# 1 Developmental exposure to pesticide contaminated food impedes bumblebee brain growth 2 predisposing adults to become poorer learners 3 Dylan B. Smith<sup>1</sup>, Andres N. Arce<sup>1</sup>, Ana Ramos Rodrigues<sup>1</sup>, Philipp H. Bischoff<sup>1</sup>, Daisy Burris<sup>1</sup>, Farah Ahmed<sup>2</sup> and Richard J. Gill<sup>1\*</sup> 4 5 <sup>1</sup>Department of Life Sciences, Imperial College London, Silwood Park, Buckhurst Road, Ascot, Berkshire, 6 SL5 7PY, UK. 7 <sup>2</sup> Core Research Laboratories, Natural History Museum, Cromwell Road, London, SW7 5BD, UK. 8 \* correspondence: r.gill@imperial.ac.uk

9 Keywords: 3D reconstruction, Bombus terrestris; Imidacloprid, micro computed tomography scanning,

10 mushroom body calyces, neonicotinoid, olfactory, segmentation, sublethal

## 11 Abstract

Understanding the risk to biodiversity from pesticide exposure is a global priority. For bees, an 12 13 understudied step in evaluating pesticide risk is understanding how pesticide contaminated foraged food brought back to the colony can affect developing individuals. Provisioning bumblebee colonies with 14 15 pesticide (neonicotinoid) treated food, we investigated how exposure during two key developmental phases (brood and/or early-adult), impacted brain growth and assessed the consequent effects on adult 16 17 learning behaviour. Using micro-computed tomography ( $\mu$ CT) scanning and 3D image analysis, we 18 compared brain development for multiple neuropils in workers 3 and 12-days post-emergence. 19 Mushroom body calyces were the neuropils most affected by exposure during either of the developmental 20 phases, with both age cohorts showing smaller structural volumes. Critically, reduced calyces' growth in 21 pesticide exposed workers was associated with lower responsiveness to a sucrose reward and impaired 22 learning performance. Furthermore, the impact from brood exposure appeared irrecoverable despite no 23 exposure during adulthood.

24

## 25 Introduction

26 Insect pollinator declines are of worldwide concern (Hallmann et al., 2017; Potts et al., 2016; Vanbergen, 27 2013), and safeguarding this important functional group and ecosystem service provider requires a deep 28 understanding of the driving factors (Gill et al., 2016; Goulson et al., 2015). Social bees, such as 29 bumblebees, honeybees and stingless bees are important insect pollinators, and the threat posed by 30 pesticide exposure is a widespread issue (Brittain and Potts, 2011; Desneux et al., 2007; Woodcock et al., 31 2017). Pesticide residues have been found inside colonies across the globe (Calatayud-Vernich et al., 2018; Daniele et al., 2018; Mitchell et al., 2017; Mullin et al., 2010; Valdovinos-Flores et al., 2017), raising 32 33 concerns as to how the prevalence of such chemicals in the environment could affect colony development 34 (Gill et al., 2012; Pohorecka et al., 2017; Whitehorn et al., 2012). For instance, controlled exposure 35 experiments and field studies investigating exposure to neonicotinoid pesticides have reported reduced 36 colony growth and sexual output (Arce et al., 2017; Baron et al., 2017; Gill et al., 2012; Leza et al., 2018; 37 Rundlöf et al., 2015; Tsvetkov et al., 2017; Whitehorn et al., 2012). Such colony level effects are likely to 38 be caused by exposure impairing worker behaviour, with cumulative effects across workers leading to a 39 functionally weakened colony (Bryden et al., 2013; Crall et al., 2018). One possibility is that neonicotinoids, 40 being a neurotoxic pesticide, affect neuronal processes important for cognitive and learning abilities 41 (Decourtye et al., 2004; Siviter et al., 2018b) translating to impaired colony tasks (Feltham et al., 2014; 42 Fischer et al., 2014; Gill and Raine, 2014).

With neurotoxic pesticide residues frequently reported in the pollen and nectar brought back by foragers
(Botias et al., 2015; Chauzat et al., 2006; David et al., 2016; Kasiotis et al., 2014; Pohorecka et al., 2012),
individuals developing and residing in the colony are likely to be chronically exposed to these compounds
(Pohorecka et al., 2017). A possibility, therefore, is that tissue development, such as the central nervous

system, is impeded. For example, honeybees reared under sub-optimal environmental conditions have 47 48 exhibited reduced brain volumetric growth and altered neuronal architecture (Groh et al., 2004; Steijven 49 et al., 2017). Impeded brain development and structural plasticity may impact on behaviours such as 50 learning ability, that require detection, assimilation and processing of sensory input from the environment 51 (Cabirol et al., 2018; Chittka, 2017; Galizia et al., 2011). Knowledge of how pesticide contaminated food 52 inside bee colonies can affect individual physiological development however, is limited (Gregorc et al., 53 2012; Wu et al., 2012, 2011). Moreover, there has been an urgent call for research linking how potential 54 pesticide induced impairment to brain development can translate to task performance as later adults (Siviter et al., 2018b; Tan et al., 2015; Tomé et al., 2012; Yang et al., 2012). We directly address this call by 55 56 investigating how developing bumblebees exposed to a neonicotinoid pesticide via treated provisioned 57 food may alter brain development and link this to effects on associative learning behaviour.

58 To experimentally test the effect of pesticide exposure on individual development we needed to first 59 consider the level of developmental plasticity in behaviour and brain growth. In social bees, worker 60 maturation can correlate with stereotyped behavioural changes (Goulson, 2010; Johnson, 2010), and 61 increased brain neuropil volumes (functional structures) (Durst et al., 1994; Galizia et al., 2011; Li et al., 62 2017; Winnington et al., 1996; Withers et al., 1993). However, it has been recognised that dissecting 63 innate effects of age (experience independent change) from co-varying cumulative increases in sensory 64 input (experience dependent change) on behaviour and brain development is difficult (Fahrbach et al., 65 1998; Jones et al., 2013; Maleszka et al., 2009; Riveros and Gronenberg, 2010). To distinguish the effects 66 of pesticide exposure from variation caused by other interacting factors, we therefore: i) attempted to 67 standardise experience and sensory input across tested workers, ii) tested workers of controlled age, and 68 iii) compared between young and old age cohorts. Furthermore, by studying two main developmental 69 stages, such as brood (larval & pupal) and early adulthood, here we reveal which development phase is 70 more vulnerable to pesticide exposure, and whether developmental plasticity in bee brains (Farris et al.,

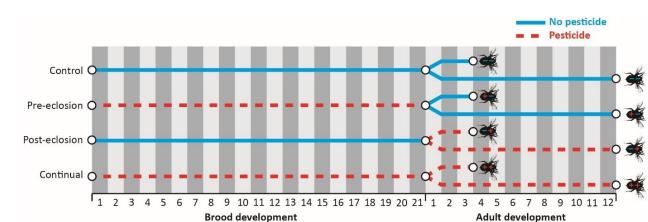
2001; Galizia et al., 2011; Riveros and Gronenberg, 2010; Withers et al., 1995) allows recovery during an
unexposed later stage.

73 Despite the functional importance of non-Apis bees (Brittain et al., 2013; Garibaldi et al., 2013; Gill et al., 74 2016) and possible differences in pesticide sensitivity relative to honeybees (Cresswell et al., 2012; Heard 75 et al., 2017; Piiroinen and Goulson, 2016; Rundlöf et al., 2015; Woodcock et al., 2016), empirical tests on 76 non-Apis bees looking at the physiological response to stress, such as postembryonic neuronal 77 development, are limited. Here we studied the response of the bumblebee, *Bombus terrestris*, a species: 78 i) that can be reared under controlled laboratory conditions; ii) for which learning performance assays 79 have been developed (Muth and Leonard, 2019; Riveros and Gronenberg, 2009; Siviter et al., 2018b; Smith 80 and Raine, 2014); iii) individuals can be exposed to pesticides within the social colony environment rather 81 than in isolation (Gill et al., 2012; Maleszka et al., 2009; Whitehorn et al., 2012). Here, we chronically 82 exposed cohorts of workers, reared inside their natal colonies, to a 5ppb concentration of the 83 neonicotinoid imidacloprid via provisioned sugar solution (40% sucrose) during two different 84 developmental stages: a) before and b) after adult eclosion from the pupal case. We investigated how the 85 link between brain growth and learning behaviour may be affected in workers exposed during: brood 86 development (pre-eclosion), early adulthood (post-eclosion) and both these developmental periods 87 (continual), comparing each to unexposed (control) workers. For each treatment we tested adult workers 88 at 3 or 12-days after eclosion (Figure 1).

Firstly, we tested worker response to sucrose and then on olfactory associative learning performance using the established proboscis extension reflex (PER) conditioning paradigm (Figure S1) (Bitterman et al., 1983; Giurfa and Sandoz, 2012; Laloi et al., 1999; Riveros and Gronenberg, 2009; Sommerlandt et al., 2014), which has previously been used to test pesticide effects on adult learning in honey bees exposed during the larval stage (Tan et al., 2015; Yang et al., 2012), and bumblebees exposed as adults (Piiroinen

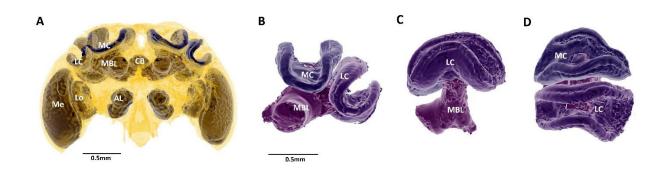
94 and Goulson, 2016; Stanley et al., 2015; Tison et al., 2017). Secondly, we employed new advances in micro-95 computed tomography ( $\mu$ CT) scanning and 3D image analysis to explore the brain *in situ* (within headcase) 96 and enable non-destructive volumetric measurements to a standardised voxel size of  $4\mu m$  (Figure 2) 97 (Baird and Taylor, 2017; Gutiérrez et al., 2018; Ribi et al., 2008; Smith et al., 2016). We segmented five 98 key neuropils: mushroom bodies (associated with learning), antennal lobes (olfaction), optic lobes -99 medullas and lobulas (vision), and central body (motor function; Table S1). For the mushroom bodies we segmented the two major components, lobes and calyces, to investigate responses of the different 100 101 functionally multisensory input and output regions, respectively (Fahrbach, 2006; Heisenberg, 2003). 102 Using a sample size exceeding any other study investigating bee brain morphology to date, we present 103 the first comparative study to directly link how chronic pesticide exposure impacts learning performance 104 by affecting bee brain development.





106Brood developmentAdult development107Figure 1. Graphic showing exposure periods for the four colony treatments (control, pre-eclosion, post108eclosion & continual) and the eight cohorts of individual workers to be tested. Blue solid line represents109untreated food (sucrose solution) and red dashed line represents pesticide-treated food. 'Brood110development' represents the larval and pupal (brood) stages of workers, with 'Adult development'111representing the number of days after eclosion from the pupal case. Individual bee symbols depict112removal of controlled aged adult workers at 3 or 12-days after eclosion for behavioural testing followed113by decapitation for μCT scanning of the brain.

114



115

Figure 2. 3D rendering of one of the studied bumblebee brains using the μCT imaging method. a, Focal
neuropils considered in this study shown in dark purple (optic lobes: medulla (Me), lobula (Lo); antennal
lobes (AL); central body (CB); mushroom body: calyces (including lateral calyx (LC) & medial calyx (MC)
and lobes (MBL)), surrounded by remaining brain tissue in transparent yellow. b-d, Isolated 3D structure
of the mushroom body which has been rotated to show b, frontal, c, lateral, and d, dorsal views.

121

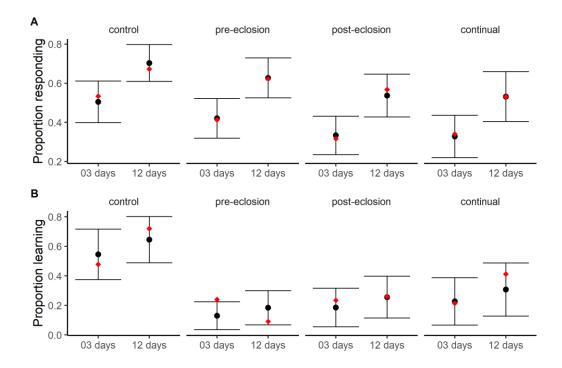
#### 122 Results

#### 123 Responsiveness

Prior to the learning assay, we confirmed whether harnessed workers (n=413; Table S2) showed a PER in 124 response to their antenna being touched by a 50% sucrose solution droplet (Figure S1). We found a 125 significantly higher proportion of 12-day compared to 3-day workers responded (GLM: age: z=-4.10, 126 127 p<0.001) which was consistent across treatments as evidenced by no age\*treatment effect and the interaction term not being retained in the model (Table S3). We also found consistent negative model 128 129 estimates for all three pesticide treatments relative to control, and detected a significantly lower 130 proportion of responsive workers from *post-eclosion* and *continual* exposed colonies (z=-2.53, p=0.011 & z=-2.40, p=0.016; Figure 3). 131

133 For the responsive workers (n=181; Table S2), we tested each worker's ability to learn to associate an 134 odour with a sucrose reward by demonstrating a PER response over ten consecutive trials (Figure S1; see 135 methods). Firstly, we categorised workers as either those exhibiting at least one response as 'learners' 136 and those showing no learnt response as non-learners. Whilst our model showed a positive estimate for 137 the effect of age, unlike responsiveness we did not detect a significant increase which was consistent 138 across treatments as evidenced by no age\*treatment effect and the interaction term not being retained 139 in the model (Table S3). However, we again detected a strong effect of pesticide exposure relative to the 140 control, with each treatment showing a significantly lower proportion of learners (GLM: pre-eclosion: z=-4.38, p<0.001; post-eclosion: z=-3.49, p<0.001; continual: z=-2.78, p<0.01; Figure 3b). 141

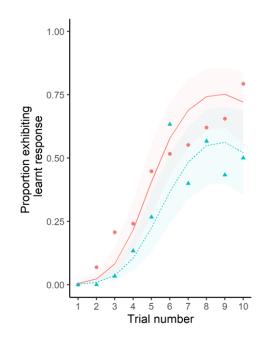
142 For all individuals classed as learners, we then looked at how the proportion of learnt responses changed 143 over the successive trials (analysis considered trials 2-10, as by definition a naïve worker cannot learn on 144 the first trial). However, because of the strong negative effect of pesticide exposure on passing the 145 responsiveness stage and proportion of learners, sample sizes for each pesticide treatment were 146 significantly reduced. Therefore, given the similarity in responses across pesticide treatments, we pooled 147 all workers from these three treatments and compared them to control whilst not distinguishing between 3-and 12-day workers. From this analysis we found that the proportion of responses increased over the 148 149 trials (GLM polynomial:  $p^1$ : t=14.26, p<0.001). This relationship, however, was non-linear and the incremental proportion decreased in rate over the consecutive trials ( $p^2$ : t=-2.48, p=0.014; Table S3). This 150 151 was primarily driven by the significant negative effect of pesticide exposure (t=-2.04, p=0.046), with 152 workers from exposed colonies showing a distinctly lower proportion of learnt responses in the latter few 153 trials relative to control (Figure 4).



155

Figure 3. Proportion of responsive workers and proportion demonstrating an olfactory associative learning response using proboscis extension reflex (PER) conditioning between treatments. A, Workers exhibiting a PER response when touching the antennae with a sucrose solution droplet prior to the PER conditioning trials; **B**, Learners (workers exhibiting at least one learnt response during the PER conditioning trials). Intersecting circular point represents the estimated model mean taken from backtransformation of the model (binomial GLM) with bars depicting the associated ±95% confidence limits. Red diamond corresponds to the mean value taken from the raw response data.

163



#### 165

Figure 4. Proportion of workers by trial exhibiting an olfactory conditioned learnt response. Workers from all three pesticide treatments were pooled (blue triangles; n=30 workers) and compared against *control* workers (red circles; n=29), with both age cohorts aggregated per treatment. Lines (blue dashed = pesticide treatment; red solid = control) represent the binomial model (LMER polynomial) estimates over the consecutive trials.

171

#### 172 Brain neuropil volumes

173 Focusing first on the mushroom body calyces, relative volumes were significantly smaller in workers from all three pesticide exposure treatments compared to control (pre-eclosion: t=-2.41, p=0.049; post-174 175 eclosion: t=-3.83, p<0.01; continual: t=-2.90, p=0.021; Table S4-5). This was consistent for both 3 and 12-176 day workers as evidenced by no effect of *age\*treatment* and the interaction term not being retained in 177 the model (Table S5). Focusing second on the relative volume of the mushroom body lobes, we again found negative model estimates for all three pesticide treatments relative to the control, however unlike 178 179 the calyces none of these comparisons were detected as significantly lower (Table S5). Analysis of the four 180 other segmented neuropils (central body, antennal lobes, lobulas and medullas) further showed that

- 181 workers from pesticide treated colonies showed no significant volumetric differences relative to *control*,
- although we did find consistent negative model estimates for the antennal lobes across all pesticide
- 183 treatments (Table S6).

184

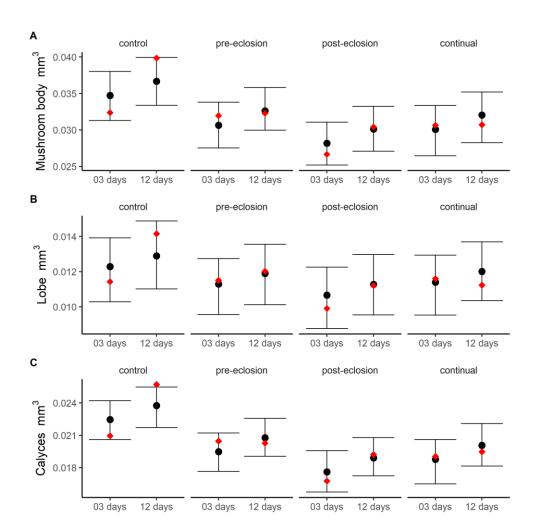


Figure 5. Volumes for A) whole mushroom body, B) mushroom body calyces and C) mushroom body lobes of bumblebee workers. These represent volumes relative to the body size of the worker (absolute volume divided by body size). Intersecting circular point represents the estimated model mean taken from back-transformation of the model (LMER) with bars depicting the associated ±95% confidence limits. Red diamond corresponds to the mean value taken from the raw response data.

#### 192 *Relationship between mushroom body calyces volume and learning score*

193 For each responsive worker that started the PER conditioned learning assay and for which we had the 194 volume of their mushroom body calyces, we took the total number of demonstrated learnt responses 195 ('learning score') and investigated the association with relative calyces' volume as the predictor variable 196 (given this was the neuropil most affected from exposure). As we did for the learning-by-trial data, we 197 pooled all workers from the three pesticide exposure treatments and compared their scores to that of 198 control workers when no distinguishing age. From this, we found a significant positive association 199 between relative volume of the calyces and learning score (t=4.51, p<0.001; Figure 6; Figure S2), but this 200 relationship was driven by *control* workers in which larger calvces equated to higher learning score. 201 Pesticide exposed workers, in contrast, showed no clear relationship as supported by the significant 202 negative volume\*treatment interaction (t=-3.96, p<0.001; Table S7-8). This finding reveals that impaired 203 functioning of the mushroom body in workers from pesticide exposed colonies is not only from reduced 204 volumetric growth, but presumably also from affected physiological composition of the tissue.

205

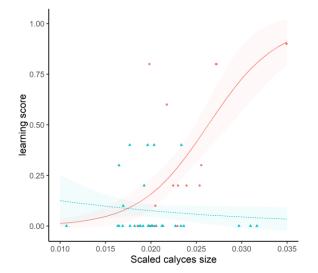


Figure 6. Relative volume of mushroom body calyces plotted against learning score. Calyces volume predicts learning score for *control* workers but not pesticide exposed workers. Workers from all three

pesticide treatments were pooled (blue triangles) and compared against *control* workers (red circles), with
 fitted lines (blue dashed = pesticide treatment; red solid = control) representing binomial model (GLM)
 estimates.

212

## 213 Discussion

Our study reveals that worker bumblebees exposed to a neurotoxic pesticide, a neonicotinoid, can affect the developmental plasticity of the brain with reduced volumetric growth manifested not only from exposure as an adult but also during brood development. The effects on adult behaviour and brain physiology from brood exposure (*pre-eclosion*) appeared irrecoverable despite no experimental provision of pesticide treated food during the 12-days of adulthood. Critically, impeded growth of the mushroom body calyces of worker brains from pesticide exposure, was associated with functional impairment as evidenced by reduced responsiveness and poorer olfactory learning behaviour.

#### 221 Pesticide exposure during early development affected responsiveness and learning

222 Neonicotinoid exposure as an adult (post-eclosion & continual) reduced the proportion of workers 223 responding to a sucrose droplet prior to the PER assay, and reduced olfactory learning performance during 224 the PER assay. These findings contribute to previous studies reporting adult neonicotinoid exposure 225 negatively affecting aspects of responsiveness in honeybees (Aliouane et al., 2009; Démares et al., 2018, 226 2016) and learning in bumblebees (Stanley et al., 2015). However, a key novelty of our study is that we 227 could compare responses from chronic exposure between age cohorts. Firstly, this revealed that young 228 (3-day) and older (12-day) workers from *post-eclosion* and *continual* exposure colonies were similarly 229 affected despite differences in the number of days of adult exposure. Additionally, despite pre-eclosion 3-230 day adults being exposed for up to 3-weeks during brood development, compared to only three days of exposure for 3-day adults from *post-eclosion* colonies, the degree of impaired learning was again similar. 231

Together these findings highlight the first 72 hours of adulthood to be critical in behavioural development,
and reveals a susceptible developmental window to environmental stress (in this case pesticide exposure)
(Sandrock et al., 2014; Wu et al., 2011), reiterating the importance of considering different life-stages
when assessing pesticide risks.

236 Secondly, workers exposed during brood development (pre-eclosion), that received no or substantially 237 lower exposure as an adult, exhibited impaired learning performance at a similar reduced level as adult 238 only exposed workers (*post-eclosion*). This indicates a lag-effect from brood exposure on adult learning, 239 highlighting the importance of considering delayed effects of pesticide exposure; a view reinforced by 240 other studies on honeybees (Apis cerang & A. melliferg) and a stingless bee (Melipong augdrifasciata 241 anthidioides) reporting larvae reared under topical or oral neonicotinoid exposure exhibited negative 242 effects on adult learning and motor function (Tan et al., 2015; Tomé et al., 2012; Yang et al., 2012). More 243 importantly, with 3- and 12-day adults from pre-eclosion colonies exhibiting a similar level of impaired 244 learning performance, this reveals that the effects from brood exposure appear irrecoverable even as an 245 unexposed adult.

246 Impaired learning in pesticide exposed workers was associated with reduced volumetric growth of the 247 mushroom body calyces

Focusing on the mushroom body calyces, 12-day adult workers from all pesticide exposure treatments possessed smaller mushroom body calyces relative to *control* colonies, and we even found differences in 3-day adult workers from *post-eclosion* and *continual* exposed colonies. With rates of mushroom body development in bumblebees considered to be at its highest during the first 72 hours of adulthood (Jones et al., 2013; Riveros and Gronenberg, 2010), this may explain why in our experiment an effect on calyces volume was detected in just 3-days of adulthood. Furthermore, our finding that 3-day adults from both 254 *pre-eclosion* and *post-eclosion* exposed colonies showed similar reductions in mushroom body volumes,

reiterates the apparent vulnerability of brain development during the first 72 hours.

256 Average volume reductions of the mushroom body calyces and lobes showed a strikingly mirrored pattern 257 to the reduced proportion of learners in each respective pesticide treatment. More tellingly, when relative 258 volume of the calyces was plotted against each respective worker's learning score, bigger relative brain 259 size equated to better learning performance in *control* workers, but this relationship was not found for 260 pesticde exposed workers. Despite some control and pesticide exposure workers possessing similar 261 relative mushroom body volumes, pesticide exposure workers demonstrated a lower learning score 262 indicating impaired neuronal functioning of this brain region. For workers from *post-eclosion* and 263 continual exposed colonies this effect could be explained by sublethal neonicotinoid concentration 264 affecting neuronal signalling given its role as a nAChE receptor agonist (Palmer et al., 2013). However, 265 with pre-eclosion workers also being affected, this indicates that neonicotinoid exposure might be 266 affecting synaptic development (i.e. proliferation & dendritic outgrowth) in the calyces. Indeed, reduced 267 microglomeruli density has been shown to occur in neonicotinoid exposed honeybees (Peng and Yang, 268 2016), and density of these structures is correlated with increased learning and memory in bees (Hourcade 269 et al., 2010; Li et al., 2017). Furthermore, the reduced learning performance in *pre-eclosion* workers in our study could stem from impeded neurogenesis where neuronal precursor cells are in some way prevented 270 271 from giving rise to Kenyon cells in the mushroom bodies, which in honeybees occurs during development 272 before eclosion from the pupal case (Fahrbach et al., 1995; Farris et al., 1999). Alternatively, the size of 273 Kenyon cells could be affected, as has been shown from exposure experiments on bumblebee cell cultures 274 (Wilson et al., 2013).

275 Mushroom body calyces were disproportionately affected over the other neuropils

276 The effect of neonicotinoid exposure was primarily localized to the mushroom body, and even within this 277 structure was manifested more heavily in the calyces than the lobes. Localised variation in plasticity has been shown in bumblebees where foraging experience increased medial but not lateral calyx volume 278 279 (Riveros and Gronenberg, 2010). The calyces act as multisensory processors fed by afferent neurons, 280 whereas the lobes predominantly function as output regions with efferent neurons, which could explain 281 why calyx volumetric variation is more tightly associated with our measure of learning performance. 282 Antennal lobes are involved in detecting and processing olfactory information (Galizia et al., 1999; 283 Hansson and Anton, 2002; Sachse et al., 1999), developmentally plastic during early adulthood (Jones et 284 al., 2013; Riveros and Gronenberg, 2010), and exhibit reduced neuronal function under nicotinic agonists 285 (Andrione et al., 2016; Barbara et al., 2008; Thany and Gauthier, 2005). Considering pesticide exposed 286 workers performed worse in the olfactory conditioning, we might therefore expect a pattern of impeded 287 growth in the antennal lobes similar to that found for the mushroom body calyces. Whilst in support of 288 this view we found consistent negative model estimates for all three pesticide treatments, unlike the 289 calyces our analysis did not detect a significant effect. Furthermore, we found no consistent reduction in 290 volume of the optic lobes (medullas & lobulas) or central body for pesticide exposed workers.

291 Possible explanations for the disproportionate effect of neonicotinoid exposure on the mushroom bodies 292 may include: i) nACh receptors targeted by neonicotinoids are found in the highest density in the 293 Kenyon cells of the mushroom bodies (Galizia et al., 2011) and so could affect Kenyon cell 294 proliferation leading to volumetric reductions; ii) The mushroom bodies, in particular the calyces, of 295 social insects have consistently been reported to be highly plastic structures due to their role in learning 296 and memory development as early adults (Cabirol et al., 2018; Farris et al., 2001; Heisenberg, 2003; Jones 297 et al., 2013; Kühn-Bühlmann and Wehner, 2006; Riveros and Gronenberg, 2010). The large amount of 298 neuronal development and re-organisation therefore increases the risk of neurotoxic exposure interfering 299 with this process; iii) Our experimental setup was stimulus deprived and not void, therefore whilst

mushroom body volumetric increase is likely to be more experience independent than dependent, we could not rule out investment in olfactory processing to compensate for a lack of visual information (Fahrbach et al., 1998; Jones et al., 2013); iv) the change in growth of non-mushroom body neuropils was simply too subtle for our  $\mu$ CT technology and/or sample sizes to detect.

304 Improved behavioural performance and mushroom body growth with age independent of experience

Our study reared workers under a stimulus deprived environment, therefore a positive effect of age is 305 306 indicative of experience independent age-enhanced learning. For both our measures of responsiveness 307 and learning we found positive estimates for age, with 12-day workers performing better on average than 308 3-day. This finding contrasts with previous bumblebee studies reporting no effect of age on aspects of 309 learning ability (Riveros and Gronenberg, 2009; Smith and Raine, 2014), but these were carried out in 310 foraging arenas whereby prior experiences could not be fully controlled. Indeed, to our knowledge, there is a lack of studies looking to identify innate age-related growth on bumblebee behaviour as 311 312 well as on brain growth. Only one histological study on the bumblebee *Bombus impatiens* by Jones 313 and colleagues has shown age-dependent volumetric growth in brain neuropils separate form 314 environmental stimuli (Jones et al., 2013). Their findings suggested c.10% increase in the mushroom body 315 calyces and lobes which interestingly is around half the increase we found in our *control* workers (just 316 over 20%), a difference that perhaps stems from variation in methodological approaches, sample sizes 317 (lower in Jones et al.) or taxonomic variation. However, together this evidence supports an innate increase in neuropil volume over the first 12-days of adulthood, which presumably is important to prepare workers 318 319 for the complex colony tasks required at this age (Maleszka et al., 2009).

320 Implications of our findings for social insects

321 Our findings that early exposure effects later adult behaviour provides a mechanistic explanation for why 322 reduced colony growth is often detected 2-3 weeks after onset of neonicotinoid exposure (Arce et al., 323 2017; Gill et al., 2012; Rundlöf et al., 2015; Siviter et al., 2018a; Tsvetkov et al., 2017; Whitehorn et al., 324 2012), and why reduced colony productivity has been correlated with neonicotinoid treated neighbouring 325 fields (Rundlöf et al., 2015; Woodcock et al., 2016). With eusocial bee colonies having overlapping 326 generations, colonies are reliant on newly emerging cohorts of workers to be effective task performers. If future generations of workers are predisposed to be inefficient functioning cohorts, this could lead to a 327 328 density dependent build-up of colony level impairment increasing the risk of colony collapse (Bryden et 329 al., 2013). Our results suggest that even if workers were to delay undertaking a task, such as foraging, in 330 attempt to developmentally recover, this strategy may be futile given we saw little adult recovery from 3 331 to 12 days of adulthood from *pre-eclosion* colonies. Importantly, these effects are unlikely to be 332 exclusively applicable to neonicotinoids as a multitude of other neurotoxic pesticides including the 333 possible neonicotinoid replacements, sulfoxamines and butenolides (Siviter et al., 2018a; Tosi and Nieh, 334 2019), are likely to end-up inside bee colonies with the potential to influence tissue development in reared 335 bees.

336

#### 337 Materials & methods

338 Animal Husbandry

Twenty-two *Bombus terrestris audax* colonies were delivered by a commercial supplier (Agralan Ltd), with colonies possessing a queen and mean (±s.e.m.) of 14.5 ± 1.1 workers on arrival (Table S9) and housed in an aerated plastic box (29 x 22.5 x 13 cm). All colonies were moved to a controlled environment (23°C; 60% humidity) red light room where they remained for the duration of the experiment. Throughout the experiment, colonies were provisioned untreated honeybee collected pollen (supplied by a commercial supplier Agralan Ltd) ad-libitum in a petri dish, and 40/60% sucrose/water solution in a gravity feeder.
Food was replenished every two days, and feeders thoroughly cleaned prior to refill (Table S10 for colony
consumption). During Phase I (days 1-21; Figure 1; Figure S3), we conducted daily checks of all newly
eclosed bees and marked each using a white paint pen (uni Posca, PC.5M 1.8-2.5mm), allowing us to
distinguish between newly eclosed workers during Phase II (day 22 onwards) from eclosed workers before
this. Colonies were checked daily for males, gynes or dead individuals which were removed and frozen at
-20°C.

351

#### 352 Experimental setup

353 On arrival, colonies were randomly assigned to the four treatments, with no significant difference in the 354 number of workers between treatments (ANOVA: F<sub>3,22</sub>=1.04, p=0.40). Mean worker thorax width was 355 similar between treatments (LMM, p>0.07) with control being 4.23mm (range=3.29-5.17), pre-eclosion 356 4.16mm (3.09-5.12), post-eclosion 4.28mm (3.14-5.63) and continual 4.33mm (3.36-5.34). Monitoring 357 overall development of workers in colonies, we implemented a fully factorial design with our colony treatments comprising a combination of two exposure phases: Phase I encompassing the majority of 358 359 brood (larval & pupal) development period and Phase II comprising the early adult development period 360 (up to 12 days). Phase I exposure period started two days after colonies arrived and lasted for 21 days 361 approximating development time from an egg or very small larva to adult eclosion (Alford, 1975; Cnaani 362 et al., 2002; Duchateau and Velthuis, 1988). This ensured that all sampled adults will have been 363 exposed/unexposed in a standardised manner during the vast majority of brood development (Figure S3). 364 On the 22<sup>nd</sup> day Phase II started, during which we checked daily for callow workers (adults recently eclosed 365 from their pupal case) and tagged each with a unique numbered Opalith tag using superglue. On tagging, 366 we randomly assigned half of the workers per colony per day to the 3-day cohort and remaining half to 367 the 12-day cohort, with tag ID used to correctly remove for testing 3 or 12 days later. Tagging period lasted

11 days to provide us with a high number of workers to test (Table S2). This window of opportunity approximates the minimum time of pupal development, in which pupae evacuate their gut and stop feeding (Cnaani et al., 2002), allowing us to standardised pesticide exposure as best possible across all tested workers. Adult workers aged 3 and 12-days after eclosion were chosen to be tested as brain development has been reported to occur both during brood and early adult stages (Farris et al., 2001; Jones et al., 2013).

374 We applied four treatments to colonies: *control* = phases I & II unexposed (n=5 colonies); *pre-eclosion* = 375 phase I exposed to Imidacloprid, phase II unexposed (n = 6); post-eclosion = phase I unexposed, phase II 376 exposed to Imidacloprid (n = 6); continual = phases I & II exposed to Imidacloprid (n = 5). The neonicotinoid 377 imidacloprid was used as: i) it is widely used across the globe (Casida, 2018; Cressey, 2017; Mitchell et al., 378 2017; Zhang, 2018); ii) it targets nAChE receptors found in insect brains (Jeschke and Nauen, 2008; Palmer 379 et al., 2013); iii) exposure has been shown to affect bee foraging and navigation known to be reliant on 380 learning ability and working memory (Feltham et al., 2014; Fischer et al., 2014; Gill and Raine, 2014; 381 Samuelson et al., 2016; Stanley and Raine, 2016). The imidacloprid treated sucrose solution provisioned 382 to the colony was made from a primary stock solution (1mg/ml) consisting of 100mg of imidacloprid 383 (powder; grade: PESTANAL<sup>®</sup>, analytical standard; brand: Fluka) dissolved in 100ml of acetone. An aliquot 384 was then added to a 40/60% sucrose/water solution to produce a 5ppb imidacloprid solution of required 385 volume. A *control* sucrose solution was made by repeating this process but a same aliquot volume of pure 386 acetone.

387

388 Assessing olfactory learning performance using proboscis extension reflex (PER) conditioning

The proboscis extension reflex (PER) conditioning paradigm we implemented was adapted from a previously reported setup on bumblebees (Stanley et al., 2015). On removal from the colony, workers were harnessed (between 13:00-14:00) using a modified 2ml centrifuge tube and a split pin yoke, under natural light in the lab and left for 2hrs to settle (Figure S1 for harness setup). All bees were then fed to
satiety using a Gilmont<sup>®</sup> syringe to present 40% sucrose solution droplets directly to the mouthparts and
left for 18 hours (overnight) in a separate controlled environment room under identical conditions as the
rearing room. For unknown reasons, 24 workers did not survive overnight and were excluded from any
data analysis. Between 08:00-09:00 the PER testing began on the remaining bees (n=413) by first testing
their PER responsiveness to a 50% sucrose solution. Immediately after, each bee was fed a small droplet
(0.8µl) of the sucrose solution for motivation 15 minutes before the start of the PER test (Figure S1).

399 PER conditioning was conducted in front of a filtered ventilation system (Expo Drills & Tools AB500 400 Extractor fan), preventing the odour coming in to contact with neighbouring harnessed bees. Each bee 401 was initially conditioned by exposure to clean air for 5 seconds, followed by scented air for 10 seconds. A 402 harnessed bee was positioned 3 cm away from a glass odour tube, with the airflow delivered at a constant 403 rate of 80 ml/second (Tetra APS – 100), which was channelled through either a 'clean' unscented odour 404 tube or diverted through a 'scented' odour tube containing a piece of filter paper (5 x 20 mm) impregnated 405 with 1 µl of lemon essential oil (Naturally Thinking Ltd.). Airflow between the clean and scented tube was 406 controlled by a solenoid valve (Nass Magnet 108-030-0257 24vAC/12vDC) connected to a Raspberry Pi 2 407 (Model B) computer to ensure each bee was exposed to a consistent amount of clean and scented air. To 408 develop an association between the lemon odour and the reward, we touched the bee's antennae with a 409 droplet of 0.8 µl of 50% sucrose 6 seconds into the 10 second odour delivery phase and allowed the bee 410 to feed.

Following trial 1, the odour and reward presentation sequence was repeated to each adult an additional nine times. The inter-trial interval (ITI) per individual was 10 minutes allowing us to conduct the PER testing in batches of up to 20 workers. We waited 15 seconds after the odour and reward presentation sequence before moving to the next individual (Smith and Burden, 2014). We recorded whether the bees showed a PER to the odour stimulus prior to or after the reward, which were defined as a learnt or non-

416	learnt response respectively. This provided a number of learnt responses achieved by each worker over
417	the nine trials, enabling us to estimate the probability per trial of workers demonstrating a learnt response
418	for each treatment. If a bee responded to the initial conditioning trial (trial 1) before the reward had been
419	presented (n=24) the individual was excluded from the experimental analyses. If a bee showed no PER
420	(did not feed) even after the reward was provided, and exhibited this over the next three consecutive
421	trials, the individual was removed from testing from that point and categorised as a non-learner.

422

#### 423 Micro-CT scanning

Linking variation in learning to differences in brain growth requires high-resolution imaging technology that can explore minute changes to soft tissue. Using traditional histological methods would have been technically challenging as it relies on physical extraction from the headcase followed by tissue fixation, dehydration, embedding and sectioning (Simmons and Swanson, 2009) increasing the risk of destructive sampling. Potentially this could have caused a greater change to neuropil volume than the experimental treatment itself. We attempted to overcome such challenges by using micro-computed tomography (μ-CT) scanning.

431 Following the PER assay, bees were humanely sacrificed by swiftly decapitating the live individual using a 432 disposable surgery scalpel and heads immediately fully submerged in a 70/30% ethanol/de-ionised water 433 solution in separate 1.5ml centrifuge tubes and stored at 5°C. Preparation of the heads followed precisely 434 the published protocol by Smith et al. (2016) with the soft brain tissue being stained for seven days with 435 phosphotungstic acid (PTA) before being CT scanned at a voxel size of  $3.5 - 4 \,\mu\text{m}$  using a Nikon Metrology 436 HMX ST 225 system (Nikon Metrology, Tring, UK). The staining and scanning methodology we employed 437 has been shown to give us confidence in the accuracy of our measurements of these complex neuropil 438 structures (Smith et al., 2016). The raw µCT data for each brain scan was reconstructed using CTPro 2.1 439 software (Nikon Metrology, Tring, UK) and processed using VG Studio Max 2.1 (Volume Graphics GmbH,

440 Heidelberg, Germany). Each 3D reconstructed scan was then re-oriented to the same optimum plane-of-441 view for visualization, and for the neuropils in question we re-sliced into a new series of 2D images. For 442 each sample, scan images were exported as 8-bit BMP image series at a standardized voxel size of 4 μm. 443 In total, 92 worker brains were  $\mu$ CT scanned, but based on staining quality and that both left and right 444 structures could be segmented (including both medial and lateral calyx for the mushroom body calyces) 445 we successfully segmented the mushroom bodies for 78 workers, central body for 88, antennal lobes for 89, medullas for 71 and lobulas for 71 (Table S4). For the purposes of comparing relative volumes 446 447 (absolute volume divided by ITD to correct for body size), bees were originally sampled to have a balanced 448 representation across treatments and age but blind of learning performance.

449

#### 450 Neuropil volume measurements

451 Segmentation and volume analysis of brain structures was carried out using the software SPIERS 2.20 452 (Serial Paleontological Image Editing and Rendering System). For segmentation, scan slices were 453 converted to binary threshold images (of white active pixels and black inactive pixels) adjusted to achieve 454 an optimum ratio of active white pixels that comprise the structure of interest, and inactive black pixels 455 for the surrounding tissue. For each component structure, looped splines were placed around the active 456 pixels at regular five slice intervals which were then interpolated across all slices between the intervals, 457 so that each structure could be defined as an independent object for 3D reconstruction and volumetric 458 calculation (for full segmentation protocol see Smith et al. 2016). This soft tissue segmentation protocol 459 has been shown to provide repeatable and precise volumetric measurements of morphological structures 460 of the bumblebee brain. To calculate absolute volume of each structure we used the voxel count function 461 in SPIERS Edit, with relative volume calculated by dividing by the inter-tegula width (standard proxy for 462 body size (Cane, 1987)). Inter-tegula width was measured using digital callipers (Workzone®) with the 463 mean of two repeated measurements used. A single value was used in our analyses for each of the

464 mushroom bodies, antennal lobes, lobulas and medullas by summing the volume of the left and right 465 paired structures.

466

#### 467 Data Analysis

468 Statistical analyses were conducted in R version 3.5.1 (R Development Core Team 2018) using RStudio 469 version 1.1.463, with models implemented using the Ime4 package (Bates et al., 2015). For all models we 470 included treatment as a fixed categorical factor. We considered measure of bee body size (ITD) as a 471 continuous variable and colony as a random factor in our models if inclusion increased the fit of the model (model comparisons were assessed by comparing the AIC) otherwise they were not retained. For 472 473 responsiveness and learning the data was analysed using the proportion of individuals showing a response 474 with a generalized linear model (glm) using a binomial distribution and included the categorical variable 475 age (3 or 12-day) as an additional fixed factor with age x treatment interaction term removed as it showed 476 no significant effect. For looking at the proportion of learners by trial we used a linear mixed effects model 477 (Imer) in which treatment consisted of two categories, control workers and pesticide workers (pooled 478 from all three pesticide treatments due to low sample sizes per treatment). We considered a second order 479 polynomial fit for trial number and individual ID as a random factor. For relative neuropil volumes we used 480 a linear mixed-effects model (LMER) that included age, ITD and colony (random effect). We used a 481 binomial generalised linear model (glm) to analyse how calyces' volume influenced the learning score, as 482 a proportion of the maximum learning score that could be achieved using calvces volume to analyse the 483 calyces volume x treatment interaction on score.

484

## 485 Author Contributions

486 RJG conceived the project; DBS & RJG designed the experiment; DBS, ARR & PHB conducted the 487 experiment; DBS & FA carried out the  $\mu$ CT scanning; DBS & DB reconstructed and segmented the brains; 488 ANA, DBS & RJG performed data analyses; DBS & RJG wrote the manuscript and ANA provided critical 489 feedback.

## 490 Acknowledgements

- 491 We thank Russell Garwood for help using SPIERS software, Dan Sykes and Amin Garbout for help with the
- 492 µCT scanning protocol, Paul Beasley for technical support, Alfredo Sánchez Tójar and Peter Graystock for
- 493 advice on da ta analysis, and Richard Abel, Mark Brown, Inti-Pedroso, Nigel Raine and Seirian Sumner for
- 494 advice on pilot work. This work was supported by NERC grants (NE/L00755X/1 & NE/P012574/1) awarded
- to RJG which funded ANA and ARR. DBS's PhD was supported by a NERC funded SSCP DTP scholarship in
- affiliation with the Grantham Institute at Imperial College London. RJG is supported by Imperial College's
- 497 Grand Challenges in Ecosystems and the Environment initiative.

498

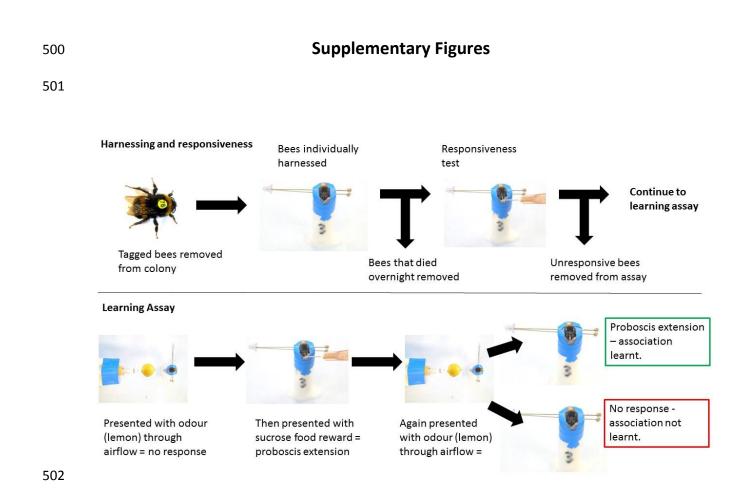


Figure S1. Proboscis extension reflex assay setup and step-by-step guide. Individuals were placed inside a plastic test tube and the tube was placed on ice for 10 minutes. Individuals were then harnessed in modified 2ml centrifuge tubes and a split pin yoke held them in place with electrical tape (blue). Harnessed bees were always placed the same distance from the air flow odour source and an extractor fan was mounted behind to remove excess odour.

508

510

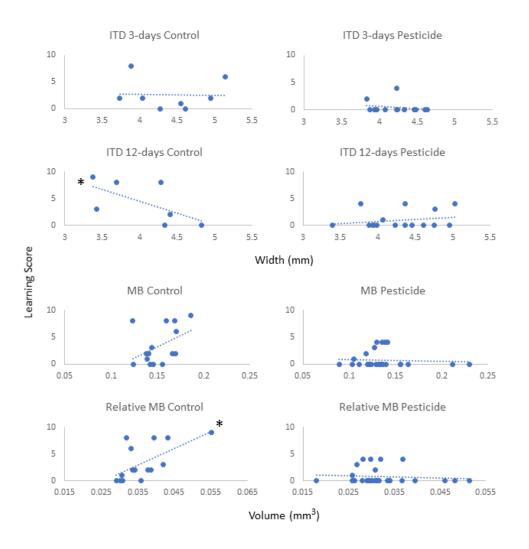
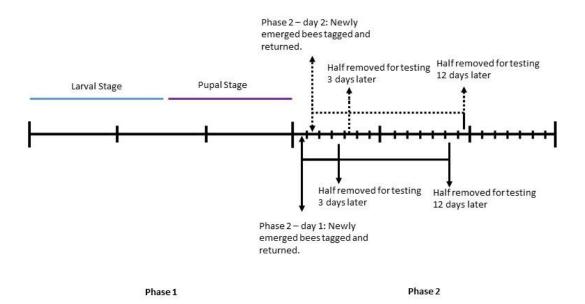


Figure S2. Predictor variables individual body size and mushroom body (MB) volumes were plotted against respective learning scores. The effect of width (mm) of the intertegula distance (ITD; proxy for body size) for 3-day and 12-day adult workers are shown. The absolute and relative (corrected for body size) whole mushroom body volumes (mm<sup>3</sup>), representing the combined volumes of the left and right hemispheres, for all workers regardless of age are shown. Filled circles represent the raw data with a dashed trend line fitted. Asterisks denote a significant negative or positive relationship ( $\alpha$ =0.05) based on a Spearman's Rank correlation analysis (detailed in Table S8).



520

521 Figure S3. Sampling for 3 and 12-day adult cohorts. During Phase I: all newly eclosed bees were marked 522 using a white paint pen, but this also represents the developmental time of workers with our sampled 523 bees having been larvae (blue) and pupae (purple) during this 21 day development period. Phase II: for 11 524 days, colonies were checked daily with any newly eclosed workers being tagged and returned to their 525 natal colony with half randomly assigned to a 3-day cohort and the other half a 12-days cohort. The respective cohorts were then removed 3 days or 12 days later. This continual tagging and sampling over 526 the 11 day period provided us with a large number of workers to test (here we provide an example for 527 528 the first two days for demonstration purposes).

529

# **Supplementary Tables**

**Table S1** Informed from honeybee studies, our study focused on five key neuropils considered to be 535 involved in the following primary functional roles.

Structure	Function	References
Mushroom Bodies	Associated with higher cognition and learning with the processing of multimodal sensory information. Possess distinct sub-compartments each with functional specialisation.	(Durst et al., 1994; Fahrbach, 2006; Farris et al., 2001; Hammer and Menzel, 1995; Heisenberg, 2003, 1998; Mobbs, 2006; Zars, 2000)
Antennal Lobes	Principle olfactory centre associated with processing of chemical stimuli.	(Hansson and Anton, 2002)
Medullas & Lobulas (optic lobes)	Processing of visual information	(Dyer et al., 2011; Lotto and Chittka, 2005; A. C. Paulk et al., 2009; Angelique C. Paulk et al., 2009; Paulk et al., 2008; Pfeiffer and Homberg, 2013)
Central Body (a.k.a. fan shaped body)	Considered involved in locomotion and orientation	(Li et al., 2009; Pfeiffer and Homberg, 2013; Strauss, 2002)

## 

539	Table S2. Number of 3- and 12-day adult workers from each colony that were prepared for tested using
540	our PER setup. 'Tagged' = number of newly emerged workers that were tagged with a unique colour and
541	numbered Opalith tag; 'Survived to Harnessing' = number of tagged workers survived to 3- or 12-days
542	post-emergence and could be harnessed for the PER assay; 'Harnessed & survived overnight' = number of
543	workers that were harnessed but also were alive inside the harness the next day (n=24 died overnight);
544	'Responsive' = number of workers that exhibited a PER response on touching the antenna with a 50%
545	sucrose solution droplet; 'Considered for PER assay' = the number of workers remaining once any workers
546	showing a PER to the lemon odour and before the sucrose provision on the 1 <sup>st</sup> trial were removed;
547	'Learner' = number of workers that showed at least one olfactory conditioned PER response over the ten
548	trials.

_		~
5	4	9

Treatment	Colony	Colony	Colony	Tagg	ed	Survive harnes		Harness surviv overn	ved	Respor	nsive	Consider PER as		Learn	ier
		3	12	3	12	3	12	3	12	3	12	3	12		
control	1	16	29	15	29	13	11	9	8		5	5	4		
control	2	18	40	17	40	16	17	9	9		7	3	7		
control	3	13	22	13	21	12	10	6	6	4	5	2	1		
control	13	18	22	18	21	17	13	7	12	4	8	1	6		
control	14	0	2	0	2	0	1	0	0	0	0	0	0		
		65	115	63	113	58	52	31	35	23	25	11	18		
pre-eclosion	4	9	7	9	7	9	5	4	2	4	2	1	0		
pre-eclosion	5	21	26	21	25	20	19	10	11	10	11	4	1		
pre-eclosion	6	9	4	8	4	8	3	3	1	3	1	1	0		
pre-eclosion	15	17	24	16	21	16	15	5	11	4	11	0	1		
pre-eclosion	16	6	12	6	12	6	8	2	6	2	6	0	1		
pre-eclosion	17	4	4	4	4	4	3	2	2	2	2	0	0		
		66	77	64	73	63	53	26	33	25	33	6	3		
post-eclosion	7	4	2	4	2	4	2	1	0	1	0	0	0		
post-eclosion	8	14	28	14	27	11	12	5	7	5	7	2	1		
post-eclosion	9	11	11	11	11	11	7	5	4	4	4	2	1		
post-eclosion	18	19	28	19	26	18	16	2	8	2	7	0	1		
post-eclosion	19	11	17	11	15	9	9	3	8	3	7	0	4		
, post-eclosion	20	4	9	4	6	4	5	2	2	2	2	0	0		
		63	95	63	87	57	51	18	29	17	27	4	7		
continual	10	15	39	15	35	11	10	4	2	3	2	0	0		
continual	11	13	16	13	15	12	8	5	2	4	2	1	1		
continual	12	6	4	5	4	5	2	3	2		2	2	2		
continual	21	14	26	13	25	13	8	3	- 8		8	0	3		
continual	23	6	18	6	17	6	4	1	3		3	0	1		
		54	103	52	96	47	32	16	17	14	17	3	7		
	L	248	390	242	369	225	188	91	114	79	102	24	35		
	Totals		638		611		413		205		181		59		

553 Table S3. Statistical comparisons of responsiveness and learning. a-b) statistical outputs from binomial 554 Generalized Linear Model in R (GLM) when analysing the proportion of workers that showed a PER response to a sucrose droplet prior to undertaking the learning assay, and the proportion of workers 555 showing at least one olfactory conditioned PER learnt response during trials 2-10 of the learning assay. 556 Exposure treatments are comparisons to *control* workers ('intercept') with age (3 versus 12-day workers) 557 558 and ITD (inter-tegula distance which is a proxy for body size) considered. For both models the interaction 559 term between pesticide treatment and age was removed as no significant effect could be detected. c) 560 statistical output from a binomial Linear Mixed Effects Model in R (LMER) when analysing the proportion 561 of workers that showed a PER response over the PER assay trials. For this model all workers from the three 562 pesticide exposure treatments were pooled to compare one cohort against *control* workers (intercept). A  $2^{nd}$  order polynomial relationship was found to be the best fit to the data. Significant differences ( $\alpha$ =0.05) 563 564 are highlighted in bold.

a) Responsive								
GLM(responsive(y/n) ~ treatment + age + ITD, family = binomial								
	Estimate	Std. Error	z value	Pr(> z				
(Intercept)	-1.279	0.959	-1.333	0.18				
pre-eclosion	-0.342	0.276	-1.242	0.21				
post-eclosion	-0.714	0.282	-2.529	0.01				
continual	-0.737	0.307	-2.400	0.01				
age	0.842	0.205	4.101	<0.00				
ITD	0.306	0.218	1.406	0.16				
b) Learners								
GLM(learner(y/n) ~ treatm	ent + age + ITD	, family = bir	omial					
(Intercept)	-3.865	1.820	-2.123	0.03				
pre-eclosion	-2.086	0.476	-4.382	<0.00				
post-eclosion	-1.663	0.476	-3.490	<0.00				
continual	-1.409	0.507	-2.778	0.00				
age	0.413	0.360	1.147	0.25				
ITD	0.953	0.413	2.308	0.02				
c) Learners by trial								
LMER(learner(y/n) ~ treatment + poly(trial,2) + (1 worker_id), family = binomial								
(Intercept)	-0.889	0.308	-2.887	0.00				
pesticide_treatments	-0.861	0.413	-2.083	0.03				
poly(trial, 2)1	48.427	5.398	8.972	<0.00				
poly(trial, 2)2	-18.909	3.965	-4.769	<0.00				

567 Table S4. Average volumes (mm<sup>3</sup>) of segmented brain neuropils using  $\mu$ CT scanning for each 568 experimental treatment and age cohort. Mean (± s.e.m.) values are based on workers across all colonies 569 and represent relative volumes (absolute volume divided by workers size). The percentage difference (% 570 diff.) of the mean of each pesticide treatment relative to the control group is provided with negative 571 values showing smaller and positive values showing larger average volumes. Based on the quality of staining and scanning, sample sizes (n) for each neuropil were limited to those that had both structures 572 573 from the left and right brain hemispheres successfully segmented. (N.B. for the central body one individual 574 (18G) was removed from the analysis as it represented an extreme outlier likely caused by a segmentation 575 error).

treatment	age	n	mean	s.e.m.	% diff.				
		total	mushroom	bodies					
control	3	9	0.1413	0.0057					
pre-eclosion	3	11	0.1402	0.0094	-0.8				
post-eclosion	3	10	0.1166	0.0051	-17.5				
continual	3	11	0.1306	0.0105	-7.5				
control	12	8	0.1607	0.0050					
pre-eclosion	12	11	0.1307	0.0076	-18.7				
post-eclosion	12	10	0.1287	0.0079	-19.9				
continual	12	8	0.1346	0.0065	-16.2				
						n	mean	s.e.m.	% diff.
	_	mush	room body	calyces		mushr	oom body l	lobes	
control	3	9	0.0916	0.0045		10	0.0496	0.0013	
pre-eclosion	3	11	0.0897	0.0037	-2.0	11	0.0504	0.0039	1.6
post-eclosion	3	10	0.0733	0.0056	-20.0	13	0.0453	0.0022	-8.7
continual	3	11	0.0813	0.0040	-11.3	13	0.0485	0.0032	-2.3
control	12	8	0.1040	0.0030		10	0.0569	0.0017	
pre-eclosion	12	11	0.0819	0.0050	-21.2	11	0.0487	0.0039	-14.3
post-eclosion	12	10	0.0813	0.0069	-21.8	11	0.0480	0.0029	-15.7
continual	12	8	0.0852	0.0039	-18.0	9	0.0486	0.0026	-14.5
		centr	al body			anteni	nal lobes		
control	3	10	0.000597	0.000031		10	0.00522	0.00020	
pre-eclosion	3	12	0.000555	0.000040	-7.0	12	0.00509	0.00020	-2.4
post-eclosion	3	13	0.000582	0.000041	-2.6	13	0.00418	0.00013	-20.0
continual	3	14	0.000586	0.000029	-1.9	14	0.00479	0.00019	-8.3
control	12	10	0.000645	0.000046		10	0.00590	0.00035	
pre-eclosion	12	9	0.000599	0.000514	-10.1	10	0.00581	0.00070	-1.5
post-eclosion	12	11	0.000584	0.000067	-9.5	11	0.00516	0.00036	-12.6
continual	12	9	0.000607	0.000023	-5.9	9	0.00467	0.00018	-20.9
		lobul	as			medul	las		
control	3	7	0.00816	0.00035		7	0.0252	0.0009	
pre-eclosion	3	11	0.00866	0.00175	6.1	11	0.0261	0.0013	3.7
post-eclosion	3	10	0.00844	0.00083	3.4	10	0.0262	0.0011	4.0
continual	3	8	0.00933	0.00199	14.4	8	0.0293	0.0029	16.5
control	12	9	0.00938	0.00115		9	0.0280	0.0014	
pre-eclosion	12	10	0.00981	0.00197	4.6	10	0.0275	0.0020	-1.8
, post-eclosion	12	8	0.00970	0.00137	3.4	8	0.0305	0.0015	9.0
continual	12	8	0.00906	0.00078	-3.4	8	0.0268	0.0009	-4.5

576

578 Table S5. Statistical comparisons of mushroom body relative volumes. Statistical outputs from a 579 Generalized Mixed Effects Model in R (LMER) for the combined volumes of the left and right a. whole 580 mushroom bodies, b. the medial and lateral calyces of the mushroom body together, c. lobes of the mushroom body for only those individuals that also had the calyxes segmented (n=78), d. the lobes of the 581 582 mushroom body for any individual that had them segmented (n=88). Exposure treatments are comparisons to control workers ('intercept') with age (3 versus 12-day workers) and ITD (inter-tegula 583 584 distance - proxy for body size) considered. For both models the interaction term between pesticide 585 treatment and age was removed as no significant effect could be detected. Significant differences ( $\alpha$ =0.05) 586 highlighted in bold black.

587

a) Total Mushroom Bodies (	1=78)				
LMER(relative MB ~ treatm	ent + age + si	ze + (1 colo	ny)		
	Estimate	Std. Error	df	t value	Pr(> t )
(Intercept)	0.058	0.008	43.244	7.732	<0.001
pre-eclosion	-0.004	0.002	5.845	-1.822	0.120
post-eclosion	-0.007	0.002	6.166	-2.873	0.027
continual	-0.005	0.002	6.348	-2.032	0.086
age	0.002	0.001	67.913	1.473	0.145
ITD	-0.006	0.002	49.038	-3.246	0.002
b) Total Mushroom Body Cal	yces (n=78)				
LMER(relative MBCalyces ~	treatment + a	age + size +	(1 colony)		
(Intercept)	0.035	0.005	40.979	7.735	< 0.001
pre-eclosion	-0.003	0.001	6.637	-2.409	0.049
post-eclosion	-0.005	0.001	7.212	-3.828	0.006
continual	-0.004	0.001	7.626	-2.896	0.021
age	0.001	0.001	69.152	1.565	0.122
ITD	-0.003	0.001	44.961	-2.919	0.005
c) Total Mushroom Body Lob	es (n=78)				
LMER(relative MBLobes ~ t	reatment + ag	ge + size + (2	1 colony)		
(Intercept)	0.024	0.003	65.458	8.182	< 0.001
pre-eclosion	-0.001	0.001	6.440	-0.890	0.406
post-eclosion	-0.001	0.001	6.184	-1.181	0.281
continual	-0.001	0.001	6.350	-0.936	0.383
age	0.001	0.000	74.263	1.137	0.259
ITD	-0.003	0.001	76.887	-4.285	<0.001
d) Total Mushroom Body Lol	oes (n=88)				
LMER(relative MBLobes ~ t	reatment + ag	ge + size + (2	1 colony)		
(Intercept)	0.024	0.003	53.979	7.295	<0.001
pre-eclosion	-0.001	0.001	5.995	-0.865	0.421
post-eclosion	-0.002	0.001	6.217	-1.387	0.213
continual	-0.001	0.001	6.261	-0.757	0.476
age	0.001	0.001	66.485	1.145	0.256
ITD	-0.003	0.001	63.112	-3.661	<0.001

Table S6. Statistical comparisons of the relative volumes of the central body, antennal lobes, lobulas and medullas. Statistical outputs are for the combined volumes of the left and right (except the central body), from a Generalized Mixed Effects Model in R (LMER). Exposure treatments are comparisons to *control* workers ('intercept') with *age* (3 versus 12-day workers) and ITD (inter-tegula distance which is a proxy for body size) considered. For both models the interaction term between pesticide treatment and age was removed as no significant effect could be detected. Significant differences ( $\alpha$ =0.05) highlighted in bold black.

596

Central Body n=88	~ troatmont : co	o . cizo . /1	(colony)		
Imer - central Body					
	Estimate	s.e.	df	t	
(Intercept)	0.000573	0.000176	64.7	3.26	0.00
Pre-eclosion	0.000074	0.000073	5.9	1.01	0.35
Post-eclosion	-0.000002	0.000072	5.6	-0.03	0.97
Continual	0.000001	0.000072	5.7	0.02	0.98
age	0.000027	0.000028	74.4	0.96	0.33
size	-0.000102	0.000038	78.4	-2.67	0.00
Antennal Lobes n=89					
lmer - antennal lob	es ~ treatment +	age + size +	(1 colony)		
(Intercept)	0.003658	0.000320	63.6	11.42	< 0.00
Pre-eclosion	-0.000034	0.000133	5.5	-0.26	0.80
Post-eclosion	-0.000208	0.000130	5.3	-1.60	0.16
Continual	-0.000176	0.000131	5.3	-1.35	0.23
age	0.000101	0.000051	73.9	1.98	0.05
size	-0.000553	0.000070	77.9	-7.96	<0.00
Lobulas n=71					
lmer - lobulas ~ tre	atment + age + s	ize + (1 colo	ony)		
(Intercept)	0.004984	0.000493	47.1	10.11	< 0.00
Pre-eclosion	0.000052	0.000140	6.4	0.37	0.72
Post-eclosion	0.000034	0.000142	7.3	0.24	0.81
Continual	0.000057	0.000146	7.4	0.39	0.70
age	0.000192	0.000083	61.5	2.30	0.02
size	-0.000693	0.000108	52.2	-6.44	<0.00
Medullas n=71	`	· · ·	ċ	· · · ·	
Imer - medullas ~ t	reatment + age +	size + (1 co	olony)		
(Intercept)	0.014020	0.001639	42.7	8.55	< 0.00
Pre-eclosion	-0.000094	0.000425	6.7	-0.22	0.83
Post-eclosion	0.000295	0.000435	8.2	0.68	0.51
Continual	0.000208	0.000447	8.0	0.46	0.65
age	0.000369	0.000287	62.5	1.29	0.20
size	-0.001826	0.000360	46.9	-5.08	<0.00

597

598

**Table S7.** Statistical output from a binomial generalised linear model in R (GLM) when analysing the final learning score achieved by each worker by the end of PER assay trials. For this model all workers from the three pesticide exposure treatments were pooled to compare one cohort against *control* workers (intercept). Significant differences ( $\alpha$ =0.05) are highlighted in bold.

604

glm(learning_score ~ relativeMBCalyces*treatment, family = binomial)							
	Estimate	Std. Error	z value	Pr(> z )			
(Intercept)	-6.992	1.416	-4.939	< 0.001			
calyces	265.823	58.981	4.507	<0.001			
pesticide	5.611	1.802	3.113	0.002			
calyces*pesticide	-321.899	81.268	-3.961	<0.001			

# 605

606

607

608

609

Table S8. Spearman's rank correlations to investigate the relationship between the response variable learning score and the predictor variables body size (ITD) or mushroom body volume. Row 1 is intertegula distance (ITD; proxy for body size) for workers regardless of age; Rows 2 & 3 are ITDs for 3 and 12-day workers (respectively); Rows 4-5 is the absolute volume of the whole mushroom bodies (left and right combined), and relative volume of the whole mushroom bodies (left and right combined) corrected for body size, respectively; Rows 6-9 also show absolute and relative volumes for the calyces and lobes separately (left and right combined). Significant differences ( $\alpha$ =0.05) are highlighted in bold red, with near

617 significant differences ( $\alpha$ =0.1) in bold black.

Row	MB Component	Con	trol	Pesticide	
	NB component	SR	р	SR	р
1	ITD	-0.509	0.053	0.020	0.919
2	ITD 3-Day	-0.123	0.772	-0.314	0.295
3	ITD 12-Day	-0.818	0.024	0.171	0.527
4	MB	0.403	0.136	-0.030	0.879
5	Relative MB	0.690	0.004	-0.126	0.516
6	MB Calyces	0.354	0.195	-0.047	0.807
7	<b>Relative MB Calyces</b>	0.622	0.013	-0.130	0.503
8	MB Lobes	0.579	0.024	0.007	0.972
9	Relative MB Lobes	0.633	0.011	-0.098	0.612

# **Table S9. Colony census at start of the experiment, and at the end of each phase.** All workers

- 620 eclosing during Phase II were tagged with a unique colour and numbered Opalith tag so the age of
- 621 each worker was known.

622

			# worke	rs eclosed
colony#	treatment	starting # workers	end of phase I	end of phase II (i.e. tagged)
1	Control	25	27	45
2	Control	17	48	58
3	Control	20	26	35
13	Control	10	28	40
14	Control	17	12	2
4	Pre-eclosion	9	17	16
5	Pre-eclosion	24	23	47
6	Pre-eclosion	18	36	13
15	Pre-eclosion	16	36	41
16	Pre-eclosion	11	19	18
17	Pre-eclosion	8	24	8
7	Post-eclosion	12	41	6
8	Post-eclosion	20	54	42
9	Post-eclosion	20	34	22
18	Post-eclosion	15	24	47
19	Post-eclosion	12	41	28
20	Post-eclosion	7	33	13
10	Continual	12	51	54
11	Continual	14	42	29
12	Continual	17	20	10
21	Continual	13	38	40
22	Continual	6	24	24

623

625

Table S10. Colony daily sucrose consumption (ml) as the experiment progressed (days 1-43). Of the

627 total provisioned sucrose (1,110ml per colony), control, pre-eclosion, post-eclosion and continual

628 treatments consumed a median (IQR) of 54 (50-56), 47 (37-57), 61 (48-69) and 51 (35-63) %

629 respectively.

Treatment	Colony #									С	onsun	nptior	n (ml)	per da	iy								
		1	3	5	7	9	11	13	15	17	19	21	23	25	27	29	31	33	35	37	39	41	43
Control	1	9	15	25	15	18	21	20	22	28	34	15	40	30	35	37	37	31	36	35	33	33	37
Control	2	14	20	29	17	20	20	29	28	27	33	13	28	28	30	30	34	26	32	35	38	37	35
Control	3	10	10	19	14	20	20	21	22	23	30	14	30	28	38	37	38	30	37	35	37	39	39
Pre-eclosion	4	7	10	13	10	13	14	15	12	16	19	9	20	17	20	16	17	16	21	21	22	25	26
Pre-eclosion	5	13	12	24	18	24	22	20	20	23	26	12	27	25	30	31	32	28	32	30	33	35	35
Pre-eclosion	6	25	21	31	24	33	33	31	31	29	39	14	41	34	39	37	40	28	35	37	35	37	45
Post-eclosion	7	13	10	29	13	17	18	16	17	20	26	12	26	37	45	46	43	36	43	48	48	54	59
Post-eclosion	8	29	22	32	21	24	24	29	23	30	37	14	30	34	39	39	38	32	42	57	42	45	52
Post-eclosion	9	18	17	30	23	28	28	35	39	35	39	21	46	42	48	53	52	36	46	52	48	55	55
continual	10	19	18	25	20	22	25	27	28	33	35	13	35	35	40	42	46	27	40	36	40	40	35
continual	11	20	15	22	24	25	25	28	26	28	32	15	35	35	39	36	46	36	38	35	35	33	32
continual	12	5	15	17	12	15	15	21	16	18	25	11	24	1	29	25	28	25	31	33	36	40	41
Control	13	17	17	17	23	21	25	21	26	24	28	15	32	30	36	33	35	38	40	39	40	40	41
Control	14	15	16	14	17	13	15	16	20	16	21	12	25	23	34	30	31	29	33	33	35	32	34
Pre-eclosion	15	17	19	23	26	25	29	24	29	25	29	17	37	28	33	31	31	30	30	26	31	29	31
Pre-eclosion	16	17	18	14	23	17	22	18	22	20	24	14	30	27	29	25	25	23	23	22	24	23	23
Pre-eclosion	17	13	14	11	19	16	21	19	22	18	20	11	25	21	23	23	23	23	23	21	21	17	19
Post-eclosion	18	15	15	15	22	18	22	29	30	19	24	14	28	23	27	36	32	30	28	32	25	27	27
Post-eclosion	19	23	23	23	29	25	31	26	36	34	36	21	38	33	38	29	34	34	36	36	34	34	33
Post-eclosion	20	15	15	20	26	20	23	22	26	21	26	13	23	22	26	25	29	30	31	28	27	26	26
Continual	21	18	19	19	26	25	29	29	37	30	39	21	42	35	39	41	43	45	46	46	47	45	46
Continual	22	13	13	10	13	10	12	9	10	7	10	4	10	7	12	12	12	12	13	15	16	9	11
Continual	23	14	15	15	21	13	16	16	19	16	19	12	24	22	25	25	25	23	25	24	25	24	26

630

631

## 634 **REFERENCES**

- 635 Alford D. 1975. Bumblebees. London: Davis-Poynter.
- Aliouane Y, El Hassani AK, Gary V, Armengaud C, Lambin M, Gauthier M. 2009. Subchronic exposure
   of honeybees to sublethal doses of pesticides: Effects on behavior. *Environ Toxicol Chem* 28:113–122. doi:10.1897/08-110.1
- Andrione M, Vallortigara G, Antolini R, Haase A. 2016. Neonicotinoid-induced impairment of odour
   coding in the honeybee. *Sci Rep* 6:38110. doi:10.1038/srep38110
- Arce AN, David TI, Randall EL, Ramos Rodrigues A, Colgan TJ, Wurm Y, Gill RJ. 2017. Impact of
   controlled neonicotinoid exposure on bumblebees in a realistic field setting. *J Appl Ecol* 54:1199–1208. doi:10.1111/1365-2664.12792
- Baird E, Taylor G. 2017. X-ray micro computed-tomography. *Curr Biol* 27:R289–R291.
   doi:10.1016/j.cub.2017.01.066
- Barbara GS, Grünewald B, Paute S, Gauthier M, Raymond-Delpech V. 2008. Study of nicotinic
   acetylcholine receptors on cultured antennal lobe neurones from adult honeybee brains.
   *Invertebr Neurosci* 8:19–29. doi:10.1007/s10158-007-0062-2
- Baron GL, Jansen VAA, Brown MJF, Raine NE. 2017. Pesticide reduces bumblebee colony initiation
  and increases probability of population extinction. *Nat Ecol Evol* 1:1308–1316.
  doi:10.1038/s41559-017-0260-1
- Bates D, Machler M, Bolker B, Walker S. 2015. Fitting linear mixed-effects models using lme4. *J Stat Softw* 67:1–48.
- 654Bitterman ME, Menzel R, Fietz A, Schäfer S. 1983. Classical conditioning of proboscis extension in655honeybees (Apis mellifera). J Comp Psychol **97**:107–119. doi:10.1037/0735-7036.97.2.107
- Botias C, David A, Horwood J, Abdul-Sada A, Nicholls E, Hill E, Goulson D. 2015. Neonicotinoid
   residues in wildflowers, a potential route of chronic exposure for bees. *Environ Sci Technol* 49:12731–12740. doi:10.1021/acs.est.5b03459
- 659 Brittain C, Potts SG. 2011. The potential impacts of insecticides on the life-history traits of bees and 660 the consequences for pollination. *Basic Appl Ecol* **12**:321–331. doi:10.1016/j.baae.2010.12.004
- Brittain C, Williams N, Kremen C, Klein A-M. 2013. Synergistic effects of non-Apis bees and honey
   bees for pollination services. *Proc R Soc B-Biological Sci* 280. doi:10.1098/rspb.2012.2767
- 663 Bryden J, Gill RJ, Mitton RAA, Raine NE, Jansen VAA. 2013. Chronic sublethal stress causes bee 664 colony failure. *Ecol Lett* **16**:1463–1469. doi:10.1111/ele.12188
- 665 Cabirol A, Cope AJ, Barron AB, Devaud JM. 2018. Relationship between brain plasticity, learning and
  666 foraging performance in honey bees. *PLoS One* **13**:e0196749.
  667 doi:10.1371/journal.pone.0196749
- 668 Calatayud-Vernich P, Calatayud F, Simó E, Picó Y. 2018. Pesticide residues in honey bees, pollen and
  669 beeswax: Assessing beehive exposure. *Environ Pollut* 241:106–114.
  670 doi:10.1016/j.envpol.2018.05.062
- 671 Cane JH. 1987. Estimation of bee size using intertegular span (Apoidea). *J Kansas Entomol Soc* 672 60:145–147. doi:10.1098/rstb.2011.0050

- 673 Casida JE. 2018. Neonicotinoids and other insect nicotinic receptor competitive modulators: Progress
- 674 and prospects. *Annu Rev Entomol* **63**:125–144. doi:10.1146/annurev-ento-020117-043042
- 675 Chauzat MP, Faucon JP, Martel AC, Lachaize J, Cougoule N, Aubert M. 2006. A survey of pesticide
  676 residues in pollen loads collected by honey bees in France. *J Econ Entomol* 99:253–262.
  677 doi:10.1093/jee/99.2.253
- 678 Chittka L. 2017. Bee cognition. *Curr Biol* **27**:R1049–R1053. doi:10.1016/j.cub.2017.08.008
- 679 Cnaani J, Schmid-Hempel R, Schmidt JO. 2002. Colony development, larval development and worker
   680 reproduction in Bombus impatiens Cresson. *Insectes Soc* 49:164–170. doi:10.1007/s00040-002 681 8297-8
- 682 Crall JD, Switzer CM, Oppenheimer RL, Ford Versypt AN, Dey B, Brown A, Eyster M, Guérin C, Pierce
  683 NE, Combes SA, de Bivort BL. 2018. Neonicotinoid exposure disrupts bumblebee nest behavior,
  684 social networks, and thermoregulation. *Science (80- )* 362:683–686.
  685 doi:10.1126/science.aat1598
- 686 Cressey D. 2017. The bitter battle over the world's most popular insecticides. *Nature* 551:156–158.
   687 doi:10.1038/551156a
- Cresswell JE, Page CJ, Uygun MB, Holmbergh M, Li Y, Wheeler JG, Laycock I, Pook CJ, de Ibarra NH,
   Smirnoff N, Tyler CR. 2012. Differential sensitivity of honey bees and bumble bees to a dietary
   insecticide (imidacloprid). *Zoology* 115:365–371. doi:10.1016/j.zool.2012.05.003
- Daniele G, Giroud B, Jabot C, Vulliet E. 2018. Exposure assessment of honeybees through study of
   hive matrices: analysis of selected pesticide residues in honeybees, beebread, and beeswax
   from French beehives by LC-MS/MS. *Environ Sci Pollut Res* 25:6145–6153. doi:10.1007/s11356 017-9227-7
- David A, Botías C, Abdul-Sada A, Nicholls E, Rotheray EL, Hill EM, Goulson D. 2016. Widespread
  contamination of wildflower and bee-collected pollen with complex mixtures of neonicotinoids
  and fungicides commonly applied to crops. *Environ Int* 88:169–178.
  doi:10.1016/j.envint.2015.12.011
- Decourtye A, Armengaud C, Renou M, Devillers J, Cluzeau S, Gauthier M, Pham-Delègue MH. 2004.
   Imidacloprid impairs memory and brain metabolism in the honeybee (Apis mellifera L.). *Pestic Biochem Physiol* **78**:83–92. doi:10.1016/j.pestbp.2003.10.001
- Démares FJ, Crous KL, Pirk CWW, Nicolson SW, Human H. 2016. Sucrose sensitivity of honey bees is
   differently affected by dietary protein and a neonicotinoid pesticide. *PLoS One* **11**:e0156584.
   doi:10.1371/journal.pone.0156584
- Démares FJ, Pirk CWW, Nicolson SW, Human H. 2018. Neonicotinoids decrease sucrose
  responsiveness of honey bees at first contact. *J Insect Physiol* 108:25–30.
  doi:10.1016/j.jinsphys.2018.05.004
- Desneux N, Decourtye A, Delpuech J-M. 2007. The sublethal effects of pesticides on beneficial
   arthropods. *Annu Rev Entomol* 52:81–106. doi:10.1146/annurev.ento.52.110405.091440
- Duchateau MJ, Velthuis HHW. 1988. Development and reproductive strategies in Bombus terrestris
   colonies. *Behaviour* 107:186–207. doi:10.1163/156853988x00340
- Durst C, Eichmüller S, Menzel R. 1994. Development and experience lead to increased volume of
   subcompartments of the honeybee mushroom body. *Behav Neural Biol* 62:259–263.

- 715 Dyer AG, Paulk AC, Reser DH. 2011. Colour processing in complex environments: Insights from the
  716 visual system of beesProceedings of the Royal Society B: Biological Sciences.
  717 doi:10.1098/rspb.2010.2412
- Fahrbach SE. 2006. Structure of the mushroom bodies of the insect brain. *Annu Rev Entomol* 51:209–
  232. doi:10.1146/annurev.ento.51.110104.150954
- Fahrbach SE, Moore D, Capaldi EA, Farris SM, Robinson GE. 1998. Experience-expectant plasticity in
   the mushroom bodies of the honeybee. *Learn Mem* 5:115–123. doi:10.1101/lm.5.1.115
- Fahrbach SE, Strande JL, Robinson GE. 1995. Neurogenesis is absent in the brains of adult honey
  bees and does not explain behavioral neuroplasticity. *Neurosci Lett* 197:145–148.
  doi:10.1016/0304-3940(95)11913-H
- Farris SM, Robinson GE, Davis RL, Fahrbach SE. 1999. Larval and pupal development of the
  mushroom bodies in the honey bee, Apis mellifera. *J Comp Neurol* 414:97–113.
- 727 doi:10.1002/(SICI)1096-9861(19991108)414:1<97::AID-CNE8>3.0.CO;2-Q
- Farris SM, Robinson GE, Fahrbach SE. 2001. Experience- and age-related outgrowth of intrinsic
   neurons in the mushroom bodies of the adult worker honeybee. *J Neurosci* 21:6395–6404.
   doi:citeulike-article-id:467293
- Feltham H, Park K, Goulson D. 2014. Field realistic doses of pesticide imidacloprid reduce bumblebee
   pollen foraging efficiency. *Ecotoxicology* 23:317–323. doi:10.1007/s10646-014-1189-7
- Fischer J, Mueller T, Spatz A-K, Greggers U, Gruenewald B, Menzel R. 2014. Neonicotinoids interfere
  with specific components of navigation in Honeybees. *PLoS One* **9**:e91364.
  doi:10.1371/journal.pone.0091364
- Galizia CG, Sachse S, Rappert A, Menzel R. 1999. The glomerular code for odor representation is
   species specific in the honeybee Apis mellifera. *Nat Neurosci* 2:473–478. doi:10.1038/8144
- Galizia GC, Galizia CG, Eisenhardt D, Giurfa M, York N, Eisenhardt ED, Galizia, G.C. and Eisenhardt, D.
  and Galizia, C.G. and Giurfa M. 2011. Honeybee neurobiology and behavior: A tribute to
  Randolf Menzel.
- 741 Garibaldi LA, Steffan-Dewenter I, Winfree R, Aizen MA, Bommarco R, Cunningham SA, Kremen C, 742 Carvalheiro LG, Harder LD, Afik O, Bartomeus I, Benjamin F, Boreux V, Cariveau D, Chacoff NP, 743 Dudenhöffer JH, Freitas BM, Ghazoul J, Greenleaf S, Hipólito J, Holzschuh A, Howlett B, Isaacs R, 744 Javorek SK, Kennedy CM, Krewenka KM, Krishnan S, Mandelik Y, Mayfield MM, Motzke I, Munyuli T, Nault BA, Otieno M, Petersen J, Pisanty G, Potts SG, Rader R, Ricketts TH, Rundlöf 745 746 M, Seymour CL, Schüepp C, Szentgyörgyi H, Taki H, Tscharntke T, Vergara CH, Viana BF, Wanger 747 TC, Westphal C, Williams N, Klein AM. 2013. Wild pollinators enhance fruit set of crops 748 regardless of Honeybee abundance. Science (80-) 339:1608–1611.
- 749 doi:10.1126/science.1230200
- Gill RJ, Baldock KCR, Brown MJF, Cresswell JE, Dicks L V., Fountain MT, Garratt MPD, Gough LA,
  Heard MS, Holland JM, Ollerton J, Stone GN, Tang CQ, Vanbergen AJ, Vogler AP, Woodward G,
  Arce AN, Boatman ND, Brand-Hardy R, Breeze TD, Green M, Hartfield CM, O'Connor RS,
  Osborne JL, Phillips J, Sutton PB, Potts SG. 2016. Protecting an ecosystem service: Approaches
  to understanding and mitigating threats to wild insect pollinators. *Adv Ecol Res* 54:135–206.
  doi:10.1016/bs.aecr.2015.10.007

bioRxiv preprint doi: https://doi.org/10.1101/690602; this version posted July 2, 2019. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under Gill RJ, Raine NE. 2014. Chronic impairment of bumblebee natural foraging behaviour induced by

- Gill RJ, Raine NE. 2014. Chronic impairment of bumblebee natural foraging behaviour induced
   sublethal pesticide exposure. *Funct Ecol* 28:1459–1471. doi:10.1111/1365-2435.12292
- Gill RJ, Ramos-Rodriguez O, Raine NE. 2012. Combined pesticide exposure severely affects individual and colony-level traits in bees. *Nature* 491:105–8. doi:10.1038/nature11585
- Giurfa M, Sandoz JC. 2012. Invertebrate learning and memory: Fifty years of olfactory conditioning of
   the proboscis extension response in honeybees. *Learn Mem* 19:54–66.
- 762 doi:10.1101/lm.024711.111
- Goulson D. 2010. Bumblebees: behaviour, ecology and conservation, 2nd ed. Oxford, UK: Oxford
   University Press.
- Goulson D, Nicholls E, Botías C, Rotheray EL. 2015. Bee declines driven by combined stress from
   parasites, pesticides, and lack of flowers. *Science (80- )* 347:1–16. doi:10.1126/science.1255957
- Gregorc A, Evans JD, Scharf M, Ellis JD. 2012. Gene expression in honey bee (Apis mellifera) larvae
   exposed to pesticides and Varroa mites (Varroa destructor). *J Insect Physiol* 58:1042–1049.
   doi:10.1016/j.jinsphys.2012.03.015
- Groh C, Tautz J, Rossler W. 2004. Synaptic organization in the adult honey bee brain is influenced by
   brood-temperature control during pupal development. *Proc Natl Acad Sci* 101:4268–4273.
   doi:10.1073/pnas.0400773101
- Gutiérrez Y, Ott D, Töpperwien M, Salditt T, Scherber C. 2018. X-ray computed tomography and its
   potential in ecological research: A review of studies and optimization of specimen preparation.
   *Ecol Evol* 8:7717–7732. doi:10.1002/ece3.4149
- Hallmann CA, Sorg M, Jongejans E, Siepel H, Hofland N, Schwan H, Stenmans W, Müller A, Sumser H,
  Hörren T, Goulson D, De Kroon H. 2017. More than 75 percent decline over 27 years in total
  flying insect biomass in protected areas. *PLoS One* 12:e0185809.
- 779 doi:10.1371/journal.pone.0185809
- Hammer M, Menzel R. 1995. Learning and memory in the honeybee. *J Neurosci*.
  doi:10.1016/j.cub.2005.09.015
- Hansson BS, Anton S. 2002. Function and morphology of the antennal lobe: New developments.
   *Annu Rev Entomol* 45:203–231. doi:10.1146/annurev.ento.45.1.203
- Heard MS, Baas J, Dorne J Lou, Lahive E, Robinson AG, Rortais A, Spurgeon DJ, Svendsen C, Hesketh
  H. 2017. Comparative toxicity of pesticides and environmental contaminants in bees: Are
  honey bees a useful proxy for wild bee species? *Sci Total Environ* 578:357–365.
- 787 doi:10.1016/j.scitotenv.2016.10.180
- Heisenberg M. 2003. Mushroom body memoir: From maps to models. *Nat Rev Neurosci* 4:266–275.
   doi:10.1038/nrn1074
- Heisenberg M. 1998. What do the mushroom bodies do for the insect brain? an introduction. *Learn Mem.* doi:10.1101/lm.5.1.1
- Hourcade B, Muenz TS, Sandoz JC, Rossler W, Devaud JM. 2010. Long-term memory leads to synaptic
   reorganization in the mushroom bodies: A memory trace in the nsect Brain? *J Neurosci* 30:6461–6465. doi:10.1523/jneurosci.0841-10.2010
- 795 Jeschke P, Nauen R. 2008. Neonicotinoids From zero to hero in insecticide chemistry. Pest Manag

- Johnson BR. 2010. Division of labor in honeybees: Form, function, and proximate mechanisms.
   Behav Ecol Sociobiol 64:305–316. doi:10.1007/s00265-009-0874-7
- Jones BM, Leonard AS, Papaj DR, Gronenberg W. 2013. Plasticity of the worker bumblebee brain in
   relation to age and rearing environment. *Brain Behav Evol* 82:250–261.
   doi:10.1159/000355845
- Kasiotis KM, Anagnostopoulos C, Anastasiadou P, Machera K. 2014. Pesticide residues in honeybees,
   honey and bee pollen by LC-MS/MS screening: Reported death incidents in honeybees. *Sci Total Environ* 485:633–642. doi:10.1016/j.scitotenv.2014.03.042
- Kühn-Bühlmann S, Wehner R. 2006. Age-dependent and task-related volume changes in the
   mushroom bodies of visually guided desert ants, Cataglyphis bicolor. J Neurobiol 66:511–521.
   doi:10.1002/neu.20235
- Laloi D, Sandoz JC, Picard-Nizou AL, Marchesi A, Pouvreau A, Taséi JN, Poppy G, Pham-Delègue MH.
   1999. Olfactory conditioning of the proboscis extension in bumble bees. *Entomol Exp Appl* 90:123–129. doi:10.1023/A:1003598301272
- Leza M, Watrous KM, Bratu J, Woodard SH. 2018. Effects of neonicotinoid insecticide exposure and
   monofloral diet on nest-founding bumblebee queens. *Proc R Soc B Biol Sci* 285:20180761.
   doi:10.1098/rspb.2018.0761
- Li L, MaBouDi H Di, Egertová M, Elphick MR, Chittka L, Perry CJ. 2017. A possible structural correlate
   of learning performance on a colour discrimination task in the brain of the bumblebee. *Proc R Soc B Biol Sci* 284:20171323. doi:10.1098/rspb.2017.1323
- Li W, Pan Y, Wang Z, Gong H, Gong Z, Liu L. 2009. Morphological characterization of single fan shaped body neurons in Drosophila melanogaster. *Cell Tissue Res*. doi:10.1007/s00441-009 0781-2
- Lotto RB, Chittka L. 2005. Seeing the light: Illumination as a contextual cue to color choice behavior
   in bumblebees. *Proc Natl Acad Sci.* doi:10.1073/pnas.0500681102
- Maleszka J, Barron AB, Helliwell PG, Maleszka R. 2009. Effect of age, behaviour and social
   environment on honey bee brain plasticity. *J Comp Physiol A Neuroethol Sensory, Neural, Behav Physiol* 195:733–740. doi:10.1007/s00359-009-0449-0
- Mitchell EAD, Mulhauser B, Mulot M, Mutabazi A, Glauser G, Aebi A. 2017. A worldwide survey of
   neonicotinoids in honey. *Science (80- )* 358:109–111. doi:10.1126/science.aan3684
- Mobbs PG. 2006. The Brain of the Honeybee Apis Mellifera. I. The Connections and Spatial
  Organization of the Mushroom Bodies. *Philos Trans R Soc B Biol Sci.*doi:10.1098/rstb.1982.0086
- Mullin CA, Frazier M, Frazier JL, Ashcraft S, Simonds R, VanEngelsdorp D, Pettis JS. 2010. High levels
   of miticides and agrochemicals in North American apiaries: Implications for Honeybee health.
   *PLoS One* 5:e9754. doi:10.1371/journal.pone.0009754
- Muth F, Leonard AS. 2019. A neonicotinoid pesticide impairs foraging, but not learning, in free-flying
   bumblebees. *Sci Rep* 9:4764.
- 835 Palmer MJ, Moffat C, Saranzewa N, Harvey J, Wright GA, Connolly CN. 2013. Cholinergic pesticides

bioRxiv preprint doi: https://doi.org/10.1101/690602; this version posted July 2, 2019. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under cause mushroom body neurorial inactivation. In honeybees. *Nat Commun* **4**:1634.

836

837 doi:10.1038/ncomms2648

- Paulk Angelique C., Dacks AM, Gronenberg W. 2009. Color processing in the medulla of the
  bumblebee (Apidae: Bombus impatiens). *J Comp Neurol*. doi:10.1002/cne.21993
- Paulk A. C., Dacks AM, Phillips-Portillo J, Fellous J-M, Gronenberg W. 2009. Visual Processing in the
   Central Bee Brain. J Neurosci. doi:10.1523/JNEUROSCI.1325-09.2009
- Paulk AC, Phillips-Portillo J, Dacks AM, Fellous J-M, Gronenberg W. 2008. The Processing of Color,
   Motion, and Stimulus Timing Are Anatomically Segregated in the Bumblebee Brain. *J Neurosci*.
   doi:10.1523/jneurosci.1196-08.2008
- Peng YC, Yang EC. 2016. Sublethal dosage of imidacloprid reduces the microglomerular density of
   honey bee mushroom bodies. *Sci Rep* 6:19298. doi:10.1038/srep19298
- Pfeiffer K, Homberg U. 2013. Organization and Functional Roles of the Central Complex in the Insect
   Brain. Annu Rev Entomol. doi:10.1146/annurev-ento-011613-162031
- Piiroinen S, Goulson D. 2016. Chronic neonicotinoid pesticide exposure and parasite stress
   differentially affects learning in honeybees and bumblebees. *Proc R Soc B Biol Sci* 283:20160246. doi:10.1098/rspb.2016.0246
- Pohorecka K, Skubida P, Miszczak A, Semkiw P, Sikorski P, Zagibajlo K, Teper D, Koltowski Z, Skubida
  M, Zdanska D, Bober A. 2012. Residues of neonicotinoid insecticides in bee collected plant
  materials from Oilseed Rape crops and their effect on bee colonies. *J Apic Sci* 56:115–134.
  doi:DOI 10.2478/v10289-012-0029-3
- Pohorecka K, Szczęsna T, Witek M, Miszczak A, Sikorski P. 2017. The exposure of honey bees to
   pesticide residues in the hive environment with regard to winter colony losses. *J Apic Sci* 61:105–125. doi:10.1515/JAS-2017-0013
- Potts SG, Imperatriz-Fonseca V, Ngo HT, Aizen MA, Biesmeijer JC, Breeze TD, Dicks L V., Garibaldi LA,
   Hill R, Settele J, Vanbergen AJ. 2016. Safeguarding pollinators and their values to human well being. *Nature* 540:220–229. doi:10.1038/nature20588
- Ribi W, Senden TJ, Sakellariou A, Limaye A, Zhang S. 2008. Imaging honey bee brain anatomy with
   micro-X-ray-computed tomography. *J Neurosci Methods* 171:93–97.
   doi:10.1016/j.jneumeth.2008.02.010
- Riveros AJ, Gronenberg W. 2010. Brain allometry and neural plasticity in the bumblebee bombus
   occidentalis. *Brain Behav Evol* **75**:138–148. doi:10.1159/000306506
- Riveros AJ, Gronenberg W. 2009. Olfactory learning and memory in the bumblebee Bombus
   occidentalis. *Naturwissenschaften* 96:851–856. doi:10.1007/s00114-009-0532-y
- Rundlöf M, Andersson GKS, Bommarco R, Fries I, Hederström V, Herbertsson L, Jonsson O, Klatt
  BKBK, Pedersen TR, Yourstone J, Smith HG, Rundlof M, Andersson GKS, Bommarco R, Fries I,
  Hederstrom V, Herbertsson L, Jonsson O, Klatt BKBK, Pedersen TR, Yourstone J, Smith HG.
  2015. Seed coating with a neonicotinoid insecticide negatively affects wild bees. *Nature*521:77-U162. doi:10.1038/nature14420
- Sachse S, Rappert A, Galizia CG. 1999. The spatial representation of chemical structures in the
  antennal lobe of honeybees: Steps towards the olfactory code. *Eur J Neurosci* 11:3970–3982.
  doi:10.1046/j.1460-9568.1999.00826.x

- Samuelson EEW, Chen-Wishart ZP, Gil RJ, Leadbeater E. 2016. Effect of acute pesticide exposure on
   bee spatial working memory using an analogue of the radial-arm maze. *Sci Rep* 6:38957.
   doi:10.1038/srep38957
- Sandrock C, Tanadini LG, Pettis JS, Biesmeijer JC, Potts SG, Neumann P. 2014. Sublethal
   neonicotinoid insecticide exposure reduces solitary bee reproductive success. *Agric For Entomol* 16:119–128. doi:10.1111/afe.12041
- Simmons DM, Swanson LW. 2009. Comparing histological data from different brains: Sources of
   error and strategies for minimizing them. *Brain Res Rev* 60:349–367.
   doi:10.1016/j.brainresrev.2009.02.002
- Siviter H, Brown MJF, Leadbeater E. 2018a. Sulfoxaflor exposure reduces bumblebee reproductive
   success. *Nature* 561:109–112. doi:10.1038/s41586-018-0430-6
- Siviter H, Koricheva J, Brown MJF, Leadbeater E. 2018b. Quantifying the impact of pesticides on
   learning and memory in bees. J Appl Ecol 55:2812–2821. doi:10.1111/1365-2664.13193
- Smith BH, Burden CM. 2014. A proboscis extension response protocol for investigating behavioral
   plasticity in insects: Application to basic, biomedical, and agricultural research. *J Vis Exp* 892 8:e51057. doi:10.3791/51057
- Smith DB, Bernhardt G, Raine NE, Abel RL, Sykes D, Ahmed F, Pedroso I, Gill RJ. 2016. Exploring
   miniature insect brains using micro-CT scanning techniques. *Sci Rep* 6:21768.
   doi:10.1038/srep21768
- Smith KE, Raine NE. 2014. A comparison of visual and olfactory learning performance in the
   bumblebee Bombus terrestris. *Behav Ecol Sociobiol* 68:1549–1559. doi:10.1007/s00265-014 1765-0
- Sommerlandt FMJ, Rossler W, Spaethe J. 2014. Elemental and non-elemental olfactory learning using
   per conditioning in the bumblebee, Bombus terrestris. *Apidologie* 45:106–115.
   doi:10.1007/s13592-013-0227-4
- Stanley DA, Raine NE. 2016. Chronic exposure to a neonicotinoid pesticide alters the interactions
   between bumblebees and wild plants. *Funct Ecol* **30**:1132–1139. doi:10.1111/1365-2435.12644
- Stanley DA, Smith KE, Raine NE. 2015. Bumblebee learning and memory is impaired by chronic
   exposure to a neonicotinoid pesticide. *Sci Rep* 5:16508. doi:10.1038/srep16508
- Steijven K, Spaethe J, Steffan-Dewenter I, Härtel S. 2017. Learning performance and brain structure
   of artificially-reared honey bees fed with different quantities of food. *PeerJ* 5:e3858.
   doi:10.7717/peerj.3858
- Strauss R. 2002. The central complex and the genetic dissection of locomotor behaviour. *Curr Opin Neurobiol.* doi:10.1016/S0959-4388(02)00385-9
- 911 Tan K, Chen W, Dong S, Liu X, Wang Y, Nieh JC. 2015. A neonicotinoid impairs olfactory learning in
  912 Asian honey bees (Apis cerana) exposed as larvae or as adults. *Sci Rep* 5:10989.
  913 doi:10.1038/srep10989
- Thany SH, Gauthier M. 2005. Nicotine injected into the antennal lobes induces a rapid modulation of
   sucrose threshold and improves short-term memory in the honeybee Apis mellifera. *Brain Res* **1039**:216–219. doi:10.1016/j.brainres.2005.01.056

- Tison L, Holtz S, Adeoye A, Kalkan Ö<sup>a</sup> in the second state of sublethal
   doses of thiacloprid and its formulation Calypso <sup>®</sup> on the learning and memory performance of
   honey bees. J Exp Biol 220:3695–3705. doi:10.1242/jeb.154518
- Tomé HVV, Martins GF, Lima MAP, Campos LAO, Guedes RNC. 2012. Imidacloprid-induced
   impairment of mushroom bodies and behavior of the native stingless bee melipona
   quadrifasciata anthidioides. *PLoS One* **7**:e38406. doi:10.1371/journal.pone.0038406
- Tosi S, Nieh JC. 2019. Lethal and sublethal synergistic effects of a new systemic pesticide,
  flupyradifurone (Sivantow), on honeybees. *Proc R Soc B* 286:20190433.
- 925 Tsvetkov N, Samson-Robert O, Sood K, Patel HS, Malena DA, Gajiwala PH, Maciukiewicz P, Fournier
   926 V, Zayed A. 2017. Chronic exposure to neonicotinoids reduces honey bee health near corn
   927 crops. *Science (80- )* 356:1395–1397. doi:10.1126/science.aam7470
- Valdovinos-Flores C, Alcantar-Rosales VM, Gaspar-Ramírez O, Saldaña-Loza LM, Dorantes-Ugalde JA.
   2017. Agricultural pesticide residues in honey and wax combs from Southeastern, Central and
   Northeastern Mexico. J Apic Res 56:667–679. doi:10.1080/00218839.2017.1340798
- 931 Vanbergen AJ. 2013. Threats to an ecosystem service: Pressures on pollinators. *Front Ecol Environ* 932 **11**:251–259. doi:10.1890/120126
- Whitehorn PR, O'Connor S, Wackers FL, Goulson D. 2012. Neonicotinoid pesticide reduces bumble
  bee colony growth and queen production. *Science (80- )* 336:351–352.
  doi:10.1126/science.1215025
- Wilson DE, Velarde RA, Fahrbach SE, Mommaerts V, Smagghe G. 2013. Use of primary cultures of
   Kenyon cells from bumblebee brains to assess pesticide side effects. *Arch Insect Biochem Physiol* 84:43–56. doi:10.1002/arch.21112
- Winnington AP, Napper RM, Mercer AR. 1996. Structural plasticity of identified glomeruli in the
  antennal lobes of the adult worker honey bee. *J Comp Neurol* 365:479–490.
  doi:10.1002/(SICI)1096-9861(19960212)365:3<479::AID-CNE10>3.0.CO;2-M
- 942 Withers GS, Fahrbach SE, Robinson GE. 1995. Effects of experience and juvenile hormone on the
  943 organization of the mushroom bodies of honey bees. *J Neurobiol* 26:130–144.
  944 doi:10.1002/neu.480260111
- Withers GS, Fahrbach SE, Robinson GE. 1993. Selective neuroanatomical plasticity and division of
  labour in the honeybee. *Nature* 364:238–240. doi:10.1038/364238a0
- 947 Woodcock BA, Bullock JM, Shore RF, Heard MS, Pereira MG, Redhead J, Ridding L, Dean H, Sleep D,
  948 Henrys P, Peyton J, Hulmes S, Hulmes L, Sárospataki M, Saure C, Edwards M, Genersch E, Knäbe
  949 S, Pywell RF. 2017. Country-specific effects of neonicotinoid pesticides on honey bees and wild
  950 bees. *Science (80- )* 356:1393–1395. doi:10.1126/science.aaa1190
- Woodcock BA, Isaac NJB, Bullock JM, Roy DB, Garthwaite DG, Crowe A, Pywell RF. 2016. Impacts of
   neonicotinoid use on long-term population changes in wild bees in England. *Nat Commun* 7:12459. doi:10.1038/ncomms12459
- Wu JY, Anelli CM, Sheppard WS. 2011. Sub-lethal effects of pesticide residues in brood comb on
   worker honey bee (apis mellifera) development and longevity. *PLoS One* 6:e14720.
   doi:10.1371/journal.pone.0014720
- 957 Wu JY, Smart MD, Anelli CM, Sheppard WS. 2012. Honey bees (Apis mellifera) reared in brood combs

958 containing high levels of pesticide Pesioues exhibit increased susceptibility to Nosema

959 (Microsporidia) infection. J Invertebr Pathol 109:326–329. doi:10.1016/j.jip.2012.01.005

- Yang EC, Chang HC, Wu WY, Chen YW. 2012. Impaired olfactory associative behavior of Honeybee
   workers due to contamination of Imidacloprid in the larval stage. *PLoS One* 7:e49472.
   doi:10.1371/journal.pone.0049472
- 2ars T. 2000. Behavioral functions of the insect mushroom bodies. *Curr Opin Neurobiol*.
   doi:10.1016/S0959-4388(00)00147-1
- 265 Zhang W. 2018. Global pesticide use: Profile, trend, cost / benefit and more. *Proc Int Acad Ecol* 266 *Environ Sci* 8:1–27.

967