

1 ***In situ* abundance and carbon fixation activity of distinct anoxygenic phototrophs in the**
2 **stratified seawater lake Rogoznica**

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25 **Abstract**

26 Sulfide-driven anoxygenic photosynthesis is an ancient microbial metabolism that contributes
27 significantly to inorganic carbon fixation in stratified, sulfidic water bodies. Methods
28 commonly applied to quantify inorganic carbon fixation by anoxygenic phototrophs, however,
29 cannot resolve the contributions of distinct microbial populations to the overall process. We
30 implemented a straightforward workflow, consisting of radioisotope labeling and flow
31 cytometric cell sorting based on the distinct autofluorescence of bacterial photo pigments, to
32 discriminate and quantify contributions of co-occurring anoxygenic phototrophic populations
33 to *in situ* inorganic carbon fixation in environmental samples. This allowed us to assign 89.3
34 $\pm 7.6\%$ of daytime inorganic carbon fixation by anoxygenic phototrophs in Lake Rogoznica
35 (Croatia) to an abundant chemocline-dwelling population of green sulfur bacteria (dominated
36 by *Chlorobium phaeobacteroides*), whereas the co-occurring purple sulfur bacteria
37 (*Halochromatium* sp.) contributed only $1.8 \pm 1.4\%$. Furthermore, we obtained two metagenome
38 assembled genomes of green sulfur bacteria and one of a purple sulfur bacterium which
39 provides the first genomic insights into the genus *Halochromatium*, confirming its high
40 metabolic flexibility and physiological potential for mixo- and heterotrophic growth.

41 **Introduction**

42 Modern anoxic phototrophic ecosystems are often seen as analogs to study the ecology of early
43 Earth environments. Sulfide-driven, anoxygenic photosynthesis is an ancient bacterial energy-
44 yielding metabolism (Brocks et al., 2005) and often dominates autotrophic carbon fixation in
45 stratified, sulfidic environments (e.g. Cohen et al., 1977). This process is primarily mediated
46 by two phylogenetically distinct groups of bacteria: i) the strictly anaerobic green sulfur
47 bacteria (GSB) of the class *Chlorobia* (Overmann, 2006), and ii) the metabolically more
48 versatile purple sulfur bacteria (PSB) affiliated with the gammaproteobacterial orders
49 *Chromatiales* and *Ectothiorhodospirales* (Imhoff, 2006a, 2006b). GSB and PSB oxidize and
50 thereby detoxify sulfide and other reduced sulfur compounds, primarily formed during sulfate-
51 dependent mineralization of organic matter in anoxic waters and sediments of marine and
52 limnic environments (Overmann and Garcia-Pichel, 2013). Although GSB and PSB basically
53 compete for the same resources, they co-occur in most photic and sulfidic environments.
54 Ecological niche partitioning between GSB and PSB is considered to be mainly based on
55 sulfide and oxygen concentrations, and light availability (Abella et al., 1980; Mas and van
56 Gemerden, 1995; Stomp et al., 2007). PSB require more light and typically form dense
57 populations at the sulfide-oxygen interface, while low-light adapted GSB often thrive beneath
58 the PSB layer (e.g. Musat et al., 2008; Llorens-Marès et al., 2015). While GSB are strictly
59 photoautotrophic, some PSB are capable of mixo- or heterotrophic, and chemotrophic growth
60 (e.g. Imhoff, 2006a, 2006b; Berg et al., 2019), which further facilitates niche partitioning.
61 Numerous studies have addressed the overall contribution of anoxygenic photosynthesis (by
62 GSB and PSB) to *in situ* dissolved inorganic carbon (DIC) fixation in sulfidic, stratified aquatic
63 environments (e.g. Camacho et al., 2001; Casamayor et al., 2001; Marschall et al., 2010; Fontes
64 et al., 2011; Morana et al., 2016). However, only few studies have investigated the relative
65 contributions of GSB and PSB populations to this process. Musat and colleagues (2008)
66 applied nanoscale secondary ion mass spectrometry (nanoSIMS) to track uptake of ^{13}C -

67 bicarbonate and ^{15}N -ammonia in single cells of GSB and PSB from the chemocline of Lake
68 Cadagno. They reported that relative abundances of PSB and GSB do not correlate with their
69 DIC fixation activity. Instead, the contribution of a rare PSB population (*Chromatium okenii*)
70 to total DIC fixation was disproportionately high. In fact, per cell DIC fixation rates in *Chr.*
71 *okenii* were up to three orders of magnitude higher than per cell DIC fixation rates of the *in*
72 *situ* dominant GSB *Chlorobium clathratiforme*. In another study at Lake Cadagno, Storelli and
73 colleagues (2013) incubated pure cultures of PSB and GSB isolated from Lake Cadagno *in situ*
74 with ^{14}C -labeled bicarbonate and quantified their contribution to DIC fixation by
75 scintillographic quantification of radioisotope incorporation into microbial biomass. Again, a
76 disproportionately high rate of DIC fixation in cells of *in situ* rare PSBs (*Candidatus Thiocystis*
77 *syntrophicum* and *Lamprocystis purpurea*) was observed, while DIC fixation by the *in situ*
78 dominant GSB *Chl. clathratiforme* was only marginal. However, these studies were somewhat
79 limited by incubations performed *ex situ* or use of pure cultures that did not necessarily
80 represent the *in situ* PSB and GSB communities. Thus, it is still unknown whether rare
81 populations of PSB in general account for high proportions of DIC fixation via anoxygenic
82 photosynthesis, or if this phenomenon is specific to Lake Cadagno.

83 Here, we present an approach combining incubations with radioisotope-labeled bicarbonate,
84 fluorescence activated cell sorting (FACS) and scintillography of the sorted populations to
85 discriminate the contribution of environmental populations of PSB and GSB to *in situ* DIC
86 fixation. Similar approaches have previously been used to quantify the activities of discrete
87 microbial populations in both pelagic and benthic environments (e.g. Zubkov et al., 2003; Vila-
88 Costa et al., 2006; Dykstra et al., 2016). These high throughput workflows enable researches
89 to directly measure the relative contribution of individual microbial groups to an overall
90 process, without relying on assumptions such as conversion factors. Using this approach, we
91 were able to show that the *in situ* numerically dominant GSB also dominate DIC fixation in
92 our model system - a stratified seawater lake (Lake Rogoznica) on the eastern Adriatic coast

93 (Croatia). The *in situ* rare PSB, on the other hand, contributed only marginally to DIC fixation
94 in Lake Rogoznica. The metagenome assembled genome of these PSB, representing the first
95 genomic dataset for the genus *Halochromatium*, illustrates their capability of mixo- and
96 heterotrophic growth.

97

98 **Results and Discussion**

99 *Chemical and hydrographical settings*

100 Lake Rogoznica is a monomictic (stratified, except for one annual mixing) seawater lake on
101 the eastern Adriatic coast (Fig. S1). At the time of sampling and experimentation (April, 2015),
102 the water column of Lake Rogoznica was stratified. The recorded chemical and hydrographical
103 water profiles were in accordance with data collected on numerous occasions in spring time
104 over the last three decades (e.g. Bura-Nakić et al., 2009; Pjevac et al., 2015). The salinity of
105 the deeper water layers was similar to the surrounding Adriatic seawater source (Lipizer et al.,
106 2014). The decreased salinity in the top 4 m of the epilimnion (Fig. 1A) was caused by rainfalls
107 during the weeks preceding sampling, since atmospheric precipitation is the only freshwater
108 source to Lake Rogoznica (Žic et al., 2013). Oxygen saturation was at or above 100% in the
109 top 4 m of the epilimnion and steadily decreased until depletion at ~9 m (Fig. 1A). The
110 chemocline, a sulfide-oxygen interface, was located at 8.5-9 m depth and accompanied by a
111 sharp peak in turbidity (Fig. 1B), indicative of a dense microbial population and the formation
112 of colloidal zero-valent sulfur (S^0 ; Kamyshny et al., 2011). S^0 was detected at 8, 9 and 10 m
113 depth, with a concentration peak of $77 \mu\text{mol l}^{-1}$ at 9 m (Fig. 1C). Thiosulfate and sulfide were
114 only detected below 8 m, and reached maximal concentrations of $12 \mu\text{mol l}^{-1}$ and 2.7mmol l^{-1} , respectively (Fig. 1C). DIC concentrations in Lake Rogoznica steadily increased with depth
115 and ranged between $3.4\text{-}5.6 \text{mmol l}^{-1}$ (Fig. 1C). These values are at least 1mmol l^{-1} higher than
116 measured and modeled DIC concentrations in Adriatic surface seawater (e.g. Cossarini et al.,
117 2015; Gemayel et al., 2015). Carbonate mineral (calcite and dolomite) dissolution could have
118

119 been an important DIC source in Lake Rogoznica, as it has been previously reported for other
120 karst and epikarst hosted surface water bodies and aquifers in Croatia and elsewhere (e.g.
121 Barešić et al., 2011; Florea et al., 2016). Alternatively, organic matter (OM) mineralization in
122 the lake sediment and anoxic water column could have resulted in DIC build-up during water
123 column stratification.

124

125 *Photosynthetic pigment analysis*

126 During stratification, Lake Rogoznica contains a high biomass of anoxygenic photosynthetic
127 bacteria in anoxic waters, while oxygenic phototrophs (i.e. cyanobacteria and algae) inhabit the
128 oxic epilimnion (Malešević et al., 2015; Pjevac et al., 2015). To assess the depth distribution
129 of phototrophic microorganisms, we analyzed the photosynthetic pigment content of a lake
130 water depth profile collected on the day of experimentation. In the upper 8 m of the water
131 column we detected mostly Chl *a*, indicative of the presence of cyanobacteria and algae
132 (oxygenic phototrophs; Scheer, 2006). Chl *a* reached a maximum of 5.0 nmol l⁻¹ at 4 m depth,
133 with a secondary peak (4.1 nmol l⁻¹) at 8 m depth (Fig. 2). No Chl *a* was detected in samples
134 collected from 9 m and 10 m depth (Fig. 2). In the epilimnion (0-8 m), small amounts of BChl
135 *a* (<5% of identified pigments), likely originating from aerobic anoxygenic species, were
136 detected (Fig. 2). The water samples collected at the chemocline (9 m) and in the hypolimnion
137 (10 m) contained 80- to 110-fold higher concentrations of BChl *a* (11.1-14.5 nmol l⁻¹) than the
138 epilimnion, reflecting the presence of GSB, PSB, and potentially other anoxygenic
139 photosynthetic bacteria (Scheer, 2006). No BChl *b*, *c* or *d* were detected. However, at the
140 chemocline (9 m) and in the hypolimnion (10 m), very high concentrations of BChl *e* (8.1-18.4
141 nmol l⁻¹), only known for low-light adapted, brown-colored GSB (Chew and Bryant, 2007),
142 were detected (Fig. 2). Interestingly, even above the anoxic chemocline (at 8 m), BChl *e*
143 concentrations (0.5 nmol l⁻¹) exceeded BChl *a* concentrations (0.1 nmol l⁻¹). Since BChl *e* is
144 only found in GSB, while BChl *a* is present in a variety of anoxygenic phototrophic bacteria,

145 the BChl *e* / BChl *a* ratio supports the previously reported dominance of GSB throughout the
146 chemocline of Lake Rogoznica (Pjevac et al., 2015).

147

148 *GSB and PSB in Lake Rogoznica*

149 To determine the community composition of anoxygenic phototrophic bacteria in Lake
150 Rogoznica, we sequenced metagenomes from chemocline (9 m) and hypolimnion (10 m)
151 samples and assessed to relative abundance of GSB and PSB in these samples via 16S rRNA
152 gene sequence read mapping and reconstruction (Fig. S2). In the chemocline (9 m), GSB-
153 related reads accounted for 13% of the total reads mapped to 16S rRNA genes, while PSB-
154 related reads accounted for 5%. In the hypolimnion (10 m) sample, the relative contribution of
155 GSB- and PSB-related reads mapped to 16S rRNA gene sequences were 17% and 2%,
156 respectively (Fig. S2).

157 From the metagenomic data, we reconstructed three nearly complete high quality metagenome
158 assembled genomes (HQ MAGs) related to anoxygenic phototrophs (Table 1). Two GSB-
159 related HQ MAGs displayed a high average nucleotide identity (>98%) to genomes of validly
160 described *Chlorobiaceae* species (Fig. S3; Table 1). The more abundant GSB in Lake
161 Rogoznica, represented by HQ MAG C10 (Table 1), was identified as a strain of *Chlorobium*
162 *phaeobacteroides* (Overmann et al., 1992). This is well in line with the high concentration of
163 BChl *e* measured in chemocline and hypolimnion samples, as BChl *e* is the dominant
164 photosynthetic pigment in the low light adapted *Chlorobium phaeobacteroides* strain BS1
165 (Overmann et al., 1992). The second GSB HQ MAG, B10, was closely related to
166 *Prosthecochloris aestuarii* DSM 271 (Gorlenko, 1970), and represents a strain of this species.
167 Based on read coverage, it was present at a much lower abundance at the time of sampling (Fig.
168 S4). The genomic potential of both HQ MAGs from GSB overlapped with the genomic
169 potential of their close relatives isolated in pure culture: both HQ MAGs show the genomic
170 potential of an obligatory anoxygenic photolithoautotroph.

171 For the PSB-related HQ MAG A10 no closely related genome sequence could be identified.
172 However, the two almost identical (1 nucleotide difference) 16S rRNA gene copies encoded in
173 HQ MAG A10 displayed 99% sequence identity to the 16S rRNA gene sequence of
174 *Halochromatium roseum* (Kumar et al., 2007) and the 16S rRNA gene sequence of the PSB
175 species isolated by FACS from Lake Rogoznica in our previous study (Pjevac et al., 2015; Fig.
176 S5). Thus, MAG A10 provides the first genomic insights into the genus *Halochromatium*
177 within the family *Chromatiaceae*. The high 16S rRNA sequence identity of 99% is in the
178 absence of a genome sequence suggestive, yet insufficient for the assignment of this MAG to
179 *Halochromatium roseum*.

180 All three validly described *Halochromatium* species - *Halochromatium salexigens* (Caumette
181 et al., 1988), *Halochromatium glycolicum* (Caumette et al., 1997), and *Halochromatium*
182 *roseum* (Kumar et al., 2007) harbor BChl *a* as their main photosynthetic pigment. They require
183 elevated NaCl concentrations (>1%) and grow photolithoautotrophically by oxidation of
184 reduced sulfur compounds. Additionally, the capability to assimilate at least some organic
185 carbon compounds (e.g. acetate, pyruvate, fumarate, succinate, malate, glycolate) during mixo-
186 or heterotrophic growth was demonstrated for all strains. Furthermore, sulfur-dependent
187 chemolithotrophic growth (both autotrophic and heterotrophic) under microaerobic conditions
188 has been demonstrated for *Halochromatium salexigens* and *Halochromatium glycolicum*
189 (Caumette et al., 1988; Caumette et al., 1997; Kumar et al., 2007). In accordance, the here
190 retrieved *Halochromatium* sp. HQ MAG A10 encodes the potential for lithotrophic growth via
191 reduced sulfur compound oxidation and CO₂ fixation via the Calvin-Benson-Bassham cycle.

192 All genes essential for phototrophic growth and chemotrophic growth with O₂ as terminal
193 electron acceptor were present (Fig. 3). Moreover, this HQ MAG encodes the capability for N₂
194 fixation and urea utilization (Fig. 4) - metabolic features that were not previously reported for
195 members of the genus *Halochromatium*. Finally, the HQ MAG A10 encodes genes for mixo-

196 or heterotrophic metabolism, including glycolate and sucrose utilization, as well as glycogen,
197 polyphosphate, and PHB storage and utilization (Fig. 4).

198

199 *FACS sorting and quantification of GSB and PSB populations*

200 To assess the contribution of PSB and GSB to DIC fixation in Lake Rogoznica, we
201 implemented a FACS-based workflow to differentially sort the GSB and PSB populations after
202 *in situ* incubations with radioisotope labeled bicarbonate. Briefly, after sorting and
203 quantification of total microbial cells based on an unspecific DNA stain, the red fluorescence
204 (FL3) of photosynthetic pigments upon excitation with green laser light (488 nm) is used as a
205 diagnostic feature to distinguish phototrophs from other microorganisms. To further
206 distinguish different pigmented populations, both forward scatter (FSC) profiles (reflecting
207 differences in cell size) and 90° side scatter (SSC) profiles (reflecting cell granularity) can be
208 used. In fact, combinations of FL3, FSC and SSC profiles have previously been used to
209 distinguish PSB and GSB from each other and other microorganisms in cultures and
210 environmental samples (Casamayor et al., 2007; Pjevac et al., 2015; Zimmermann et al., 2015).
211 Here, we opted for identification, sorting and quantification of GSB and PSB populations based
212 on FL3 versus SSC profiles, as this yielded the clearest distinction between populations (Fig.
213 4).

214 In the chemocline samples, we detected a low abundance ($1.0 \pm 1.0\%$ of all cells) population
215 displaying a FL3 vs. SSC profile resembling that of the PSB previously selectively sorted from
216 Lake Rogoznica (Pjevac et al., 2015). Another, significantly more abundant ($18.4 \pm 1.4\%$)
217 population was identified on FL3 vs. SSC profiles in the chemocline (9 m) samples (Fig. 4).
218 Taking in account the absence of oxygenic phototrophs at and below the chemocline and the
219 matching relative abundance of GSB in chemocline samples (Fig. S2), we identified this
220 population as GSB (Fig. 4). In the hypolimnion sample (10 m), the abundance of the PSB was
221 too low for sorting, while the GSB accounted for $19.7 \pm 6.4\%$ of all cells (Fig. 4).

222

223 *DIC fixation by anoxygenic phototrophic bacteria in Lake Rogoznica*

224 Previous studies, performed in other aquatic stratified environments, combined *in situ*
225 incubations with isotopically labeled bicarbonate with incubations in the dark or in the presence
226 of oxygenic photosynthesis inhibitors (e.g. 3-(3',4'-dichlorophenyl)-1,1'-dimethyl urea -
227 DCMU) to distinguish DIC fixation by anoxygenic phototrophs, oxygenic phototrophs and
228 chemolithoautotrophs (e.g. Camacho et al., 2001; Casamayor et al., 2001; García-Cantizano et
229 al., 2005; Marschall et al., 2010; Fontes et al., 2011; Morana et al., 2016). Our approach, on
230 the other hand, requires only *in situ* ¹⁴C-bicarbonate incubations followed by FACS and
231 scintillography to quantify the contributions of individual anoxygenic phototrophic populations
232 to DIC fixation. Using this method, cell specific rates of *in situ* DIC fixation were determined
233 by scintillography for total cell sorts (i.e. all SYBR Green-I stained cells), as well as GSB and
234 PSB specific cell sorts (i.e. subpopulations of all SYBR Green-I stained cells differentially
235 based on their FL3 vs. SSC profile).

236 Average *in situ* per cell DIC fixation rates were similar in epilimnion (15.1 ± 6.0 amol C cell⁻¹
237 h⁻¹) and chemocline (12.4 ± 2.8 amol C cell⁻¹ h⁻¹) samples, while average *in situ* per cell DIC
238 fixation rates in hypolimnion samples (1.2 ± 1.2 amol C cell⁻¹ h⁻¹) were not higher than the
239 background signal measured in dead controls (0.7 ± 0.1 amol C cell⁻¹ h⁻¹) (Fig. 5). Average *in*
240 *situ* per cell DIC fixation rates of the chemocline-dwelling GSB population from Lake
241 Rogoznica (60.4 ± 14.7 amol C cell⁻¹ h⁻¹; Fig. 5) were in a similar range to rates calculated for
242 the *Chl. clathratiforme* (GSB) cells from Lake Cadagno (Musat et al., 2008). The average *in*
243 *situ* per cell DIC fixation rates of the chemocline-dwelling PSB population (27.1 ± 11.1 amol
244 C cell⁻¹ h⁻¹; Fig. 5) were in the same order of magnitude as rates determined for chemocline-
245 dwelling GSB in this and other studies (Musat et al., 2008; Storelli et al., 2013). However,
246 these values are up to three orders of magnitude lower than rates determined for PSB
247 populations in Lake Cadagno (Musat et al., 2008; Storelli et al., 2013). Notably, DIC fixation

248 rates of the multiple, co-occurring PSB populations in Lake Cadagno also varied by up to 100-
249 fold (Musat et al., 2008; Storelli et al., 2013).

250 By integrating average cell abundances with average *in situ* per cell DIC fixation rates, we
251 determined that the chemocline-dwelling GSB population was responsible for almost the entire
252 (89.3 ±7.6%) *in situ* DIC fixation measured in this study, whereas the low abundant PSB
253 population contributed only 1.8 ±1.4%. This is in contrast to the disproportionately high
254 contribution of PSB to *in situ* DIC fixation reported for the chemocline of Lake Cadagno
255 (Musat et al., 2008; Storelli et al., 2013).

256 Finally, the hypolimnion-dwelling GSB population (2.1 ±1.1 amol C cell⁻¹ h⁻¹) in Lake
257 Rogoznica did not assimilate significant amounts of labeled DIC (Fig. 5). This result is
258 consistent with previously observed low DIC uptake by anoxygenic phototrophs in the
259 hypolimnion (e.g. Guerrero et al., 1985; Camacho and Vicente 1998) and can be attributed to
260 low light quality through self-shading.

261

262 *Conclusions*

263 We show that anoxygenic photosynthesis-driven DIC fixation in Lake Rogoznica in spring
264 2015 was primarily mediated by a highly abundant chemocline-dwelling GSB population,
265 whereas the co-occurring PSB contributed only marginally to DIC fixation. Our radioisotope-
266 labeling and FACS-based workflow facilitated unprecedented insights into DIC fixation by
267 environmental populations of anoxygenic phototrophs. Our results contrast previous findings
268 from Lake Cadagno and illustrate that high contributions of rare PSB populations to DIC
269 fixations cannot be generalized for other stratified lakes. Further research is needed to obtain a
270 more representative and comprehensive picture of PSB vs. GSB mediated DIC fixation in
271 environments inhibited by anoxygenic phototrophs. Furthermore, the here presented workflow
272 can be readily adapted for analyze various co-occurring phototrophic microbial populations,
273 and also used to investigate other aspects of the *in situ* metabolism of physiologically flexible

274 PSB. For example, photoheterotrophic activity, a prominent metabolic feature encoded in the
275 here recovered *Halochromatium* MAG and other PSB genomes (Berg et al., 2019; Luedin et
276 al., 2019), could be quantified by tracking the incorporation of radiolabeled organic substrates
277 into environmental PSB populations.

278

279 **Experimental procedures**

280 *Study site and sample collection*

281 Lake Rogoznica (Fig. S1) is a small (~1 ha, 15 m max depth) seawater lake on the Dalmatian
282 peninsula Gradina, Croatia (43°32'N, 15°58'E). It is located at approximately 100 m distance
283 from the seashore, exchanging water with the Adriatic Sea through subsurface channels in the
284 porous limestone characteristic to Croatian coastal areas (e.g. Žic et al., 2013). The lake water
285 is rich in nutrients and dissolved organic carbon (DOC). The hypolimnion is highly sulfidic
286 during stratification (Pjevac et al., 2015). On April 2nd, 2015 hydrographical parameters
287 (dissolved oxygen concentration [O₂], temperature [T], conductivity [Sm], turbidity [t] and pH)
288 of a vertical lake profile were recorded with a Hydrolab DS5 multiparameter water quality
289 probe (OTT Hydromet, Germany). Water samples for DIC and reduced sulfur species (RSS)
290 analyses were collected with a vertically lowered 5-liter Niskin bottle (General Oceanics, USA)
291 in 1 meter intervals between the surface [~10-60 cm] and 10 m water depth. At 7 m
292 (epilimnion), 9 m (chemocline) and 10 m (hypolimnion) below surface, larger sample volumes
293 were collected for activity incubations and metagenome sequencing.

294

295 *Sample preservation and chemical analyses*

296 Samples for DIC analyses were collected in 1.5 ml amber borosilicate vials (Zinsser Analytic
297 Qualyvials); 15 µl saturated mercury chloride solution (2.7 mol l⁻¹) were added and vials were
298 closed with polytetrafluoroethylene (PTFE) coated septa and screw caps. DIC was measured
299 by infrared spectroscopy (Analytik Jena multi N/C 2100s) as CO₂ liberated from the original

300 sample in a 10% phosphoric acid trap. Samples for reduced sulfur species analyses were
301 collected in 1.5 ml amber borosilicate vials (Zinsser Analytic Qualyvials) and preserved with
302 ZnCl_2 (final concentration 4%). Sulfide concentrations were determined
303 spectrophotometrically in ZnCl_2 fixed samples as described before (Cline, 1969). Zero-valent
304 sulfur was converted to thiosulfate via sulfitolysis (Jørgensen et al., 1979; Ferdelman et al.,
305 1991). Briefly, ZnCl_2 fixed samples were buffered with HEPES (pH 8) and reacted with a 2%
306 sulfite solution at 70°C for 12 h to yield thiosulfate. Parallel samples with no added sulfite were
307 used as a control for background thiosulfate concentrations. The thiosulfate products were then
308 derivatized with monobromobimane as described in Zopfi et al., (2004). Derivatized samples
309 were analyzed on an Acquity H-class UPLC system (Waters Corporation, USA) equipped with
310 an Acquity UPLC BEH C8 column and controlled with the Empower III software. Sample
311 temperature was maintained at 4°C, column temperature was 40°C, and the sample injection
312 volume used was 1 μl . The mobile phase consisted of acetic acid (0.25 v/v), pH 3.5 (A), and
313 100% UPLC-grade methanol (B). The following gradient conditions were used at a constant
314 flow rate of 0.65 ml min^{-1} : start, 5% B; 0-4 min, ramp of curve 6 (linear) to 10% B; 4-5 min,
315 ramp of curve 6 to 95% B; 5-6 min, 95% B; 6.01 min, 5% B; 8 min, 5% B; injection of the next
316 sample. The fluorescence detector was set to an excitation wavelength of 380 nm and an
317 emission wavelength of 480 nm.

318

319 *Pigment analysis*

320 One l (0-7 m) or 1.5 l (8-10 m) water samples were collected onto 47 mm glass fiber filters
321 (type GF/F, nominal pore size 0.7 μm) using low vacuum. The filters were folded, gently dried
322 using a paper towel, and stored at -80°C. In the laboratory, the filters were homogenized in 7
323 ml acetone: methanol mixture (7:2) using a glass Teflon tissue homogenizer. The filter debris
324 was removed by centrifugation (10,000 g). Pure extracts were analyzed using the Prominence-
325 *i* LC-2030C HPLC system (Shimadzu, Japan). Pigments were separated on a heated (40°C)

326 Luna 3 μ C8(2) 100 Å column (Phenomenex Inc., USA) with binary solvent system A: 20% 28
327 mmol l⁻¹ ammonium acetate + 80% methanol, B: 100% methanol. Chlorophyll (Chl) *a*, BChl
328 *a*, and BChl *e* were detected at 665 nm, 770 nm, and 655 nm, respectively. The HPLC system
329 was calibrated using 100% methanol extracts of *Synechocystis* sp. PCC6803, *Rhodobacter*
330 *sphaeroides* and *Chlorobium phaeobacteroides* with known concentrations of Chl *a*, BChl *a*,
331 and BChl *e*, respectively.

332

333 *DNA extraction and metagenome sequencing*

334 DNA for metagenome sequencing was extracted from 6 polycarbonate membrane filters (type
335 GTTP; 0.2 μ m pore size, Whatman, UK) per depth (9 m and 10 m) by a phenol/chloroform-
336 based protocol (Massana et al., 1997). Briefly, filters were incubated with lysozyme (1 g l⁻¹) at
337 37°C for 45 min, and proteinase K (0.2 g l⁻¹) and SDS (1%) at 55°C for 1 h. Extraction was
338 performed twice with 750 μ l of phenol: chloroform (CHCl₃): isoamyl alcohol (IAA) (25 : 24 :
339 1, pH 8) solution and once with 750 μ l of CHCl₃ : IAA (24 : 1) solution. The aqueous phases
340 of all extractions for each depth were collected and pooled. Finally, 1/10 volume of sodium
341 acetate was added to the pooled aqueous phase and DNA was precipitated for 30 min at -20°C
342 with 1 ml of isopropanol. After a centrifugation step (20 min at 4°C and 20,000 g) the DNA
343 pellet was washed with 500 μ l of 70% ethanol. After a second centrifugation step (5 min at 4°C
344 and 20,000 g) the pellet was dissolved in 80 μ l of deionized (MQ) water and stored at -20°C
345 until further processing. Aliquots of the DNA extracts were sent to the Max Planck Genome
346 Centre (MP-GC, Cologne) for paired end library preparation and metagenome sequencing on
347 the Illumina HiSeq 2500. Paired end metagenome reads were quality trimmed at a phred score
348 of 15 using the bbdduk function of the BBMap package (v. 35.82;
349 <https://sourceforge.net/projects/bbmap/>). Small subunit (SSU) rRNA gene sequences were
350 reconstructed from the quality trimmed metagenomic reads and classified against the SILVA
351 SSU rRNA gene database using PhyloFlash (v.3.0; Gruber-Vodicka et al., 2019). A

352 metagenome co-assembly of both samples was performed with Spades v. 3.11.1 (Bankevich et
353 al., 2012) and binned with Metwatt v. 3.5.3 (Strous et al., 2012). Three HQ MAGs classified
354 as anoxygenic phototrophs were identified and polished by iterative re-assembly for 10
355 iterations as described in Mußmann et al., 2017. The phylogenetic affiliation, completeness
356 and redundancy of the HQ MAGs was assessed with MIGA (Rodriguez-R et al., 2018) and the
357 HQ MAGs were automatically annotated in RAST (Aziz et al., 2008). For phylogenetic
358 placement of the *Halochromatium* sp. A10 HQ MAG, the 16S rRNA gene sequences encoded
359 in HQ MAG A10, alongside with a selection of 16S rRNA gene sequences from
360 *Chromatiaceae* isolates and environmental samples were aligned against the SILVA r132 SSU
361 database using the SINA aligner (Pruesse et al., 2012). The alignment was uploaded to the IQ-
362 TREE webserver (Trifinopoulos et al., 2016) for phylogenetic tree reconstruction using default
363 parameters, and the resulting tree was visualized in iTOL (Letunic and Bork, 2016).

364

365 *Nucleotide Accession Numbers*

366 Metagenome assembled genomes (MAGs) are available under NCBI genome
367 accessions/RAST project IDs POWB00000000/2049430.3 (*Halochromatium* sp. A10),
368 POWC00000000/290513.3 (*Prosthecochloris* sp. B10) and POWD00000000/290513.4
369 (*Chlorobium* sp. C10), respectively, while the metagenome raw reads can be retrieved from
370 ENA under the study accession number PRJEB26778.

371

372 *Carbon assimilation experiments*

373 Samples for *in situ* ¹⁴C-bicarbonate assimilation experiments were collected with a 5-liter
374 Niskin bottle at 7 m, 9 m and 10 m water depth as described above. Sample aliquots were
375 transferred to 6 ml exetainer screw cap vials (Labco Limited) directly from the Niskin bottles.
376 A sterile rubber tube and a slow flow rate were used during transfer to avoid oxygen intrusion
377 to anoxic chemocline and hypolimnion samples. Four exetainer vials were overflowed and filled

378 at each depth. From one exetainer per depth, 325 μl sample were removed with a syringe
379 through the screw cap septum, and replaced with 325 μl 37% formaldehyde solution, resulting
380 in a final formaldehyde concentration of 2%. Thereafter, 0.10 mmol l^{-1} (7 m), 0.12 mmol l^{-1} (9
381 m) or 0.15 mmol l^{-1} (10 m) ^{14}C -labeled sodium bicarbonate (specific activity 56 mCi mmol^{-1})
382 was added to all 4 exetainers of each depth sample with a glass syringe. The sample-filled
383 exetainers were mounted vertical to carrier rings, and lowered to *in situ* depths of 7 m, 9 m or
384 10 m, respectively. Incubations were performed at ambient light and temperature conditions
385 between 8 am and 2 pm, to assure light availability for phototrophic activity. After 6 h of
386 incubation samples were retrieved, fixed with 2% formaldehyde (final concentration) as
387 described above, and kept at 4°C until further processing.

388

389 *FACS and scintillation counting of sorted cells*

390 Prior to flow cytometry, cells were stained with SYBR Green I (Marie et al., 1997) and large
391 suspended particles were removed by filtration through 5 μm pore-size filters (Sartorius) to
392 avoid clogging of the flow cytometer. Flow sorting was performed using a BD FACSCalibur
393 flow cytometer equipped with a cell sorter and a 15 mW argon ion laser exciting at 488 nm
394 (Becton Dickinson, UK). Autoclaved Milli-Q water was used as sheath fluid. Cell sorting was
395 done at a low flow rate of $12 \pm 3 \mu\text{l min}^{-1}$ or a medium flow rate of $35 \pm 5 \mu\text{l min}^{-1}$ with single
396 cell sort mode to obtain highest purity. The event rate was adjusted with a fluorescence
397 threshold and sorting was performed on a rate of approximately 25-250 particles s^{-1} . SYBR
398 Green I stained cells were identified on scatter dot plots of green fluorescence (filter FL1
399 530/30) versus 90° light scatter. Bacteriochlorophyll fluorescence was identified on scatter dot
400 plots of red fluorescence (filter FL3 650LP) versus 90° light scatter. For subsequent
401 measurements 25,000-100,000 cells were sorted and filtered onto 0.2 μm polycarbonate filters
402 (GTTP, Millipore). Collected cell batches on polycarbonate filters were directly transferred
403 into 6 ml scintillation vials and mixed with 5 ml UltimaGold XR (Perkin Elmer) scintillation

404 cocktail. Radioactivity of sorted cell batches was measured in a liquid scintillation counter (Tri-
405 Carb 2900, Perkin Elmer). The abundance of PSB and GSB in ^{14}C -carbon assimilation
406 experiments is given as relative percentage of gated cells identified as PSB/GSB out of all
407 SYBR green-stained cells concurrently counted using flow cytometry.

408

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584 **Tables and figures**

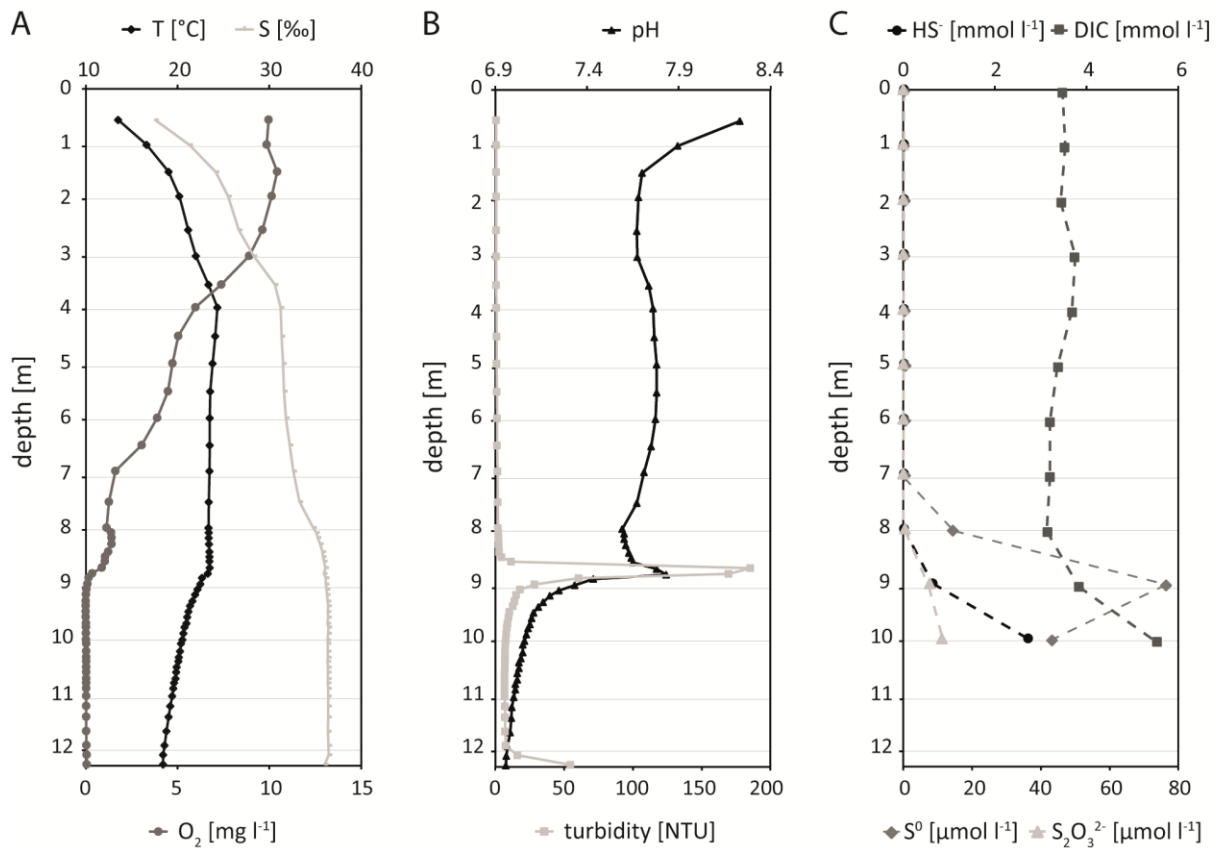
585 **Table 1:** Genomic features of recovered metagenome assembled genomes (MAGs) related to
586 anoxygenic phototrophs.

MAG	Contigs	CDS¹	Genome size [Mb]	GC content	Completeness	Contamination
<i>Halochromatium</i> sp. (A10)	43	3,913	4.19	60.8%	82.9%	1.8%
<i>Prosthecochloris aestuarii</i> spp. (B10)	32	2,070	2.26	50.1%	92.8%	0.9%
<i>Chlorobium phaeobacteroides</i> spp. (C10)	40	2,009	2.13	44.1%	91.9%	0%

587 ¹coding domain sequences

588

589



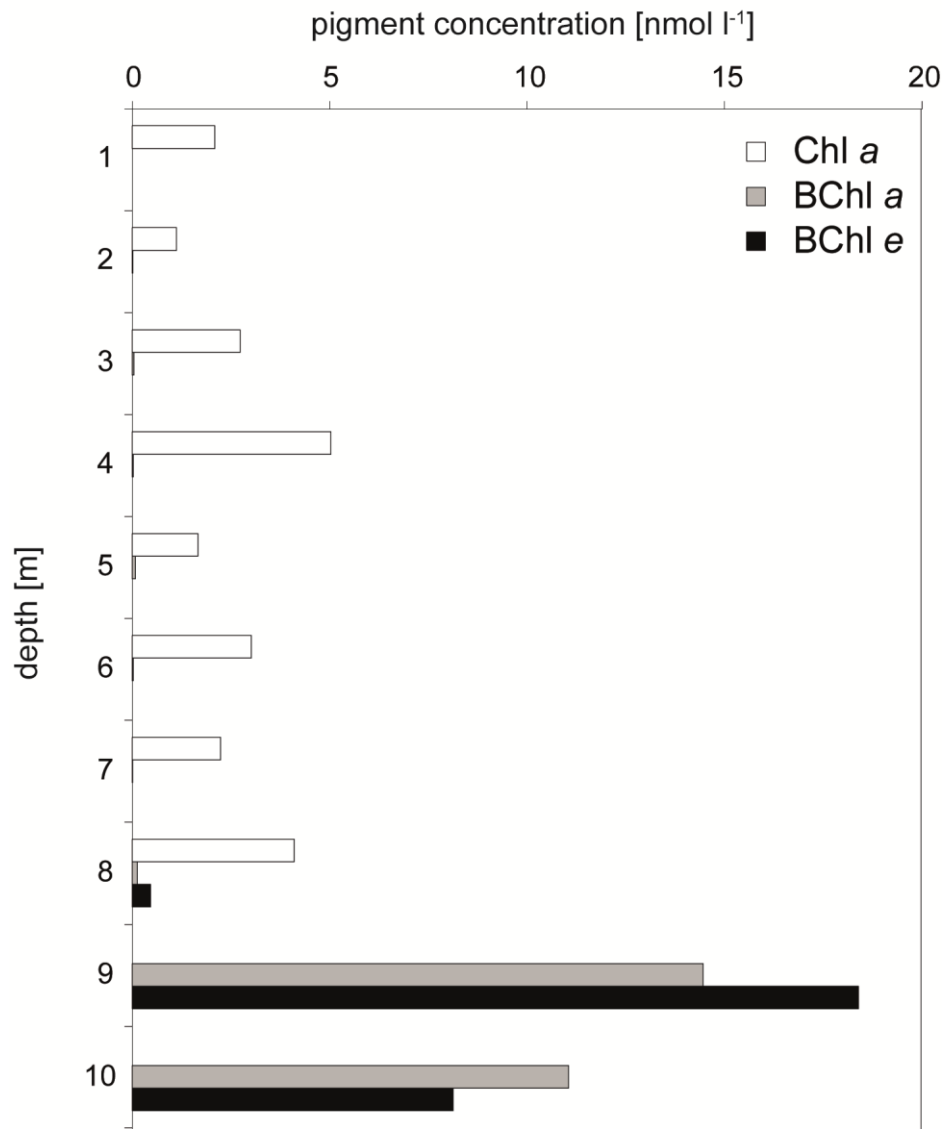
590

591 **Figure 1.**

592 Hydrographical profile and reduced sulfur speciation in a depth profile of Lake Rogoznica
593 collected in April 2015. From left to right: **A** Temperature (T) (°C), salinity (S) (‰) (top x
594 axis), and dissolved oxygen (O₂) concentration (mg l⁻¹) (bottom x axis) profiles. **B** pH (top x
595 axis) and turbidity (NTU) (bottom x axis) profiles. **C** Sulfide (HS⁻) concentration (mmol l⁻¹),
596 dissolved inorganic carbon (DIC) concentration (mmol l⁻¹) (top x axis), elemental sulfur (S⁰)
597 concentration (μmol l⁻¹), and thiosulfate (S₂O₃²⁻) concentration (μmol l⁻¹) (bottom x axis)
598 profiles.

599

600



601

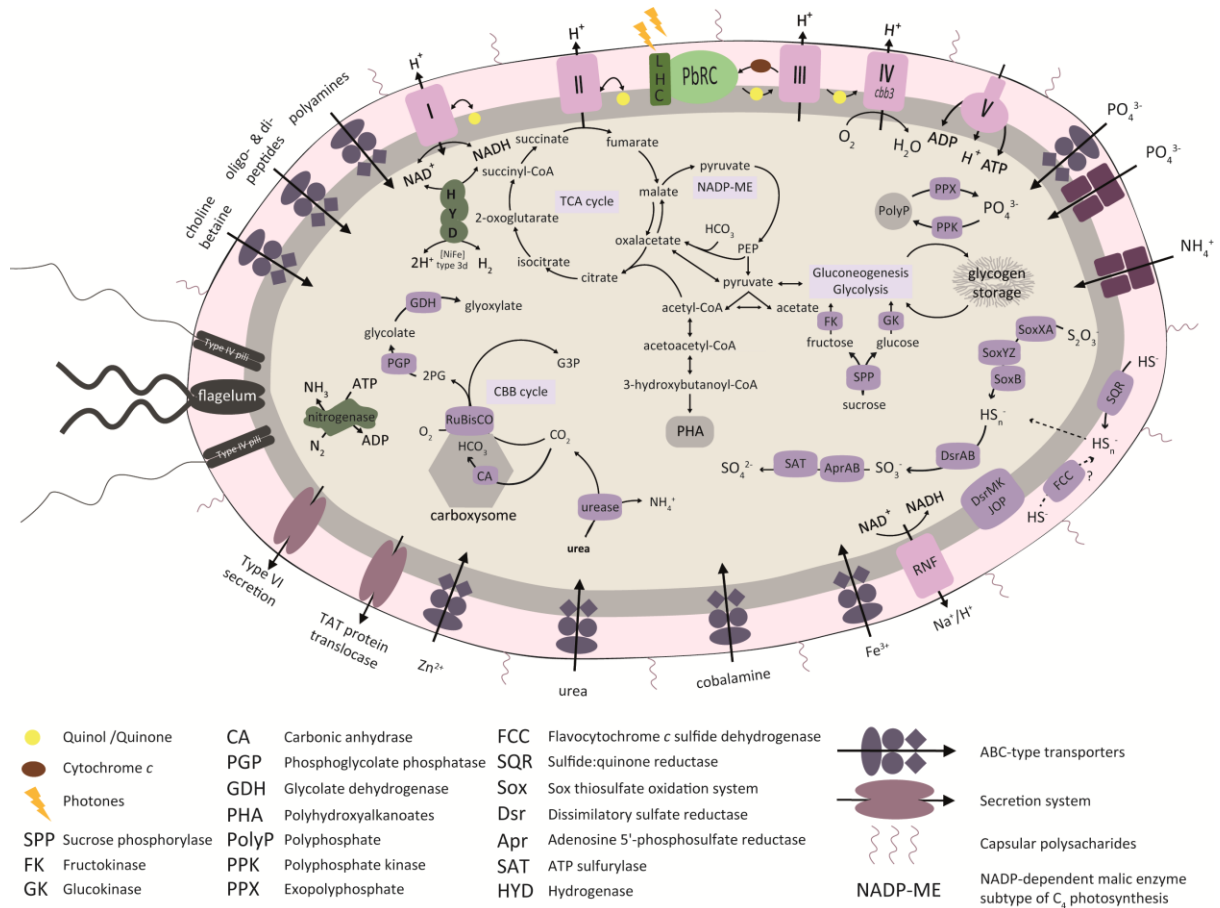
602 **Figure 2.**

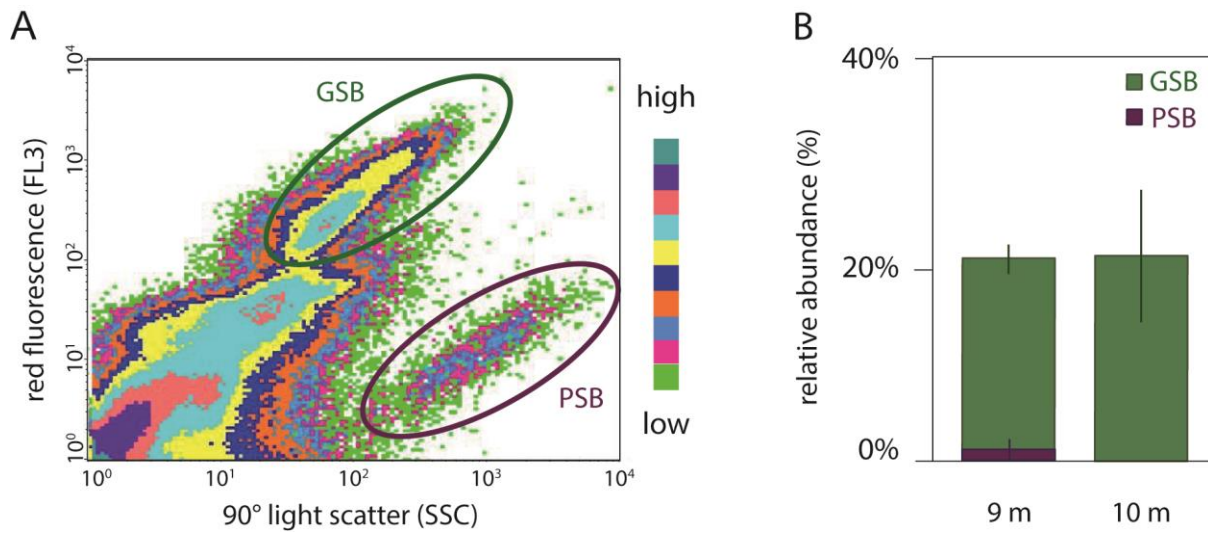
603 Depth profile of Chlorophyll *a* (Chl *a*), Bacteriochlorophyll *a* (BChl *a*), and

604 Bacteriochlorophyll *e* (BChl *e*) concentrations (nmol l⁻¹) in Lake Rogoznica.

605

606





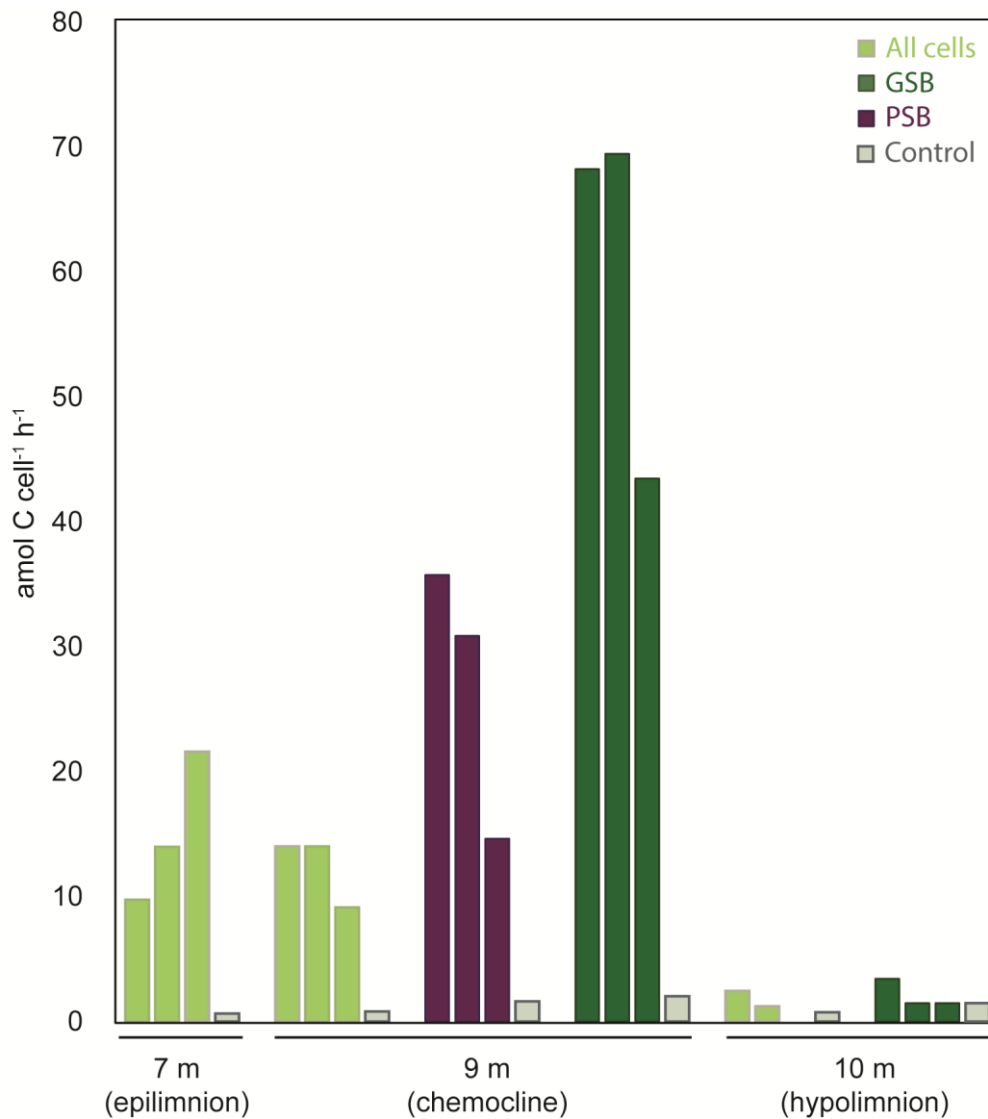
616

617 **Figure 4.**

618 **A** An exemplary sort plot of red fluorescence (FL3; y-axis) versus 90° light scatter (SSC; x-
619 axis) recorded during FACS of a ¹⁴C-bicarbonate incubated chemocline (9 m) sample. The
620 populations identified as GSB (green circle) and PSB (red circle) are highlighted. The color
621 scale bar indicates event frequencies. **B** Relative abundance of sorted GSB (green) and PSB
622 (purple) populations in the chemocline (9 m) and hypolimnion (10 m) samples.

623

624



625

626 **Figure 5.**

627 Averaged cell specific carbon assimilation rates (in $\text{amol C cell}^{-1} \text{ h}^{-1}$) plotted separately for

628 each sorted population from epilimnion (7 m), chemocline (9 m) and hypolimnion (10 m)

629 samples. Sorts of all cells (100,000 cells/sort) are shown in light green, sorts of GSB (50,000

630 cells/sort) are shown in dark green, and sorts of PSB (25,000 cells/sort) are shown in purple.

631 Averaged cell specific carbon assimilation rates (in $\text{amol C cell}^{-1} \text{ h}^{-1}$) for sorts of all cells

632 (100,000 cells/sort) in dead controls are depicted in gray for each depth.