

1 **Edgy conservation: Canadian plants at-risk are overwhelmingly range-edge populations**  
2 **and under-studied**

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4 S. Klemet-N'Guessan<sup>1\*†</sup>, R. Jackiw<sup>2\*</sup>, C.G. Eckert<sup>2\*</sup>, A. L. Hargreaves<sup>1\*††</sup>

5 1. Department of Biology, McGill University, 1205 Dr Penfield Ave, Montreal, Quebec,  
6 Canada, H3A 1B1

7 2. Department of Biology, Queen's University, 116 Barrie St., Kingston, ON, Canada, K7L  
8 3N6

9  
10 \*SKN and RJ made equal contributions, as did CGE and ALH as senior authors.

11 †Current address: Department of Biology, Trent University, 2140 East Bank Drive,  
12 Peterborough, Ontario, Canada, K9L 1Z8.

13 ††Corresponding Author: [anna.hargreaves@mcgill.ca](mailto:anna.hargreaves@mcgill.ca); 514 398 7401

14

15 **ABSTRACT**

16 As biodiversity declines toward the poles, high-latitude countries will contain the poleward range  
17 edge of many species, potentially focusing national conservation toward range-edge populations  
18 whose global conservation value remains contentious. Using the >200 vascular plants assessed  
19 for protection in Canada, we ask whether national species-conservation rankings are biased  
20 toward range-edge populations and supported by adequate research. Of 192 plant taxa deemed  
21 at-risk in Canada, 77% were only found in Canada at the northernmost 20% or less of their  
22 range. Higher threat categories had more peripheral taxa, and the mismatch between national and  
23 global threat rankings was greater for peripheral vs. non-peripheral taxa. Almost half (43%) of  
24 Canadian at-risk plants had not been studied in the peer-reviewed, conservation-relevant  
25 literature, 57% had not been studied in Canada, and peripheral populations received even less  
26 research effort than non-peripheral taxa. Only 5% of 7-9 conservation-relevant studies assessed  
27 at-risk populations in the context of their geographic range—information that is critical to  
28 establishing their relative conservation value. Thus, flora conservation in Canada is largely the  
29 conservation of edge populations, yet edge populations themselves and the geographic context  
30 that makes them unique are understudied, a research gap we must close to improve evidence-  
31 based conservation.

32

33 **Keywords:** species distributions, range limits, wildlife conservation, cold-edge populations,  
34 endangered species, botanical diversity

35

## 36 INTRODUCTION

37 The latitudinal gradient in species diversity poses an interesting conservation dilemma for high-  
38 latitude countries. As species drop out toward higher latitudes, large polar countries—  
39 particularly in the northern hemisphere where landmass is greater—are likely to contain the  
40 poleward range edge of many species. All else being equal, species that occur in a jurisdiction  
41 only at the edge of their range (hereafter ‘peripheral species’) will occupy less area and so have  
42 fewer individuals than non-peripheral species. This potentially makes peripheral species more  
43 likely to be locally rare and deemed ‘at risk’ even if they have robust populations in their range  
44 core (Lesica and Allendorf 1995; Glass et al. 2017). Northern jurisdictions must therefore decide  
45 how much of their conservation resources to spend protecting peripheral versus more endemic  
46 species (Bunnell et al. 2004). This allocation is tricky as the relative conservation value of  
47 peripheral populations remains controversial (e.g. Lesica and Allendorf 1995; Bunnell et al.  
48 2004; Glass et al. 2017)

49 On one hand, peripheral populations may be less important to species’ overall persistence  
50 and harder to conserve than populations closer to the range core (Lesica and Allendorf 1995). In  
51 the absence of dispersal barriers, range edges are thought to arise from declines in the abundance  
52 and quality of habitat, such that peripheral populations are predicted to be less fit, smaller, and  
53 more isolated than central populations (Brown et al. 1996; Eckert et al. 2008). Small populations  
54 are more prone to genetic drift, which can lead to the fixation of deleterious mutations and  
55 interfere with selection of favourable mutations, thereby eroding population fitness and adaptive  
56 potential (Brown et al. 1996). This is particularly true of poleward populations that arose from  
57 post-glacial range expansions, as expansion additionally reduces genetic diversity and fitness via  
58 successive founder effects and ‘surfing’ of deleterious alleles (Waters et al. 2013; Pironon et al.

59 2017). Small, isolated populations are also more prone to extinction from demographic and  
60 environmental stochasticity, and it has often proven challenging to maintain or increase  
61 population sizes at species range edges (Bunnell et al. 2004). For these reasons, some biologists  
62 and policy makers have argued against allocating limited conservation resources to edge  
63 populations (Hunter and Hutchinson 1994; Bunnell et al. 2004).

64         Despite their demographic vulnerability, peripheral populations may harbour distinctive  
65 genetic diversity important for preserving species' evolutionary potential (Hunter and  
66 Hutchinson 1994; Hampe and Petit 2005) and hence capacity for response to rapid  
67 environmental change. If peripheral populations occupy different environments than core  
68 populations (Lesica and Allendorf 1995), they will experience novel selection (Mayr 1954;  
69 Chevin and Lande 2011). Transplant experiments reveal that peripheral populations can be  
70 locally adapted to range-edge conditions (Sexton et al. 2011; Hargreaves and Eckert 2019),  
71 contributing uniquely to intra-species diversity. Dispersal ability is also strongly selected on at  
72 range edges (Phillips et al. 2010; Williams et al. 2016), and both adaptation to local habitat and  
73 increased dispersal ability can prime peripheral populations for colonizing habitat beyond the  
74 species current range (Darling et al. 2008; Hargreaves and Eckert 2014; Hargreaves et al. 2015).  
75 Poleward populations are increasingly valuable in this regard, as they are geographically poised  
76 to initiate poleward range shifts under climate warming (Gibson et al. 2009).

77         As the above controversy suggests, range position alone may be insufficient to determine  
78 a population's conservation value. Edge populations are not always smaller or more isolated than  
79 range-centre populations, nor does range position predict genetic diversity, genetic quality, or  
80 demographic stability consistently enough to guide conservation policy (Eckert et al. 2008;  
81 Pironon et al. 2017). Empirical evidence for local adaptation in peripheral populations is

82 decidedly mixed, though still poorly tested (Hargreaves et al. 2014). Finally, even ‘super edge  
83 populations’ with high performance at many sites in many years may be hard to identify without  
84 large experiments, as ecologically important adaptations can be masked in benign years or by  
85 poor maternal provisioning or inbreeding (Sexton et al. 2011; Hargreaves and Eckert 2019).  
86 Assessing the conservation value of specific edge populations therefore requires empirical  
87 research, yet earlier syntheses suggest peripheral populations are under-studied compared to core  
88 populations (Eckert et al. 2008; Sexton et al. 2009).

89         Canada is an excellent case for studying peripheral species’ representation in  
90 conservation policy and research in high-latitude countries. Canada is the world’s second largest  
91 country, spanning almost 10 million km<sup>2</sup> and more than 41° of latitude—as much latitude as  
92 separates Canada from the Equator. Canada’s biodiversity is clustered at its southern border  
93 (Coristine et al. 2018), and many species only occur in Canada at the northern tip of a much  
94 larger range (Fig. 1). Canadian conservation legislation seems especially likely to protect such  
95 peripheral species, as both the initial conservation assessment body (Committee on the Status of  
96 Endangered Wildlife in Canada; COSEWIC) and federal Species at Risk Act (SARA) largely  
97 ignore populations outside Canada when determining national conservation priority. Previous  
98 estimates suggest that ~75% of Canada’s at-risk species only occur in Canada at their northern  
99 range edge (Yakimowski and Eckert 2007; Gibson et al. 2009). Yet in 2012, three species were  
100 denied federal protection because the government deemed they were ‘only’ peripheral  
101 populations (Fraser 2000; SARA 2012), logic that could jeopardize the protection of many  
102 species in Canada.

103         Here we explore the relationships between peripherality (whether a taxon occurs in a  
104 jurisdiction only at its geographic range edge), conservation priority, and research effort. We

105 consider vascular plants as they are of high conservation value globally but often receive  
106 disproportionately little conservation funding (Raven 1987; Schemske et al. 1994); are relatively  
107 diverse in Canada (e.g. 3600 species compared to ~150 mammals); and make up the largest  
108 proportion (~23%) of taxa assessed by COSEWIC. Using 214 plant taxa with a published  
109 COSEWIC assessment and range map, we designated each as peripheral (northernmost 20% or  
110 less of its range in Canada) or non-peripheral (>20% of its range in Canada). We then tested (i)  
111 whether the proportion of peripheral taxa increases from low to high COSEWIC threat rankings  
112 (not-at-risk, special concern, threatened, endangered) and (ii) whether Canadian and international  
113 threat rankings differ more for peripheral vs. non-peripheral taxa using standardized NatureServe  
114 rankings. We surveyed the research on at-risk taxa to (iii) test whether the conservation-relevant  
115 research effort was evenly spread between peripheral and non-peripheral taxa and (iv) assess  
116 how range-wide studies of at-risk taxa could inform conservation.

117

## 118 **METHODS**

### 119 *Canadian conservation designation*

120 We considered the 214 vascular plant taxa assessed by COSEWIC as of July 2017 for which a  
121 published COSEWIC assessment was available (one not-at-risk species, *Hackelia ciliata*, was  
122 excluded due to lack of published assessment). Eligibility for protection under Canada's Species  
123 at Risk Act (SARA) is determined by the COSEWIC Vascular Plant subcommittee using  
124 quantitative criteria established by the International Union for the Conservation of Nature  
125 (IUCN). Initial COSEWIC assessments consider only populations in Canada (Gärdenfors 2001;  
126 COSEWIC 2017), but can be adjusted by secondary assessments that consider the likelihood of

127 genetic replenishment from core populations, local adaptations in Canadian populations,  
128 Canadian habitat's importance to the taxon's persistence, and the condition of populations  
129 outside Canada (IUCN 2018). COSEWIC assessments use the 'best biological information on a  
130 species', including scientific studies, community knowledge, and Aboriginal traditional  
131 knowledge (SARA 2010). The final decision to nationally protect a taxon under SARA rests with  
132 the federal government after considering the socioeconomic implications (SARA 2010); we use  
133 COSEWIC rather than SARA designations throughout as they more closely reflect biology.

#### 134 *Taxa characterization*

135 For each of the 214 taxa (including species, subspecies, varieties, and populations), we recorded  
136 its COSEWIC designation. 192 taxa were designated at-risk, i.e. of special concern (may become  
137 threatened or endangered due to biological constraints and other threats: 21%), threatened (likely  
138 to become endangered: 22%), or endangered (facing imminent extirpation or extinction: 47%).  
139 The remaining assessed taxa were designated 'not-at-risk' (do not face an imminent extirpation  
140 or extinction risk given the current circumstances: 10%). For four species with separate listings  
141 for two geographically distinct populations (*Eleocharis geniculata*, *Psilocarphus brevissimus*,  
142 *Smilax rotundifolia*, *Solidago speciosa*), we included both listings as separate data points. For  
143 each taxon, we determined its taxonomic group (angiosperm, gymnosperm, or pteridophyte  
144 (ferns and allies)), lifespan (annual/biennial, perennial, or mixed), growth form (herb/graminoid,  
145 woody shrub, or tree), habitat (aquatic or terrestrial), and general location in Canada. This  
146 information was generally in COSEWIC reports, otherwise we used SARA and plants USDA  
147 databases (SARA Act 2011; USDA 2018).

148 (i) Proportion of peripheral taxa from lowest to highest COSEWIC threat ranking

149 To assess whether peripherality is associated with COSEWIC threat ranking, we designated each  
150 taxon as either ‘peripheral’ (only the northernmost 20% or less of their distribution in Canada;  
151 Fig. 1), or ‘non-peripheral’. *Micranthes spicata* had  $\leq 20\%$  of its eastern range edge in the Yukon  
152 with the rest in Alaska; it was deemed non-peripheral. Peripheral designation was determined  
153 visually from range maps in COSEWIC reports (214 taxa), or from the USDA database (USDA  
154 2018) if the COSEWIC report lacked a map (*Pedicularis furbishiae*). For four taxa close to the  
155 20% threshold, we checked visual estimates by digitizing COSEWIC maps and calculating the  
156 proportion range area in Canada (P. Caissy & A.L. Hargreaves, unpublished data).

157 Analyses were done in R 3.5.1 (R Development Core Team 2018). We tested whether the  
158 proportion of peripheral taxa differed among the four COSEWIC threat rankings using a  $4 \times 2$   
159 Fisher’s exact test (*fisher.test*). As it did, we tested which ranks differed from each other using a  
160 pairwise post-hoc comparison holding the overall  $\alpha = 0.05$  (*pairwiseNominalIndependence*,  
161 rcompanion package version 2.1.7 (Mangiafico 2018)). To test whether the pattern of  
162 peripherality vs. threat ranking differed among habitats, life spans, or growth forms, we ran one  
163 binomial generalized linear model (GLM) per grouping, with peripherality as a binomial  
164 response (yes/no) and threat rank and group as interacting categorical predictors (peripheral  $\sim$   
165 threat\_rank  $\times$  group, *glm*, logit link function, Fig. S1). Models of life span excluded 5 taxa with  
166 ‘mixed’ life spans as there were no ‘not-at-risk’ taxa in this category. Group levels are described  
167 above (*Taxa characterization*). Significance of GLM predictors was assessed using type III tests  
168 (*Anova*, car package version 2.1.4 (Fox and Weisberg 2018)). Finally, we tested whether  
169 peripheral taxa were more endangered overall by converting the four COSEWIC categories to



170 numeric ranks (1-4) and testing whether this differed between peripheral and non-peripheral taxa  
171 using a non-parametric two-sample Wilcoxon test (*wilcox.test*).

172 *(ii) Canadian vs. global NatureServe threat rankings*

173 We compared Canadian vs. global threat rankings from NatureServe (i.e. ‘rounded global status’;  
174 NatureServe 2018); NatureServe ranks use consistent criteria so national and global rankings are  
175 directly comparable. Ranks range from 1 (most threatened) to 5 (least threatened, NatureServe  
176 2018). While NatureServe also considers taxa below the species level, Canadian or global ranks  
177 were missing for seven taxa (three at-risk and four not-at-risk), so  $n = 207$  taxa (full list of taxa’s  
178 COSEWIC, SARA and NatureServe ranks in Dryad data). We calculated the ‘rank discrepancy’  
179 for each taxon as ‘Global rank – Canadian rank’. Values are numeric but only range from 0 (i.e.  
180 no difference) to 4 (i.e. taxon ranked as 5 (least threatened) globally and 1 (most threatened) in  
181 Canada), so are non-parametric. We tested whether rank discrepancy differed between peripheral  
182 and non-peripheral taxa using a two-sample Wilcoxon test.

183         Although Nature Conservancy and COSEWIC ranks use different numbers of categories  
184 (five and four, respectively), both are derived from IUCN criteria. To test whether the two bodies  
185 ranked taxa consistently, we converted COSEWIC ranks to numeric ranks from 4 (not-at-risk) to  
186 1 (endangered); as NatureServe did not give any species a 5 (least threatened) in Canada, both  
187 NatureServe Canadian ranks and numeric COSEWIC ranks varied from 1 to 4. We used a paired  
188 Wilcoxon test to assess whether COSEWIC and NatureServe ranks for Canadian populations  
189 differed overall (the five taxa with two distinct COSEWIC designations were each counted  
190 twice). As NatureServe and COSEWIC ranks did differ significantly (Fig. S2a), we tested  
191 whether the difference between NatureServe – COSEWIC Canadian ranks was greater for  
192 peripheral taxa (unpaired Wilcoxon test; Fig. S2b), and whether it altered conclusions re. the

193 ‘Global rank – Canadian rank’ (paragraph above) by redoing that analysis replacing NatureServe  
194 Canadian ranks with COSEWIC ranks (Fig. S3).

195 *(iii) Research effort on plant taxa considered at-risk in Canada*

196 To assess the research effort on plants at-risk in Canada, we searched ISI Web of Science for  
197 studies on each at-risk taxon using its Latin name, common name, and synonyms listed in its  
198 COSEWIC assessment. While COSEWIC evaluated some specific subspecies, varieties or  
199 populations, we found few studies for these smaller designations so considered all taxa at the  
200 species level ( $n = 189$  at-risk species). We narrowed results to ecological or evolutionary studies  
201 by including a second search term “*\*ecolog\* OR evolution\* OR population\* OR demograph\**  
202 *OR genetic\* OR conservation\* OR fitness*”. We discarded studies that did not present data on the  
203 taxon of interest (e.g. only mentioned it in key words), yielding 2940 studies.

204 We categorized the resulting studies on their conservation and geographic relevance. We  
205 categorized conservation relevance as: 1—no data on natural populations or that could be used  
206 by COSEWIC; 2—data from natural populations but not relevant to COSEWIC (e.g. how much  
207 a plant species contributed to a herbivore’s diet); or 3—data of potential use to COSEWIC and  
208 conservation (e.g. performance, local adaptation or genetic diversity of natural populations);  
209 ultimately we only considered this third category in analyses (below). We also classified whether  
210 studies sampled wild Canadian populations, wild populations in the USA, both (providing a  
211 wider geographic context for the at-risk Canadian populations), or neither (sampled populations  
212 outside Canada/USA or no specific population). Studies that investigated more than one at-risk  
213 taxon were counted for each taxon included.

214 We then assessed the distribution of research effort potentially relevant to conservation  
215 (category 3). We first tested whether species that were peripheral or not peripheral in Canada

216 (binary predictor) differed in the probability that they had been studied anywhere in their range  
217 (yes/no response; binomial GLM), or differed in the number of studies per species from  
218 anywhere in their range (over-dispersed count response, negative binomial GLM, *glm.nb*, MASS  
219 package version 7.3.45 (Venables and Ripley 2013)). We then reran both analyses considering  
220 only studies that included Canadian populations. For negative binomial models, we assessed the  
221 significance of peripherality by comparing the model with peripherality as a predictor ( $\#studies \sim$   
222  $peripheralYN$ ) to the null model ( $\#studies \sim 1$ ) using a likelihood ratio test with significance  
223 evaluated using the Chi squared distribution (*anova*,  $df = 1$ ).

#### 224 *(iv) Conservation insights from comparing edge vs core populations*

225 Studies that sample populations from both Canada and the USA should reveal the most about the  
226 relative conservation value of peripheral populations in Canada. We therefore read each such  
227 study and noted examples where the results yielded insights relevant to conservation that could  
228 not have been gleaned from smaller-scale sampling.

## 229 **RESULTS**

230 Of 192 at-risk plant taxa in Canada, 77% were peripheral, with only the northernmost 20% or  
231 less of their distribution in Canada. Many occurred only in Ontario or British Columbia (68% of  
232 148 peripheral taxa vs. 30% of 44 non-peripheral taxa; Fig. 1). Thus, these provinces are  
233 disproportionately responsible for conserving at-risk plants, and peripheral species in particular.

234

#### 235 *(i) Proportion of peripheral taxa from lowest to highest COSEWIC threat ranking*

236 The percentage of plant taxa that only occur peripherally in Canada increased with increasing  
237 endangerment ( $\chi^2_{df=3} = 17.7$ ,  $P = 0.0005$ ). The percentage of peripheral taxa was significantly

238 higher in the Endangered vs. Special concern rank (post-hoc  $P = 0.0007$ ), and the differences  
239 between Endangered and Not at risk and between Threatened and Special concern were almost  
240 significant (post-hoc  $P = 0.055$ ; Fig. 2). The pattern of peripherality vs. threat ranking was  
241 consistent between aquatic vs. terrestrial habitats, herbaceous/graminoid vs. woody growth  
242 forms, and annual/biennial vs. perennial life spans (Fig. S1). Overall, peripheral taxa were  
243 significantly more endangered than non-peripheral taxa (Wilcoxon test  $W = 2827$ ,  $P < 0.001$ ).

244

245 *(ii) Canadian vs. global NatureServe threat rankings*

246 For the 207 taxa in our dataset with both Canadian and global NatureServe ranks (157 peripheral,  
247 50 not peripheral), the disparity between Canadian and global ranks was greater for taxa that are  
248 peripheral in Canada compared to taxa that are not (Wilcoxon test,  $W = 6756$ ,  $P < 0.001$ ; Fig. 3).

249 Of the 203 plant taxa that NatureServe considered at-risk (ranks 1 to 3) in Canada, 67% were  
250 considered secure (ranks 4 or 5) across their global range; most of these at-risk-nationally but  
251 secure-globally taxa are peripheral in Canada (125 peripheral, 14 non-peripheral; Fig. 3).

252 NatureServe tended to consider taxa more threatened than COSEWIC, but this difference  
253 reflected higher threat rankings for non-peripheral taxa (Fig. S2); the greater discrepancy  
254 between national and global threat ranks for peripheral taxa (Fig. 3) remains if one compares  
255 NatureServe global ranks and COSEWIC ranks directly (Fig. S3).

256

257 *(iii) Research effort on plant species considered at-risk in Canada*

258 We found 709 conservation-relevant studies on plant species considered at-risk (by COSEWIC)  
259 in Canada. Almost half (43%) of Canada's 189 at-risk plant species had not been studied in peer-  
260 reviewed work that could inform conservation (Fig. 4a). Though this does not preclude the

261 existence of studies in the non-refereed literature or journals not indexed on Web of Science, it  
262 suggests that the ‘best biological information’ is sparse for these taxa. Species peripheral in  
263 Canada and species not peripheral in Canada did not differ in the likelihood that they had been  
264 studied somewhere in their range (43% vs 45%; binomial GLM: peripheral\_designation  $\chi^2_{df=1} =$   
265 0.10,  $P = 0.75$ ) or in the number of studies overall (negative binomial GLM:  
266 peripheral\_designation  $\chi^2_{df=1} = 0.89$ ,  $P = 0.34$ ; Fig. 4a).

267         Of the 709 conservation-relevant studies, only 188 included Canadian populations; more  
268 than half (57%) of plant species at-risk in Canada had not been studied in Canada (Fig. 4b).  
269 Compared to non-peripheral species, species that are peripheral in Canada were as likely to have  
270 been studied in Canada (57% vs 55%; binomial GLM: peripheral\_designation  $\chi^2_{df=1} = 0.16$ ,  $P =$   
271 0.69) but had significantly fewer studies per species that included Canadian populations  
272 (negative binomial GLM: peripheral\_designation  $\chi^2_{df=1} = 7.27$ ,  $P = 0.0070$ ; Fig. 4b).

273

#### 274 (iv) Conservation insights from comparing edge vs core populations

275 Only 42 (5%) of 709 studies on plants at-risk in Canada studied Canadian populations in the  
276 context of their geographic range (i.e. included both Canadian and USA populations). These  
277 studies covered 21 (11%) of 189 at-risk species, and almost half were on one of three species:  
278 Whitebark pine (*Pinus albicaulis*; 8 studies), Butternut (*Juglans cinerea*, 7 studies), and Juniper  
279 sedge (*Carex juniperorum*, 3 studies). Studies that took a range-wide perspective provided  
280 unique insights into the conservation of peripheral populations (Table 1). These include whether  
281 edge populations differ from core populations genetically or demographically, or in their traits or  
282 habitat. For instance, populations of Deerberry decreased in size and frequency toward the  
283 species northern range edge in Canada, but nevertheless were as productive and genetically

284 diverse as core populations, and showed evidence of local adaptation and high dispersal ability  
285 (Yakimowski and Eckert 2007, 2008). Thus the demographic and genetic value of these  
286 populations was not predicted by their peripherality, size, or spatial isolation.

## 287 **DISCUSSION**

288 Our results show that conservation of plants in Canada is fundamentally the conservation of  
289 range-edge populations. Three quarters of nationally at-risk plant taxa only occur in Canada at  
290 the northernmost 20% or less of their North American range, in line with earlier estimates for all  
291 at-risk taxa combined (Gibson et al. 2009). While many not-at-risk plants are also peripheral in  
292 Canada (Fig. 2), peripheral taxa had higher threat rankings overall and a greater discrepancy  
293 between their Canadian vs. global threat ranking (Fig. 3). Higher national threat ranks for  
294 peripheral taxa could reflect real increased risk per population, since human activity is also  
295 highest in southern Canada (Coristine and Kerr 2011), or the precautionary principle since  
296 Canada has little control over the fate of populations outside its borders. Either way, northern-  
297 edge populations are geographically poised to initiate northward range shifts under climate  
298 warming, so may be critical for Canada's future biodiversity (Gibson et al. 2009). Moreover,  
299 northward range shifts will bring new species to Canada's southern border, requiring an  
300 informed policy on conserving peripheral populations. Unfortunately, conservation of peripheral  
301 taxa has been debated in the absence of much relevant scientific evidence.

302 While peripheral populations are prioritized for conservation in Canada, they do not seem  
303 prioritized for peer-reviewed research that could inform their conservation. Based on our Web of  
304 Science survey, more than half of the plant species with at-risk populations in Canada had not  
305 been studied in Canada in a way that could guide their conservation, and peripheral plants had

306 significantly fewer studies that included Canadian populations than non-peripheral plants (Fig.  
307 4). Even the best-studied at-risk peripheral plant (*Cirsium pitcheri*) had only 8 studies that  
308 included Canadian populations. While this could reflect difficulty in obtaining permits or  
309 adequate sample sizes, taxonomic bias is likely. For example, one bird species that is both  
310 peripheral and at-risk in Canada had almost 50 studies that included Canadian populations  
311 (Marbled murrelet; Web of Science search May 2019). Thus peripheral plants appear  
312 systematically under-studied in Canada.

313 Conservation-relevant studies that include both Canadian and US populations illustrate  
314 the value of studying peripheral populations directly and in a broad geographical context.  
315 However, most such studies have assessed neutral genetic diversity and population structure  
316 (Table 1). Conservation would particularly benefit from studies that assess characteristics  
317 important for long-term persistence and range expansion, such as habitat preferences, population  
318 demography and dispersal ability (Schemske et al. 1994). Future genetic work could move  
319 beyond neutral variation to evaluating the adaptive diversity likely to be important in responding  
320 to environmental change (Shaw and Etterson 2012), and local adaptation through which range-  
321 edge populations may contribute uniquely to species' biodiversity (Yeaman et al. 2016). Whether  
322 researchers will close these knowledge gaps depends partially on how government agencies  
323 incentivise (i.e. fund) and remove barriers to (i.e. permit) research on at-risk peripheral  
324 populations. Unfortunately, the “peripherality issue” is not currently highlighted in federal  
325 programs that fund species-at-risk research in Canada (e.g. Government of Canada 2019).

326 We hope that exposing the lack of peer-reviewed study inspires future work on at-risk  
327 edge populations, but recognize that amassing this work will take time, and that some of the most

328 informative types of study, e.g. large reciprocal transplants, will be impossible with endangered  
329 taxa. We therefore suggest that we have an excellent and potentially under-used body of research  
330 that could inform Canadian conservation: the already extensive theory and empirical research on  
331 species range edges (Sexton et al. 2009; Pironon et al. 2017). While this research clearly shows  
332 that edge populations can vary significantly from one another in demography (Sagarin et al.  
333 2006) and adaptation (Hargreaves and Eckert 2019), it also reveals broad scale patterns that can  
334 be predictive, e.g. that poleward range edges are often dispersal limited whereas high-elevation  
335 edge populations are often population sinks (Halbritter et al. 2013; Hargreaves et al. 2014). For  
336 countries like Canada whose biodiversity is disproportionately comprised of range-edge  
337 populations, leveraging the species-distribution literature could help inform conservation,  
338 helping meet our commitments to protecting current and future biodiversity.

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### 346 **AUTHOR CONTRIBUTIONS**

347 CGE designed and supervised the original study, which RJ conducted as an undergraduate  
348 project (2014). SKN updated the database in 2017, ALH did the statistics, and SKN and ALH  
349 co-wrote the current manuscript with edits from CGE and RJ.



350 **Table 1.** Studies with wide geographic coverage can shed light on the conservation needs and  
 351 value of peripheral populations. Key findings taken from studies of plant taxa deemed at-risk in  
 352 Canada that sampled both Canadian and USA populations. All taxa are peripheral in Canada or  
 353 threatened throughout their range.

<b>Response</b>		
Finding for edge populations	Species	Reference
<b>Genetic uniqueness</b>		
Genetically differentiated from core populations (neutral variation)	Branched bartonia	Ciotir et al. 2013
	Green dragon	Boles et al. 1999
	Spalding's Campion	Lesica et al. 2016
Greatest population differentiation	Butternut	Hoban et al. 2010
Second highest cold tolerance	Whitebark pine	Bower and Aitken 2006
<b>Genetic diversity</b>		
Not lower neutral genetic diversity than core populations	Heartleaf plantain	Mymudes and Les 1993
	Deerberry	Yakimowski and Eckert 2008
	Green dragon	Boles et al. 1999
	Golden paintbrush	Godt et al. 2005
	Cucumber tree	Budd et al. 2015
	Whitebark pine	Liu et al. 2016
<b>Population demography</b>		
Not smaller than core populations	Golden paintbrush	Godt et al. 2005
Smaller & more isolated than core populations	Deerberry	Yakimowski and Eckert 2007
<b>Performance &amp; ecology</b>		
Similar habitat to core populations	Small whorled pogonia orchid	Mehrhoff 1989
Lowest seed viability & distinct dormancy patterns	Green dragon	Yang et al. 1999

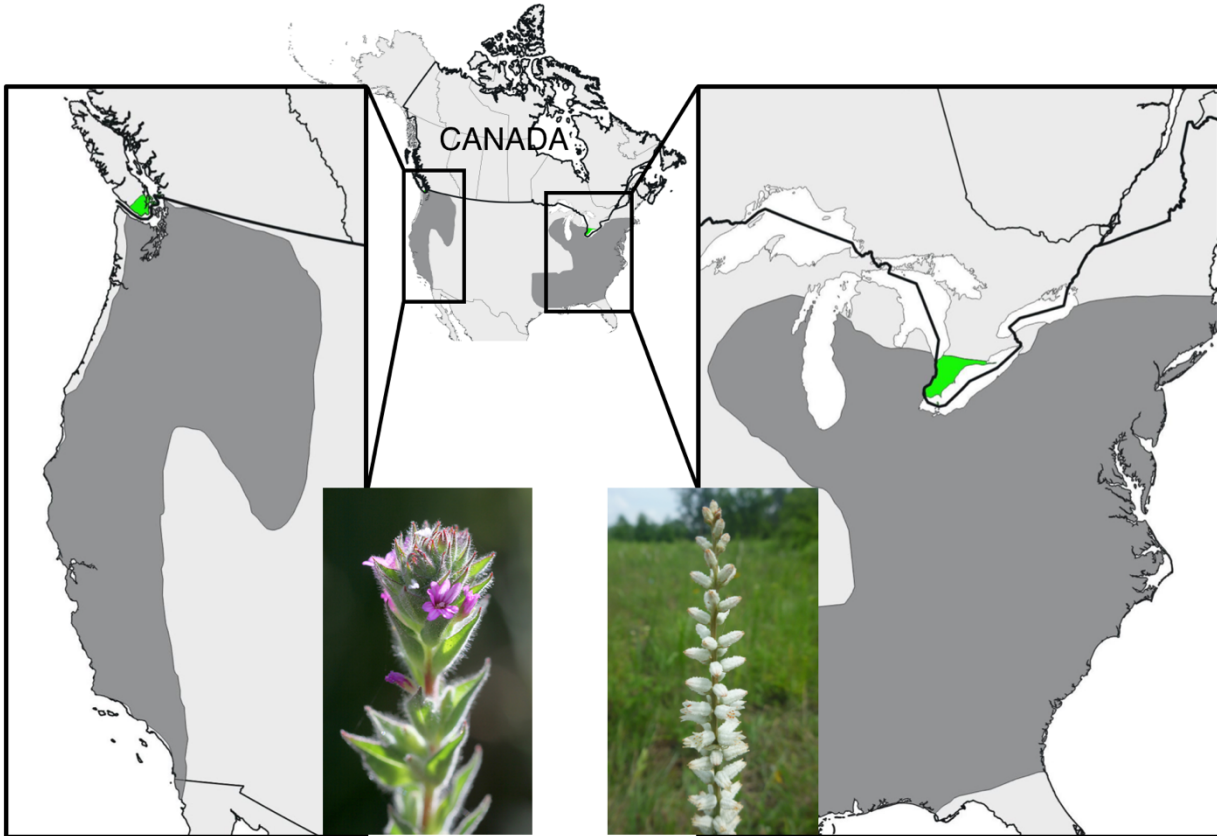
Higher seed mass & equal  
sexual reproduction /  
productivity as core  
populations

Deerberry

Yakimowski and Eckert 2007

- 
- 354 Peripheral: Branched bartonia = *Bartonia paniculata ssp. paniculata*; Cucumber tree = *Magnolia*  
355 *acuminata*; Deerberry = *Vaccinium stamineum*; Golden paintbrush = *Castilleja levisecta*; Green  
356 dragon = *Arisaema dracontium*; North American ginseng = *Panax quinquefolius*; Small whorled  
357 pogonia orchid = *Isotria medeoloides*; Heartleaf plantain = *Plantago cordata*; Spalding's  
358 campion = *Silene spaldingii*
- 359 Not peripheral: Butternut = *Juglans cinerea*; Whitebark pine = *Pinus albicaulis*  
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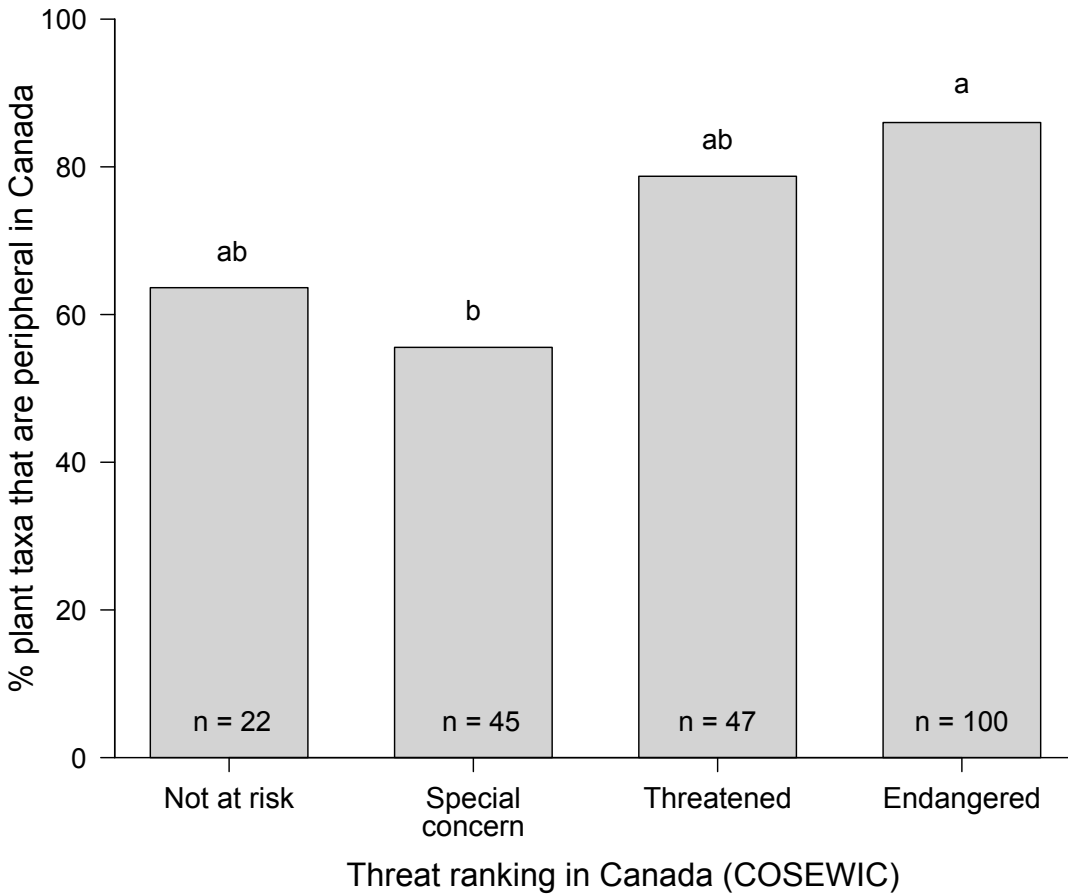
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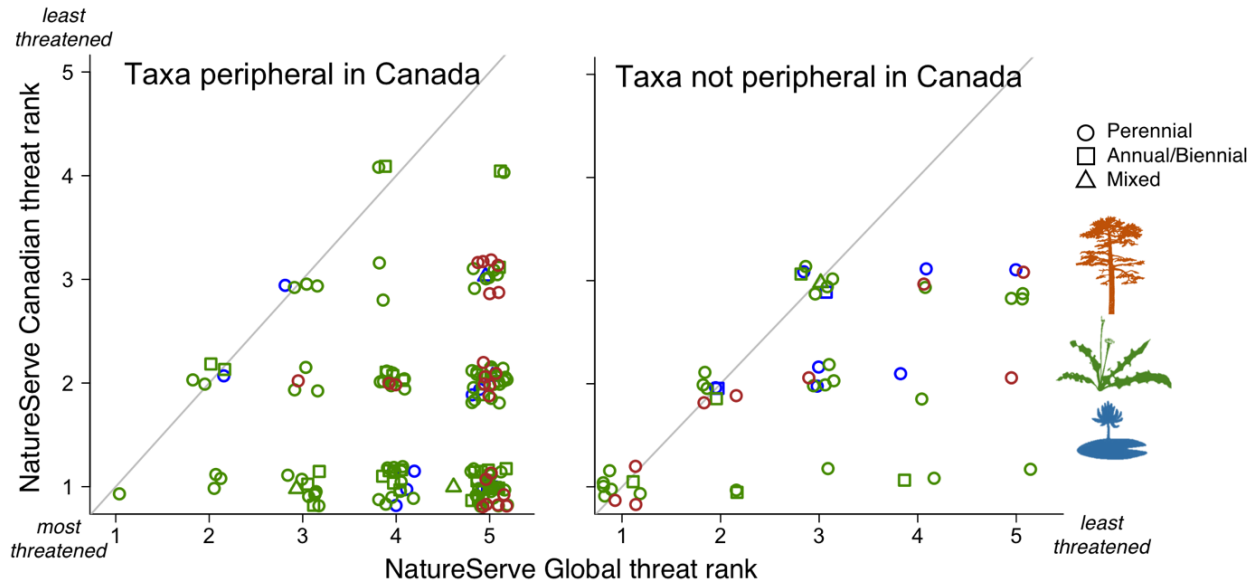
363 **Figure 1. Two examples of ‘peripheral’ species that only occur in Canada at the**  
364 **northernmost 20% or less of their range.** Left is Dense spike-primrose (*Epilobium*  
365 *densiflorum*, Onagraceae), right is Colicroot (*Alettris farinosa*, Nartheciaceae), both endangered  
366 in Canada. Green polygons show each species’ Canadian distribution, dark grey polygons show  
367 its distribution outside Canada. As was the case for 70% of plant taxa considered at risk in  
368 Canada, these species occur in the Canadian biodiversity hotspots of southern British Columbia  
369 (left) or southern Ontario (right). Distribution maps were modified from COSEWIC reports  
370 (COSEWIC 2005, 2015) by Pascale Caissy. Photo credits: Debra L Cook (left); Victoria  
371 McPhail (right).

372



373

374 **Figure 2. Higher threat rankings include more peripheral taxa.** The percentage of peripheral  
375 taxa (those with  $\leq 20\%$  of their northernmost range in Canada) varied among COSEWIC threat r  
376 ankings (total  $n = 214$  plant taxa). Threat ranks are from least threatened (Not at risk) to most thr  
377 eatened (Endangered). The number of taxa per rank is shown at the base of each bar. Contrasting  
378 letters indicate significant differences among categories; the differences between Endangered an  
379 d Not at risk and between Threatened and Special Concern were almost significant ( $P = 0.055$ ).



380

381

382

383 **Figure 3. Discrepancy between Canadian and global threat ranks is greater for peripheral**

384 **species.** Diagonal lines indicate Canadian populations have the same threat ranking as the taxon

385 globally. Taxa that are peripheral in Canada (left,  $n = 157$  taxa) have a greater mismatch between

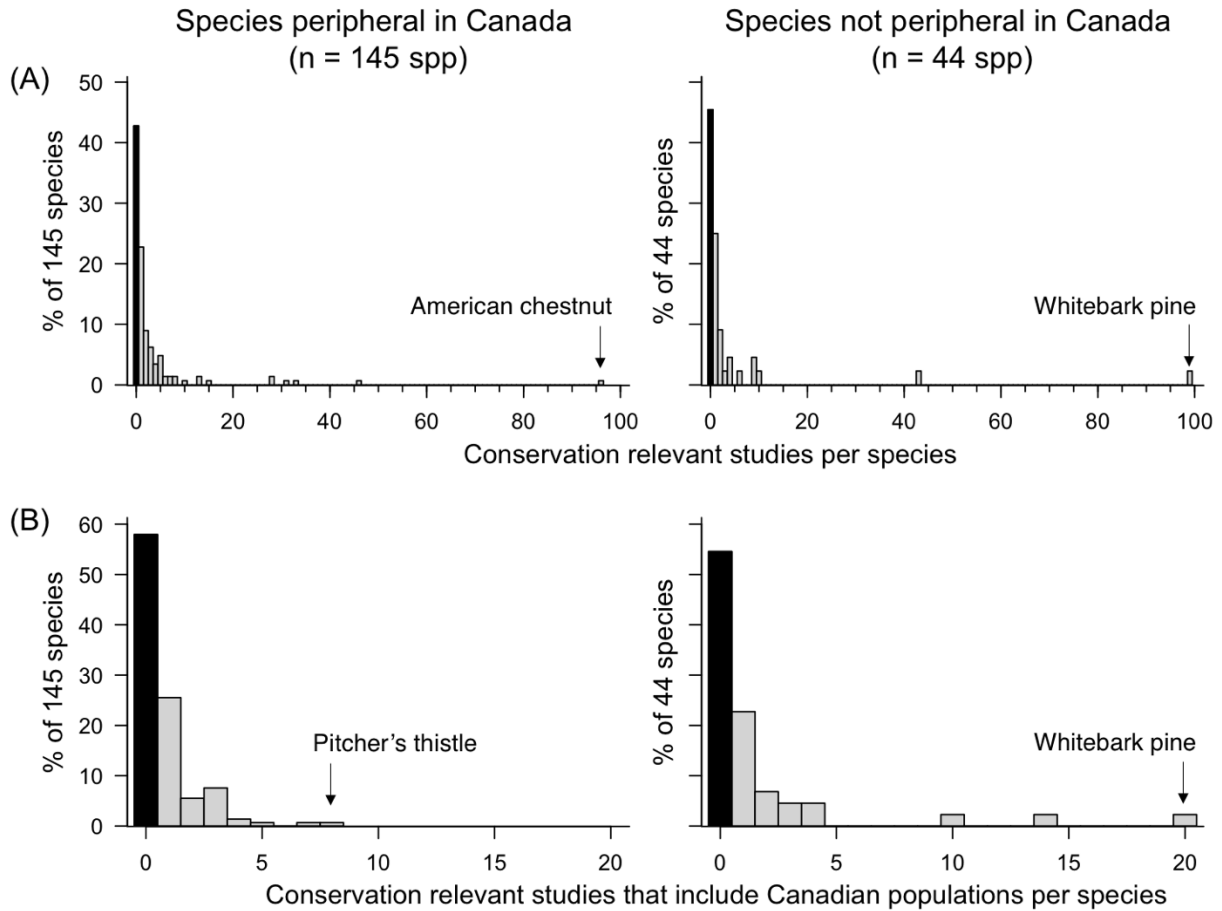
386 their Canadian and global threat ranks (more taxa listed as threatened in Canada but secure

387 globally) than taxa with >20% of their range in Canada (right,  $n = 50$  taxa). Point shape indicates

388 lifespan; colour indicates growth form and habitat (brown = woody shrub or tree, green = non-

389 woody terrestrial plant, blue = aquatic plant).

390



391

392

393 **Figure 4. Distribution of conservation-relevant research on plant species at-risk in**

394 **Canada.** (A) Top panels show distribution of all 702 ecological or evolutionary studies that

395 could inform conservation; (B) bottom panels show 188 studies from (A) that included data from

396 wild Canadian populations. Left panels = peripheral species ( $\leq 20\%$  of their range in Canada);

397 right panels = non-peripheral species  $>20\%$  of their range in Canada). Black bars indicate species

398 for which we found no peer-reviewed studies.

399

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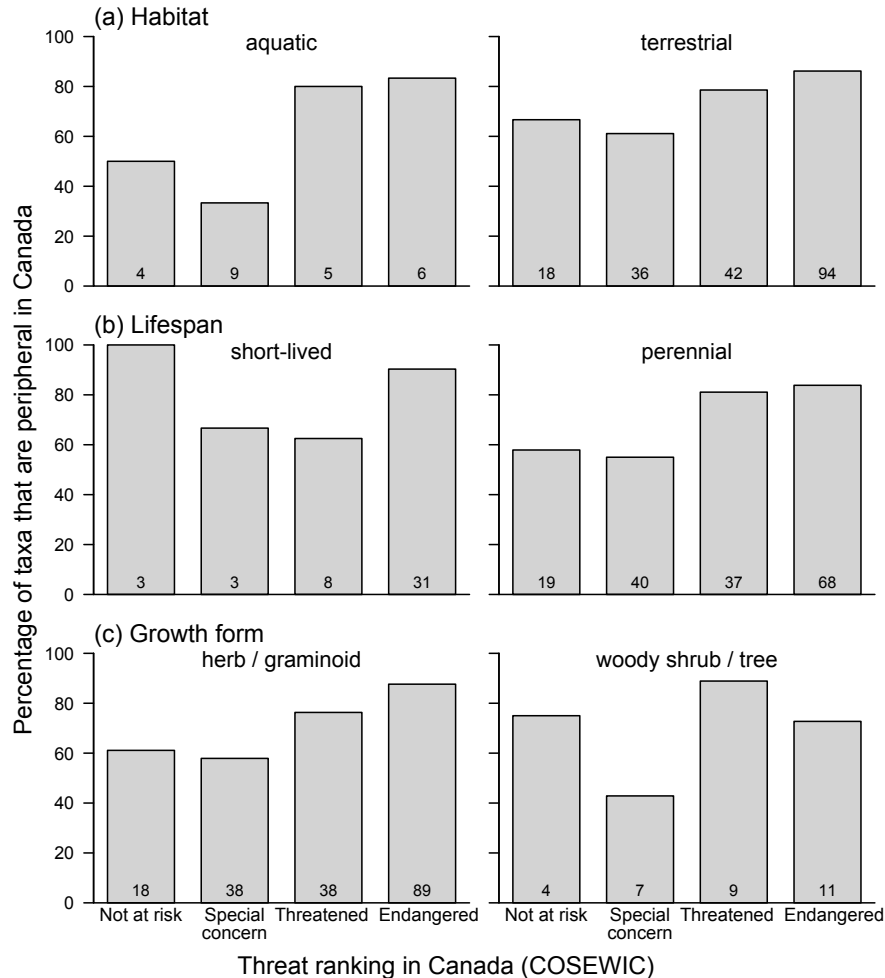
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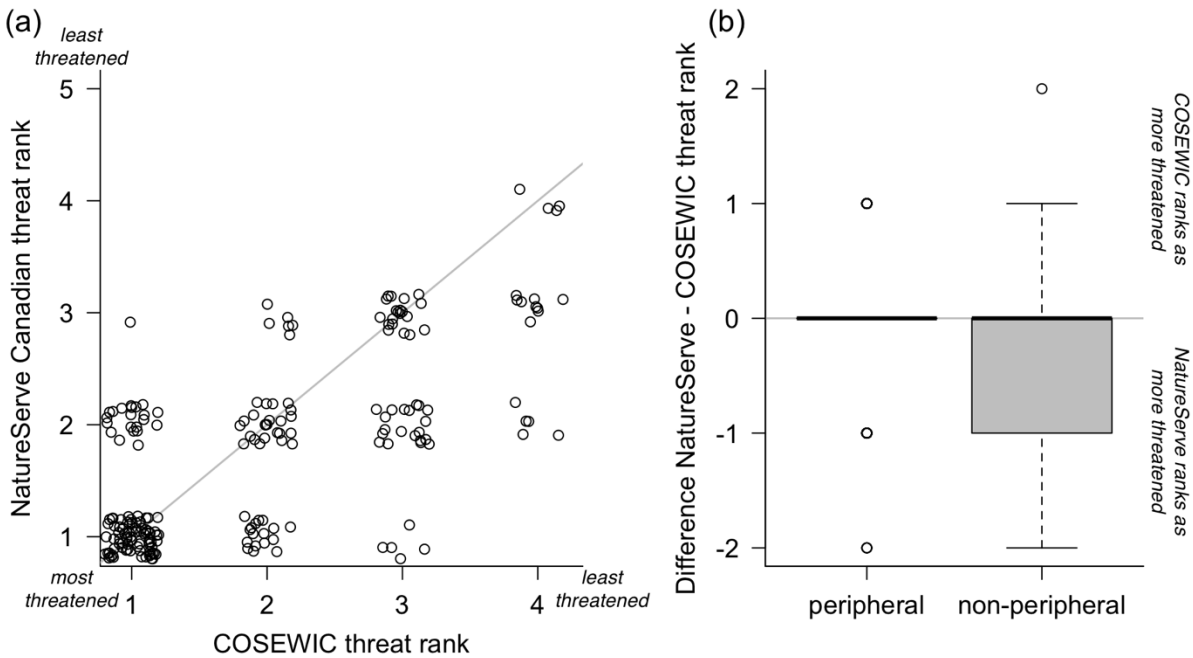
## Supplementary Material for

### Edgy conservation: Canadian at-risk plants are overwhelmingly range-edge populations and under-studied

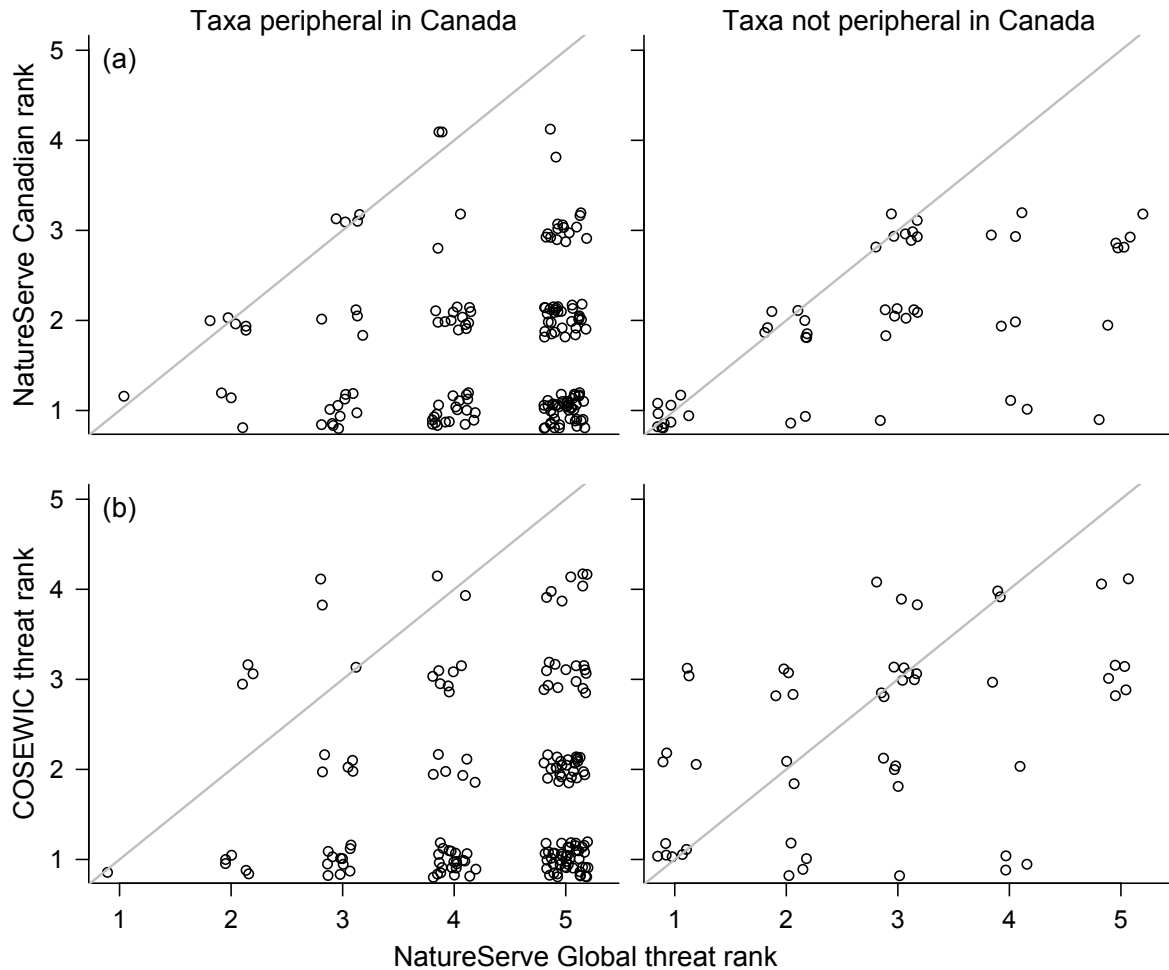


**Figure S1. The relation between peripherality and threat status does not differ between habitats, lifespans, or growth forms.** Each panel was assessed with a binomial GLM:

proportion peripheral taxa  $\sim$  threat\_rank  $\times$  group, where group was habitat, lifespan, or growth form. The interaction term was non-significant in full models (threat  $\times$  habitat:  $\chi^2_{df=3} = 0.99$ ,  $P = 0.80$ ; threat  $\times$  lifespan:  $\chi^2_{df=3} = 4.60$ ,  $P = 0.20$ ; threat  $\times$  growth form:  $\chi^2_{df=3} = 2.92$ ,  $P = 0.41$ ), so was dropped from final models. In all final models (peripheral  $\sim$  threat\_rank + group), the proportion of peripheral taxa did not differ among groups (habitat:  $\chi^2_{df=1} = 1.68$ ,  $P = 0.19$ ; lifespan:  $\chi^2_{df=1} = 0.51$ ,  $P = 0.47$ ; growth form:  $\chi^2_{df=1} = 0.20$ ,  $P = 0.65$ ), but differed significantly with COSEWIC status (more peripheral taxa in the endangered vs. special concern rank, as per main results Fig. 2). Threat categories are listed from least to most threatened, a taxon is considered peripheral if it occurs in Canada at the northernmost 20% or less of its range, and numbers at bottom give the number of taxa per bar (total n is lower for (b) as 5 taxa with a mixed lifespan (annual/perennial or biennial/perennial) were excluded).



**Figure S2. NatureServe ranks Canadian plant populations as more nationally-imperiled than does COSEWIC.** NatureServe ranks range from 1 (most threatened) to 5 (least threatened), but none of the plant taxa assessed by COSEWIC were given a rank of 5. COSEWIC categorical ranks were converted to numeric ranks from 1 = endangered to 4 = not at risk. (a) Even assuming equivalent scales (i.e. NatureServe 4 equivalent to COSEWIC 4), NatureServe ranks of Canadian populations differed from COSEWIC ranks (paired Wilcoxon test  $W = 1078$ ,  $P = 0.0006$ ;  $n = 207$  taxa) in that NatureServe tended to rank taxa as more nationally threatened than COSEWIC. (b) The difference between NatureServe and COSEWIC rankings in (a) was driven primarily by non-peripheral taxa, which showed a bigger discrepancy (unpaired Wilcoxon test  $W = 4788$ ,  $P = 0.008$ ).



**Figure S3. Peripheral taxa have a greater discrepancy between Canadian and global threat ranks whether one uses NatureServe Canadian ranks (a) or COSEWIC ranks (b).** Threat rankings are described in Fig. S2 (1=most threatened). (a) Same data as shown in Fig. 3: the difference in NatureServe Global – NatureServe Canadian threat ranks is greater for peripheral vs non-peripheral species. (b) The difference between NatureServe Global – COSEWIC Canadian ranks is also greater for peripheral vs. non-peripheral species (Wilcoxon test,  $W=7072$ ,  $P < 0.001$ ).