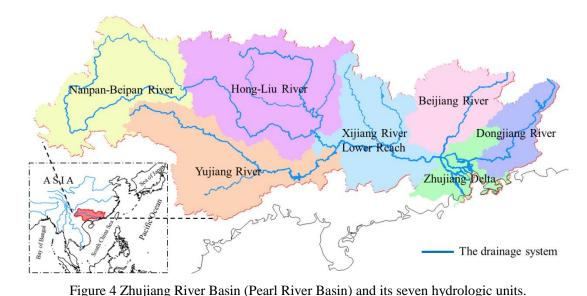
198 2.2 Study area: the Zhujiang River Basin

In DVS framework, to value ES, one needs three sets of data with spatial-temporal consistency: total ES volume, ES consumption and TVPH. Moreover, to discuss the new insights of DVS into ES value, such as the scale-specificity of DVS in valuing ES and the efficiency of DVS in valuing the ES fluxes, the case study area should be nested divided and each subdivision has the same data sets, the data on ES fluxes among subdivisions should be available. The water provision in Zhujiang River Basin meets these conditions.

205

206 Here, we presented a case study of valuing the water provision in Zhujiang River Basin based on 207 DVS framework. As water resources are indispensable to human survival and local proximal, 208 following DVS, we analyze the value of water provision based on total water resources volume, 209 water consumption and TVPH. Moreover, as water resources are directional flows, we delineate 210 the donors and recipients of water resources fluxes (passing-by water), and then value each fluxes 211 based on the volume of passing-by water, water consumption and TVPH of recipients. In Zhujiang 212 River Basin, the Zhujiang River Water Resources Bulletin (1) provides the total water resources 213 volume, water consumption and gross domestic product (GDP, here indicates TVPH) of the overall river basin, (2) divides the Zhujiang River Basin into seven hydrologic units: the 214 215 Nanpan-Beipan River, the Hong-Liu River, the Yujiang River, the Xijiang River Lower Reach, the

- Beijiang River, the Dongjiang River and the Zhujiang Delta (Figure 4), (3) provides the total water 216
- resources volume, water consumption and GDP of each hydrologic units and (4) provides the data 217
- for calculating the volume of passing-by water among hydrologic units, and the water 218 219 consumption and GDP of recipients (PRWRC, 2017).
- 220
- 221





223

224

225

226 As a case study for demonstrating the application of DVS, in valuing the water provision in 227 Zhujiang River Basin, we only consider the volume of water resources rather than its environment 228 and temporal distribution for brevity. All data that we use are available from the Zhujiang River 229 Water Resources Bulletin provided by the Pearl River Water Resources Commission of the

230 Ministry of Water Resources (PRWRC) (http://www.pearlwater.gov.cn/xxcx/szygg/).

231

3. Results 232

233 3.1 The value of the water provision at river basin scale

234 Firstly, we valued the water provision at river basin scale by taking the Zhujiang River Basin as 235 the accounting unit. In 2015, as the total volume of water resources in Zhujiang River Basin is 236 410.04 billion m³; the water usage 60.43 billion m³ and the GDP 8.5105 trillion yuan (PRWRC, 2017), the OSV of the water provision in Zhujiang River Basin is 8.5105 trillion yuan which indicates that 237 238 the water usage of 60.43 billion m³supports the GDP of 8.5105 trillion yuan; the OCV of the water provision in Zhujiang River Basin is 23.5096 trillion yuan bits which indicates that the total water 239 240 provision of 410.04 billion m³ provides the optional capacity for supporting the TVPH (measured 241 by GDP) as much as 23.5096 trillion yuan bits (Table 3).

242

Table 3 The Output Support Value (OSV) and Optional Capacity Value (OCV) of water provision

	Total volume (e9 m ³)	Usage (e9 m ³)	GDP (e12 yuan)	OSV (e12 yuan)	<i>OCV</i> (e12 yuan bits)
Zhujiang River Basin	410.04	60.43	8.5105	8.5105	23.509

provided and experienced by Zhujiang River Basin at river basin scale in 2015

246 247

248 **3.2** The value of the water provision at sub-basin scale

Secondly, we valued the water provision at sub-basin scale by taking seven major hydrologic units of the Zhujiang River Basin as the accounting units. For each hydrologic unit, both the passing-by water and the local water yield provide the optional capacity for supporting the TVPH (measured by GDP). Following the principle of local priority, in 2015, the Hong-Liu River, Xijiang River Lower Reach and Zhujiang Delta respectively receive passing-by water 37.40 billion m ³ 200.28 billion m ³ and 360.20 billion m ³ which provide OCV 0.1935 trillion yuan bits, 1.0332 trillion yuan bits and 18.7463 trillion yuan bits, respectively (Table 4).

256 257

Table 4 The Output Support Value (OSV) and Optional Capacity Value (OCV) of water provision

experienced by each hydrologic unit of Zhujiang River Basin at sub-basin scale in 2015

	Total volume (e9 m ³)	Usage (e9 m ³)	GDP (e12 yuan)	OSV (e12 yuan)	<i>OCV</i> (e12 yuan bits)
Nanpan-Beipan River	40.00	4.77	0.5817	0.5817	1.7846
Hong-Liu River	126.37	9.78	0.5175	0.5175	1.9103
Passing-by water	37.40				0.1935
Yujiang River	44.14	7.86	0.5240	0.5240	1.3045
Xijiang River Lower Reach	79.92	10.79	0.5709	0.5709	1.6492
Passing-by water	200.28				1.0332
Beijiang River	61.68	5.26	0.3757	0.3757	1.3344
Dongjiang River	27.40	4.64	0.8436	0.8436	2.1614
Zhujiang Delta	30.53	17.33	5.0971	5.0971	4.1641
Passing-by water	360.20				18.7463
Sum	1007.92	60.43	8.5105	8.5105	34.2816

²⁶⁰

261

Distributing the OCV of the passing-by water into every hydrologic unit which exports water resources, we calculated the OCV of the water provision provided by each hydrologic unit, including local service and output service (Table 5). The result showed the OCV of the point-to-point water provision fluxes (Table 5).

266

267

268 Table 5 The Optional Capacity Value (OCV) of water provision provided by each hydrologic unit

		OSV	(e12 yuan bits)		
	Local service	Output to Hong-Liu River	Output to Xijiang River Lower Reach	Output to Zhujiang Delta	Sum
Nanpan-Beipan River	1.7846	0.1935	0.1929	1.9465	4.117
Hong-Liu River	1.9103		0.6311	6.3671	8.908
Yujiang River	1.3045		0.2091	2.1099	3.623
Xijiang River Lower Reach	1.6492			3.8976	5.546
Beijiang River	1.3344			3.0857	4.420
Dongjiang River	2.1614			1.3396	3.501
Zhujiang Delta	4.1641				4.164

269 of Zhujiang River Basin in 2015

271

272 3.3 The value variation of the water provision from 2006 to 2015 at sub-basin scale

273 Thirdly, we compared the OSV and OCV of the water provision provided by each hydrologic unit

in 2006 and 2015. Results showed that the OCV of the water provision of both the local and the

exported services provided by each hydrologic unit increase with the economic growth (Table 6).

Table 6 The Output Support Value (OSV) and Optional Capacity Value (OCV) of local and passing-by water provision provided by each hydrologic unit of Zhujiang

River Basin in 2006 and 2015

	GDPWater resources(e12 yuan)(e9 m 3)		Water usage OSV (a12 yrsp)			OCV (e12 yuan bits)								
			(e9 m ³)		(e9 m ³)		OSV (e12 yuan)		Local service		Output service		Su	ım
	2006	2015	2006	2015	2006	2015	2006	2015	2006	2015	2006	2015	2006	2015
Nanpan-Beipan River	0.1577	0.5817	29.66	40.00	4.62	4.77	0.1577	0.5817	0.4231	1.7846	0.6269	2.3329	1.0500	4.1176
Hong-Liu River	0.1481	0.5175	87.35	126.37	10.34	9.78	0.1481	0.5175	0.4560	1.9103	1.7373	6.9982	2.1934	8.9085
Yujiang River	0.1442	0.5240	36.95	44.14	7.29	7.86	0.1442	0.5240	0.3374	1.3045	0.6981	2.3190	1.0355	3.6235
Xijiang River Lower Reach	0.1765	0.5709	70.74	79.92	11.82	10.79	0.1765	0.5709	0.4555	1.6492	1.2427	3.8976	1.6982	5.5468
Beijiang River	0.1055	0.3757	64.20	61.68	5.36	5.26	0.1055	0.3757	0.3780	1.3344	1.1734	3.0857	1.5514	4.4201
Dongjiang River	0.2837	0.8436	38.32	27.40	4.20	4.64	0.2837	0.8436	0.9051	2.1614	0.6961	1.3396	1.6011	3.5010
Zhujiang Delta	1.7566	5.0971	34.26	30.53	19.50	17.33	1.7566	5.0971	1.4286	4.1641			1.4286	4.1641
Sum	2.7723	8.5105	361.49	410.04	63.13	60.43	2.7723	8.5105	4.3836	14.3085	6.1745	19.9731	10.5581	34.2816

284 4. Discussion

285 4.1 New insights of DVS

286 DVS provides a new perspective to understand the value of ES. In 2015, in Zhujiang River Basin 287 at the river basin scale, the OSV of the water provision (8.5105 trillion yuan) means that a definite 288 volume water usage (60.43 billion m 3 supports a definite TVPH (8.5105 trillion yuan, measured 289 by GDP) (Table 3). The OCV of the water provision (23.5096 trillion yuan bits) means that the volume of total water provision (410.04 billion m³) provides the optional capacity for supporting 290 291 the TVPH (23.5096 trillion yuan bits, measured by GDP) (Table 3). DVS makes the importance 292 and value of ES easy be understood through direct value (OSV) and indirect value (OCV) in the 293 perspective different from traditional exchange value.

294

DVS provides new insight into the relationship between the ES value and the social and economic development (Table 2). From 2006 to 2015, the OSV and OCV of the water provision in seven accounting units of Zhujiang River Basin increase with their TVPH (here measured by GDP) (Table 6). This is consistency with the point in previous researches (Costanza et al., 2014). Conclusively, the social and economic development depends on the ES and the ES value increases with the social and economic development. Based on this, we would not simply make the social and economic development against with the ecological conservation.

302

303 DVS provides new insight into how to understand and quantify the value of ES spatial subsidies (i.e. 304 ES fluxes from one region to another region). The spatial subsidies of ES increase (1) the ES OCV 305 experienced by the recipient accounting units and (2) the ES OCV provided by the donor 306 accounting units, and then (3) the social and economic development of recipient accounting units 307 increase the ES OCV provided by the donor accounting units. In Zhujiang River Basin, according 308 to the passing-by water, (1) the recipients receive more OCV (Table 4), (2) the donors provide more 309 OCV (Table 5), and (3) the economic growths in recipients promote the increase of the OCV provided by donors (Table 6). Based on this framework, we could evaluate the ES value variation in different 310 311 scenarios of ecological conservation and natural resources exploitation in a region, and then 312 optimize the plans for ecological conservation and natural resources exploitation.

313

314 DVS provides a new method to quantify payments for ecosystem services (PES). ES spatial subsidies 315 provide OCV to recipients (accounting unit), then recipients should provide corresponding ecological compensation to donors (Yang et al., 2019). The compensation standard could reference to the 316 317 OCV of ES spatial subsidies. In 2015, the hydrologic unit of Nanpan-Beipan River provides the OCV of the water provision 1.7846 trillion yuan bits, 0.1935 trillion yuan bits, 0.1929 trillion yuan 318 319 bits, 1.9465 trillion yuan bits to the Nanpan-Beipan River, Hong-Liu River, Xijiang River Lower 320 Reach and Zhujiang Delta respectively (Table 5). Then, if a water resources protection program in 321 Nanpan-Beipan River costs 4.1176 billion yuan, the Nanpan-Beipan River, Hong-Liu River, Xijiang

River Lower Reach and Zhujiang Delta should pay 1.7846 billion yuan, 0.1935 billion yuan, 0.1929
billion yuan, 1.9465 billion yuan respectively.

324

325 DVS provides a new framework for ES management. Following DVS, one could (1) identify the features of an individual ES, (2) identify the accounting units appropriately, (3) collect 326 327 corresponding data sets, and then (4) calculate the OSV and OCV of this ES (Table 2). Based on 328 the ES OCV, one could discuss the ES management in each region and the PES among regions 329 (Yang et al., 2019). If one values a set of ES in a region with limited space and resources, one 330 could estimate the overall ES value in this region following DVS (Figure 3), and then discriminate 331 between different services to decide what should be managed and what should be traded to satisfy 332 human needs/ wants for human wellbeing. In this case study, we valued the water provision in 333 Zhujiang River Basin using the total water provision volumes, water consumptions and TVPH (indicated by GDP) at basin scale and sub-basin scale respectively. This method could be used in 334 335 any other river basins and regions.

336

In DVS, there are intrinsic consistency among the OSV at different scales and intrinsic 337 338 incommensurability among the OCV at different scales. OSV of an ES means the TVPH supported by the consumed ES, and the TVPH and consumed ES in a certain space-time condition are 339 constant, which is independent of spatial and temporal scales, so OSV are intrinsically consistent 340 341 at different scales. In contrast, OCV of an ES means the optional capacity of supporting the TVPH 342 provided by the total ES, and the spatial subsidies of an ES among accounting units increase the 343 total volume of ES in every accounting unit, their OCV are intrinsically incommensurable at 344 different scales. Moreover, the spatial distributions of the TVPH, ES and ES consumption are 345 mismatched, and OCV of an ES is calculated by these three indicators non-linearly (Table 2), so 346 their OCV (even only considering the local services) are intrinsically incommensurable at different 347 scales.

348

349 4.2 Advantages of DVS

DVS provides a simple and general assessment scheme for valuing ES. Firstly, DVS avoids the 350 351 practical difficulties in traditional ES valuation approaches, such as the difficulty of re-estimating a large amount of unit values and the difficulty of accurately valuing each individual ES. Secondly, 352 353 the misestimate of ES value in DVS is traceable and easy-to-correct, although DVS does not completely avoid the underestimate or overestimate on individual ES led by insufficient 354 355 information and misunderstanding on an ES. Thirdly, DVS provides a general approach and a 356 general unit to value ES, avoids the difficulties for transforming and aggregating incommensurable data which come from the values of different type's ES. Fourthly, DVS avoids 357 358 the amplification of underestimate or overestimate on the overall ES value of a region (or the 359 Earth), although DVS do not completely avoid the miss-counting or double-counting of ES types

360 led by insufficient information.

361

362 DVS avoids the practical difficulties in traditional ES valuation approaches. The value of ES 363 depends on the level of socioeconomic development (Costanza et al., 2014). So, following the unit value based approach, one should value the 2011 ES using the 2011 unit values (Costanza et al., 364 365 2014), value the 2050 ES using the 2050 unit values (Kubiszewski et al., 2017), rather than value the 2011 ES using the 1997 unit values (Costanza et al., 2014), value the 2050 ES using the 2011 366 unit values (Kubiszewski et al., 2017). However, re-estimating all unit values is a heavy work, 367 368 especially, considering the non-linearity in ecosystem services (Barbier et al., 2008; Koch et al., 369 2009) and the spatial heterogeneity of unit values (de Groot et al., 2012; Crossman et al., 2013; 370 Drakou et al., 2015). In the primary data based approach, the difficulty for accurately valuing each 371 individual ES is remained (Guo et al., 2001; Dai et al., 2016), although there are some technique innovations in recent years (Coscieme et al., 2014). In DVS, the OSV of an ES is determined by 372 the TVPH and ES consumption; the OCV of an ES is determined by total ES volume, ES 373 374 consumption and OSV (Table 2). In this case study, the TVPH was indicated by GDP which is the 375 only indicator of indicating the TVPH in all current social and economic indexes. If someone gets 376 another indicator which could indicate the TVPH more accurately, he could use that indicator to replace GDP in his study. This replacement has no impact on the assessment scheme of DVS. In 377 this work, the freedom was evaluated by the average uncertainty of selecting ES consumption 378 379 from the total volume of this ES. The average uncertainty was described by log base 2 which 380 indicated the uncertainty in a binary decision (Ulanowicz, 1986). If someone believes that another 381 indicator could indicate the average uncertainty more appropriately, he could use that one to replace log base 2 in his study. This replacement does not change the assessment scheme of DVS. 382 So, DVS reflects the level of economic development and avoids the difficulties in unit value based 383 384 approach and primary data based approach.

385

386 DVS makes the misestimate of ES value traceable and easy-to-correct. In traditional approaches, 387 such as direct market valuation methods, revealed preference methods and stated preference 388 methods, the valuation of different types of ES depends on different rules and their underestimate 389 and overestimate are difficult to identify and correct (Pascual et al., 2010). The ES valuation 390 method in DVS (Table 2) makes the underestimate or overestimates of individual ES value to be 391 traceable and easy-to-correct, although DVS does not completely avoid the underestimate or 392 overestimate on individual ES led by insufficient information and misunderstanding on an ES.

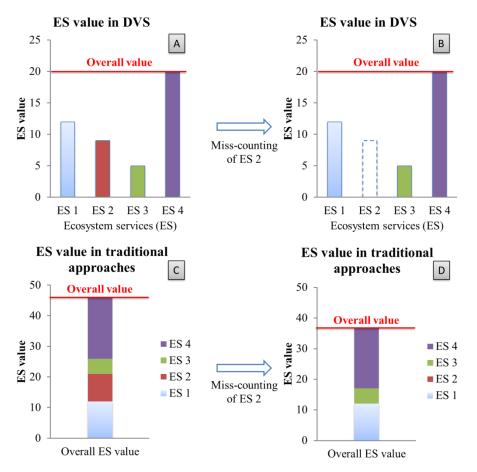
393

394 DVS provides a general approach and a general unit to value ES. There is a great difficulty in 395 aggregating the non-monetary values of ES because the non-monetary ES values could be 396 qualitative or quantitative, and are valued by diverse units/ indicators for different ES. For 397 resolving this problem, someone tries the Multi-Criteria Decision Analysis (MCDA) so as to

transform the incommensurable data (i.e., monetary and non-monetary values) into non-monetary, 398 dimensionless values, and then to mathematically incorporate non-monetary features into the 399 aggregation (Martin and Mazzotta, 2018). However, one needs to provide some assumptions if he/ 400 401 she wants to transform a set of incommensurable data and aggregate them. Different researchers 402 provide different assumptions, and their aggregating results are sensitive to their assumptions 403 (Martin and Mazzotta, 2018). Moreover, the transforming and aggregating approaches are very complex. In DVS, we calculate the OCV of all types' ES using a general approach and a general 404 unit. So, DVS avoids the difficulties for transforming and aggregating incommensurable data 405 406 which come from the values of different type's ES.

407

408 DVS avoids the amplification of underestimate or overestimate on the overall ES value of a region 409 (or the Earth). The common problem of valuing the total economic value of ES in a region (or the Earth) in both the unit value based approach and the primary data based approach is the amplified 410 underestimate or overestimate caused by the miss-counting or double-counting of ES types (Fu et 411 al., 2011; Stoeckl et al., 2014; van Ree et al., 2017), as their aggregating approaches are counting 412 413 each value of all types' ES (Costanza et al., 2014; Dai et al., 2016) (Figure 5). Stoeckl et al. (2014) 414 try to consider the complexity and non-linearity of ecosystems and use statistical techniques to 415 identify and control their overlapping benefits, but their methods and estimates still have some limitations. In DVS, the total value of ES in a region (or the Earth) takes the maximum one from 416 the values of all types' ES (Figure 3, 5). So, the severities of underestimate or overestimate on the 417 overall ES value in a region (or the Earth) led by miss-counting or double-counting of ES types in 418 ES valuation would be relieved in DVS, although DVS do not completely avoid the miss-counting 419 420 or double-counting of ES types led by insufficient information (Figure 5). 421



423

Figure 5 The schematic diagrams for the severities of underestimate on the overall ES value in a
region (or the Earth) led by the miss-counting of ES 2 in two ES valuation approaches. The
miss-counting of ES 2 does not lead a severe underestimate on the overall ES value in DVS (from
A to B), but does in traditional approaches (from C to D). The values in these figures are all
fictitious.

- 429
- 430

431 **5. Conclusions**

In this paper, we proposed a dual value system (DVS), a new assessment scheme to value ES.
DVS includes two types of ES value: (1) Output Support Value (OSV) means the TVPH supported
by the consumed ES, and (2) Optional Capacity Value (OCV) means the optional capacity of
supporting the TVPH provided by the total ES. The overall value of ES in a region (or the Earth)
takes the maximum one from the values of all types' ES.

437

438 DVS provides new insight to understand and value ES:

(1) DVS makes the importance and value of ES easy be understood in new perspective;

440 (2) DVS concisely shows the relationship between ES value and the social and economic441 development;

442 (3) DVS provides a tool for a new PES approach;

443	(4) DVS provides a new framework for ES management;
444	(5) DVS has the scale-specificity.
445	
446	DVS overcomes the limitations on current ES valuation approaches:
447	(1) DVS avoids the practical difficulties in traditional ES valuation approaches;
448	(2) DVS makes the misestimate of ES value traceable and easy-to-correct;
449	(3) DVS provides a general approach and a general unit to value ES;
450	(4) DVS avoids the amplification of underestimate or overestimate on the overall ES value of
451	a region.
452	
453	DVS provides a useful quantifying framework for coordinating ecological protection and
454	ecological compensation.
455	
456	Acknowledgments
457	We thank Li Yongchi, Li Zhaolei, Yang Yuanyuan and Zhao Bin for grateful assistance that helps
458	to improve the manuscript. This research has not received any specific grant from funding
459	agencies in the public, commercial, or not-for-profit sectors.
460	
461	References
462	Barbier E B, Koch E W, Silliman B R, Hacker S D, Wolanski E, Primavera J, Granek E F, Polasky S,
463	Aswani S, Cramer L A, Stoms D M, Kennedy C J, Bael D, Kappel C V, Perillo G M E, Reed D J.
464	2008. Coastal ecosystem-based management with nonlinear ecological functions and values.
465	Science, 319: 321-323 https://doi.org/10.1126/science.1150349
466	Boumans R, Costanza R, Farley J, Wilson M A, Portela R, Rotmans J, Villa F, Grasso M. 2002.
467	Modeling the dynamics of the integrated earth system and the value of global ecosystem services
468	using the GUMBO model. Ecol Econ, 41: 529-560 https://doi.org/10.1016/S0921-8009(02)00098-8
469	Chaisson E. 2002. Cosmic Evolution: The Rise of Complexity in Nature. Cambridge, Mass.: Harvard
470	University Press
471	Coscieme L, Pulselli F M, Marchettini N, Sutton P C, Anderson S, Sweeney S. 2014. Emergy and
472	ecosystem services: A national biogeographical assessment. Ecosyst Serv, 7: 152-159
473	https://doi.org/10.1016/j.ecoser.2013.11.003
474	Costanza R. 2008. Ecosystem services: Multiple classification systems are needed. Biol Conserv, 141:
475	350-352 <u>https://doi.org/10.1016/j.biocon.2007.12.020</u>
476	Costanza R, DArge R, DeGroot R, Farber S, Grasso M, Hannon B, Limburg K, Naeem S, ONeill R V,
477	Paruelo J, Raskin R G, Sutton P, VandenBelt M. 1997. The value of the world's ecosystem services
478	and natural capital. Nature, 387: 253-260 <u>https://doi.org/10.1038/387253a0</u>
479	Costanza R, de Groot R, Sutton P, van der Ploeg S, Anderson S J, Kubiszewski I, Farber S, Turner R K.
480	2014. Changes in the global value of ecosystem services. Global Environ Chang, 26:152-158 19

481	https://doi.org/10.1016/j.gloenvcha.2014.04.002
482	Costanza R, Groot R D, Braat L, Kubiszewski I, Fioramonti L, Sutton P, Farber S, Grasso M. 2017.
483	Twenty years of ecosystem services: How far have we come and how far do we still need to go?
484	Ecosyst Serv, 28: 1-16 https://doi.org/10.1016/j.ecoser.2017.09.008
485	Crossman N D, Burkhard B, Nedkov S, Willemen L, Petz K, Palomo I, Drakou E G, Martin-Lopez B,
486	McPhearson T, Boyanova K, Alkemade R, Egoh B, Dunbar M B, Maes J. 2013. A blueprint for
487	mapping and modelling ecosystem services. Ecosyst Serv, 4: 4-14
488	https://doi.org/10.1016/j.ecoser.2013.02.001
489	Dai Z, Puyang X, Han L. 2016. Using assessment of net ecosystem services to promote sustainability
490	of golf course in China. Ecol Indic, 63: 165-171 https://doi.org/10.1016/j.ecolind.2015.11.056
491	de Groot R, Brander L, van der Ploeg S, Costanza R, Bernard F, Braat L, Christie M, Crossman N,
492	Ghermandi A, Hein L, Hussain S, Kumar P, McVittie A, Portela R, Rodriguez L C, Ten Brink P,
493	van Beukeringh P. 2012. Global estimates of the value of ecosystems and their services in monetary
494	units. Ecosyst Serv, 1: 50-61 https://doi.org/10.1016/j.ecoser.2012.07.005
495	Drakou E G, Crossman N D, Willemen L, Burkhard B, Palomo I, Maes J, Peedell S. 2015. A
496	visualization and data-sharing tool for ecosystem service maps: Lessons learnt, challenges and the
497	way forward. Ecosyst Serv, 13: 134-140 https://doi.org/10.1016/j.ecoser.2014.12.002
498	Farley J, Costanza R. 2010. Payments for ecosystem services: From local to global. Ecol Econ, 69:
499	2060-2068 https://doi.org/10.1016/j.ecolecon.2010.06.010
500	Fisher, B., Turner, R.K. and Morling, P., 2009. Defining and classifying ecosystem services for
501	decision making. ECOL ECON, 68:643-653. https://doi.org/10.1016/j.ecolecon.2008.09.014
502	Fisher B, Turner R K, Morling P. 2009. Defining and classifying ecosystem services for decision
503	making. Ecol Econ, 68: 643-653 https://doi.org/10.1016/j.ecoser.2017.11.008
504	Fu B, Su C, Wei Y, Willett I R, Lue Y, Liu G. 2011. Double counting in ecosystem services valuation:
505	causes and countermeasures. Ecol Res, 26: 1-14 https://doi.org/10.1007/s11284-010-0766-3
506	Guo Z W, Xiao X M, Gan Y L, Zheng Y J. 2001. Ecosystem functions, services and their values - a
507	case study in Xingshan County of China. Ecol Econ, 38: 141-154
508	https://doi.org/10.1016/S0921-8009(01)00154-9
509	Haines-Young R, Potschin M. 2013. Common International Classification of Ecosystem Services
510	(CICES): Consultation on Version 4, August-December 2012. EEA Framework Contract No
511	EEA/IEA/09/003. http://www.cices.eu or http://www.nottingham.ac.uk/cem.
512	Hayha T, Franzese P P. 2014. Ecosystem services assessment: A review under an ecological-economic
513	and systems perspective. Ecol Model, 289: 124-132
514	https://doi.org/10.1016/j.ecolmodel.2014.07.002
515	Hayha T, Franzese P P, Paletto A, Fath B D. 2015. Assessing, valuing, and mapping ecosystem
516	services in Alpine forests. Ecosyst Serv, 14: 12-23 https://doi.org/10.1016/j.ecoser.2015.03.001
517	Jiang W. 2017. Ecosystem services research in China: A critical review. Ecosyst Serv, 26: 10-16
518	https://doi.org/10.1016/j.ecoser.2017.05.012
519	Kenter J O, Jobstvogt N, Watson V, Irvine K N, Christie M, Bryce R. 2016. The impact of information,
520	value-deliberation and group-based decision-making on values for ecosystem services: Integrating

521	deliberative monetary valuation and storytelling. Ecosyst Serv, 21: 270-290
522	https://doi.org/10.1016/j.ecoser.2016.06.006
523	Koch E W, Barbier E B, Silliman B R, Reed D J, Perillo G M E, Hacker S D, Granek E F, Primavera J
524	H, Muthiga N, Polasky S, Halpern B S, Kennedy C J, Kappel C V, Wolanski E. 2009. Non-linearity
525	in ecosystem services: temporal and spatial variability in coastal protection. Front Ecol Environ, 7:
526	29-37 https://doi.org/10.1890/080126
527	Koellner T, Geyer R. 2013. Global land use impact assessment on biodiversity and ecosystem services
528	in LCA. Int J Life Cycle Ass, 18: 1185-1187 https://doi.org/10.1007/s11367-013-0580-6
529	Kubiszewski I, Costanza R, Anderson S, Sutton P. 2017. The future value of ecosystem services:
530	Global scenarios and national implications. Ecosyst Serv, 26: 289-301
531	https://doi.org/10.1016/j.ecoser.2017.05.004
532	Mancini M S, Galli A, Coscieme L, Niccolucci V, Lin D, Pulselli F M, Bastianoni S, Marchettini N.
533	2018. Exploring ecosystem services assessment through Ecological Footprint accounting. Ecosyst
534	Serv, 30: 228-235 <u>https://doi.org/10.1016/j.ecoser.2018.01.010</u>
535	Martin D M, Mazzotta M. 2018. Non-monetary valuation using Multi-Criteria Decision Analysis:
536	Sensitivity of additive aggregation methods to scaling and compensation assumptions. Ecosyst Serv,
537	29: 13-22 https://doi.org/10.1016/j.ecoser.2017.10.022
538	Millennium Ecosystem Assessment. 2005. Ecosystems and Human Well-Being: Synthesis. Washington,
539	DC: Island Press
540	Norton B, Costanza R, Bishop R C. 1998. The Evolution of Preferences: Why 'Sovereign' Preferences
541	May Not Lead to Sustainable Policies and What to Do about It. Ecol Econ, 24: 193-211
542	https://doi.org/10.1016/S0921-8009(97)00143-2
543	Ouyang Z, Wang X, Miao H. 1999. A primary study on Chinese terrestrial ecosystem services and their
544	ecological-economic values (in Chinese). Acta Ecol Sin 19: 607-613
545	http://www.ecologica.cn/stxb/ch/reader/view_abstract.aspx?file_no=990504&flag=1
546	Pascual U, Balvanera P, D áz S, Pataki G R, Roth E, Stenseke M, Watson R T, Ba Ak Dessane E, Islar
547	M, Kelemen E, Maris V, Quaas M, Subramanian S M, Wittmer H, Adlan A, Ahn S, Al-Hafedh Y S,
548	Amankwah E, Asah S T, Berry P, Bilgin A, Breslow S J, Bullock C, Cáceres D, Daly-Hassen H,
549	Figueroa E, Golden C D, Gómez-Baggethun E, González-Jim énez D, Houdet J L, Keune H, Kumar
550	R, Ma K, May P H, Mead A, O Farrell P, Pandit R, Pengue W, Pichis-Madruga R, Popa F, Preston
551	S, Pacheco-Balanza D, Saarikoski H, Strassburg B B, van den Belt M, Verma M, Wickson F, Yagi
552	N. 2017. Valuing nature's contributions to people: the IPBES approach. Curr Opin Env Sust, 26-27:
553	7-16 https://doi.org/10.1016/j.cosust.2016.12.006
554	Pascual U, Muradian R, Brander L, Gómez-Baggethun E, Mart n-López B, Verma M, Armsworth P,
555	Christie M, Cornelissen H, Eppink F. 2010. The economics of valuing ecosystem services and
556	biodiversity. In: TEEB Ecological and Economic Foundation. 2010. The Economics of Ecosystems
557	and Biodiversity: The Ecological and Economic Foundations. London/ Washington: Earthscan
558	PRWRC: Pearl River Water Resources Commission of the Ministry of Water Resources. 2017.
559	Zhujiang River Water Resources Bulletin 2015 (in Chinese). Guangzhou: Pearl River Water
560	Resources Commission of the Ministry of Water Resources

561	http://www.pearlwater.gov.cn/xxcx/szygg/
562	SCPRC: State Council of the People's Republic of China. 2014. The State Council approved the
563	Development Planning of Zhujiang-Xijiang Economic Zone (in Chinese). Beijing: State Council
564	of the People's Republic of China
565	http://www.gov.cn/zhengce/content/2014-07/16/content_8933.htm
566	Stoeckl N, Farr M, Larson S, Adams V M, Kubiszewski I, Esparon M, Costanza R. 2014. A new
567	approach to the problem of overlapping values: A case study in Australia's Great Barrier Reef.
568	Ecosyst Serv, 10: 61-78 https://doi.org/10.1016/j.ecoser.2014.09.005
569	TEEB Synthesis. 2010. Mainstreaming the Economics of Nature: a Synthesis of the Approach,
570	Conclusions and Recommendations of TEEB. London/ Washington: Earthscan
571	Ulanowicz R E. 1986. Growth and Development: Ecosystems Phenomenology. New York:
572	Springer-Verlag
573	van Ree C C D F, van Beukering P J H, Boekestijn J. 2017. Geosystem services: A hidden link in
574	ecosystem management. Ecosyst Serv, 26: 58-69 https://doi.org/10.1016/j.ecoser.2017.05.013
575	Yang H, Zhao B, Chen J. 2019. Quantifying the Payments for Ecosystem Services among hydrologic
576	units in Zhujiang River Basin, China based on the indicator of Optional Capacity Value. bioRxiv:
577	616680 https://doi.org/10.1101/616680
578	