Anthropocene refugia: integrating history and predictive modelling to

assess the space available for biodiversity in a human-dominated world

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1. Abstract

During periods of strong environmental change, some areas may serve as refugia, where components of biodiversity can find protection, persist and potentially expand from should conditions again become favourable. The refugia concept has previously been used in the context of climatic change, to describe climatically stable areas in which taxa survived past Quaternary glacial-interglacial oscillations, or where they might persist in the future under anthropogenic climate change. However, with the recognition that Earth has entered the Anthropocene, an era in which human activities are the dominant driving force on ecosystems, it is critical to also consider human pressures on the environment as factors limiting species distributions. Here, we present a novel concept, Anthropocene refugia, to refer to areas that provide spatial and temporal protection from human activities and that will remain suitable for a given taxonomic unit in the long-term. It integrates a deep-time perspective on species biogeography that provides information on the natural rather than current-day relictual distribution of species, with spatial information on modern and future anthropogenic threats. We define the concept and propose a methodology to effectively identify and map realised and potential current and future refugia, using examples for two megafauna species as a proof of concept. We argue that identifying Anthropocene refugia will improve biodiversity conservation and restoration by allowing better prediction of key areas for conservation and potential for re-expansions today and in the future. More generally, it forms a new conceptual framework to assess and manage the impact of anthropogenic activities on past, current and future patterns of species distributions.

Keywords: Anthropocene, ecosystem restoration, refuge, rewilding, shifting baseline, species distribution modelling

2. Introduction

While species distribution patterns have strong signatures tied to natural biotic and abiotic process, it is becoming increasingly difficult to ignore the role that humans play in shaping the composition of species assemblages across landscapes. With 95% of land having at least some degree of modification by human activities [1], the extent of wilderness areas have declined dramatically, and so have the opportunities to protect and conserve viable populations of many species in their natural habitat. Meanwhile, anthropogenic climate change is predicted to cause range shifts, range contractions and changes in elevational distributions in many organisms [2,3], challenging our approach to biodiversity conservation [4]. Human impacts on the global environment have become so pervasive that a new geological epoch has been proposed: the Anthropocene [5]. Identifying the space available for biodiversity protection and recovery in this human-dominated world is a challenge that requires a comprehensive understanding of the interactions between species' natural biogeographic patterns and the spatial distribution of anthropogenic pressures.

The terms 'refuge' or 'refugia' are commonly used to refer to areas where components of biodiversity retreat to, and persist in, under increasing environmental stress, with the potential to re-expand once the stress decreases [6,7]. These concepts have previously been applied in the context of past or contemporary climatic change, to identify areas that are relatively buffered from climatic changes and where components of biodiversity have persisted in the past or may be able to persist in the future. Such areas can act as sources of recolonization when environmental conditions improve, often have long-lasting imprints on species distributions [8], and have therefore become a central focus in much biogeographic research [6]. However, previous uses of the concept have not accounted for the impact of anthropogenic pressures, other than climate change, on species' biogeographic patterns [6]. With the recognition that human activities are now a major force driving global ecosystems [9], there is a need for incorporating a wider range of human pressures, beyond anthropogenic climate change, as drivers limiting the persistence of species in human-dominated landscapes.

Here, we build on previous uses of the term to introduce a novel concept - Anthropocene refugia - which

refers to areas allowing the long-term survival and persistence of organisms that are sensitive to human

activities and providing sources for broader recovery once pressures have lessened. It intersects knowledge

on the potential distribution of the organism of interest, incorporating prehistoric and historic data on taxa's

past distribution, with spatially explicit information on current and future anthropogenic pressures. The

outcome is the identification of areas where an organism can persist, or be restored to, in the Anthropocene,

given its ecology, vulnerability to human activities and the predicted changes in suitability of these areas.

This concept will contribute to a much-needed development of more proactive approaches to nature

conservation to overcome the ongoing species and ecosystem declines and the long-term reductions in

biodiversity this entails [10,11]. We define this novel conceptual framework and suggest a methodology to

identify Anthropocene refugia in practice, identifying the sources of information and available material that

can be used for this exercise. We also discuss the applications for species' conservation, management and

restoration, using examples for two megafauna species as case studies and proof of concept.

3. Anthropocene refugia: a conceptual framework

The concept of refugia was originally used to study the response of organisms to past periods of glacial-

interglacial oscillations of the late Quaternary [6], allowing a greater understanding of observed patterns of

species' biogeography, evolution and demography [e.g. 12–14]. More recently, the term of climatic refugia

has been applied to contemporary landscapes, referring to locations projected to harbour remnants of

present-day climates, which may serve as safe havens for biodiversity under future climate change [7]. In

conservation planning, identifying these climate change refugia can help prioritise management efforts in

the face of contemporary climate change [15,16].

These definitions, however, only consider climate as the driver of change in species distribution and

abundance, both in evolutionary and ecological timeframes [16]. While changes in species' biogeography

during the Quaternary were largely driven by glacial-interglacial oscillations [17], the more recent use of

climate change refugia fails to incorporate anthropogenic pressures that, along with climate change, can

affect the distribution of species in an increasingly human-dominated world. The concept of Anthropocene

refugia overcomes this limitation by incorporating climate change and a wide range of anthropogenic

pressures into the identification of refugia. It designates a spatial entity that answers to two qualities: being

ecologically suitable for the biodiversity unit considered and having relatively low levels of human pressure

that allow its long-term (through several generations) persistence in this area (Figure 1). We differentiate

realised and potential refugia, based on whether the refugia is within the taxa's current range, or within the

area where it could potentially persist in the future.

Anthropocene refugia are different from "remnants", patches of suitable habitat for species intolerant to a

human-modified landscape [18], or "coolspots", areas across species distributions that are free from

anthropogenic threats that the species is sensitive to [19], which are both snapshots in time of where

species can persist within their current range. In the Anthropocene refugia concept, a strong emphasis is

made on identifying the potential range of species, taking into account the natural rather than current-day

relictual niche of the species, and the stability of suitable habitat through time. Niche truncation caused by

past local extirpations biases our understanding of natural habitat-species relationships. By

underestimating the extent of potential suitable habitat, this can affect our understanding of areas

available for conservation today [20] and provide unreliable forecasts of distribution changes under future

climate change [21]. This case of shifted baseline can be avoided by incorporating prehistoric and historic

data on the taxon's occurrence into estimates of potential distribution. As such, this concept bridges

perspectives on taxa's long-term biogeography with current and future predictions of distribution patterns

in human-dominated landscapes.

Anthropocene refugia can be considered both conceptually and as a tool for conservation and

management. As a concept, Anthropocene refugia allows higher-level thinking about the implication and

consequences of human activities on species distribution and about the space available for biodiversity in

the Anthropocene. Similar to the concept of glacial-interglacial refugia, which allows for a critical

examination of current biogeographic patterns with an historical perspective, the identification of

Anthropocene refugia will equip the next generations of ecologists with a tool to understand how

biogeographic patterns emerge in a human-dominated world. It combines the protection perspective with

the restoration (reintroduction or rewilding) perspective, in contrast to more classic conservation

approaches which focus on the former [22].

As a tool, Anthropocene refugia can be used in categorisation, decision making, and inference for

conservation planning. The approach is complementary to previous efforts to identify priority regions for

establishing protected areas or remaining wilderness areas [1,23], but acknowledges that species

persistence does not only rely on the existence of formally protected or pressure-free areas [24]. Although

protected areas are most certainly a critical component of large-scale conservation planning efforts, some

species have demonstrated their ability to tolerate substantial levels of human pressures and survive

outside of formally protected areas [25,26]. By enabling persistence, acting as wildlife corridors and serving

as a source for future recolonization, refugia are critical to support biodiversity protection in a changing

world. Compared to previous efforts to map human footprint [27] or degree of human modification [1],

which provide information on the current distribution of areas that have low human pressure,

Anthropocene refugia considers a finer taxonomic resolution, by intersecting these type of information

with distribution data for given taxa. Because individual taxonomic units – populations or species – are

often regarded as the fundamental unit of conservation, this makes it a valuable new tool to inform

conservation management.

Figure 1. Schematic representation of the Anthropocene refugia concept. Spatially-explicit data on taxa distribution (point data, range maps or predicted distributions) and anthropogenic pressures are combined to map the distribution of realised and potential Anthropocene refugia.

4. Identifying Anthropocene refugia

In practice, mapping Anthropocene refugia requires geographic information on the long-term distribution of the taxa of interest and of the relevant anthropogenic pressures (Figure 1). We provide suggestions below for the type of data that could be included and identify some online and open-access datasets from which this information can be retrieved (Supplementary Information).

Taxa distribution data

Distribution of taxa can be mapped from 1) point data, the locations where a species has been recorded, 2)

geographic ranges, the geographic boundaries of the area where a species is known to occur [28] and 3)

predicted distributions, the areas where a species is likely to be present as inferred from the suitability of

environmental conditions [29]. We refer the reader to existing reviews to understand the attributes,

strength and limitations of these different data types in the context of conservation [29–33].

To be useful in mapping Anthropocene refugia, point data need to be transformed into areas, e.g. by

intersecting them with geographic units (grid squares or administrative boundaries) or by converting them

into geographic ranges (extent of occurrence EOO or area of occupancy AOO [31]), using interpolation and

expert knowledge. Predicted distributions that relate species occurrence with environmental conditions

based on mechanistic (process-based) or correlative (statistical) niche modelling enable the extrapolation of

incomplete point locality data, and the interpolation of habitat suitability measure throughout the range

[34,35]. They can provide a more realistic outcome by minimising both commission and omission errors

(erroneous indication that a species is present or absent, respectively), but require following high standard

practices regarding the choice of species distribution data or mechanistic links describing species' niche,

environmental predictors, and the building and evaluation of the model [36]. Mechanistic and correlative

niche models can also be used to predict changes in taxa distribution under future scenarios of climate

change [37,38]. From these, hypotheses about the future extent of occurrence of organisms can be derived

to inform conservation interventions [e.g. 39], and serve as a basis for mapping Anthropocene refugia.

Niche-truncation caused by past or current anthropogenic impacts may bias our understanding of species'

biogeography and ecological requirements [20,21]. This, in turn can lead to underestimating the potential

distribution of species and the area available for their conservation [20,40] and bias forecasts of species

distribution under future climate change [21]. The distribution data used in the mapping should be

informative of the natural rather than current-day relictual range of species, to counter the undesirable

effects of the shifting baseline syndrome [41]. This may require a deeper investigation of past local

extinctions and truncations in species-environment relationship, which may be informed by integrating

knowledge or data on the historical distribution of the species [42].

Long-term biodiversity data, i.e. pre-historical and historical occurrence records collected from

archaeological, museum, written and oral sources, can fill gaps in knowledge and inform the natural

distribution of organisms, i.e. the range they could occupy today in the absence of human impacts. Fossil

records of the late Quaternary can provide information on biodiversity changes as a result of climatic changes

and anthropogenic impacts over the last millenia, while historical descriptions and museum specimens

collected by literate travellers and naturalists are informative of the more recent past, particularly of the

period following European expansion. Finally, traditional ecological knowledge accumulated through many

generations of close interactions between people and the natural world can provide local information on the

state of ecosystems from the relatively recent past [43]. Ultimately, the temporal cut-off for long-term

biodiversity data should be taxon-specific and reflect the timeline and geography of known or suspected

impacts as well as the objectives of the study.

Mechanistic or correlative niche modelling can be used to extrapolate habitat suitability and delineate the

potential range of the species [e.g. 39,56]. If past climate data are available, a multi-temporal calibration

approach to identify a taxon's potential niche from past and present occurrence records can be used to

project the range in current environmental conditions, assuming that the niche has remained stable through

time [45]. For data-poor regions or taxa, expert knowledge can also bring a useful approximation to the

potential natural range of the organism. Finally, a combination of different approaches can prove the most

efficient to infer the biogeographic history of species and deduce their current potential range [46,47].

In general, the use of long-term data in ecological analyses is increasing, with positive outcomes to improve

our understanding of natural species-habitat relationships and reduce the pervasive effect of the shifting

baseline syndrome. As an example, Lentini et al. reconstructed the historical distribution of kākāpō *Strigops*

habroptilus using a combination of Holocene fossil records, mid-Holocene climate data and species

distribution modelling, identifying areas suitable for reintroductions for this endangered bird species [44].

Laliberte and Ripple used historical data to identify patterns of range contraction and expansions in 43

species of North American carnivores and ungulates, and showed reduced persistence of species in areas of

higher human influence following Euro-American settlement [48]. These types of studies can form the basis

of Anthropocene refugia mapping for the taxa they are considering.

All data types are prone to error, incompleteness and biases (temporal, geographical, environmental and

taxonomic), which can affect interpretation of distribution patterns. Biases in the Holocene fossil records are

examined by Crees et al. in this special issue [49] and discussion on the biases and errors in data extracted

from the historical literature and museum specimens can be found in [50–52]. Historical data, while powerful

in their potential to highlight neglected aspects of species' biogeography, require a rigorous methodology for

their collection and interpretation to avoid the common pitfalls in historical ecology [53]. Before including

long-term data in analyses of species distribution, sources of errors and biases should be thoroughly

investigated and corrected for, and the remaining uncertainties acknowledged.

We list examples of online repositories for global species distribution data, including archaeological

records, in Supplementary Information. Each of these data sources need to be critically evaluated for their

relevance to answer the biological questions, according to criteria such as resolution, scale or

representativeness of natural range.

Anthropogenic pressures

The anthropogenic pressures to include in the mapping are those that directly or indirectly have a negative

impact on the persistence of the organism of interest and will vary according to the taxa considered and

the spatial scale of the study. For example, while land mammals are mostly affected by habitat loss,

degradation and harvesting [54], reptiles and amphibians appear to be primarily affected by agriculture and

biological resource use, urban development, natural system modification and invasive species [55], and

megafauna species across taxonomic groups are mainly threatened by direct harvesting [56]. Salafsky et al.

[57] proposed 11 categories of threats in a unified classification that is being used by the International

Union for the Conservation of Nature (IUCN) Red List [58]. It can prove a valuable resource, along with a

literature review or expert knowledge, to identify the set of threat that is relevant to the organism of

interest.

Recent studies have also used primary data sources on human stressors to build indices of global

cumulative impact of human activities on the environment. The Human Footprint project uses for example

data on 1) built environments, 2) population density, 3) electric infrastructure, 4) crop lands, 5) pasture

lands, 6) roads, 7) railways, and 8) navigable waterways to map the direct and indirect human pressures on

the environment globally in 1993 and 2009 [27]. In a recent study, Kennedy et al. [1] provide a cumulative

measure of human modification of terrestrial lands based on 13 anthropogenic stressors classified in five

major categories: 1) human settlement, 2) agriculture, 3) transportation, 4) mining and energy production,

and 5) electrical infrastructure, for a median year of 2016. Spatially explicit global datasets for these two

projects are available online [59,60] and are relevant resources that can be used for mapping

Anthropocene refugia.

New and emerging threats from wildlife trade, changing land use patterns, and the increase in global human

populations, may increase pressure on ecosystems in the future. In a study projecting the global risk of

habitat conversion in the future, Oakleaf et al. estimate that 20% of the remaining global natural land could

become converted in the next decades [61]. On the other hand, the implementation of new protected areas

and farmland abandonment may provide additional space for biodiversity in some areas [62,63]. In order to

identify which areas will allow the long-term persistence of a given organism, these future dynamics need to

be incorporated, as much as possible, into the mapping of Anthropocene refugia. While it may prove difficult

to obtain robust predictions for the direction of these changes for all relevant anthropogenic pressures,

forecasts of future spatial patterns of human activities under different scenarios for some pressure types are

available (e.g. [61,64–66]) and can be used as a first approximation to understand how the distribution of

anthropogenic pressures will change in the future.

With the growing development of remote-sensing technologies [67], global environment modelling and

systematic surveys, spatially-explicit information on anthropogenic pressures is increasingly made available

through publicly available online repositories. We provide a non-exhaustive list of these spatial datasets for

different types of human pressures in the Supplementary Information, which can serve as a basis for more

case-specific listing of potential data sources. If spatial information for some of the identified threats is

unavailable, proxy variables that correlate with the variable of interest can be considered for replacement.

The methodology can also be revisited to include additional pressures as more data become available in the

future.

Mapping Anthropocene refugia

Assigning relative pressure scores to each human pressure variable and intersecting these with species

distribution data will result in a taxa-specific map of anthropogenic pressure intensity within the range of

the species. Anthropocene refugia lie at the intersection of areas that remains suitable for the organism

through time and areas with low levels of observed and predicted anthropogenic pressure. A distinction is

made between "realised" refugia that are currently occupied by the organism of interest and predicted to

remain suitable in the future, and "potential" refugia, i.e. areas where the taxa is not currently extant but

that are within its potential future range under scenarios of future climate change.

Using thresholds to convert maps of continuous value into binary maps highlighting the specific areas that

could act as Anthropocene refugia may be useful for management and decision-making purposes. This

approach however leads to loss of valuable information and should be used with caution. If necessary, an

option is to calibrate the threshold based on the level of threat that the organism is currently able to

sustain within its range [68], the rationale being that an organism will possibly be able to persist in the long-

term within areas of lower or similar level of anthropogenic pressure than what it is currently experiencing,

with the assumption that the tolerance of humans for its presence will remain similar. This is a conservative

approach as refugia could be underestimated for some species that are actually able to sustain higher levels

of anthropogenic pressures than what is currently observed. To reflect uncertainty and provide a better

decision tool, we recommend testing and reporting variability from using various thresholds. A lower limit

on the size of contiguous areas that qualify as refugia for each taxa can also be applied, to meet the

requirements for these areas to sustain viable populations [69, but see 70].

In Box 1, we illustrate the Anthropocene refugia concept by mapping realised and potential Anthropocene

refugia for two megafauna species, the American bison *Bison bison* and tiger *Panthera tigris*. Both species

have undergone massive declines in range and abundance in the past and their persistence is dependent on

ongoing conservation programs. Their major ecological roles also make then good candidates for trophic

rewilding [71]. While these maps primarily have an illustrative purpose, we briefly discuss potential

implications in terms of conservation strategies in Box 1. Details on the methodology to produce these

maps are available in the Supplementary Information.

5. Applications in conservation

The concept of Anthropocene refugia fills a gap left open by other conservation approaches that only

identify climate change refugia or remnant populations to prioritize conservation efforts. The approach is

conceptually similar to that proposed by Allan et al. in a recent study, which combined cumulative human

impact with current distributions of terrestrial vertebrates to identify hotspots of impacted species richness

and coolspots of unimpacted species richness [19]. However, instead of focusing on species' current range

of occurrence, our approach offers a complementary framework in which information from the past is used

to understand the potential distribution of a given taxon in the future. We are thus offering a tool to not

only define threat mitigation strategies, but to also more fully identify restoration options under future

global change and changing anthropogenic pressures. This framework is coherent with the recent

designation of 2021–2030 as the "decade of ecosystem restoration" by the United Nations General

Assembly and the increasing emphasis put on rewilding as a restoration tool [72]. Identifying those refugia

is arguably a challenging exercise and the definition itself is open to further discussion and refinement.

With this in mind, we believe there is a high potential for Anthropocene refugia to inform contemporary

conservation and restoration, and we suggest possible applications below.

Realised refugia

Conservation management and restoration within realised Anthropocene refugia is important to maintain

existing populations in areas that are predicted to remain suitable for the taxa in the long-term and to

promote self-managing, biodiverse ecosystems. This is likely to be facilitated by the lower need for human

intervention in areas where taxa are already extant and where the level of anthropogenic pressure is not

expected to increase beyond its tolerance level in the near future. This can help identify priority regions for

establishing new protected areas, and inform strategies to directly mitigate the threats driving species'

declines [19]. It can feed into decision making for conservation management and conservation planning,

e.g. following a similar framework as those proposed for climate change refugia [15,16]. This can also be

used to identify candidate sites for reinforcements, i.e. the release of an organism into an existing

population of conspecifics to enhance population viability [73]. Nonetheless, in a context of fragmented

distributions, there is a possibility that populations within these refugia represent sink populations that

would become extinct if they are no longer accessible to dispersers [74], or refugee populations confined to

suboptimal habitats, with consequences of decreased fitness and density [75]. The viability of population in

these refuge areas could be assessed using information from paleo- and historical ecology as well as field

studies [76]. Identifying corridors between refugia and maintaining connectivity between these areas is

thus important to allow natural dispersal and improve population persistence [4]. These areas can also be

affected by ecological imbalances (e.g. from the loss of ecological interactions [77]), causing a need to

restore their functionality, e.g. through trophic rewilding [71]. From a socio-ecological perspective, the

implications for local populations of focusing conservation actions on these areas should also be considered

[78].

Potential refugia

Mapping potential Anthropocene refugia can form the basis of a more in-depth evaluation of candidate sites

for reintroduction and introduction efforts. For many threatened species, conservation success will rely on

the expansion of current relict distributions through reintroductions, a strategy that can be guided by the

mapping of Anthropocene refugia. By selecting sites that match the biotic and abiotic needs of the focal

species in the long-term, this can also inform restoration strategies such as trophic rewilding, an approach

promoting self-regulating ecosystems through the introduction of species to restore top-down trophic

interactions and associated trophic cascades [71]. It also opens the possibility to change our approach to

conservation interventions and translocations to allow the emergence of novel ecosystems that will be robust

to future habitat shifts and changing anthropogenic impacts [79]. This may involve planning corridors to allow

colonisation through natural dispersal or implementing measures to introduce species outside of their

historical range, an approach called assisted colonisation [80].

Identifying areas that are most likely to be recolonised in the near future is also important to anticipate the

complex impacts this could have on ecosystems as well as the interactions with society. The socio-ecological

implications of some species' recolonising part of their historical range or occupying new areas should be

carefully considered, in particular for species often involved in human-wildlife conflicts, such as large

carnivores [81]. Unified socio-ecological approaches that explicitly acknowledge competing perspectives

between wildlife conservation and social and governance contexts can be applied to provide new insights

into management options for sustainable human–wildlife coexistence [82].

323 6. Challenges and opportunities in mapping Anthropocene refugia

The historical pressures that caused a taxa's local extinction must no longer exist locally for an area to be

considered a suitable candidate for restoration. As some threats can be difficult to explicitly include in the

mapping exercise, a thorough investigation of past and existing impacts from human activities is necessary

before actual conservation recommendations can be proposed.

The highly dynamic state that characterizes the Anthropocene hinders our ability to make predictions for the

direction and intensity of changes in the global environment beyond the next few decades. Future global

warming, with temperatures predicted to exceed Quaternary levels, will have potential knock-on effects on

the entire biosphere and unpredictable consequences for the survival of individual species [83]. This

challenges our ability to predict the location of stable Anthropocene refugia, a significant limitation that will

not be easily resolved. In this context of uncertainty, focusing conservation efforts on areas that are likely to

remain suitable in the near future might be our best option to increase the chances of survival for sensitive

taxa and maintain a genetic diversity that will allow potential future recovery. It also gains time for

conservationists and managers to develop long-term solutions for the survival of populations and species.

Opportunities to improve the mapping of long-term refugia will increase as we gain a better understanding

of taxa distributions' response to climate change and with the future release of spatial data on predicted

changes in anthropogenic pressures. As these are integrated into the analysis, one can move from a paradigm

focused on places to one focused on dynamics.

Another important consideration is the scale at which Anthropocene refugia should be identified and

managed. The geopolitical scale used in conservation has an enormous influence on the identification of

management areas, with for example low congruence between the network of conservation areas identified

at the broad regional *vs* the fine local scale [84]. In practice, the definition of conservation targets is often

site-specific. For example, the Natura 2000 network of protected areas, one of the pillar of the European

Community conservation policy, emphases the management of specific sites [85]. Finer-scale refugia maps

based on local-scale and high-resolution distribution information could thus prove relevant for the

identification of land units that will form the basis of management decisions in this context. However, this

site-specific approach does not always provide an efficient prioritisation in relation to broader biodiversity

concerns [86,87]. Smaller-scale analyses may lead to suboptimal prioritisation with respect to the value for

global biodiversity and cost effectiveness, calling for a wider historical and geographical context to

contextualise management measures. This demonstrates how conservation strategies can be identified over

a wide range of scale and how identifying Anthropocene refugia at both a fine and broad resolution can prove

useful for conservation. Ultimately, decisions relative to scale will depend on the objectives of each project

and the availability of relevant datasets. The rationale for these choices needs to be explicit and argued, in

order to withstand scrutiny and allow for future refinement.

Box 1. Illustration of Anthropocene refugia mapping for the American bison *Bison bison* (A to D) and Tiger *Panthera tigris* (E to H). Maps A-C and E-G are based on the taxa-specific index of current anthropogenic threat, on which we overlaid the areas with a high probability of being converted into cropland or built areas in the future (predicted future transformed areas) for maps C and G. A and E show the current (IUCN) ranges where each species is extant today, B and F show the current-potential distributions, and C and G show the potential future range under climate change. D and H show the locations of realised and potential Anthropocene refugia (AR) for the two species, calculated by applying a threshold on the threat index within the potential future range of the species equal to the 95% percentile of the threat level currently experienced by the species. Grey areas are predicted to be climatically suitable for the species in the future but have a value of threat index above the reported threshold. See Supplementary Information for details about the methodology and maps of AR using different thresholds.

American bison. The American bison currently persist in very fragmented and highly managed populations in the United States and Canada. The areas within the northernmost part of the range, which are within the future predicted range for the species and with low levels of threats now and in the future, are realised refugia for the species. Conservation actions could focus on managing and reinforcing populations in these areas, e.g. by enforcing protection and building corridors between these populations. The climatically suitable area is also predicted to expand in Canada and Alaska under future climate change, and land in these areas is not predicted to be transformed in the near future, making these regions potential refugia. These are good candidate areas to restore bison populations through reintroductions or assisted colonisation. On the contrary, the southern part of the current potential range has high level of threats and the range of the species is predicted to contract and fragment under future climate change. Conservation and restoration in this area will thus require greater efforts.

Tiger. The current range of the tiger is very fragmented and located in areas surrounded with high anthropogenic pressures (e.g. in India). Some of these areas represent realised refugia because they are predicted to remain within the range of the species under future climate change and the anthropogenic pressures are not expected to increase (e.g. protected areas in India, Bhutan, China and southeast Asia and in the Russian Far East). Threat levels in northeast China, Russia, Mongolia and Kazakhstan are low and the climate suitability is predicted to remain stable. These are potential candidate areas for restoration, although a more thorough investigation of historical threats that drove the species' local extinction in these areas would be needed to shed light on their current suitability. Other areas are part of the current potential range of the species, but may become unsuitable in the future because of predicted range shifts under climate change (e.g. Borneo) or a predicted increase in land-conversion (Southeast Asia and part of Indonesia). These elements could be taken into account for establishing conservation priorities to ensure the species' long-term persistence.

7. Conclusion

We here introduce a novel concept – Anthropocene refugia – to account for the role of anthropogenic pressures in defining realised and potential refugia for biodiversity in a human-dominated planet. Our main intention is to call for a better consideration of the full range of anthropogenic pressures, beyond climate change, to identify refugia in the Anthropocene, while at the same time assessing re-expansion possibilities. We also emphasise the importance of considering a long-term perspective in defining taxa's potential distribution, to overcome the shifting baseline syndrome that affects our understanding of natural biogeographical patterns, limiting many assessments of restoration options. We highlight key possible applications of the concept with special emphasis on its potential to inform restoration approaches such as trophic rewilding, as well as nature management more generally, in the face of contemporary climate change. It is our hope to see a larger range of potential applications of this approach discussed and implemented in the future. Despite the many challenges, we argue that the Anthropocene refugia concept has the potential to bridge important gaps in our perspectives on the past (the distribution and ecology of wildlife prior to severe human pressure), present (where wildlife occur now in a human-dominated world) and future (where society will allow wildlife to exist in a reconciling world), thus representing both an important concept to reflect on our coexistence with wildlife on this planet, and an integral component of the conservation and restoration toolbox to protect and promote biodiversity in the Anthropocene.

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References

1. Kennedy CM, Oakleaf JR, Theobald DM, Baruch-Mordo S, Kiesecker J. 2019 Managing the middle: A shift in conservation priorities based on the global human modification gradient. *Glob. Change Biol.* (doi:10.1111/gcb.14549)

2. Freeman BG, Lee-Yaw JA, Sunday JM, Hargreaves AL. 2018 Expanding, shifting and shrinking: The impact of global warming on species' elevational distributions. *Glob. Ecol. Biogeogr.* (doi:10.1111/geb.12774)

3. Lenoir J, Svenning J-C. 2015 Climate-related range shifts - a global multidimensional synthesis and new research directions. *Ecography* **38**, 15–28. (doi:10.1111/ecog.00967)

4. Heller NE, Zavaleta ES. 2009 Biodiversity management in the face of climate change: A review of 22 years of recommendations. *Biol. Conserv.* **142**, 14–32. (doi:10.1016/j.biocon.2008.10.006)

5. Crutzen PJ. 2002 Geology of mankind. *Nature* **415**, 27.

6. Bennett KD, Provan J. 2008 What do we mean by 'refugia'? *Ice Age Refug. Quat. Extinctions Issue Quat. Evol. Palaeoecol.* **27**, 2449–2455. (doi:10.1016/j.quascirev.2008.08.019)

7. Keppel G, Van Niel KP, Wardell-Johnson GW, Yates CJ, Byrne M, Mucina L, Schut AGT, Hopper SD, Franklin SE. 2012 Refugia: identifying and understanding safe havens for biodiversity under climate change. *Glob. Ecol. Biogeogr.* **21**, 393–404. (doi:10.1111/j.1466-8238.2011.00686.x)

8. Graham CH, Moritz C, Williams SE. 2006 Habitat history improves prediction of biodiversity in rainforest fauna. *Proc. Natl. Acad. Sci.* **103**, 632–636. (doi:10.1073/pnas.0505754103)

9. Vitousek PM, Mooney HA, Lubchenco J, Melillo JM. 1997 Human Domination of Earth's Ecosystems. *Science* **277**, 494–499. (doi:10.1126/science.277.5325.494)

10. Davis M, Faurby S, Svenning J-C. 2018 Mammal diversity will take millions of years to recover from the current biodiversity crisis. *Proc. Natl. Acad. Sci.* **115**, 11262–11267. (doi:10.1073/pnas.1804906115)

11. Svenning J-C. 2018 Proactive conservation and restoration of botanical diversity in the Anthropocene's "rambunctious garden". *Am. J. Bot.* **105**, 963–966. (doi:10.1002/ajb2.1117)

12. Svenning J-C, Normand S, Kageyama M. 2008 Glacial refugia of temperate trees in Europe: insights from species distribution modelling. *J. Ecol.* **96**, 1117–1127. (doi:10.1111/j.1365-2745.2008.01422.x)

13. Hewitt GM. 2004 Genetic consequences of climatic oscillations in the Quaternary. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* **359**, 183–195. (doi:10.1098/rstb.2003.1388)

14. Wallis GP, Waters JM, Upton P, Craw D. 2016 Transverse Alpine Speciation Driven by Glaciation. *Trends Ecol. Evol.* **31**, 916–926. (doi:10.1016/j.tree.2016.08.009)

15. Morelli TL *et al.* 2016 Managing Climate Change Refugia for Climate Adaptation. *PLOS ONE* **11**, e0159909. (doi:10.1371/journal.pone.0159909)

16. Keppel G, Mokany K, Wardell-Johnson GW, Phillips BL, Welbergen JA, Reside AE. 2015 The capacity of
refugia for conservation planning under climate change. *Front. Ecol. Environ.* **13**, 106–112.
(doi:10.1890/140055)

17. Stewart JR, Lister AM, Barnes I, Dalén L. 2010 Refugia revisited: individualistic responses of species in
space and time. *Proc. R. Soc. B Biol. Sci.* **277**, 661–671. (doi:10.1098/rspb.2009.1272)

18. Keppel G, Wardell-Johnson GW. 2012 Refugia: keys to climate change management. *Glob. Change Biol.*
18, 2389–2391. (doi:10.1111/j.1365-2486.2012.02729.x)

19. Allan JR, Watson JEM, Marco MD, Possingham HP, Venter O. 2019 Hotspots of human impact on
threatened terrestrial vertebrates. *PLoS Biol.* , 18.

20. Monsarrat S, Novellie P, Rushworth I, Kerley GIH. 2019 Shifted distribution baselines: neglecting long-
term biodiversity records risks overlooking potentially suitable habitat for conservation management.
bioRxiv , 565929. (doi:10.1101/565929)

21. Faurby S, Araújo MB. 2018 Anthropogenic range contractions bias species climate change forecasts.
Nat. Clim. Change **8**, 252–256. (doi:10.1038/s41558-018-0089-x)

22. Margules CR, Pressey RL. 2000 Systematic conservation planning. *Nature* **405**, 243–253.

23. Allan JR, Venter O, Watson JEM. 2017 Temporally inter-comparable maps of terrestrial wilderness and
the Last of the Wild. *Sci. Data* **4**, 170187. (doi:10.1038/sdata.2017.187)

24. Dudley N, Jonas H, Nelson F, Parrish J, Pyhälä A, Stolton S, Watson JEM. 2018 The essential role of
other effective area-based conservation measures in achieving big bold conservation targets. *Glob.*
Ecol. Conserv. **15**, e00424. (doi:10.1016/j.gecco.2018.e00424)

25. Galán-Acedo C, Arroyo-Rodríguez V, Andresen E, Verde Arregoitia L, Vega E, Peres CA, Ewers RM. 2019
The conservation value of human-modified landscapes for the world's primates. *Nat. Commun.* **10**.
(doi:10.1038/s41467-018-08139-0)

26. López-Bao JV, Kaczensky P, Linnell JDC, Boitani L, Chapron G. 2015 Carnivore coexistence: Wilderness
not required. *Science* **348**, 871. (doi:10.1126/science.348.6237.871-b)

27. Venter O *et al.* 2016 Global terrestrial Human Footprint maps for 1993 and 2009. *Sci. Data* **3**, 160067.
(doi:10.1038/sdata.2016.67)

28. Gaston KJ. 1991 How Large Is a Species' Geographic Range? *Oikos* **61**, 434. (doi:10.2307/3545251)

29. Rondinini C, Wilson KA, Boitani L, Grantham H, Possingham HP. 2006 Tradeoffs of different types of
species occurrence data for use in systematic conservation planning: Species data for conservation
planning. *Ecol. Lett.* **9**, 1136–1145. (doi:10.1111/j.1461-0248.2006.00970.x)

30. Boitani L, Maiorano L, Baisero D, Falcucci A, Visconti P, Rondinini C. 2011 What spatial data do we need
to develop global mammal conservation strategies? *Philos. Trans. R. Soc. B Biol. Sci.* **366**, 2623–2632.
(doi:10.1098/rstb.2011.0117)

31. Gaston KJ, Fuller RA. 2009 The sizes of species' geographic ranges. *J. Appl. Ecol.* **46**, 1–9.
(doi:10.1111/j.1365-2664.2008.01596.x)

32. Bombi P, Luiselli L, D'Amen M. 2011 When the method for mapping species matters: defining priority
areas for conservation of African freshwater turtles: Prioritization of African turtles. *Divers. Distrib.* **17**,
581–592. (doi:10.1111/j.1472-4642.2011.00769.x)

33. Rodríguez JP, Brotons L, Bustamante J, Seoane J. 2007 The application of predictive modelling of
species distribution to biodiversity conservation. *Divers. Distrib.* **13**, 243–251. (doi:10.1111/j.1472-
4642.2007.00356.x)

34. Elith J, Leathwick JR. 2009 Species Distribution Models: Ecological Explanation and Prediction Across
Space and Time. *Annu. Rev. Ecol. Evol. Syst.* **40**, 677–697.
(doi:10.1146/annurev.ecolsys.110308.120159)

35. Kearney M, Porter W. 2009 Mechanistic niche modelling: combining physiological and spatial data to
predict species' ranges. *Ecol. Lett.* **12**, 334–350. (doi:10.1111/j.1461-0248.2008.01277.x)

36. Araújo MB *et al.* 2019 Standards for distribution models in biodiversity assessments. *Sci. Adv.* **5**,
eaat4858. (doi:10.1126/sciadv.aat4858)

37. Pearson RG, Dawson TP. 2003 Predicting the impacts of climate change on the distribution of species:
are bioclimate envelope models useful? *Glob. Ecol. Biogeogr.* **12**, 361–371.

38. Kearney MR, Wintle BA, Porter WP. 2010 Correlative and mechanistic models of species distribution
provide congruent forecasts under climate change: Congruence of correlative and mechanistic
distribution models. *Conserv. Lett.* **3**, 203–213. (doi:10.1111/j.1755-263X.2010.00097.x)

39. Jarvie S, Svenning J-C. 2018 Using species distribution modelling to determine opportunities for trophic
rewilding under future scenarios of climate change. *Philos. Trans. R. Soc. B Biol. Sci.* **373**, 10.
(doi:http://dx.doi.org/10.1098/rstb.2017.0446)

40. Nüchel J, Bøcher PK, Xiao W, Zhu A-X, Svenning J-C. 2018 Snub-nosed monkeys (Rhinopithecus):
potential distribution and its implication for conservation. *Biodivers. Conserv.* **27**, 1517–1538.
(doi:10.1007/s10531-018-1507-0)

41. Soga M, Gaston KJ. 2018 Shifting baseline syndrome: causes, consequences, and implications. *Front.*
Ecol. Environ. **16**, 263–263.

42. Turvey ST, Crees JJ, Di Fonzo MMI. 2015 Historical data as a baseline for conservation: reconstructing
long-term faunal extinction dynamics in Late Imperial–modern China. *Proc. R. Soc. B Biol. Sci.* **282**,
20151299. (doi:10.1098/rspb.2015.1299)

43. Eckert LE, Ban NC, Frid A, McGreer M. 2018 Diving back in time: Extending historical baselines for
yelloweye rockfish with Indigenous knowledge. *Aquat. Conserv. Mar. Freshw. Ecosyst.* **28**, 158–166.
(doi:10.1002/aqc.2834)

44. Lentini PE, Stirnemann IA, Stojanovic D, Worthy TH, Stein JA. 2018 Using fossil records to inform
reintroduction of the kakapo as a refugee species. *Biol. Conserv.* **217**, 157–165.
(doi:10.1016/j.biocon.2017.10.027)

45. Nogués-Bravo D. 2009 Predicting the past distribution of species climatic niches. *Glob. Ecol. Biogeogr.*
18, 521–531. (doi:10.1111/j.1466-8238.2009.00476.x)

46. Faurby S, Svenning J-C. 2015 Historic and prehistoric human-driven extinctions have reshaped global
mammal diversity patterns. *Divers. Distrib.* **21**, 1155–1166. (doi:10.1111/ddi.12369)

47. Hoban S, Dawson A, Robinson JD, Smith AB, Strand AE. 2019 Inference of biogeographic history by
formally integrating distinct lines of evidence: genetic, environmental niche and fossil. *Ecography* ,
ecog.04327. (doi:10.1111/ecog.04327)

48. Laliberte AS, Ripple WJ. 2004 Range Contractions of North American Carnivores and Ungulates.
BioScience **54**, 123. (doi:10.1641/0006-3568(2004)054[0123:RCONAC]2.0.CO;2)

49. Crees JJ, Collen B, Turvey ST. 2019 Bias, incompleteness, and the "known unknowns" in the Holocene
faunal record. *Philos. Trans. R. Soc. B Biol. Sci.* **This issue**.

50. Monsarrat S, Kerley GIH. 2018 Charismatic species of the past: Biases in reporting of large mammals in
historical written sources. *Biol. Conserv.* **223**, 68–75. (doi:10.1016/j.biocon.2018.04.036)

51. Monsarrat S, Boshoff AF, Kerley GIH. 2018 Accessibility maps as a tool to predict sampling bias in
historical biodiversity occurrence records. *Ecography* (doi:10.1111/ecog.03944)

52. Newbold T. 2010 Applications and limitations of museum data for conservation and ecology, with
particular attention to species distribution models. *Prog. Phys. Geogr.* **34**, 3–22.
(doi:10.1177/0309133309355630)

53. Pooley S. 2018 Descent with modification: Critical use of historical evidence for conservation. *Conserv.*
Lett. , e12437. (doi:10.1111/conl.12437)

54. Schipper J *et al.* 2008 The status of the world's land and marine mammals: diversity, threat, and
knowledge. *Science* **322**, 225–230. (doi:10.1126/science.1165115)

55. Böhm M *et al.* 2013 The conservation status of the world's reptiles. *Biol. Conserv.* **157**, 372–385.
(doi:10.1016/j.biocon.2012.07.015)

56. Ripple WJ *et al.* 2019 Are we eating the world's megafauna to extinction? *Conserv. Lett.* , e12627.
(doi:10.1111/conl.12627)

57. Salafsky N *et al.* 2008 A Standard Lexicon for Biodiversity Conservation: Unified Classifications of
Threats and Actions: *Classifications of Threats & Actions*. *Conserv. Biol.* **22**, 897–911.
(doi:10.1111/j.1523-1739.2008.00937.x)

58. IUCN. 2018 The IUCN Red List of Threatened Species. Version 2018-2. http://www.iucnredlist.org.

59. Venter O *et al.* 2016 Data from: Global terrestrial Human Footprint maps for 1993 and 2009.
(doi:10.5061/dryad.052q5.2)

60. Christina M. Kennedy, James Oakleaf, David M. Theobald, Sharon Baruch-Mordo, Joseph Kiesecker.
2018 Global Human Modification. (doi:10.6084/m9.figshare.7283087.v1)

61. Oakleaf JR, Kennedy CM, Baruch-Mordo S, West PC, Gerber JS, Jarvis L, Kiesecker J. 2015 A World at
Risk: Aggregating Development Trends to Forecast Global Habitat Conversion. *PLOS ONE* **10**, e0138334.
(doi:10.1371/journal.pone.0138334)

62. Queiroz C, Beilin R, Folke C, Lindborg R. 2014 Farmland abandonment: threat or opportunity for
biodiversity conservation? A global review. *Front. Ecol. Environ.* **12**, 288–296. (doi:10.1890/120348)

63. Ceaușu S, Hofmann M, Navarro LM, Carver S, Verburg PH, Pereira HM. 2015 Mapping opportunities
and challenges for rewilding in Europe. *Conserv. Biol.* **29**, 1017–1027. (doi:10.1111/cobi.12533)

64. Hewson J, Crema S, González-Roglich M, Tabor K, Harvey C. 2019 New 1 km Resolution Datasets of
Global and Regional Risks of Tree Cover Loss. *Land* **8**, 14. (doi:10.3390/land8010014)

65. van Vuuren DP, Lucas PL, Hilderink H. 2007 Downscaling drivers of global environmental change:
Enabling use of global SRES scenarios at the national and grid levels. *Glob. Environ. Change* **17**, 114–
130. (doi:10.1016/j.gloenvcha.2006.04.004)

66. Seto K, Güneralp B, Hutyra LR. 2015 Global Grid of Probabilities of Urban Expansion to 2030.

67. Kerr JT, Ostrovsky M. 2003 From space to species: ecological applications for remote sensing. *Trends*
Ecol. Evol. **18**, 299–305. (doi:10.1016/S0169-5347(03)00071-5)

68. Escobar LE, Awan MN, Qiao H. 2015 Anthropogenic disturbance and habitat loss for the red-listed
Asiatic black bear (Ursus thibetanus): Using ecological niche modeling and nighttime light satellite
imagery. *Biol. Conserv.* **191**, 400–407. (doi:10.1016/j.biocon.2015.06.040)

69. Pe'er G *et al.* 2014 Toward better application of minimum area requirements in conservation planning.
Biol. Conserv. **170**, 92–102. (doi:10.1016/j.biocon.2013.12.011)

70. Wintle BA *et al.* 2019 Global synthesis of conservation studies reveals the importance of small habitat
patches for biodiversity. *Proc. Natl. Acad. Sci.* **116**, 909–914. (doi:10.1073/pnas.1813051115)

71. Svenning J-C *et al.* 2016 Science for a wilder Anthropocene: Synthesis and future directions for trophic
rewilding research. *Proc. Natl. Acad. Sci.* **113**, 898–906.

72. Perino A *et al.* 2019 Rewilding complex ecosystems. *Science* **364**, eaav5570.
(doi:10.1126/science.aav5570)

73. Seddon PJ, Griffiths CJ, Soorae PS, Armstrong DP. 2014 Reversing defaunation: Restoring species in a
changing world. *Science* **345**, 406–412. (doi:10.1126/science.1251818)

74. Fahrig L, work(s): GMR. 1994 Conservation of Fragmented Populations. *Conserv. Biol.* **8**, 50–59.

75. Kerley GIH, Kowalczyk R, Cromsigt JPGM. 2012 Conservation implications of the refugee species
concept and the European bison: king of the forest or refugee in a marginal habitat? *Ecography* **35**,
519–529. (doi:10.1111/j.1600-0587.2011.07146.x)

76. Han H *et al.* 2019 Diet Evolution and Habitat Contraction of Giant Pandas via Stable Isotope Analysis.
Curr. Biol. **29**, 664-669.e2. (doi:10.1016/j.cub.2018.12.051)

77. Valiente-Banuet A *et al.* 2015 Beyond species loss: the extinction of ecological interactions in a
changing world. *Funct. Ecol.* **29**, 299–307. (doi:10.1111/1365-2435.12356)

78. Burgess ND *et al.* 2007 Correlations among species distributions, human density and human
infrastructure across the high biodiversity tropical mountains of Africa. *Biol. Conserv.* **134**, 164–177.
(doi:10.1016/j.biocon.2006.08.024)

79. Hobbs RJ, Higgs E, Harris JA. 2009 Novel ecosystems: implications for conservation and restoration.
Trends Ecol. Evol. **24**, 599–605. (doi:10.1016/j.tree.2009.05.012)

80. IUCN/SSC. 2013 Guidelines for Reintroductions and Other Conservation Translocations. , viiii + 57 pp.

81. Bruskotter JT, Shelby LB. 2010 Human Dimensions of Large Carnivore Conservation and Management:
Introduction to the Special Issue. *Hum. Dimens. Wildl.* **15**, 311–314.
(doi:10.1080/10871209.2010.508068)

82. Ceauşu S, Graves RA, Killion AK, Svenning J-C, Carter NH. 2019 Governing trade-offs in ecosystem
services and disservices to achieve human–wildlife coexistence. *Conserv. Biol.* **33**, 543–553.
(doi:10.1111/cobi.13241)

83. Moritz C, Agudo R. 2013 The Future of Species Under Climate Change: Resilience or Decline? *Science*
341, 504–508. (doi:10.1126/science.1237190)

84. Erasmus BFN, Freitag S, Gaston KJ, Erasmus BH, Jaarsveld AS van. 1999 Scale and conservation planning
in the real world. *Proc. R. Soc. Lond. B Biol. Sci.* **266**, 315–319. (doi:10.1098/rspb.1999.0640)

85. EC—European Commission. 2000 Managing Natura 2000 sites: the provisions of Article 6 of the
'Habitats' Directive 92/43/EEC. Office for Official Publications of the European Communities:
Luxembourg.

86. Albuquerque FS, Assunção-Albuquerque MJT, Cayuela L, Zamora R, Benito BM. 2013 European Bird
distribution is "well" represented by Special Protected Areas: Mission accomplished? *Biol. Conserv.*
159, 45–50. (doi:10.1016/j.biocon.2012.10.012)

87. Bladt J, Strange N, Abildtrup J, Svenning J-C, Skov F. 2009 Conservation efficiency of geopolitical
coordination in the EU. *J. Nat. Conserv.* **17**, 72–86. (doi:10.1016/j.jnc.2008.12.003)

578