Plant respiration: controlled by photosynthesis or biomass?

List of Authors: Alessio Collalti[1,2], Mark G. Tjoelker[3], Günter Hoch[4], Annikki Mäkelä[5], Gabriele Guidolotti[6], Mary Heskel[7], Giai Petit[8], Michael G. Ryan[9,10], Giorgio Matteucci[1], Giovanna Battipaglia[11], I. Colin Prentice[12,13,14*]

Affiliations:

1. Institute for Agriculture and Forestry Systems in the Mediterranean, National Research Council of Italy (CNR-ISAFOM), 87036, Rende (CS), Italy (alessio.collalti@cnr.it, giorgio.matteucci@cnr.it)

2. Division on Impacts on Agriculture, Forests and Ecosystem Services, Foundation Euro-Mediterranean Centre on Climate Change (CMCC-IAFES), 01100, Viterbo, Italy

3. Hawkesbury Institute for the Environment, Western Sydney University, Penrith, NSW 2751, Australia. (M.Tjoelker@westernsydney.edu.au)

4. Department of Environmental Sciences – Botany, University of Basel, Schönbeinstrasse 6, Basel 4056, Switzerland. (guenter.hoch@unibas.ch)

5. Institute for Atmospheric and Earth System Research (INAR), Faculty of Science & and Faculty of Agriculture and Forestry, P.O. Box 27 (Latokartanonkaari 7) FI-00014 University of Helsinki, Finland. (annikki.makela@helsinki.fi)

6. Institute of Research on Terrestrial Ecosystem, National Research Council of Italy (CNR-IRET), 00015, Monterotondo Scalo (RM), Italy. (gabriele.guidolotti@cnr.it)

7. Department of Biology, Macalester College, Saint Paul, MN USA 55105. (mheskel@gmail.com)

8. Department of Land, Environment, Agriculture and Forestry, University of Padova, 35020 Legnaro (PD), Italy. (giai.petit@unipd.it)

9. Natural Resource Ecology Laboratory, Colorado State University, Fort Collins, CO 80523-

1499, USA. (mryan@rams.colostate.edu)

10. USDA Forest Service, Rocky Mountain Experiment Station, Fort Collins, CO 80526, USA

11. Department of Environmental, Biological and Pharmaceutical Sciences and Technologies,

University   of   Campania   «L.   Vanvitelli»,   81100,   Caserta,   Italy.

(giovanna.battipaglia@unicampania.it)

12. AXA Chair of Biosphere and Climate Impacts, Imperial College London, Department of

Life   Sciences,   Silwood   Park   Campus,   Buckhurst   Road,   Ascot   SL5   7PY,   UK

(c.prentice@imperial.ac.uk)

13. Department of Biological Sciences, Macquarie University, North Ryde, NSW 2109,

Australia

14. Ministry of Education Key Laboratory for Earth System Modeling, Department of Earth

System Science, Tsinghua University, Beijing 100084, China

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*Contact information: Imperial College London, Department of Life Sciences, Silwood Park

Campus, Buckhurst Road, Ascot SL5 7PY, UK. Tel. +44 207 594 2354, email:

c.prentice@imperial.ac.uk

Abstract

Two simplifying hypotheses have been proposed for whole-plant respiration. One links

respiration to photosynthesis; the other to biomass. Using a first-principles carbon balance model

with a prescribed live woody biomass turnover, we show that if turnover is fast, the

accumulation of respiring biomass is low and respiration depends primarily on photosynthesis;

while if turnover is slow, the accumulation of respiring biomass is high and respiration depends

primarily on biomass. But the first scenario is inconsistent with evidence for substantial

carryover of fixed carbon between years, while the second implies far too great an increase in

respiration during stand development – leading to depleted carbohydrate reserves and an

unrealistically high mortality risk. These two mutually incompatible hypotheses are thus both

incorrect. Respiration is *not* linearly related either to photosynthesis or to biomass, but rather it is

more strongly controlled by recent photosynthates (and reserve availability) than by total

biomass.

INTRODUCTION

The amount of carbon that accumulates in actively growing stands of vegetation depends on the balance of photosynthesis (gross primary production, P) and whole-plant respiration (R). The difference between these fluxes is net primary production (P_n). Most annual P_n is allocated to structural growth (G), but some is stored as non-structural carbohydrates (NSC, mostly starch and sugars), some is released back to the atmosphere in the form of biogenic volatile organic compounds (BVOCs), and some is exuded to the rhizosphere (Chapin *et al.* 2006). The fraction of P that accumulates in biomass, and the fraction that returns to the atmosphere through plant metabolism, are crucial quantities that determine the sign and magnitude of the global climate-carbon feedback – which remains one of the greatest sources of uncertainty in the global carbon cycle (Friedlingstein *et al.* 2014). But despite many ecophysiological studies aiming to understand P_n and R dynamics during stand development a general understanding is still lacking.

Some authors have hypothesized a constant P_n:P (carbon use efficiency, equivalent to $1 - (R{:}P)$) ratio, with R tightly constrained by P irrespective of biomass, climate, tree species and stand age (e.g. Gifford 2003; Van Oijen *et al.* 2010). Waring *et al.* (1998, W98 hereafter) indicated a universal P_n:P of ~ 0.5. Since R depends on the matter produced by photosynthesis, Gifford (2003) suggested that these two processes must be balanced over the longer term – making R proportional to P, consistent with W98. He argued that prescribing P_n (or R) as a constant fraction of P could be a simpler, and potentially more accurate, alternative to explicit, process-based modelling of R. A number of land vegetation models (reviewed in Collalti & Prentice 2019) have adopted this simplification.

An alternative hypothesis, grounded in metabolic scaling theory, suggests that R should scale with biomass following a power law, $Y = a\,X^b$ (West *et al.* 1999). According to some studies (e.g. Reich *et al.* 2006, R06 hereafter), R (Y) scales isometrically ($b \sim 1$) with whole-plant carbon (C) or nitrogen (N) contents (X), and this scaling is similar within and among different species, and irrespective of environmental and climatic conditions – which might influence the

normalization constant (a), but not the exponent (b). Isometric scaling of R with biomass was

assumed in the traditional view of forest dynamics set out e.g. by Kira & Shidei (1967) and

Odum (1969). In the absence of major disturbances, if R increases in parallel with biomass, then

P_n necessarily declines – because P cannot increase indefinitely, but rather stabilizes at canopy

closure. Mori *et al.* (2010) however indicated that biomass and R are isometrically related only in

young trees, tending towards $b \sim {}^3/_4$ in mature trees. A general value of ${}^3/_5$ has also been

proposed (Michaletz *et al.* 2014). But however it is interpreted, this scaling hypothesis implies

that R depends on biomass, and is related to P only to the extent that P and biomass vary

together.

Many terrestrial biosphere models simulate plants respiration more explicitly, thus partially

uncoupling R from P. The most widely used (and observationally supported) method, also

adopted here, divides R into growth (R_G) and maintenance (R_M) components (McCree 1970). R_G

is considered to be a fixed fraction of new tissue growth, independent of temperature, the

fraction varying only with the cost of building the compounds constituting the new tissue

(Penning de Vries 1972). Temperature, substrate availability and the demand for respiratory

products are considered to control R_M (Cannell & Thornley 2000). Several studies have

investigated the effects of short- and long-term changes in temperature on R_M, mostly at the leaf

level (e.g. Heskel *et al.* 2016; Huntingford *et al.* 2017). The nature of the temperature responses

and the acclimation of R_M are important and much-discussed issues, but they are not considered

further here. In contrast, the effects on respiration of woody biomass (the substrate), its

accumulation, and the transition rate of respiring sapwood into non-respiring heartwood, have

received relatively little attention (Tjoelker *et al.* 1999; Kuptz *et al.* 2011). These latter processes

are the focus of this review.

The fixed-ratio hypothesis of W98 and the scaling hypothesis of R06 could both be used – at

least in principle, across the twenty orders of magnitude variation in plant mass – to estimate R

and P_n without the need for explicit process-based modelling of R (McMurtrie *et al.* 2008; Price

et al. 2010). However, they yield quite different results, and both hypotheses have been subject

to criticism (e.g. Medlyn & Dewar 1999; Mäkelä & Valentine 2001; Kozłowski & Konarzewski

2005; O'Connor *et al*. 2007; Keith *et al*. 2010; Agutter & Tuszynski 2011; Price *et al.* 2012;

Collalti *et al*. 2018; Collalti & Prentice 2019). To our knowledge, there has been no previous

attempt to compare these hypotheses directly, and their consequences for forest carbon balance

during stand development, in a consistent modelling framework. We attempt to fill this gap by

providing illustrative simulations of the long term trajectories of R, P_n and $P_n{:}P$. The simulations

are based on the first principles of mass balance implemented in a process-based,

ecophysiological model that has been tested against detailed time-series observations in a real

forest stand. Our aim is to provide insight into the consequences of different assumptions about

live woody biomass turnover (live woody biomass is the metabolically active fraction of

sapwood: see Supporting Information). We show how different (extreme) assumptions about the

live woody turnover map on to the alternative hypotheses, while seeking an answer to the pivotal

question: is R a function of photosynthesis alone (W98), or of biomass alone (R06)? Or are

neither correct?.

METHODS

Theoretical framework

A general equation describing autotrophic respiration (R) is:

$$R = P - P_n = P - (G + G_{\mathrm{R}}) \tag{1}$$

139   where P and P_n are gross and net primary production, G is structural and litter biomass

140   production and G_{R} is the flux to NSC reserves and secondary compounds including, exudates

141   and BVOCs (all in g C ground area^{-1} time^{-1}). If R is further decomposed into growth (R_{G}) and

142   maintenance (R_{M}) respiration (McCree 1970), then:

143
$$R = R_{\mathrm{G}} + R_{\mathrm{M}} = g_{\mathrm{R}}\mathrm{G} + m_{\mathrm{R}}W_{live} \qquad (2)$$

144   where g_{R} and m_{R} are the growth and maintenance respiration coefficients (i.e. respiratory CO_2

145   released per unit biomass produced by growth and by the maintenance of the existing biomass:

146   both, per unit time and unit mass), and W_{live} is living biomass (Amthor 2000). W_{live} can be

147   broken down further:

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$$W_{\mathrm{live}} = W_{\mathrm{live\_woody}} + W_{\mathrm{green}} \qquad (3)$$

where $W_{\mathrm{live\_woody}}$ and W_{green} are the biomass of live woody pools (living cells in stem,

branches and coarse roots) and non-woody tissues (leaves and fine roots), respectively. Because

plant tissues require N as a component of the enzymes that sustain metabolic processes, living

biomass is often expressed in nitrogen units, g N ground area^{-1} (Cannell & Thornley 2000),

while respiration is expressed in carbon units. Then m_{R} is in units of g C g N^{-1} time^{-1}. Temporal

changes in $W_{\mathrm{live\_woody}}$ can be summarized by first-order biochemical kinetics:

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$$\underbrace{\frac{\mathrm{d}W_{\mathrm{live\_woody}}}{\mathrm{d}t}}_{\text{relative change}} = \underbrace{\varphi \cdot G_{\mathrm{live\_woody}}}_{\text{incoming flux}} - \underbrace{W_{\mathrm{live\_woody}} \cdot \tau}_{\text{outgoing flux}} \qquad (4)$$

where $G_{\mathrm{live\_woody}}$ is the part of G allocated to live woody, φ converts carbon to nitrogen

content (g N g C^{-1}), and τ is the live woody turnover rate per unit time (t). A similar expression

can be written for W_{green}. The first term on the right-hand side of equation (4) represents the

"incoming" flux of new living cells; while second term represents the "outgoing" flux of living

cells that die and become metabolically inactive. But while W_{green} may be only a small fraction of

total forest biomass, not changing much after canopy closure, $W_{\mathrm{live\_woody}}$ (as also total W)

becomes large during forest development and is potentially a strong driver of R (Reich *et al.*

2008). Note that this general approach including a turnover rate parameter (τ) is equally valid for

any mass-, area- or volume-based analyses (Thornley $\&$ Cannell 2000).

Setting $\tau = 1$ year^{-1} in equation (4) would imply a tight coupling between the previous year's

growth and the current year's respiration flux – as suggested by Gifford (2003) – and yields a

close approximation to the W98 assumption of a fixed ratio between P_n and P, thus cancelling

any effect of biomass accumulation. The implication of a one-year-lag between carbon fixation

and respiration in woody compounds is consistent with the findings of Amthor (2000), Kagawa

et al. (2006a, b), Gough *et al*. (2008, 2009) and Richardson *et al*. (2013, 2015) of a physiological

asynchrony by about one year between P.

Alternatively, setting $\tau = 0.1$ year^{-1} would imply that most new sapwood cells live for many

years, and would closely approximate the R06 assumption of proportionality between R and

biomass. Thus, the amount of respiring biomass is regulated by the amount of substrate that is

produced each year, forming new sapwood, versus the amount that is converted into non-living

tissues and no longer involved in metabolism; the balance of these processes being controlled by

τ (see proofs-of-concept in Fig. 1a and b, and Table 1 for elaboration).

Because carbon supply (photosynthesis) and carbon metabolic demand (respiration) are not

necessarily synchronized, the model assumes that temporary carbon imbalances between P and R

(implying $P_n < 0$, Roxburgh *et al.* 2005) are met by the remobilization or recycling of NSC

stored during previous year(s) – so long as the NSC pool is not completely emptied (the carbon

starvation hypothesis; McDowell *et al.* 2008). A full description of the modelled NSC dynamics

is provided in Box 1.

Simulation set-up

The logic described above was implemented in the process-based forest growth model 3D-

CMCC-CNR and applied, as a case study, to a temperate deciduous forest. Additional model

description can be found in Supporting Information, Collalti *et al.* (2014, 2016, 2018), and

Marconi *et al.* (2017).

Very limited data are available on the turnover rate τ of live cells in sapwood, which is often

either guessed or inferred by model calibration (e.g. White *et al.* 2000). We carried out ten

simulations with τ varied in arbitrary 0.1 yr^{-1} steps, from $\tau = 1$ yr^{-1} (100% of turnover, all the

previous year live cells of sapwood becomes non-respiring heartwood in the current year), down

to $\tau = 0.1$ yr^{-1} (only 10% of the previous year's live cells of sapwood biomass dies). Thus, we

started with the largest prior distribution for τ, assuming that values outside this range are not

functionally possible (Table 1). We are unaware of any studies reporting changes in τ with age or

biomass; we have therefore assumed that τ is constant in time.

The standard configuration of 3D-CMCC-CNR assigns $\tau = 0.7$ yr^{-1} (Collalti *et al.* 2019) and

this same value has been used by several authors in various modelling contexts (e.g. Bond-

Lamberty *et al.* 2005; Tatarinov *& Cenciala 2006). Other models have applied different values

(see Box 2). Zaehle *et al.* (2005), Poulter *et al.* (2010) and Pappas *et al.* (2013) found that τ is a

critical parameter for both LPJ-DGVM and LPJ-GUESS. We are not aware of similar sensitivity

analyses for other models. Leaf and fine root turnover rates are assumed here to be 1 yr^{-1},

appropriately for deciduous trees (Pietsch *et al.* 2005). The model parameter for maximum age (a

parameter that controls, among other things, an age-effect on leaf conductance: Kirschbaum,

2000; Smith *et al.* 2001) was set arbitrarily large, to avoid building in prior assumptions about

age effects. Age- and size-effects are therefore considered synonymous (Mencuccini *et al.* 2005).

A stochastic background whole-tree mortality rate (1% of trees removed each year) was retained

and included in equation (4) to ensure a realistic self-thinning (Smith *et al.* 2001; Kirschbaum,

2005). All other parameters were left unchanged from the standard model configuration.

Test site and model run

The model was applied to simulate the development of an even-aged stand of European beech

(*Fagus sylvatica* L.; Sorø, Denmark; Wu *et al.* 2013, Reyer *et al.* 2019:

http://doi.org/10.5880/PIK.2019.008) in daily time steps from 1950 to 2100. The reasons for

choosing this stand are: (a) the extensive literature on European beech, allowing key parameter

values to be assigned with confidence; (b) the exceptional quantity of data available at the Sorø

site for initializing and evaluating the model; and (c) because the trees are deciduous we can

assume a complete annual turnover of leaves and fine roots and, therefore, more easily

disentangle the contributions of W_{green} and W_{live_woody}. Deciduous species are also expected to

have greater within-season variability in $P_n{:}P$, and greater asynchrony in carbon supply and

demand, than evergreen species (Dietze *et al.* 2014; Martínez-Vilalta *et al.* 2016). However, the

model results are based on general principles and expected to apply more generally.

We simulated forest development up to 2100, consistent with the common economic rotation

length for this species in northern Europe. After canopy closure, modelled leaf area index (LAI)

and the relative amounts of leaf and fine-root biomass became stable or even slightly decrease,

as is usually observed (Yang *et al.* 2011, 2016). Therefore, changes in modeled R and its

components R_M and R_G, could be attributed to changes in the total amount of living woody

biomass and the costs of its maintenance.

In 1950 the stand was aged 30 years with an average tree diameter at breast height of ~ 6 cm

and a density of 1326 trees ha^{-1}. Model state variables were initialized using species-specific

functional and allometric-relationships from the literature, and previous model applications at

this site (Collalti *et al.* 2016, 2018; Marconi *et al.* 2017). Model sensitivity to parameter values

and their uncertainties have been assessed in depth in a previous work (see Collalti *et al.* 2019,

especially their Fig. 2 and Table 3). Management, in the form of thinning, occurred at the site

only up to 2014. After that year, only stochastic mortality was accounted for in the model. Live

wood was initialized at 15% of sapwood biomass (as the fraction of current year sapwood:

Pietsch *et al.* 2005) and assigned a C:N ratio of 48 g C g N^{-1}, not changing with increasing

biomass (Ceschia *et al.* 2002; Damesin 2003). The minimum concentration of NSC was assumed

to be 11% of sapwood dry mass (Hoch *et al.* 2003; Genet *et al.* 2010; Martínez-Vilalta *et al.*

2016). Daily meteorological forcing variables were obtained as historical ensemble means from

five Earth System Models up to 2005 provided by the Inter-Sectoral Impact Model

Intercomparison Project (ISI-MIP, Warszawski *et al.* 2014). Data for the period 1995–2005 were

then randomly repeated up to 2100. Additional simplifying assumptions were made in order to

focus specifically on the effects of increases in tree size, as follows: no disturbances (whether

herbivory or management) after 2014; no effect of changes in soil N availability; and,

importantly, to avoid possible confounding of temperature effects on R_M with other warming

effects, a stable (1995–2005) climate and atmospheric CO_2 concentration (~ 380 μmol mol^{-1}).

Exports of carbon to exudates and BVOCs were neglected.

RESULTS

Data-model agreement

The standard model configuration satisfactorily reproduced P, R, P_n and the ratio P_n:P when

compared to independent, site-level, carbon balance data (Wu *et al.* 2013) for the period 2006–

2010 (Fig. 2, Table 2), corresponding to a stand age of ~ 85–90 yrs. P was in agreement with

eddy covariance data, while R was slightly underestimated compared to values in Wu *et al.*

Consequently, the model overestimated the average P_n:P ratio by 14% compared to Wu *et al.*

However, Wu *et al.* argued that the values of R they obtained (by subtracting modelled

heterotrophic from measured ecosystem respiration) may have been somewhat overestimated,

given also the ample standard deviation (±143 g C m^{-2} yr^{-1}). The model results are otherwise in

good agreement with Wu *et al.* for woody carbon stocks (both above- and below-ground), annual

wood production (the sum of carbon allocated to stems, branches and coarse roots), and annual

above- and below-ground litter production (the sum of carbon allocated to leaves and fine roots)

(Table 2). Modelled respiration of the woody compartments, leaf and total (above- and below-

ground) respiration, and NSC pool and fluxes, are all compatible with values reported by

previous investigations, and within the ranges of total, wood and leaf respiration, and P_n:P ratios

reported for European beech (e.g. Barbaroux & Brèda 2002; Barbaroux *et al.* 2002; Knohl *et al.*

2003; Granier *et al.* 2008; Davi *et al.* 2009; Genet *et al.* 2010; Guidolotti *et al.* 2013). A model

validation forced by measured climate at this site is also described in previous papers (Collalti *et*

al. 2016; 2018; Marconi *et al.* 2017).

The effect of varying τ

The simulations produced a broad spectrum of diverging trajectories, ranging from an

approximately constant P_n:P ratio (for large τ) to a constantly decreasing P_n:P ratio (for small τ)

(Fig. 1a). For $\tau = 1$ yr^{-1}, P_n:P stays close to 0.5. For $\tau \leq 0.2$ yr^{-1} P_n:P eventually falls below the

lower limit of commonly observed values (0.22; Collalti & Prentice 2019) and the physiological

limit of 0.2 proposed by Amthor (2000). Figure 2 also shows the effects of varying τ in

determining different trajectories for P_n (Fig. 2b) and R (Fig. 2c) and consequent differences in

the partitioning between R_M and R_G (Fig. S2) with modelled R, at the end of simulations, ranging

from ~ 800 g C m^{-2} yr^{-1}, giving P_n ~ 900 g C m^{-2} yr^{-1} and P ~ 1700 g C m^{-2} yr^{-1} (Fig. S1 and

for NSC flux Fig. S3b) consistent with a steady-state between R or P_n and P, to two cases ($\tau =$

0.1, 0.2 yr^{-1}) in which trees die from starvation.

The model did not generate any consistent power-law relationship between R and biomass

either for b ~ 1 (i.e. R06), or for ~ $3/4$ (Mori *et al.* 2010), or for ~ $2/3$ (Makarieva *et al.* 2005),

and for ~ $3/5$ (Michaletz *et al.* 2014) (Table S1). The simulations indicated b ~ 1 initially,

shifting with increasing tree size to b ~ 0.74 for $\tau = 0.1$ yr^{-1} ($R^2 = 0.99$, $n = 117$) or 0.19 for $\tau = 1$

yr^{-1} ($R^2 = 0.84$, $n = 150$; 'n' corresponds to years of simulation). For the relation between R and

whole-plant N, again the simulations indicated b ~ 1 initially, shifting to b ~ 0.82 for $\tau = 0.1$ yr^{-1}

($R^2 = 0.99$, $n = 117$) or 0.27 for $\tau = 1$ yr^{-1} ($R^2 = 0.82$, $n = 150$) (Figs. 3a and 3c). Note that the

highest b values corresponded to simulations which ended because the trees died.

DISCUSSION

R is not entirely determined by P

A constant $P_n \colon P$ ratio, obtained on the model by setting $\tau = 1$ yr^{-1}, conflicts with observations

that show a substantially slower turnover rate of living cells. Parenchyma cells within secondary

xylem are very often more than a year old, and can be up to 200 years old (Spicer & Holbrook

2007). The constant ratio hypothesis is also contrary to the evidence that much of the recently-

fixed assimilate pool is at first stored as reserves, and only later used for metabolism or growth

(Schiestl-Aalto *et al.* 2015, 2019). Indeed, some report decoupling between growth (which

would imply some CO_2 released for both R_G and subsequently R_M) and photosynthesis – with

growth ceasing long before photosynthesis – because of their different sensitivities to

environmental drivers. Kagawa *et al*. (2006a) reported that up to 43%, and according to Gough

et al. (2009) up to 66%, of annual photosynthetates are used during the year(s) after they have

been fixed. Gaudinski *et al*. (2009), Malhi (2012), and Delpierre *et al*. (2016) all found negative

correlations between annual carbon inflows and above- or below-ground wood growth. Chen *et*

al. (2013) found that R does not scale isometrically with P. Some suggested that R_G could be

supplied exclusively by recent photosynthates while R_M by the previously stored ones (Lötscher

et al. 2004). Along the same lines, Maier *et al*. (2010), Kuptz *et al*. (2011) and Lynch *et al*.

(2013) found that both R_G and R_M are not completely satisfied by recent assimilates, and that

some current R_M can be derived from woody tissues constructed in previous years. Litton *et al*.

(2007) and Yang *et al*. (2016) both found low correlations between respiration and the annual

production of woody compounds. Many studies have also reported little variation in the CO_2

efflux from sapwood in relation to tree-ring age, despite a stepwise decrease in the fraction of

living cells towards the centre of the stems (e.g. Ceschia *et al.* 2002; Spicer *&* Holbrook 2007;

Pallardy 2010). These various observations imply that some carbon is fixed one year and used

for the tree's own growth and metabolism in the next or subsequent years, and that the inner

sapwood contains a population of living cells formed in previous years.

These observations are all incompatible with the hypothesis of a tight coupling of R and P

(alone), and with model results obtained by assuming complete turnover of live cells in sapwood

during a single year.

R is not entirely determined by biomass

On the other side of the ledger, model simulations indicate that low τ values (≤ 0.2 yr^{-1}) can

lead to excessively high respiration burdens, impossibly low P_n:P ratios (< 0.2), and ultimately

carbon starvation when all NSC is consumed and whole-tree R_M or growth can no longer be

sustained (Fig. 2). This model result is quantitatively dependent on the values adopted for C:N

ratio and the minimum NSC-pool which increases with tree size, but is consistent with the idea

that P_n:P ratios ≤ 0.2 are not physiologically sustainable (Amthor, 2000). Amthor described the

0.65–0.2 bounds as those reflecting maximum-growth with minimum-maintenance expenditure

(0.65) and minimum-growth with maximum physiologically-sustainable maintenance costs (0.2).

Such a minimum P_n:P value agrees also with Keith *et al.*'s (2010) considerations that trees

require always some annual biomass (litter) production in order to survive. With such low τ,

simulated woody R_M:$R > 90\%$, a value much higher than those (56–65%) reported by Amthor &

Baldocchi (2001) (Supporting Information Fig. S2), and this initiates a spiral of decline, whereby

neither P nor NSC drawdown are sufficient to avoid a long term carbon imbalance (Supporting

Information, Figs. S3a and b; Wiley *et al.* 2017; Weber *et al.* 2018).

Slightly higher τ values (from 0.3 to 0.5 yr^{-1}) were found to limit woody biomass increase

because of high NSC demand, leading to a shift in the allocation of assimilates to refill NSC at

the expense of growth, and P_n:P values close to 0.2 (Fig. 2a). Values of $\tau > 0.5$ did not show

such behaviour and allowed structural and non-structural compounds to accumulate in parallel,

while P_n gradually declined and eventually levelled off. This scenario allows structural biomass

accumulation to continue even in older trees, as observed (Stephenson *et al*. 2014).

Scaling relationships

We simulated forest dynamics from juvenile up to very large, mature trees. R06's results

supporting isometric scaling were based on measurements of seedlings and 6- to 25-year-old

trees with, presumably, very little heartwood. Mori *et al*. (2010) analysed single trees located

outside forests, so neither of these studies is strictly comparable with ours. Some other studies

have suggested that the scaling slope of approximately 1 for whole-plant mass may be valid early

in stand development, but that the exponent may eventually become smaller than $^3/_4$, a

phenomenon that has been termed 'ontogenic drift' (Makarieva *et al.* 2008). Piao *et al*. (2010) in

a global analysis also found a low correlation, and a low scaling exponent, between R and whole-

plant biomass ($b \sim 0.21$, corresponding to $\tau \sim 0.9$ in our simulations). Piao *et al*. (2010) argued

that, for large mature trees, an increasing fraction of woody C and N biomass is composed of

metabolically inactive heartwood, and concluded that a linear-relationship between respiration

and whole-plant biomass should not be expected, while a curvilinear-relationship at the small

end of the size spectrum seemed more appropriate (Kozłowski & Konarzewski 2005). Li *et al*.

(2005) also found no evidence for an isometric or $^3/_4$ power scaling relationship, indicating

instead a range between 1.14 and 0.40, decreasing with plant size. The only approximately

isometric relationship we found in our simulations – across all τ used – was between R and the

living components of biomass C (b in the range of 0.8 to 1, with R^2 always > 0.93) and biomass

N ($b \sim 0.9$, with R^2 always > 0.97) (Figs. 3b and 3d; Makarieva *et al.* 2005; Kerkhoff & Enquist

2006; Gruber *et al.* 2009).

Overall, none of these findings are compatible with a tight isometric-relationship of R to

whole-plant C (or N) biomass and model results lend mechanistic support to this view.

R is determined by P, biomass and reserve availability

Plants store large amounts of non-structural carbohydrates (e.g. up to rebuild the whole leaf

canopy one to more than four times, Hoch *et al*. 2003) and, when need, plants can actively buffer

the asynchronies between carbon demand (i.e. R and G) and supply (i.e. P) by tapping the pool

of non-structural carbon resources (see Fig. S3 in Supporting Information for NSC trends).

Several lines of evidence and a growing body of literature support the view of a (at least) quasi-

active sink of NSC, that is, NSC competes with growth, while it controls R (and including other

nonmetabolic functions, see Hartmann *&* Trumbore 2016), in a feed-forward circular

mechanism. Schuur *&* Trumbore (2006), Carbone *et al*. (2007) and Lynch *et al*. (2013) all

reported that plant-respired CO_2 is a mixture of old and new assimilated carbohydrates.

Likewise, Vargas *et al*. (2009), Carbone *et al*. (2013), Richardson *et al*. (2013); Muhr *et al*.

(2013, 2016), and Solly *et al*. (2018) found that NSC up to 17 year old, remobilized from

parenchyma cells, can be used for growth or metabolism.

Aubrey *&* Teskey (2018) found that carbon-starved roots and whole-tree saplings die before

complete NSC depletion, but the threshold NSC level at which this happens remains largely

unknown or uncertain for most species. These thresholds are likely to vary among tissues (Weber

et al. 2018), species (Hoch *et al*. 2003), phenotypes, habit and wood anatomy (Dietze *et al.*

2014), and to increase with tree size (Sala *et al.* 2012). Others have reported that trees cannot

draw down NSC to zero because of limitations in carbohydrate remobilization and/or transport

(Hartmann *&* Trumbore 2016; Wiley *et al.* 2017). A minimum NSC level, which has been found

to proportionally increase with size biomass, may also be required to maintain a safety margin

and a proper internal functioning of trees (e.g. osmoregulation), regardless of whether growth is

limited by carbon supply (Woodruff *&* Meinzer 2011; Sala *et al*. 2011, 2012; Martínez-Vilalta *et*

al. 2016; Huang *et al*. 2019). Genet *et al*. (2010) found shifts during ontogeny in carbon

allocation from biomass growth to reserves regardless of seasonal fluctuations, habitat and

climate and Palacio *et al*. (2012) found that in repeatedly defoliated trees for 11 yrs, and left to

recover for other 6 yrs, reduced growth but similar stem NSC concentration when compared to

other undefoliated trees.

It has further been suggested that a considerable fraction of NSC (mostly starch) in the inner-

part of wood may become compartmentalized and sequestered away from sites of phloem

loading, and thus no longer accessible for both tissue-growth and respiration (Sala *et al.* 2012).

Root exudation to mycorrhizal fungi and secondary metabolites (not accounted here) could

accelerate NSC depletion (Pringle 2016) and potentially a risk of carbon starvation even for

values of τ well above 0.2.

Overall, asynchrony between (photosynthetic) source and (utilization) sink implies some

degree of uncoupling of R, and consequently P_n (and growth), from P and biomass. Carbon

demand for metabolism and growth can be mediated by tapping the pool of NSC but only to the

extent and to the amount that it is accessible and (surely) usable by plants.

Implications

It has been suggested that the observed decline of P_n during stand development cannot be

exclusively caused by increasing respiration costs with tree size (Tang *et al.* 2014). The idea,

implicit in the growth and maintenance respiration paradigm – that R_M is a 'tax' that must be paid

first and which ultimately controls growth – has also been criticized (Gifford 2003). While this

paradigm has some weaknesses (Thornley 2011; Box 2), and is not changed much over the last 50

years despite some theoretical and experimental refinements, no other general and similarly

promising mechanistic approach to the modelling of whole-plant respiration has been proposed.

Any successful modelling approach for plant respiration must allow plants to steer a middle course

between tight coupling to photosynthesis (inconsistent with many observations) and dependence on

ever-increasing biomass (risking carbon starvation and death) and the buffering capacity of reserves

during carbon imbalances. It seems likely that plants strive to keep an optimal quantity of living

cells that can effectively be sustained by photosynthesis or, when necessary, by drawing on NSC

and down regulating allocation to (nonphotosynthetic) metabolically active pools as to minimize

maintenance costs (Makarieva *et al.* 2008). This suggests active control on carbon use efficiency

and NSC usage by plants and thus on the turnover of the living cells.

Other processes, including hydraulic and nutrient limitations, may be in play (Carey *et al.*

2001; Xu *et al.* 2012). Malhi *et al.* (2015) argued for a link between high whole-plant mortality

rates and high forest productivity as ecophysiological strategies that favour rapid growth may also

result in fast turnover of trees. However, Spicer *&* Holbrook (2007) noted that metabolic activity

does not decline with cell age and Mencuccini *et al.* (2005) that effects of 'age as such' (including

cellular senescence and apoptosis) are probably not responsible for declining P, but are linked to the

functional and structural consequences of increasing plant size.

To reduce the large uncertainty surrounding this issue, it will be necessary on the one hand to

use models that explicitly account for the turnover of biomass and the reserves; and on the other

hand, to carry out experimental and field measurements of the dynamics of living cells in wood and

reserves availability. These processes have a direct bearing on the stocks and fluxes that drive the

carbon balance of forests.

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REFERENCES

Agutter P.S., Tuszynski J.A. (2011). Analytic theories of allometric scaling, *The Journal of*
Experimental Biology, 214: 1055–1062.
Amthor J. (2000). The McCree-de Wit-Penning de Vries-Thornley Respiration Paradigms: 30 Years
Later. *Annals of Botany*, 86: 1–20.
Amthor J., Baldocchi D. (2001). Terrestrial higher plant respiration and net primary production. In
Terrestrial Global Productivity, ed. J. Roy, B. Sauger, and H. A. Mooney, 33–59. Sand Diego,
CA: Academic Press
Arora V., Boer G. (2005). A parameterization for leaf phenology in the terrestrial ecosystem
component of climate models. *Global Change Biology*, 11 (1): 33–59.
Aubrey D.P., Teskey R.O. (2018). Stored carbohydrates can maintain root respiration for extended
periods. *New Phytologist*, 218: 142–152.
Barbaroux C., Brèda N. (2002). Contrasting distribution and seasonal dynamics of carbohydrate
reserve in stem wood of adult ring-porous sessile oak and diffuse-porous beech trees. *Tree*
Physiology, 22: 1201–1210.
Barbaroux C., Brèda E., Dufrêne E. (2002). Distribution of above-ground and below-ground
carbohydrate reserves in adult trees of two contrasting broad-leaved (*Quercus cerris* and *Fagus*
sylvatica). *New Phytologist*, 157: 605–615.
Bonan G.B., Levis S., *et al*. (2003). A dynamic global vegetation model for use with climate
models: concepts and description of simulated vegetation dynamics. *Global Change Biology*, 9:
1543–1566
Bond-Lamberty B., Gower S., *et al*. (2005). Reimplementation of the Biome-BGC model to
simulate successional change. *Tree Physiology*, 25: 413–424.
Campioli M, Verbeeck H, *et al*. (2013). Can decision rules carbon allocation for years with
contrasting and extreme weather conditions? A case study for three temperate beech forests.
Ecological Modelling, 263: 42–55.
Cannell M., Thornley J. (2000). Modelling the Components of Plant Respiration: Some Guiding
Principles. *Annals of Botany*, 85: 45–54.
Carbone M.S., Czimizik CI, McDufee K.E., Trumbore S.E. (2007). Allocation and residence time
of photosynthetic products in a boreal forest using a low-level ^{14}C pulse-chase labelling
technique. *Global Change Biology*, 13: 466–477.
Carbone M.S., Czimczik C.I., Keenan T.F., *et al*. (2013). Age, allocation and availability of non-
structural carbon in mature red maple trees. *New Phytologist,* 200: 1145–1155.
Carey E.V., Sala A., Keane R., *et al*. (2001). Are old forests underestimated as global carbon sinks?
Global Change Biology, 7: 339–344
Chapin III F.S., Woodwell G.M., *et al*. (2006). Reconciling Carbon-cycle Concepts, Terminology,
and Methods. *Ecosystems*, 9(7): 1041–1050
Chen G., *et al*. (2013). Allocation of gross primary production in forest ecosystems: allometric
constraints and environmental responses. *New Phytologist*, 200: 1176–1186.
Ceschia E., Damesin C., *et al*. (2002). Spatial and seasonal variations in stem respiration of beech
trees (*Fagus sylvatica*). *Annals of Forest Science*, 59: 801–812
Collalti A., Marconi S., Ibrom A., *et al*. (2016). Validation of 3D-CMCC Forest Ecosystem Model
(v.5.1) against eddy covariance data for 10 European forest sites. *Geoscientific Model*
Development, 9: 479–504.

Collalti A., Trotta C., Keenan T.F., *et al*. (2018). Thinning can reduce losses in carbon use efficiency and carbon stocks in managed forests under warmer climate. *Journal of Advances in Modeling Earth System*, doi: 10.1029/2018MS001275.

Collalti A., Thornton P.E., Cescatti A., *et al*. (2019). The sensitivity of the forest carbon budget shifts across processes along with stand development and climate change. *Ecological Applications,* 29(2): 1–18, https://doi.org/10.1002/eap.1837.

Collalti A., Prentice I.C. (2019). Is NPP proportional to GPP? Waring's hypothesis twenty years on. *Tree Physiology,* https://doi.org/10.1093/treephys/tpz034.

Damesin C., Ceschia E., *et al*. (2002). Stem and branch respiration of beech: from tree measurements to estimations at the stand level. *New Phytologist*, 153: 159–172

Damesin C. (2003). Respiration and photosynthesis characteristics of current-year stems of *Fagus sylvatica*: from the seasonal pattern to an annual balance. *New Phytologist*, doi: 10.1046/j.1469-8137.2003.00756.x

Davi H., Barbaroux C., Francois C., Dufrêne E. (2009). The fundamental role of reserves and hydraulic constraints in predicting LAI and carbon allocation in forests. *Agricultural and Forest Meteorology*, 149: 349–361.

Delpierre N., *et al*. (2016). Wood phenology, not carbon input, controls the interannual variability of wood growth in a temperate oak forest. *New Phytologist*, 210: 459–470

DeLucia E., Drake J., Thomas R., *et al*. (2007). Forest carbon use efficiency: Is respiration a constant fraction of gross primary production? *Global Change Biology*, 13, 1157–1167.

Dewar R.C., Medlyn B.E., McMurtrie R.E.. (1999). Acclimation of the respiration/photosynthesis ratio to temperature: insights from a model. *Global Change Biology*, 5: 615-622.

Dietze M., Sala A., *et al*. (2014). Nonstructural Carbon in Woody Plants. *Annual Review of Plant Biology*, 65: 667–87.

Doughty C.E., Metcalfe D., Girardin C., *et al*. (2015). Drought impact on forest carbon dynamics and fluxes in Amazonia. *Nature*, 519: 78–82.

Dufrêne E., Davi E., *et al*. (2005). Modelling carbon and water cycles in a beech forest. Part I: Model description and uncertainty analysis on modelled NEE. *Ecological Modelling*, 185: 407-436.

Enquist B.J., *et al*. (2007). Biological scaling: Does the exception prove the rule? *Nature*, doi: 10.1038/nature05548

Fatichi S., Leunzinger S., K☐rner C. (2014). Moving beyond photosynthesis: from carbon source to sink-driven vegetation modeling. *New Phytologist*, 201: 1086–1095.

Friedlingstein P., Joel G., *et al*. (1999). Toward an allocation scheme for global terrestrial carbon models. *Global Change Biology*, 5: 755–770.

Friedlingstein P., Meinshausen M., *et al*. (2014).Uncertainties in CMIP5 Climate Projections due to Carbon Cycle Feedbacks. *Journal of Climate*, 27: 511–526.

Friend A.D., Lucht W., Rademacher T.T., Keribin R., Betts, R., *et al*. (2014). Carbon residence time dominates uncertainty in terrestrial vegetation responses to future climate and atmospheric CO_2. *Proceedings of the National Academy of Sciences*, 111(9): 3280–3285.

Furze M.E., *et al*. (2019). Whole-tree nonstructural carbohydrate storage and seasonal dynamics in five temperate species. *New Phytologist*, 221: 1466–1477. doi: 10.1111/nph.15462.

Gaudinski J.B., Torn M.S., *et al*. (2009). Use of stored carbon reserves in growth of temperate tree
roots and leaf buds: analyses using radiocarbon measurements and modelling. *Global Change*
Biology, 15: 992–1014.
Genet H., Bréda N., Dufrêne E. (2010). Age-related variation in carbon allocation at tree and stand
scales in beech (*Fagus sylvatica* L.) and sessile oak (*Quercus petraea* (Matt.) Liebl.) using a
chronosequence approach. *Tree Physiology*, 30: 177–192.
Gifford R. (2003). Plant respiration in productivity models: conceptualisation, representation and
issues for global terrestrial carbon-cycle research. *Functional Plant Biology*, 30: 171–186.
Goulden M.L., *et al*. (2011). Patterns of NPP, GPP, respiration, and NEP during boreal forest
succession. *Global Change Biology*, 17(2): 855–871.
Gough C.M, Vogel C.S., *et al*. (2008). Multi-year convergence of biometric and meteorological
estimates of forest carbon storage. *Agricultural and Forest Meteorology*, 148: 158–170.
Gough C.M., Flower C.E., *et al*. (2009). Whole-ecosystem labile carbon production in a north
temperate deciduous forest. *Agricultural and Forest Meteorology*, 149: 1531–1540.
Granier A., Breda N., *et al*. (2008). Ten years of fluxes and stand growth in a young beech forest at
Hesse. *Annals of Forest Science*, 65: 704. doi:10.1051/forest:2008052.
Gruber A., Wieser G., Oberhuber W. (2009). Intra-annual dynamics of stem CO_2 efflux in relation
to cambial activity and xylem development in *Pinus cembra*. *Tree Physiology*, 26: 641–649.
Guidolotti G., *et al*. (2013). Effect of environmental variables and stand structure on ecosystem
respiration components in a Mediterranean beech forest. *Tree Physiology*, 33: 960–972
Hartmann H., Trumbore S. (2016). Understanding the roles of nonstructural carbohydrates in forest
trees – from what we can measure to what we want to know. *New Phytologist*, 211: 386–403
Heskel M.A., O'Sullivan O., *et al*. (2016). Convergence in the temperature response of leaf
respiration across biomes and plant functional types. *Proceedings of the National Academy of*
Sciences, www.pnas.org/cgi/doi/10.1073/pnas.1520282113
Hoch G., Richter A., K□rner C. (2003). Non-structural carbon compounds in temperate forest trees.
Plant, Cell and Environment, 26: 1067–1081.
Huang J., Hammerbacher A., Weinhold A., *et al*. (2019). Eyes on the future – evidence for trade-
offs between growth storage and defense in Norway spruce. *New Phytologist*, 222: 144–158
Huntingford C., Atkin O.K., *et al*. (2017). Implications of improved representations of plant
respiration in a changing climate. *Nature Communications*, doi:10.1038/s41467-017-01774-z.
Kagawa A., Sugimoto A., Maximov T.C. (2006a). Seasonal course of translocation, storage and
remobilization of ^{13}C pulse-labeled photoassimilate in natural growing *Larix gmelinii* saplings.
New Phytologist, 171: 793–804.
Kagawa A., Sugimoto A., Maximov T.C. (2006b). $^{13}CO_2$ pulse-labelling of photoassimilates reveals
carbon allocation within and between tree rings. *Plant, Cell & Environment*, 29: 1571–1584
Keith, H., Mackey, B., *et al*. (2010). Estimating carbon carrying capacity in natural forest
ecosystems across heterogeneous landscapes: addressing sources of error. *Global Change*
Biology, 16: 2971-2989.
Kerkhoff A.J., Enquist B.J. (2005). Ecosystem allometry: the scaling of nutrients stocks and
primary productivity across plant communities. *Ecology Letters*, 9: 419–427
Kira T., Shidei T. (1967) Primary production and turnover of organic matter in different forest
ecosystems of the Western Pacific. *Japanese Journal of Ecology*, 17:70–87.

Kirschbaum M.U.F. (2000). Forest growth and species distribution in a changing climate. *Tree*
Physiology, 20: 309-322.
Kirschbaum M.U.F. (2005). A model analysis of the interaction between forest age and forest
responsiveness to increasing CO_2 concentration. *Tree Physiology*, 25: 953–963.
Knohl A., Schulze E.-D., *et al*. (2003). Large carbon uptake by an unmanaged 250-year-old
deciduous forest in Central Germany. *Agricultural and Forest Meteorology*, 118: 151–167
Kozłowski T.T. (1992). Carbohydrate sources and sinks in woody plants. *Botanical Review*, 58:
107–222
Kozłowski T.T., Konarzewski M. (2005). West, Brown and Enquist's model of allometric scaling
again: the same questions remain. *Functional Ecology*, 19: 739–743
Krinner G., Viovy N., *et al*. (2005). A dynamic global vegetation model for studies of the coupled
atmosphere-biosphere system. *Global Biogeochemical Cycles*, doi:10.1029/2003GB002199.
Kuptz D., Fleischman F., *et al*. (2011). Seasonal patterns of carbon allocation to respiratory pools in
60-yr-old deciduous (*Fagus sylvatica*) and evergreen (*Picea abies*) trees assessed via whole-
tree stable carbon isotope labelling. *New Phytologist*, 191: 160–172.
Li H.T., Han X.G., Wu J.G. (2005). Lack of evidence for ¾ scaling of metabolism in terrestrial
plants. *Journal of Integrative Plant Biology*, doi.org/10.1111/j.1744-7909.2005.00167.x
Litton C., Raich J., Ryan M.G. (2007). Carbon allocation in forest ecosystems. *Global Change*
Biology, 13: 2089–2109.
Lynch D.J., Matamala R., Iversen C.M., Norby R.J., Gonzalez-Meler M.A. (2013). Stored carbon
partly fuels fine-root respiration but is not used for production of new fine roots. *New*
Phytologist, 199: 420–430.
Machado J.-L., Reich P.B. (2006). Dark respiration rate increases with plant size in saplings of
three temperate tree species despite decreasing tissue nitrogen and nonstructural carbohydrates.
Tree Physiology, 26: 915–923.
Malhi Y. (2012). The productivity, metabolism and carbon cycle of tropical forest vegetation.
Journal of Ecology, 100: 65–75.
Malhi Y., Doughty C.E., *et al*. (2015). The linkages between photosynthesis, productivity, growth
and biomass in lowland Amazonian forests. *Global Change Bio*logy, 21: 2283–2295.
Maier C.A., Johnsen K.H., *et al*. (2010). Relationships between stem CO_2 efflux, substrate supply,
and growth in young loblolly pine trees. *New Phytologist*, 185: 502–513.
Makarieva A.M., Gorshkov V.G., Li B.-L. (2005). Revising the distributive network models of
West, Brown & Enquist (1997) and Banavar, Maritan & Rinaldo (1999). Metabolic inequity of
living tissues provides clues for the observed allometric scaling rules. *Journal of Theoretical*
Biology, 237, 291–307.
Makarieva A.M., Gorshkov V.G., Li B.-L., *et al*. (2008). Mean mass-specific metabolic rates are
strikingly similar across life's major domains: Evidence for life's metabolic optimum.
Proceedings of the National Academy of Sciences, doi/10.1073/pnas.0802148105.
Mäkelä A., Valentine H. (2001). The ratio of NPP to GPP: evidence of change over the course of
stand development. *Tree Physiology*, 21: 1015–1030.
Marconi S., Chiti T., *et al*. (2017). The Role of Respiration in Estimation of Net Carbon Cycle:
Coupling Soil Carbon Dynamics and Canopy Turnover in a Novel Version of 3D-CMCC
Forest Ecosystem Model. *Forests*, 8: 220–227.

Martínez-Vilalta J., Sala A., *et al*. (2016). Dynamics of non‑structural carbohydrates in terrestrial
plants: a global synthesis. *Ecological Monographs*, 86(4): 495–516
McDowell N., Pockman W.T., *et al*. (2008). Mechanisms of plant survival and mortality during
drought: why do some plants survive while others succumb to drought? *New Phytologist*, 178:
719–739.
McCree K. (1970). An equation for the rate of respiration of white clover plants grown under
controlled conditions. In S. I., *Prediction and measurement of photosynthetic productivity* (pp.
221–229). Wageningen: The Netherlands: Centre for Agricultural Publishing and
Documentation.
McMurtrie R.E., Norby R.J., *et al*. (2008). Why is plant-growth response to elevated CO_2 amplified
when water is limiting, but reduced when nitrogen is limiting? A growth–optimisation
hypothesis. *Functional Plant Biology*, 35: 521–534
Medlyn, B., & Dewar, R. (1999). Comment on the article by R. H. Waring, J. J. Landsberg and M.
Williams relating net primary production to gross primary production. *Tree Physiology*, 19 (2):
137–138.
Meir P., Grace J. (2002). Scaling relationship for woody tissue respiration in two tropical rain
forests. *Pant, Cell and Environment*, 25: 963–973
Merganicova K., Merganic J., Lehtonen A., *et al*. (2019). Forest carbon allocation modelling under
climate change. *Tree Physiology* (under review).
Mencuccini M., Martínez-Vilalta J. *et al*. (2005). Size-mediated ageing reduces vigour in trees.
Ecology Letters, 8: 1183–1190
Mori S., Yamaji K., *et al*. (2010). Mixed-power scaling of whole-plant respiration from seedlings to
giant   trees.   *Proceedings   of   the   National   Academy   of   Sciences,*
www.pnas.org/cgi/doi/10.1073/pnas.0902554107
Muhr J., Angert A., *et al*. (2013). Carbon dioxide emitted from live stems of tropical trees is several
years old. *Tree Physiology*, 33: 743–752
Muhr J., Messier C., Delagrange S., *et al*. (2016). How fresh is maple syrup? Sugar maple trees
mobilize carbon stored several years previously during early springtime sap-ascent. *New*
Phytologist, 209: 1410–1416.
O'Connor M.P., Kemp S.J., Agosta S.J., *et al*. (2007). Reconsidering the mechanistic basis of the
metabolic theory of ecology. *Oikos*, 116: 1058–1072.
Odum   E.P.   (1969).   The   Strategy   of   Ecosystem   Development.   *Science*,   doi:
10.1126/science.164.3877.262
Oleson K.W., Lawrence D.M., *et al*. (2013). Technical Description of version 4.5 of the
Community Land Model (CLM). Ncar Technical Note NCAR/TN-503+STR, National Center
for Atmospheric Research, Boulder, CO, 422 pp.
Palacio S., Hernández R., Maestro-Martínez M., Camarero J.J. (2012). Fast replenishment of initial
carbon stores after defoliation by the pine processionary moth and its relationship to the re-
growth ability of trees. *Trees – Structure and Function*, 26: 1627–1640
Pallardy S.G., (2010). Physiology of Woody Plants, New York, USA: Academic Press.
Pappas C., Fatichi S., *et al*. (2013). Sensitivity analysis of a process-based ecosystem model:
Pinpointing parameterization and structural issues. *Journal of Geophysical Research*, 118: 502–
528

Piao S., Luyssaert S., *et al.* (2010). Forest annual carbon cost: a global-scale analysis of autotrophic
respiration. *Ecology*, 91 (3): 652–661.
Pietsch S.A., *et al.* (2005) BGC-model parameters for tree species growing in central European
forests. *Forest Ecology and Management*, 211: 264–295.
Poulter B., Hattermann F., *et al.* (2010). Robust dynamics of Amazon dieback to climate change
with perturbed ecosystem model parameters. *Global Change Biology*, 16: 2476–2495.
Price C.A., Gillooly J.F., *et al.* (2010). The metabolic theory of ecology: prospects and challenges
for plant biology. *New Phytologist*, 188: 696–710
Price C.A., Wietz J.S., *et al.* (2012). Testing the metabolic theory of ecology. *Ecology Letters*, 15:
1465–1474.
Reich P.B., Tjoelker M.G., *et al.* (2006). Universal scaling of respiratory metabolism, size and
nitrogen in plants. *Nature*, doi:10.1038/nature04282.
Reich P.B., Tjoelker M.G., *et al.* (2008). Scaling of respiration to nitrogen in leaves, stems and
roots of higher land plants. *Ecology Letters*, doi: 10.1111/j.1461-0248.2008.01185.x
Reyer C., Silveyra Gonzalez R. Dolos K., *et al.* (2019). The PROFOUND database for evaluating
vegetation models and simulating climate impacts on forests. *V. 0.1.12. GFZ Data Services*.
http://doi.org/10.5880/PIK.2019.008
Richardson A.D., Carbone M.S., *et al.* (2013). Seasonal dynamics and age of stemwood
nonstructural carbohydrates in temperate forest trees. *New Phytologist*, 206: 590–597.
Richardson A.D., Carbone M.S., *et al.* (2013). Distribution and mixing of old and new nonstructural
carbon in two temperate trees. *New Phytologist*, 197: 850–861.
Roxburgh S.H., Berry S.L., Buckley T.N., Barnes B., Roderick M.L. (2005). What is NPP?
Inconsistent accounting of respiratory fluxes in the definition of net primary production.
Functional Ecology, 19: 378–382.
Sala, A., Fouts W., Hoch G. (2011). Carbon storage in trees: does relative carbon supply decrease
with tree size? *In* Size- and Age-Related Changes in Tree Structure and Function. Eds. Meinzer
F.C., Lachenbruch B. and Dawson T.E.. Springer, Dordrecht, pp. 287–306.
Sala A., Woodruff D.R., Meinzer F.C. (2012). Carbon dynamics in trees: feast or famine? *Tree
Physiology*, 32: 764–775.
Sala A., Mencuccini M. (2014). Plump trees win under drought. *Nature Climate Change*, 4: 666–
667.
Sato H., Itoh A., Kohyama T. (2007). SEIB–DGVM: A new Dynamic Global Vegetation Model
using a spatially explicit individual-based approach. *Ecological Modelling*, 200: 279–307
Schiestl-Aalto P., Kulmala L., *et al.* (2015). CASSIA – a dynamic model for predicting intra-annual
sink demand and interannual growth variation in Scots pine. *New Phytologist*, 206: 647–659.
Schiestl-Aalto P., Ryhti K. Mäkelä A*., et al.* (2019). Analysis of the NSC Storage Dynamics in Tree
Organs Reveals the Allocation to Belowground Symbionts in the Framework of Whole Tree
Carbon Balance. *Front. For. Glob. Change*, 2:17. doi: 10.3389/ffgc.2019.00017
Schuur E.A.G., Trumbore S.E. (2006). Partitioning sources of soil respiration in boreal black spruce
forest using radiocarbon. *Global Change Biology*, 12: 165–176.
Schwalm C.R., Ek A.R. (2004). A process-based model of forest ecosystems driven by
meteorology. *Ecological Modelling*, 179: 317–348.

Sitch S., Smith B., *et al*. (2003). Evaluation of ecosystem dynamics, plant geography and terrestrial
carbon cycling in the LPJ dynamic global vegetation model. *Global Change Biology*, 9: 161–
185.
Smith B., Prentice I.C., Sykes M.T. (2001). Representation of vegetation dynamics in the modelling
of terrestrial ecosystems: comparing two contrasting approaches within European climate
space. *Global Ecology and Biogeography*, 10: 621–637
Smith A.M., Stitt M. (2007). Coordination of carbon supply and plant growth. *Plant, Cell and
Environment*, 30: 1126–1149.
Solly E.F., Brunner I., Helmisaari H.-S., *et al*. (2018). Unravelling the age of fine roots of temperate
and boreal forests. *Nature Communications*, DOI: 10.1038/s41467-018-05460-6
Spicer R., Holbrook N.M. (2007). Parenchyma cell respiration and survival in secondary xylem:
does metabolic activity decline with cell age?. *Plant, Cell and Environment*, 30: 934–943
Stephenson N.L., Das A.J., *et al*. (2014). Rate of accumulation increases continuously with tree
size. *Nature*, doi:10.1038/nature12914.
Tang J., Luyssaert S., Richardson A.D., Kutsch W., Janssens I.A. (2014). Steeper declines in forest
photosynthesis than respiration explain age-driven decreases in forest growth. *Proceedings of
the National Academy of Sciences*, doi/10.1073/pnas.1320761111.
Tatarinov F., Cenciala E. (2006). Application of BIOME-BGC model to managed forests 1.
Sensitivity analysis. *Forest Ecology and Management*, 237: 267–279.
Tjoelker M.G., Oleksyn J., Reich P.B. (1999). Acclimation of respiration to temperature and CO_2 in
seedlings of boreal tree species in relation to plant size and relative growth rate. *Global Change
Biology*, 49: 679–691.
Thornley J.H.M., Cannell M.G.R. (2000). Modelling the components of plant respiration:
representation and realism. *Annals of Botany*, 85: 55–67.
Thornley J.H.M. (2011). Plant growth and respiration re-visited: maintenance respiration defined –
it is an emergent property of, not a separate process within, the system – and why the
respiration: photosynthesis ratio is conservative. *Annals of Botany*, 108: 1365–1380
Thornton P.E., Law B.E., *et al*. (2002). Modeling and measuring the effects of disturbance history
and climate on carbon and water budgets in evergreen needleleaf forests. *Agricultural and
Forest Meteorology*, 113: 185–222.
Turner M., Beer C., *et al*. (2017). Evaluation of climate-related carbon turnover processes in global
vegetation models for boreal and temperate forests. *Global Change Biology*, doi:
10.1111/gcb.13660
Van Oijen M., Rougier J., Smith R. (2005). Bayesian calibration of process-based models: bridging
the gap between models and data. *Tree Physiology*, 25: 915–927.
Van Oijen M., Schapendonk A., Hoglind M. (2010). On the relative magnitudes of photosynthesis,
respiration, growth and carbon storage in vegetation. *Annals of Botany*, 105: 739–797.
Vargas R., Trumbore S.E., Allen M.F. (2009). Evidence of old carbon used to grow new fine roots
in a tropical forest. *New Phytologist*, 182: 710–718.
Waring R.H., Landsberg J.L., Williams M. (1998). Net primary production of forests: a constant
fraction of gross primary production? *Tree Physiology*, 18: 129–134.
Warnant P., *et al*. (1994). CARAIB: A global model of terrestrial biological productivity. *Global
Biochemical Cycles*, 8(3): 255–270.

Warszawski L., Frieler K., *et al*. (2014). The Inter-Sectoral Impact Model Intercomparison Project
(ISI-MIP): Project Framework. *Proceedings of the National Academy of Sciences*, doi:
10.1073/pnas.1312330110.
Weber R., Schwendener A., *et al*. (2018). Living on next to nothing: tree seedlings can survive
weeks with very low carbohydrate concentration. *New Phytologist*, doi: 10.1111/nph.14987.
Wertin T.M., Teskey R.O. (2008). Close coupling of whole-plant respiration to net photosynthesis
and carbohydrates. *Tree Physiology*, 28: 1831–1840.
West G.B., Brown J.H., Enquist B.J. (1999). A general model for the structure and allometry of
plant vascular systems. *Nature*, 400, 664–667
White M., Thornton P.E., *et al.* (2000). Parameterization and Sensitivity Analysis of the BIOME-
BGC Terrestrial Ecosystem Model: Net Primary Production Controls. *Earth Interactions*, 4: 1-
85.
Wiley E., Helliker B. (2012). A re-evaluation of carbon storage in trees lends greater support for
carbon limitation to growth. *New Phytologist*, 195: 285–289.
Wiley E., *et al*. (2017). Dying piece by piece: carbohydrate dynamics in aspen (*Populus
tremuloides*) seedlings under severe carbon stress. *Journal of Experimental Botany*, 68(18):
5221–5232.
Woodruff D.R., Meinzer F.C. (2011). Water stress, shoot growth and storage of nonstructural
carbohydrates along a tree height gradient in a tall conifer. *Plant, Cell & Environment*, 11:
1920–1930.
Wu J., Larsen K., *et al*. (2013). Synthesis on the carbon budget and cycling in a Danish, temperate
deciduous forest. *Agricultural and Forest Meteorology*, 181: 94–107.
Xu C.-Y., *et al*. (2012). Age-related decline of stand biomass accumulation is primarily due to
mortality and not to reduction in NPP associated with individual tree physiology, tree growth or
stand structure in a *Quercus*-dominated forest. *Journal of Ecology*, 100: 428–440
Yang Y., *et al*. (2011). Carbon and nitrogen dynamics during forest stand development: a global
synthesis. *New Phytologist*, 190: 977–989
Yang J., He Y., Aubrey D.P., *et al*. (2016). Global patterns and predictors of stem CO_2 efflux in
forest ecosystems. *Global Change Biology*, 22: 1433–1444.
Yoda K., Kira T., Ogawa H., Hozumi K. (1963). Self-thinning in overcrowded pure stands under
cultivated and natural conditions. *J. Biol. Osaka City Univ*. 14, 107–129.
Zaehle S., Sitch S., *et al*. (2005). Effects of parameter uncertainties on the modelling of terrestrial
biosphere dynamics. *Global Biogeochemical Cycles*, 19, doi:10.1029/2004GB002395
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777 **Tables**

778 **Table 1** Underlying modelling assumptions adopted in the analysis

τ level	Corresponding underlying assumption	Reference
$\tau = 0.1$	Low turnover rate, which implies only accumulation of respiring biomass ($i.e. R \propto biomass$)	Reich *et al.* 2006
$\tau = 1$	High turnover rate, with death of cells annually equalling live cell production ($i.e. R \propto P$)	Waring *et al.* 1998
$0.1 < \tau < 1$	Intermediate turnover rate	e.g. White *et al.* 2000, see Box 2
$\tau = 0$	Functionally impossible, because it would imply no mortality of cells	-
$\tau > 1$	Physically impossible, because turnover would exceed the number of available living cells	-

780 **Table 2** Model validation (averages for the years 2006–2010), in brackets standard deviation

781 (when available). Literature data come from Wu *et al.* (2013).

2006-2010 mean (± S.D)	$P_n{:}P$	P	R	P_n	$P_n$woody	$P_n$litter	Above Ground-R	Below Ground-R	Above Ground woody stocks	Below Ground woody stocks	Total woody stocks
Units	unitless			g C m^{-2} yr^{-1}					g C m^{-2}		
Literature	0.37	1881 (± 127)	1173 (± 143)	708 (± 65)	307 (± 57)	401	872	301	9885 (± 279)	1848 (± 160)	11733 (± 281)
Modelled	0.45 (± 0.02)	1706 (± 52)	937 (± 30)	768 (± 60)	309 (± 56)	314 (± 9)	635 (± 20)	302 (± 9)	8993 (± 278)	1954 (± 545)	10948 (± 333)

Figures

Fig. 1 Proofs-of-concept for total R_M (a) and live woody biomass accumulation (b) over the

course of forest development (time) and increases in size, assuming different live wood turnover

(τ, yr^{-1}) rate values, from 1 (W98) to 0.1 yr^{-1} (R06) and including two intermediate values at 0.5

and 0.7 yr^{-1} (e.g. White *et al*. 2000). R_{M_green} (i.e. leaf and fine root R_M) was assumed constant

over-time and arbitrarily equal to 2. Summing up R_{M_wood} and R_{M_green} gives the total R_M. Initial

woody biomass was arbitrarily considered equal to 10, new annual live wood was also arbitrarily

considered equal to 10, $m_R = 0.2$; ($R_M = (W_{live\_woody} + W_{green}) \cdot m_R$; see Eq. 3). (b) Initial woody

biomass was arbitrarily considered equal to 10, new annual live wood was arbitrarily considered

equal to 10. The model is: $W_{live\_woody\,(t+1)} = W_{live\_woody\,(t)} + \Delta Win_{live\_woody\,(t+1)} - \Delta Wout_{dead\_woody\,(t+1)}$

(see equation 4 in the main text).

Fig. 2 Model results for (a) $P_n{:}P$ ratio (dimensionless), (b) net primary production (P_n, g C m^{-2}

yr^{-1}) and, (c) autotrophic respiration (R, g C m^{-2} yr^{-1}) performed with varying τ (coloured lines).

The beginning of simulations correspond to 1950 (stand age 30 years); the end of simulations

correspond to 2100 (stand age 180 years). The dark-pointed red line can be considered as a

mechanistic representation of W98's fixed $P_n{:}P$ ratio ($\tau = 1$ yr^{-1}) while the dark pink line

approximates R06's scaling relationship between R and biomass ($\tau = 0.1$ yr^{-1}). Orange dotted

lines represent Amthor's (2000) (A00) '*allowable*' range for the $P_n{:}P$ ratio (0.65 to 0.2). The red

dots give the average measured values (Wu *et al*. 2013) at the site for (a) $P_n{:}P$ ratio, (b) P_n and

(c) R. Vertical bars represent the standard deviation with horizontal bars representing the period

2006–2010 (stand age ~ 85–90 years). The shaded area represents the overall uncertainty of

model results.

Fig. 3 Regression analyses between whole-plant autotrophic respiration (R, g C m^{-2} yr^{-1}) and (a)

whole-plant carbon (W; g C m^{-2}), (b) carbon in living pools ($W_{live\_woody} + W_{green}$; g C m^{-2}), (c)

whole-plant nitrogen (W; g N m^{-2}) and (d) nitrogen in living pools ($W_{\text{live\_woody}}$ + W_{green}; g N m^{-2}).

Different colours represent different τ values as described in the legend panels (with $\tau = 0.1$ yr^{-1},

$n = 117$; with $\tau = 0.2$ yr^{-1}, $n = 149$; otherwise $n = 150$).

 BOX 1

The function and dynamics of non-structural carbohydrates

NSC is a surprisingly poorly known component of the whole-tree carbon balance, and

commonly disregarded in models (Schiestl-Aalto *et al*. 2019; Merganicova *et al*. *under*

review). However, the ability of trees to prioritize storage over growth depends on the role of

NSC in allowing temporal asynchrony between carbon demand and carbon supply (Fatichi *et*

al. 2014). Such imbalances are assumed to be buffered by drawing down NSC reserves.

Recent studies support this assumption, showing that during periods of negative carbon

balance (for example during the dormant season, periods of stress, or natural or artificially

induced defoliation episodes) NSC is remobilized and transported from the sites of phloem

loading, while during periods of positive carbon balance plants preferentially allocate recently

assimilated carbon to replenish NSC. Only afterwards is "new" carbon used to sustain growth

(Weber *et al*. 2018; Huang *et al*. 2019). Because ultimately plant survival depends more on

metabolic carbon demands than on growth, some have argued that all positive carbon flows

should be used to replenish NSC at the expense of growth until a minimum NSC pool size

(30–60% of the seasonal maximum, Martínez-Vilalta *et al*. 2016) is reached ('active' storage:

Sala *et al.* 2012), thus maintaining a safety margin against the risk of carbon starvation (Wiley

& Helliker 2012; Huang *et al*. 2019). Note that this assumption departs from the notion that

NSC is a mere reservoir for excess supply of carbon relative to growth demand ('passive'

storage: Kozłowski 1992). In the model, carbon allocation to all tree structural and non-

structural pools is computed here daily and is controlled by functional-constraints due to

direct and lagged C-requirements (Huang *et al*. 2019). It is assumed that a minimum NSC

threshold level concentration (11% of sapwood dry mass: Genet *et al*. 2010) has to be

maintained for multiple functions including osmoregulation, cell turgor, vascular integrity,

tree survival (reviewed in Hartmann & Trumbore 2016) and organ-specific phenology (leaf

and fine-root formation). The greater the sapwood mass, the greater the minimum NSC

threshold must be (Dietze *et al*. 2014). For deciduous trees, four phenological phases are

distinguished: (i) the *dormant* phase, where R is fuelled by NSC-consumption; (ii) the *leaf*

onset phase, when leaf and fine root production consume NSC (unless the carbon balance is

positive, in which case new assimilates are used) until the predicted maximum annual LAI is

reached; (iii) the *full growing* phase, when new assimilates are allocated to stem, coarse roots,

branch, and fruits, and only into the NSC pool if this is below its minimum level; (iv) the *leaf*

fall phase, when all assimilates are allocated to the NSC reserve pool while some (~ 10%)

NSC is relocated from falling leaves and dying fine roots (Campioli *et al*. 2013). For

evergreen species the model follows a simpler schedule consisting of a first maximum growth

phase, when the model allocates NSC to foliage and fine roots up to peak LAI, and a second

full growing phase, when the model allocates to all of the pools (Kuptz *et al*. 2011). Such

patterns of whole-tree seasonal NSC dynamics have been confirmed by Furze *et al*. (2018)

and a similar phenological and carbon allocation scheme has been adopted by other models

(e.g. Krinner *et al*. 2005; Arora *&* Boer 2005).

BOX 2

Turnover rates and other uncertainties in models

Most vegetation models assume a fixed rate of sapwood turnover, τ. However, lack of

information on this parameter has been shown to be an important source of uncertainty in the

modelled carbon balance of vegetation stands (Goulden *et al.* 2011; Malhi 2012; Collalti *et al.*

2019). Values adopted in current models include: $\tau = 0.7$ yr^{-1} in CLM (Oleson *et al.* 2013),

Forest v.5.1 (Schwalm & Ek 2004), 3D-CMCC-CNR (Collalti *et al.* 2019) and Biome-BGC

(Thornton *et al.* 2002); $\tau \sim 0.75$ yr^{-1} in CASTANEA (Dufrêne *et al.* 2005); $\tau = 0.85$ yr^{-1} in LPJ-

GUESS (Smith *et al.* 2001); $\tau = 0.95$ yr^{-1} in SEIB-DGVM (Sato *et al.* 2007), LPJ-DGVM (Sitch

et al. 2003) and NCAR-LSM (Bonan *et al.* 2003); and $\tau \sim 1$ yr^{-1} in CARAIB (Warnant *et al.*

1994), PnET (Whythers *et al.* 2013), and ORCHIDEE (Krinner *et al.* 2005).

Additional sources of uncertainty include the lack of consideration of a size- or age-related

decline in the ratio of living to dead cells (suggesting a declining τ) (Damesin *et al.* 2002;

Ceschia *et al.* 2002), the effect of changes in climate (which could temporarily increase τ to

reduce maintenance costs in favour of growth: Doughty *et al.* 2015), changes in tissue N and

NSC concentrations (Machado & Reich 2006; Thurner *et al.* 2017), and, a probable, genetically

controlled down-regulation of basal respiration rates with the ageing of cells (Carey *et al.* 2001;

Wiley *et al.* 2017). Moreover, both τ and basal respiration rates (g_R and m_R) are likely to vary

among different tree biomass pools (Reich *et al.* 2008). Respiratory carbon losses per unit plant

mass may also change to sustain growth as an acclimatory response to carbon demand due to

increasing plant size, and perhaps with changing climate (Smith & Stitt 2007). These hypotheses

are all grounded in theory, but are supported by very limited observations (Friend *et al.* 2014;

Thurner *et al.* 2017).