Partitioning plant spectral diversity into alpha and beta components

- 3 Etienne Laliberté^{1,2*}, Anna K. Schweiger^{1,2}, Pierre Legendre²
- 4
- 5 ¹Institut de recherche en biologie végétale, Université de Montréal, 4101 Sherbrooke Est,
- 6 Montréal, Québec H1X 2B2, Canada
- 7 ²Département de sciences biologiques, Université de Montréal, C. P. 6128, succursale Centre-
- 8 ville, Montréal, Québec H3C 3J7, Canada
- 9 *Corresponding author. Email: <u>etienne.laliberte@umontreal.ca</u>
- 10 Phone: +1 514 343 6132
- 11
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25 **ABSTRACT**

26 Plant spectral diversity — how plants differentially interact with solar radiation — is an integrator 27 of plant chemical, structural, and taxonomic diversity that can be remotely sensed. We propose 28 to measure spectral diversity as spectral variance, which allows the partitioning of the spectral 29 diversity of a region, called spectral gamma (v) diversity, into additive alpha (α ; within 30 communities) and beta (β ; among communities) components. Our method calculates the 31 contributions of individual bands or spectral features to spectral y-, β -, and α -diversity, as well 32 as the contributions of individual plant communities to spectral diversity. We present two case 33 studies illustrating how our approach can identify "hotspots" of spectral α -diversity within a 34 region, and discover spectrally unique areas that contribute strongly to β -diversity. Partitioning 35 spectral diversity and mapping its spatial components has many applications for conservation 36 since high local diversity and distinctiveness in composition are two key criteria used to 37 determine the ecological value of ecosystems.

38 INTRODUCTION

39 Major environmental changes, including land-use change, climate change, and invasive species 40 are altering the Earth's biodiversity. The rapid rate and broad extent of those changes far 41 exceed our capacity to monitor them via field-based sampling alone. This calls for the 42 development of new remote sensing approaches that can provide rapid estimates of biodiversity 43 over broad regions (Pereira et al. 2013; Turner 2014; Bush et al. 2017). For terrestrial plants, 44 imaging spectroscopy is emerging as the most promising remote sensing method for estimating 45 biodiversity (Féret & Asner 2014; Wang & Gamon 2019). This is because its high spectral 46 resolution allows plant species to be discriminated from one another, while also enabling the 47 determination of ecologically important foliar functional traits (Asner & Martin 2009; Ustin et al. 48 2009). 49 For every pixel of an aerial image, imaging spectroscopy measures reflected solar radiation in

50 tens to hundreds of contiguous, narrow (~10 nm wide) wavelength bands, usually covering all or 51 part of the visible to shortwave infrared range (400–2500 nm) of the electromagnetic spectrum. 52 Leaf "spectral signatures" of plants provide unique expressions among species of how solar 53 radiation interacts with photosynthetic pigments, water, proteins, as well as structural and 54 chemical defense compounds, and thus represent the evolution of plant adaptations to different 55 environmental conditions (Cavender-Bares et al. 2016; McManus et al. 2016). At the crown 56 scale, these spectral signatures are further influenced by architectural traits due to scattering of 57 photons within canopies (Asner 1998; Ollinger 2010). Therefore, plant spectral diversity is 58 emerging as an integrator of plant chemical, structural and taxonomic diversity that can be 59 remotely sensed (Cavender-Bares et al. 2017; Schweiger et al. 2018; see also Appendix S1 in 60 Supporting Information).

61 One of the most influential conceptual developments in community ecology has been 62 Whittaker's (1960, 1972) suggestion to partition biodiversity across space into α , β , and γ 63 components. Originally, α diversity was defined as the species diversity within communities, and 64 β as the variation in species composition *among* communities; together, α - and β -diversities 65 iointly determined *v*-diversity, which is the species diversity across an entire region of interest. In 66 this paper, we transpose this foundational ecological concept from species diversity to spectral 67 diversity (Fig. 1). This requires that the spatial resolution of the imagery matches the size of the 68 object of interest (Woodcock & Strahler 1987), meaning that pixels should be approximately 69 equal or smaller than the size of an average canopy plant. At such fine spatial resolutions, the 70 relationship between spectral and taxonomic diversity is strongest (Wang et al. 2018a) and 71 imaging spectroscopy can provide direct, spatially explicit estimates of plant alpha (α ; within 72 community) diversity (Féret & Asner 2014; Wang et al. 2018b), and can detect changes in plant 73 community composition across landscapes (Draper et al. 2019). The ability to generate wall-to-74 wall, high-resolution maps of canopy plant diversity across entire regions brings tremendous 75 benefits for biodiversity science and conservation (e.g., Asner et al. 2017); however, conceptual 76 and methodological challenges remain, especially with regard to β -diversity estimation (Rocchini 77 et al. 2010, 2018).

78 Spectral diversity is sometimes called spectral heterogeneity or spectral variability (Rocchini et 79 al. 2010), and has been defined as spatial variation in spectral reflectance (Rocchini et al. 2010; 80 Ustin & Gamon 2010; Gholizadeh et al. 2018; Wang & Gamon 2019). Intuitively, spectral 81 diversity can be conceptualized as multivariate dispersion, for which there are various statistical 82 measures highlighting different aspects of spectral diversity. For example, Wang et al. (2018a) 83 used the average coefficient of variation (CV) of each band for a set of pixels, whereas Rocchini 84 et al. (2010) used the mean distance from the spectral centroid; we note that the latter has also 85 been proposed as a measure of functional diversity in multivariate trait space (Laliberté &

86 Legendre 2010). However, none of the currently used metrics allow the partitioning of spectral 87 diversity into its α (within communities) and β (among communities) components (Fig. 1). 88 Here we propose to use the *spectral variance* among image pixels as a measure of spectral 89 diversity. Our approach builds on that of Legendre and De Cáceres (2013) for species inventory 90 data, adapts it to spectral data, and extends it to jointly consider α -, β -, and y-diversity. Casting 91 spectral diversity as spectral variance has a number of benefits: 92 1. the classical partitioning of sums of squares allows us to partition spectral γ -diversity into 93 additive spectral α - and β -diversity components (Fig. 1), from which the relative 94 importance of local and regional processes regulating spectral diversity across a region 95 of interest can be inferred; 96 2. it allows us to estimate the contributions of individual plots or communities to spectral β -97 diversity, highlighting areas that are spectrally distinct within the broader region; 98 it allows us to calculate the contributions of individual bands or spectral features to 99 spectral γ -, β - or α -diversity (Fig. 2), providing information about the underlying biological 100 traits driving spectral diversity; 101 4. it is easily implemented in software packages in a computationally efficient way, which is 102 important when dealing with high-volume image data; 103 5. it provides a direct link to other statistical procedures based on least squares (e.g., 104 MANOVA, multiple linear regression, canonical redundancy analysis, K-means 105 partitioning). 106 After describing the theory behind our spectral diversity partitioning approach, we illustrate it 107 using a simulation. We then apply our method to imaging spectroscopy data collected over the

108 Bartlett Experimental Forest by the National Ecological Observatory Network (NEON) Airborne

109 Observation Platform (AOP; Kampe *et al.* 2010). The R code and data for our analyses are

110 available online (<u>https://github.com/elaliberte/specdiv</u>).

111 PARTITIONING SPECTRAL DIVERSITY

112 Size and shape of spatial units

113 Partitioning spectral y-diversity into its α and β components first requires defining the extent of 114 the region of interest (Fig. 1). Delineating the region of interest is relatively straightforward since 115 it corresponds to the region over which imagery is acquired or a subset thereof. Delineating the 116 size and shape of communities across the region of interest, however, is more difficult. What 117 constitutes an ecological community has been the subject of considerable debate (see review 118 by Ricklefs 2008). Generally, a community is defined as "a group of organisms representing 119 multiple species living in a specified place and time" (Vellend 2010). This definition implies that 120 a community must be larger than the size of an individual organism, but how much larger will 121 depend on the objectives of the study. For the purpose of this work, we focus on communities of 122 canopy plants, because these are the organisms that can be seen in aerial images. We use 123 "community" in the sense of "sampling unit" in vegetation surveys, which can be defined as the 124 area in which the species composition of the vegetation type of interest is adequately 125 represented (Mueller-Dombois & Ellenberg 1974).

Setting the size of a community to the size of typical inventory plot for a given ecosystem type facilitates interpretation as this is the sampling unit that field ecologists are familiar with. For example, forest inventory plots often measure 20 m × 20 m (Fig. 1), which is large enough to include several canopy trees. However, we recognize that setting fixed and regularly shaped boundaries to delineate communities is artificial (Ricklefs 2008), and point out that community size and shape can be changed in the analysis.

132 Spectral gamma (y) diversity

133 Let **Y** = $[y_{ij}]$ be a matrix containing the positions, along the p axes defining the spectral space 134 (column vectors $\mathbf{y}_1, \mathbf{y}_2, \dots, \mathbf{y}_p$ of \mathbf{Y}), of *n* pixels (row vectors $\mathbf{x}_1, \mathbf{x}_2, \dots, \mathbf{x}_n$ of \mathbf{Y}) in a region of 135 interest (Fig. 2). We use indices i and j to denote rows (pixels) and columns (axes) of matrix Y. 136 respectively. The p axes could be all or a subset of the original spectral bands, a set of 137 vegetation indices calculated from selected spectral bands (Bannari et al. 1995), or a set of p 138 uncorrelated spectral features extracted using dimensionality reduction methods such as 139 principal component analysis (PCA). We use PCA in this section and in our case studies and 140 point out the mathematical relationships between the principal components (PCs) and spectral 141 variation below. We use the general term variation for sums of squares (SS, an abbreviation for 142 "sum of the squared deviations from the mean"), and reserve the term *variance* when talking 143 about spectral diversity (SD). 144 We refer to the total spectral diversity of the entire region as spectral y-diversity (SD_y). SD_y is

measured by the total variance of **Y**, or Var(**Y**). This is done by first computing for every pixel and spectral feature y_{ij} the squared deviations s_{ij} from the average pixel (across the whole region) in terms of spectral reflectance, i.e. the column means of **Y**:

$$s_{ij} = \left(y_{ij} - \overline{y}_j\right)^2. \tag{1}$$

148 The total sum of squares (SS) of matrix **Y** is calculated by summing all s_{ij} :

$$SS_{\gamma} = \sum_{i=1}^{n} \sum_{j=1}^{p} s_{ij}.$$
 (2)

149 Contrary to SS_{γ} , SD_{γ} is scaled by the number of pixels in the region, such that SD_{γ} of regions 150 containing different numbers of pixels can be compared with one another:

$$SD_{\gamma} = Var(\mathbf{Y}) = SS_{\gamma}/(n-1).$$
 (3)

We note that for calculating the joint SD_{γ} of adjacent regions, their SS_{γ} statistics can be added and divided by the total number of pixels minus one, but their region-level SD_{γ} statistics cannot be added directly.

One might be interested in determining the individual contribution of the *j*th spectral feature to SS_{γ}. We call this the *feature contribution to spectral* γ *-diversity* or FCSD_{γ , *j*} (Fig. 2), which can be calculated from the sum of squares of the *j*th feature:

$$SS_{\gamma,j} = \sum_{i=1}^{n} s_{ij}.$$
(4)

157 Dividing $SS_{\gamma,j}$ by (n - 1) gives the variance of the *j*th feature, or $Var(\mathbf{y}_j)$. FCSD_{γ,j} can then be 158 calculated as:

$$FCSD_{\gamma,j} = Var(\mathbf{y}_j) / Var(\mathbf{Y}) = SS_{\gamma,j} / SS_{\gamma}.$$
(5)

- 159 If the *p* features are principal components from PCA scaling type 1, then the $FCSD_{y,j}$ values 160 correspond to their relative eigenvalues. We note that $FCSD_{y,j}$ cannot be mapped because the 161 contribution of each spectral feature applies to the region as a whole.
- 162 Likewise, one might wish to estimate the individual contribution of the *i*th pixel within the region 163 to SD_{γ}. We refer to this as the *local contribution to spectral* γ *-diversity*, or LCSD_{γ ,*i*}, which is 164 calculated as:

$$LCSD_{\gamma,i} = SS_{\gamma,i}/SS_{\gamma}$$
(6)

165 where

$$SS_{\gamma,i} = \sum_{j=1}^{p} s_{ij}.$$
(7)

We note that $LCSD_{\gamma,i}$ indices are important visual elements in PCA ordination: each $LCSD_{\gamma,i}$ value corresponds to the squared distance from one pixel to the centroid in the *p*-dimensional PCA ordination plot. In addition, the $LCSD_{\gamma,i}$ can be plotted on maps since one value is associated with every pixel in the image. Doing so indicates which pixels are most spectrally 170 dissimilar from the mean pixel of the region in spectral feature space. We note that the $SS_{y,i}$ and

- 171 LCSD_{y,i} indices are additive. The indices from adjacent pixels within an area of interest, for
- example an individual tree, can be added up, such that their sums represent the local
- 173 contributions of the area of interest to SS_{γ} and SD_{γ}. LCSD indices are also useful when
- 174 computed at the community scale (i.e. $LCSD_{\beta}$), because they then correspond to the ecological
- 175 concept of β -diversity; see "Spectral beta (β) diversity" below.

176 Partitioning the total sum of squares

- 177 Partitioning the sum of squares forms the basis of a series of classic statistical approaches
- 178 based on least squares, such as the analysis of variance (ANOVA). From these methods, it is
- 179 well known that the total sum of squares of a matrix **Y** (SS_{total}) can be partitioned into additive

among-group (SS_{among}) and within-group (SS_{within}) components:

$$SS_{total} = SS_{among} + SS_{within}.$$
 (8)

In linear regression analysis, we talk about the SS explained by the regression equation and the
residual variation. These two components sum to the total sum of squares.

Using the same indices as in the previous section, the ANOVA relationship can be expressedas:

$$\sum_{i=1}^{n} \sum_{j=1}^{p} (y_{ij} - \bar{y}_j)^2 = \sum_{k=1}^{q} \sum_{j=1}^{p} m \left(\hat{y}_{kj} - \bar{y}_j \right)^2 + \sum_{i=1}^{m} \sum_{k=1}^{q} \sum_{j=1}^{p} (y_{ij} - \hat{y}_{kj})^2$$
(9)

where *q* is the number of groups, \hat{y}_{kj} is the arithmetic mean of the *j*th variable (column) for the *k*th group:

$$\hat{y}_{kj} = \left(\sum_{i=1}^{m} y_{ijk}\right)/m \tag{10}$$

and where *m* is the number of samples (rows) in each group *k*; an important assumption here is
that *m* is equal in each group. The proof of this theorem can be found in standard statistics
textbooks and is therefore not shown here.

- 190 In the next two sections, we apply Equation 9 to partition the total sum of squares of a region
- 191 SS_{γ} into additive among- (β) and within-group (α) components from which spectral β and α -
- 192 diversity can be calculated directly.

193 Spectral beta (β) diversity

194 Let us divide **Y** into *q* groups of *m* spatially contiguous pixels, where each group corresponds to 195 a local community (e.g., a vegetation survey plot); n = q m. Here, we assume that each of these communities corresponds to a square of equal area, which is \sqrt{m} pixels wide; with this setup, 196 197 each community is represented by the same number of pixels. We will present later our 198 suggestion to use a rarefaction procedure to handle situations where *m* differs among groups. 199 Spectral β -diversity, or SD_{β}, represents the degree to which the *q* communities within a region 200 differ from each other in terms of spectral composition. We note that SD_{β} is a non-directional 201 measure of β -diversity sensu Anderson et al. (2011). To calculate SD_{β}, we first compute the 202 squared deviations ski of the kth community from the average pixel of the region in terms of 203 spectral reflectance, i.e. the column means of **Y** across all variables *j*:

$$\mathbf{s}_{kj} = \left(\hat{y}_{kj} - \bar{y}_j\right)^2 \tag{11}$$

where \hat{y}_{kj} is the arithmetic mean of the *k*th community (i.e. the community centroid) for the *j*th spectral feature (Equation 10).

206 The sum of squares associated with each community k (SS_{β,k}) is:

$$SS_{\beta,k} = \sum_{j=1}^{p} m \, s_{kj}. \tag{12}$$

207 The total sum of squares of the β component (SS $_{\beta}$) is:

$$SS_{\beta} = \sum_{k=1}^{q} SS_{\beta,k} \tag{13}$$

208 from which SD_{β} is calculated as:

$$SD_{\beta} = SS_{\beta}/(n-1).$$
(14)

209 The contribution of each community *k* to SD_{β} , which we call the *local contribution to spectral* β -

210 *diversity* (LCSD_{β,k}), can be computed by the following ratio of sum of squares:

$$LCSD_{\beta,k} = SS_{\beta,k} / SS_{\beta}.$$
 (15)

- Finally, one can compute the *feature contribution to spectral* β -diversity or FCSD_{β ,j} of the *j*th
- 212 spectral feature as:

$$FCSD_{\beta,j} = SS_{\beta,j} / SS_{\beta}$$
(16)

213 where

$$SS_{\beta,j} = \sum_{k=1}^{q} m \, s_{kj}.\tag{17}$$

214 We note here that $LCSD_{\beta,k}$ can be mapped because each community *k* has its own $LCSD_{\beta}$

value. On the other hand, $FCSD_{\beta,j}$ or SD_{β} cannot be mapped because they refer to the region as a whole.

217 Spectral alpha (α) diversity

218 Spectral α -diversity, or SD_{α}, is the degree to which neighbouring pixels within a local community

- 219 differ spectrally from each other. Contrary to SD_{β} and SD_{γ} , which apply to the entire region, SD_{α}
- is defined at the community level. Therefore, we denote SD_{α} by the index *k*, $SD_{\alpha,k}$, since it is
- measured for each community k. To calculate $SD_{\alpha,k}$, we first compute for every pixel and

spectral feature per community y_{ijk} the squared deviations s_{ijk} from the mean pixel spectrum of

the *k*th community for each spectral feature or column of **Y**:

$$s_{ijk} = (y_{ijk} - \hat{y}_{kj})^2.$$
 (18)

224 The sum of squares associated with the *j*th spectral feature of community *k* is:

$$SS_{\alpha,jk} = \sum_{i=1}^{m} s_{ijk},$$
(19)

and the total sum of squares for community *k* is:

$$SS_{\alpha,k} = \sum_{j=1}^{p} \sum_{i=1}^{m} s_{ijk}.$$
(20)

SD_{*a,k*} is obtained by dividing SS_{*a,k*} by (m - 1), where *m* is the number of pixels within one

227 community, to make it comparable with other communities differing in their numbers of pixels:

$$SD_{\alpha,k} = SS_{\alpha,k}/(m-1).$$
(21)

228 The total sum of squares of the α -component for all q communities within the entire region is:

$$SS_{\alpha} = \sum_{k=1}^{q} SS_{\alpha,k}.$$
 (22)

Importantly, following Equations 8 and 9, SS_{α} and SS_{β} are linked to SS_{γ} by the relationship:

$$SS_{\gamma} = SS_{\beta} + SS_{\alpha}.$$
 (23)

230 Therefore, SS_{α} and SS_{β} can be used directly to determine the relative importance of the α and β

231 components to spectral γ -diversity.

Finally, the *contribution of the j*th feature *to the spectral* α *-diversity* of the *k*th community, which we call FCSD_{α , jk}, can be computed as:

$$FCSD_{\alpha,jk} = SS_{\alpha,jk} / SS_{\alpha,k}.$$
(24)

These $FCSD_{\alpha,jk}$ values can be mapped and give us useful information about the origin of

235 spectral α -diversity across different communities.

236 CASE STUDY 1: SIMULATED REGIONS

237 To illustrate our approach, we first use leaf spectra data to simulate imagery (Appendix S2). 238 This removes much of the complexity associated with real imagery, where one has to deal with 239 much higher numbers of pixels, varying illumination and sensor viewing geometry, and 240 presence of shaded and non-vegetated pixels. We simulated two regions with equal spectral v-241 diversity, but contrasting spectral β - and α -diversities (Fig. 3). Each region is composed of 25 × 242 25 pixels, populated with leaf-level spectra of three temperate tree species (i.e. Populus 243 deltoides W. Bartram ex Marshall subsp. deltoides Marsh, P. tremuloides Michaux, and Betula 244 alleghaniensis Britton) measured in the field on 15 individual plants (Fig. S1). These 25 × 25 245 pixels regions are equally split into 25 communities, each composed of 5 × 5 pixels. 246 For both scenarios, we calculated the SS across the entire region (SS_v) , partitioned SS_v into its 247 β and α components, and calculated spectral γ - β -, and α -diversity (Fig. 4a). As spectral 248 features (columns of Y) we used the first three PCs (using type-I scaling in PCA), which 249 together explained >97% of the total variation in spectral reflectance. As expected, spectral v-250 diversity was equal for both scenarios (Table 1), whether expressed as the total sum of squares 251 $(SS_v = 1.66)$, or standardized by n - 1 pixels $(SD_v = 0.0027)$. In addition, in the high β -diversity 252 scenario, spectral variation among communities (SS_{β}, ~84%) largely exceeded spectral 253 variation within communities (SS_a, ~16%), whereas in the low β -diversity scenario SS_{β} was 254 much lower (~5%) than SS_{α} (~95%) (Table 1). 255 Next, we determined the local contributions of individual communities to spectral β -diversity

(LCSD_{β}). In the high β -diversity scenario (Fig. 4b, left panel), communities 12 and 21 (numbered as in Fig. 3) contributed the most to spectral β -diversity. These were the only two plots (out of 25) containing spectra of *Populus tremuloides*. In other words, these two plots had the most distinctive spectral composition compared to other communities. By contrast, in the low β -

260 diversity scenario, community 16 was the most spectrally distinct community, something that 261 could not be easily detected by examining this scenario visually (Fig. 3, right panel). As 262 illustrated here, it is important to note that a region with low SD_{β} can still have individual 263 communities showing high LCSD_{β} values, because LCSD_{β} values are *proportions* of SD_{β}. 264 We then estimated the contributions of individual spectral features to spectral diversity (FCSD) 265 for each scenario. For spectral y-diversity (total variance of the region), FCSD_y declined 266 progressively from the first to the third PC (Table 1). As mentioned previously, $FCSD_{\nu}$ values are 267 equal to the relative PCA eigenvalues of the spectral feature. Likewise, for spectral β -diversity, 268 the contribution from the first to subsequent PCs decreased in both scenarios (Table 1). The 269 relative contributions of individual spectral features to β -diversity were fairly similar in both 270 regions, even though they differed considerably in spectral β -diversity. For spectral α -diversity, 271 however, FCSD_a values differed noticeably among the two scenarios (Fig. 4c–d). In the low α -272 diversity scenario (Fig. 4c-d, left column), PC 2 contributed more strongly to the spectral α -273 diversity of most communities than PC 1, whereas the opposite was true for the high α -diversity 274 scenario (Fig. 4c–d, right column). The FCSD_{α} values were not expected to decrease in a 275 monotonic way since α -diversity is orthogonal to y-diversity and the PCs are those of y-, not of 276 *α*-diversity.

We note that SS, SD and LCSD indices are exactly the same whether using the original spectral bands or all PCs, because PCA type-I scaling preserves the Euclidean distance among objects (e.g., image pixels in spectral space). Conversely, the equations developed in this paper hold and can be used directly with the original band data. However, FCSD values would change when using the original spectral bands instead of PCs, since this would then indicate the relative contributions of individual spectral bands (instead of PCs) to spectral diversity.

283 CASE STUDY 2: NEON IMAGERY

284 Next, we applied our method for partitioning spectral diversity to imaging spectroscopy data 285 collected by NEON's Airborne Observation Platform (AOP: Kampe et al. 2010) over the Bartlett 286 Experimental Forest (https://www.neonscience.org/field-sites/field-sites-map/BART). In this case 287 study, we used a scene measuring 280 m (east-west) x 1000 m (north-south), acquired in 288 August 2017. Spectral data were processed to surface reflectance and subsampled to 1-m pixel 289 size by NEON. Our workflow is illustrated in Figure 2. 290 For spectral diversity calculations we selected a community (i.e. plot) size of 40 m × 40 m, which 291 is the base plot size used by NEON. We used rarefaction to standardize the number of pixels 292 per community used for analysis. We used a normalized difference vegetation index (NDVI) 293 threshold of >0.8 to identify the minimum number of vegetated pixels across all plots in the 294 image (termed $m_{\rm min}$), which was 1474 (= 92% of the 1600 pixels per community). We randomly 295 selected m_{\min} pixels per plot, and applied our spectral diversity partitioning approach to all 296 selected pixels. The rarefaction was repeated 30 times and results were averaged across all 30 297 repeats (Fig. 2). Alternatively, one could take the median value instead of the mean if 298 distributions are skewed.

Our analyses revealed that spectral α -diversity in this forested landscape accounted for 77% of the spectral γ -diversity, whereas β -diversity accounted for the remaining 23% (Table 1). In other words, there is considerably more spectral diversity within individual 40 m × 40 m communities than among communities in this forest. Figure 5 illustrates how spectral diversity is spatially structured. Two areas contribute strongly to spectral β -diversity (LCSD $_{\beta}$, darker colours in Fig. 5). The tree communities in these areas are more spectrally dissimilar from the average community than communities with lower LCSD $_{\beta}$ (lighter colours in Fig. 5).

For completeness, we evaluated the effects of shadows and community size on spectral diversity calculations (Appendix S3). We found that removing shadows had little influence on spectral diversity (Figs. S4–S6) and that results remained remarkably stable for plots ranging from 20 m × 20 m (400 m²) to 140 m x 140 m (almost 2 ha) in size (Figs. S7–S9).

310 **DISCUSSION**

311 In this paper, we proposed a new method for partitioning plant spectral y-diversity (i.e. the 312 spectral diversity of a region) into additive α - (within community) and β -diversity (among 313 community) components. Our approach builds on a method for partitioning β -diversity initially 314 designed for community data (Legendre & Cáceres 2013), adapts it to spectral data and, 315 importantly, extends it to include α , β and γ components. Partitioning spectral diversity can bring 316 new insights and generate new hypotheses about the origins and maintenance of plant spectral 317 diversity across regions. For instance, high spectral β -diversity could result from turnover in 318 plant species and/or functional trait composition across environmental gradients (e.g., soil 319 properties, hydrology), whereas high spectral α -diversity might result from local biotic 320 interactions among co-occurring plants (e.g., resource partitioning, conspecific negative density 321 dependence). Mapping spectral indices such as LCSD_{β} and SD_{α} could be used as a biodiversity 322 "discovery tool" to design targeted field sampling campaigns to test such hypotheses, e.g., by 323 comparing community composition and diversity in areas with high and low LCSD_{β} and SD_{α} 324 values, respectively (Fig. 5).

Partitioning spectral diversity allows the determination of the spectral features contributing most strongly to spectral α -, β - or γ -diversity (FCSD), which helps in understanding the underlying biological traits driving spectral diversity at different spatial scales. In our case studies, the spectral features were principal components (PCs), which are linear combinations of the original wavelength bands. As such, the individual contributions of all wavelength bands to each

330 spectral feature can be retrieved. The bands, in turn, can be linked to specific plant properties, 331 since the biophysical and biological causes of spectral variation across spectral regions and for 332 specific absorption features of molecules are reasonably well understood (Gates et al. 1965; 333 Curran 1989; Asner 1998; Kokaly et al. 2009; Ustin et al. 2009). Identifying the traits contributing 334 most strongly to spectral α -diversity might inform us about how co-occurring species are 335 partitioning resources at the local scale, whereas identifying the traits contributing most strongly 336 to spectral β -diversity might reveal important mechanisms driving changes in community 337 composition across environmental gradients.

338 Partitioning plant spectral diversity and mapping its spatial components has applications in 339 biodiversity management. Indeed, managers often need to estimate the ecological value of 340 different ecosystems over large regions, for example to prioritize conservation or restoration 341 efforts. However, access to field data might be limited. Using imaging spectroscopy data, our 342 approach of partitioning spectral diversity allows the identification of areas with high spectral α -343 diversity, which likely coincide with local "hotspots" of taxonomic and/or functional trait diversity. 344 Further, high LCSD_{β} values indicates areas with rare spectral composition, i.e., containing 345 communities that are most spectrally dissimilar from the average community within the region of 346 interest. Given that species spectral dissimilarity is linked to their functional and phylogenetic 347 dissimilarity (Schweiger et al. 2018), spectrally rare communities can be expected to have rare 348 taxonomic and/or functional composition, either because they harbor uncommon species, or 349 rare combinations of common species.

Our approach measures spectral variance directly (Fig. 2), which is in contrast to other studies that have prior to deriving biodiversity metrics first translated remotely-sensed spectra into plant species (e.g., Féret & Asner 2013), "spectral species" (Féret & Asner 2014), or plant functional traits (e.g., Dahlin *et al.* 2013; Schneider *et al.* 2017). While spectral diversity does not isolate any particular facet of plant biodiversity (e.g., taxonomic, chemical, structural), it integrates all of

these facets (Schweiger *et al.* 2018; Appendix S1). From a practical perspective, casting spectral diversity as spectral variance depends on fewer user decisions compared to other approaches (e.g., selecting the number of clusters for classifying spectral species, selecting the plant traits and modelling approach to predict traits from spectra). This makes spectral diversity easily comparable across different regions. Therefore, maps of SD_{α} and LCSD_{β} could be ideal candidates for biodiversity products from remotely sensed spectral imagery.

361 **Comparison with other approaches**

362 Much of the interest in measuring spectral diversity from remote sensing data stems from the 363 spectral variation hypothesis (Palmer et al. 2002), stating that the spatial variation in spectral 364 reflectance expresses overall variation of the environment. As areas of high environmental 365 variation often harbour more species than areas with low environmental variation, spectral 366 variation across space can potentially uncover botanically interesting areas (Palmer et al. 2002). 367 However, spectral diversity has been predominantly used to investigate relationships between 368 plant spectra and taxonomic units at the α - and γ -diversity scale, whereas the β component has 369 received less attention (Rocchini et al. 2018).

370 Historically. Landsat satellites were instrumental for spurring large-scale biodiversity studies. 371 Early sensors contained few spectral bands; thus, a large body of literature deals with using 372 NDVI for predicting and mapping taxonomic diversity (Gould 2000; see review by Pettorelli et al. 373 2005). Recent advances in sensor technology, particularly increased spectral resolution, have 374 led to a variety of approaches to calculate spectral α -diversity (Rocchini *et al.* 2010). This 375 includes metrics such as the standard deviation or coefficient of variation of spectral indices 376 (Oindo & Skidmore 2002), or spectral bands among pixels (Hall et al. 2010; Gholizadeh et al. 377 2018; Wang et al. 2018a), the convex hull volume of pixels in spectral feature space (Dahlin 378 2016), the mean distance of pixels from the spectral centroid (Rocchini et al. 2010), the number

of spectrally distinct clusters or spectral species in ordination space (Féret & Asner 2014), and
diversity metrics based on dissimilarity matrices among species spectra or image pixels
(Schweiger *et al.* 2018). Of these, our method is most similar to the mean distance to the
spectral centroid (Rocchini *et al.* 2010). The difference is that we square the individual distances
to the spectral centroid; doing so allows us to partition sums of squares into additive
components (Equation 9).

385 Fewer studies have considered spectral β -diversity (Rocchini *et al.* 2018). One approach for 386 studying β -diversity using spectra has been to combine ordination scores of species inventories 387 with spectral data in multivariate models to predict the positions of pixels with unknown species 388 composition in species-ordination-space (Schmidtlein et al. 2007). This method and some of its 389 variants (Rocchini *et al.* 2018) do not measure spectral β -diversity *per se*, but instead use 390 spectra to estimate changes in community composition across the landscape. Rao's guadratic 391 entropy has been suggested as a measure of spectral β -diversity, based on the dissimilarity 392 among image pixels within a moving window (Rocchini et al. 2018). However, a moving window 393 approach expresses spectral β -diversity for many small sub-regions independently from one 394 another and does not estimate the spectral β -diversity of the region as a whole. Another 395 approach for studying spectral β -diversity has been to measure the pairwise dissimilarity in the 396 composition of spectral species among mapping units, and to re-project those pairwise 397 dissimilarities onto an RGB colour space (Féret & Asner 2014). This method yields a useful map 398 showing changes in spectral composition across the region, similar to our mapping of the first 399 three PCs in Figure 5, but it does not calculate spectral diversity.

400 Methodological considerations

A number of methodological aspects should be considered before applying our approach to
imaging spectroscopy data. These include: (1) the choice of a brightness normalization

procedure, (2) whether all or a subset of the wavelength bands, or spectral features, should be
used, (3) masking non-vegetated pixels, or not, (4) determining community size, and (5)
deciding on the scaling type (i.e. type I or II; Legendre & Legendre 2012) if using PCA as a
spectral feature extraction method. We discuss these methodological points in detail in
Appendix S4.

408 CONCLUSION

409 Plant spectral diversity is emerging as an integrator of chemical, structural, and taxonomic 410 aspects of plant biodiversity, which can be remotely sensed (Cavender-Bares et al. 2017). 411 Partitioning plant spectral diversity using our approach can help us to better understand and 412 generate new hypotheses about the origins of, and the processes that drive, biodiversity 413 variation across regions. Given the rapid and broad extent of current environmental changes, 414 remote sensing of plant biodiversity over large regions is more important than ever (Turner 415 2014; Wang & Gamon 2019). Our approach can identify local α -diversity hotspots as well as 416 unique areas contributing strongly to β -diversity – two central facets of biodiversity. 417 Our approach is timely since current technological developments in high-resolution UAV 418 imaging spectroscopy will make this technology more accessible to ecologists in the coming 419 years (Aasen et al. 2018; Arroyo-Mora et al. 2019). For example, the Canadian Airborne 420 Biodiversity Observatory (www.caboscience.org) is developing UAV spectroscopy to understand 421 how plant biodiversity is responding to major environmental changes across Canada. We 422 anticipate that a growing number of ecologists will embrace this transformative technology for 423 mapping plant biodiversity. In fact, a wealth of moderate-resolution imaging spectroscopy data 424 are already freely available for a wide range of ecosystems across the United States as part of 425 the NEON program (www.neonscience.org). Partitioning spectral diversity could become a 426 useful tool for remotely sensing plant biodiversity from these new data sources.

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556 **TABLES**

- 557 **Table 1** Partitioning spectral diversity into additive components for the two simulated regions
- 558 (Figs. 3–4) and for the NEON imagery (Fig. 5).

		High β , low α	Low β , high α	NEON imagery
		21222324251617181920111213141567891012345	21 22 23 24 25 16 17 18 19 20 11 12 13 14 15 6 7 8 9 10 1 2 3 4 5	
Sums of squares (SS)				
SSγ		1.66	1.66	740.24
SS_{β} (% of SS_{γ})		1.40 (84.4%)	0.08 (4.8%)	168.85 (22.8%)
SS_{α} (% of SS_{γ})		0.26 (15.6%)	1.58 (95.2%)	571.39 (77.2%)
Spectral diversity (SD)				
SDγ		0.0027	0.0027	0.0029
SDβ		0.0022	0.00013	0.00065
\overline{SD}_{α}		0.00043	0.0026	0.0022
Feature contribution to SD (FCSD)				
FCSD _y	PC 1	0.800	0.800	0.690

	PC 2	0.157	0.157	0.154
	PC 3	0.023	0.023	0.090
FCSD _β	PC 1	0.906	0.859	0.520
	PC 2	0.079	0.108	0.329
	PC 3	0.013	0.019	0.092
$\overline{\text{FCSD}}_{\alpha}$	PC 1	0.259	0.798	0.733
	PC 2	0.513	0.158	0.108
	PC 3	0.050	0.022	0.092

560 **FIGURE LEGENDS**

561 **Figure 1** Partitioning plant spectral y-diversity into additive β and α components. A region of 562 interest is split into a number of communities of a specific size and shape (here, 20 m × 20 m 563 squares, representing standard forest inventory plots). Spectral y-diversity refers to the total 564 spectral diversity in the entire region calculated from pixel-level reflectance. The β component 565 corresponds to spectral diversity among communities, with similar colours sharing more similar 566 spectral composition. The α component refers to spectral diversity *within* individual 567 communities. The left-most panel is a true colour (red-green-blue, RGB) image of an area of 568 Bartlett Experimental Forest; colours for the other panels were obtained using the reflectance of 569 different wavelength bands (R = 779 nm, G = 639 nm, B = 2301 nm), followed by linear 570 stretching.

571 Figure 2 Overview of our proposed workflow for partitioning plant spectral diversity. In our 572 NEON case study, spectral data pre-processing included removing atmospheric water 573 absorption bands (wavelengths between 1340–1455 nm and 1790–1955 nm) and noisy regions 574 of the spectrum (wavelengths <400 nm and >2400 nm), and applying a Savitzky-Golay filter 575 (order = 3, size = 7) to every pixel in the image to remove high-frequency noise. We masked all 576 pixels with normalized difference vegetation index (NDVI) values <0.8, and brightness-577 normalized all spectra (Feilhauer et al. 2010). Then, we performed a PCA with type I-scaling, 578 and visually inspected the first 17 PCs which together accounted for >99% of the total spectral 579 variation among all pixels (Fig. S3). Only the first five PCs showed meaningful biological spatial 580 patterns and were retained for spectral diversity measurements; PCs 6-17 were excluded 581 based on visual inspection as they expressed artefacts from image acquisition and processing 582 (Fig. S3). For illustration purposes, in the diversity partitioning analysis (bottom panel) we show 583 communities composed of only three pixels, whereas in fact we used a community size of 40 ×

584 40 pixels in our NEON case study. For abbreviations see text. *The shade mask is for illustrative 585 purposes and is not applied to the PCs shown in the middle panel.

586 **Figure 3** Two simulated landscapes of equal spectral y-diversity, but with contrasting spectral β -587 diversity and α -diversity. Left: high spectral β -diversity but low α -diversity. Right: low spectral β -588 diversity but high α -diversity. Each landscape is composed of 25 communities (numbered black 589 squares), each composed of 5×5 pixels (smaller coloured squares). The size of each pixel is 590 equivalent to the size of an individual plant and their colour corresponds to one of the 15 leaf 591 spectra (= 3 species × 5 individuals) shown in Figure S1. These colours were set by mapping 592 the scores of the first three principal components (PC) for each spectrum to a red-green-blue 593 (RGB) scale (PC 1 = green, PC 2 = red, PC 3 = blue). We generated the high spectral β -594 diversity but low spectral α -diversity scenario (left panel) by randomly assigning (with 595 replacement) pixels within each community with individual spectra from single species (Fig. S1, 596 bottom row). We selected species identity per community at random using the following 597 probabilities: 0.60 (Betula alleghaniensis, green hues), 0.35 (Populus deltoides, blue hues) and 598 0.05 (*Populus tremuloides*, red hues). In this scenario, spectral β -diversity was high and spectral 599 α -diversity low because interspecific spectral variation (particularly between *Betula* and the two 600 Populus species) was higher than intraspecific spectral variation (Fig. S2). Next, to reduce 601 spectral β -diversity and increase α -diversity while holding γ -diversity constant, we moved the 602 pixels of the left panel to randomly selected positions in the right panel. 603 **Figure 4** (a) Spectral α -diversity (SD_{α}), (b) local contribution to spectral β -diversity (LCSD_{β}), and

604 (c-d) feature contribution to spectral α -diversity (FCSD_{α}) for the first two spectral features (i.e.

605 first two principal components of the brightness-normalized reflectance data) of each community

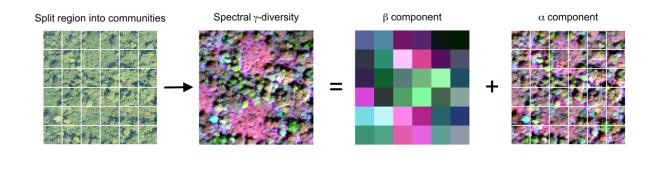
606 in the two simulated regions. PC = principal component.

607 **Figure 5** Partitioning spectral diversity using imaging spectroscopy data acquired by the

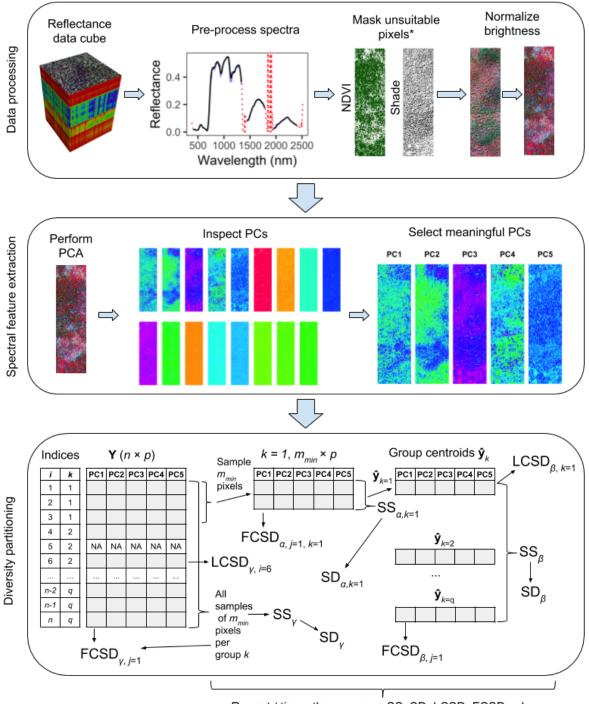
- 608 National Ecological Observatory Network (NEON) over the Bartlett Experimental Forest site.
- From left to right: (1) true colour (red-green-blue, RGB) image with 0.1 m ground resolution, (2)
- 610 false colour image with 1 m resolution based on the first three principal components (PCs) of
- 611 the spectral image cube (PC1 = red, PC2 = green, PC3 = blue), (3) local contribution to spectral
- 612 β-diversity (LCSD_β), and (4) spectral α-diversity (SD_α) of forest communities, each measuring 40
- 613 m × 40 m. For panels 3 and 4, light hues correspond to low, dark hues to high values of $LCSD_{\beta}$
- 614 coefficients and SD_{α} values, respectively.

FIGURES

617 Figure 1



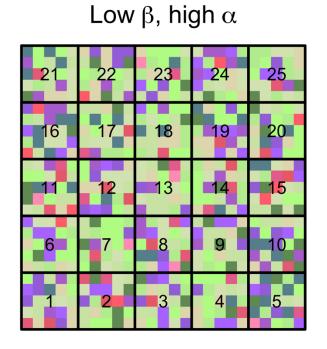
620 Figure 2



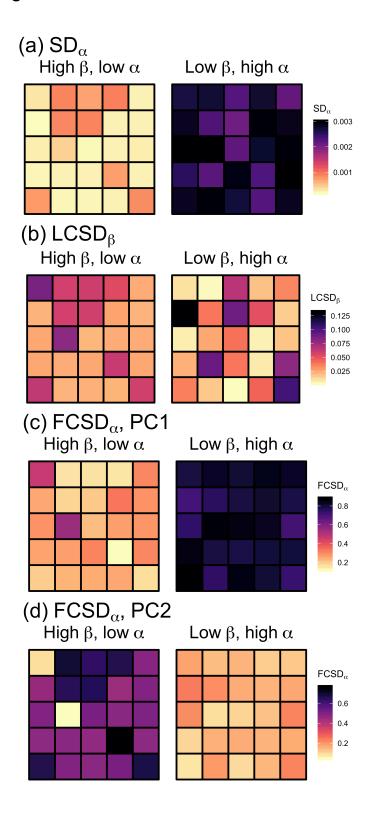
Repeat t times then average SS, SD, LCSD, FCSD values

622 Figure 3

High β , low α



625 Figure 4



627 Figure 5

