

**An environmental DNA tool for monitoring the status of the Critically Endangered Smalltooth Sawfish,
Pristis pectinata, in the Western Atlantic**

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1 **Abstract**

2 The Critically Endangered Smalltooth Sawfish, *Pristis pectinata*, was once widespread in the tropical and
3 subtropical waters of the Atlantic Ocean, but following substantial declines over the past century, the core
4 population is currently confined to southwest Florida in the U.S. and the Bahamas. Recent research and verified
5 public encounter reports suggests that this core population may be stabilizing and, potentially, expanding into
6 formerly occupied areas of their historic range in the Western Atlantic; however, the status of this species in non-
7 core waters is not well understood. Environmental DNA (eDNA) methods provide a relatively cost effective and
8 rapid assessment tool for monitoring species occurrence in aquatic habitats. Here, we have developed an eDNA tool:
9 a species-specific Droplet Digital™ PCR (ddPCR™) assay targeting a 100-base pair portion of the mitochondrial
10 NADH dehydrogenase subunit 2 gene in *P. pectinata*, with the ability to reliably detect as little as 0.25 pg of target
11 DNA. The assay was validated by collecting and analyzing a water sample from known *P. pectinata* nursery habitat
12 in Florida, which was found to contain an average of 11.54 copies of target DNA/μL (SE = 0.72) in the reaction.
13 The assay was then further tested by placing a juvenile sawfish in an *ex situ* tank and analyzing water samples
14 collected at time intervals. The implementation of this eDNA tool into field surveys will provide additional, reliable
15 data to assess species recovery and aid in prioritizing localities beyond the core range in which to focus research and
16 education initiatives.

17

18 **Introduction**

19 Sawfishes are among the most threatened families of marine fishes worldwide (Dulvy et al. 2014), with all five
20 species listed as Critically Endangered or Endangered on the International Union for Conservation of Nature (IUCN)
21 Red List of Threatened Species (Dulvy et al. 2016). All sawfishes have undergone global declines in range and
22 abundance due to direct exploitation, bycatch in fisheries, and habitat loss (Dulvy et al. 2016). These threats are
23 exacerbated by their life history traits (e.g., late maturity, low fecundity, and long life spans), which leave sawfishes
24 susceptible to overexploitation, and makes population recovery a slow process (Stevens et al. 2000; Carlson and
25 Simpfendorfer 2015).

26 The Critically Endangered Smalltooth Sawfish, *Pristis pectinata*, is thought to have experienced the largest
27 global range contraction of all sawfishes and is currently found in less than 20% of its former range (Dulvy et al.
28 2016). Once widespread in the tropical and subtropical waters of the Atlantic Ocean, remaining core population(s)

29 are thought to be limited to the U.S. and the Bahamas (Carlson et al. 2013), making these populations of global
30 conservation significance (Dulvy et al. 2016). Within U.S. waters, *P. pectinata* were historically found from Texas
31 to the Carolinas (Brame et al. 2019) but saw substantial losses in both range and abundance over the past century,
32 with the current population restricted to southwest Florida (SWFL) by the 1980's (Norton et al. 2012).

33 Due to the dramatic declines in range and abundance, *P. pectinata* was listed as Endangered in 2003 under
34 the U.S. Endangered Species Act of 1973 by the National Marine Fisheries Service (NMFS) (NMFS 2003), and a
35 Species Recovery Plan (SRP) was developed to promote recovery and long-term viability of the species in U.S.
36 waters (NMFS 2009; 2018). One characteristic of full species recovery is re-establishment in some or all of the
37 former range (Akçakaya et al. 2018); therefore, the SRP for *P. pectinata* designated 15 recovery regions throughout
38 their historic range in U.S. waters, wherein recovery efforts should occur if species presence is confirmed (NMFS
39 2009; 2018). As a result of over 15 years of U.S. federal and state protections, scientific advances in the
40 understanding of the biology and ecology of the species, and public education initiatives, the core population of *P.*
41 *pectinata* in SWFL is believed to be stabilizing (NMFS 2018). One line of evidence for this potential stabilization is
42 the emergence of relatively recent sawfish encounter reports within formerly occupied parts of their historic range in
43 U.S. waters, including in designated recovery regions (NMFS 2018); however, the status of *P. pectinata* in these
44 non-core areas is unknown.

45 Traditional survey methods for monitoring the status of rare species can be expensive and time-consuming
46 (Lewison et al. 2004). Environmental DNA (eDNA) methods provide a relatively cost effective and rapid
47 assessment tool for monitoring species occurrence in aquatic habitats (Rees et al. 2014; Evans et al. 2017). Water
48 provides a medium for traces of DNA recently shed by organisms (e.g., cellular debris, skin cells, blood, feces,
49 urine), which can be collected and analyzed via genetic assays (Jerde et al. 2011). EDNA has been shown to be as
50 sensitive, and sometimes more sensitive in rare species detections compared to survey methods such as
51 electrofishing (Evans et al. 2017), Baited Remote Underwater Video systems (BRUVs) and underwater visual
52 censuses (UVCs) (Boussarie et al. 2018), and traditional net surveys (Thomsen et al. 2012). EDNA methods also
53 negate the need to capture and handle the target species, making it an ideal tool to assess the presence or absence of
54 a threatened species (Rees et al. 2014). EDNA has been used in targeted, single species detections for a growing
55 number of threatened elasmobranchs, including the Endangered Maugean Skate, *Zearaja maugeana* (Weltz et al.
56 2017), the Vulnerable Great White Shark, *Carcharodon carcharias* (Lafferty et al. 2018), the Vulnerable Chilean

57 Devil Ray, *Mobula tarapacana* (Gargan et al. 2017), and the Critically Endangered Largetooth Sawfish, *Pristis*
58 *pristis* (Simpfendorfer et al. 2016).

59 Here, we develop and validate an eDNA assay to detect the presence of *P. pectinata* DNA in water
60 samples, for use as a tool for monitoring their recovery in the Western Atlantic. This tool will allow scientists and
61 managers to better understand the status of *P. pectinata* in non-core areas and provide quantitative baseline data
62 from which to measure progress towards recovery. Such data can also aid in prioritizing recovery regions in which
63 to focus research and education initiatives, playing an important role in adaptive management strategies as *P.*
64 *pectinata* expands into its former range.

65

66 ***Pristis pectinata* eDNA assay development**

67 Field and laboratory controls

68 To reduce the risk of contamination by exogenous DNA or cross-contamination between samples, rigorous controls
69 were used throughout all stages of this research (see Ficetola et al. 2016; Goldberg et al. 2016; Port et al. 2016;
70 Schweiss et al. In press). All water collection bottles and filtering equipment were cleaned prior to each use using a
71 combination of two methods of sterilization; cleaning with 10% bleach followed by either autoclaving at 120°C for
72 20 min or exposure to UV light for 20 min, depending on the materials. To prevent contamination between the
73 stages of sample processing, water filtration, DNA extractions, and PCR amplifications were conducted in
74 physically isolated laboratories. Furthermore, water samples were filtered in laboratories where contemporary *P.*
75 *pectinata* tissue had never been present (see Deiner et al. 2015). During water filtration and DNA extraction,
76 designated sterile forceps for each sample were used to handle used filters and gloves were changed between
77 samples to prevent cross-contamination between samples (see Pilliod et al. 2013; Goldberg et al. 2016). During
78 DNA extractions and PCR, aerosol barrier pipette tips were used to prevent cross-contamination between samples
79 (Schweiss et al. In Press). Additionally, no positive samples were included in any PCRs due to the risk of
80 contamination from the positive itself, as per ancient DNA (aDNA) PCR protocols (see Mulligan 2005).

81 To test for the possibility of contamination, negative control samples were incorporated into water sample
82 collection and each stage of laboratory processing and analyzed through to PCRs, which were conducted in
83 replicates of five (Jerde et al. 2011; Bakker et al. 2017). To test for contamination in the field, 3 L of autoclaved DI
84 water were brought onto the boat and stored in three sterile 1 L Nalgene® bottles on ice until filtration. To test for

85 contamination during filtration, 3 L of autoclaved DI water were filtered and processed through to PCR. Negative
86 controls for DNA extractions contained no particulate matter or filters, and PCR negatives contained no DNA
87 template. Analysis of all negative control samples, using the optimized protocols described below, found no
88 evidence of target DNA across all PCR replicates.

89

90 Water collection, filtration, and DNA extraction

91 Three liter (L) water samples were collected for all aspects of this study using three sterile, 1 L high-density
92 polyethylene Nalgene[®] bottles, which were kept on ice in pre-cleaned marine coolers until filtration, which occurred
93 within 24 hours of collection. All water samples were vacuum-filtered using 47 millimeter (mm) 0.8 micron (µm)
94 nylon filters, and used filters were rolled and preserved in 95% ethanol at room temperature. Total eDNA was
95 extracted from filters using the QIAGEN[®] DNeasy[®] Blood & Tissue Kit following the Goldberg et al. (2011)
96 protocol incorporating QIAshredder[™] spin columns. The qualities of DNA extracts were visualized using 2%
97 agarose gels and the quantities of DNA were assessed using Thermo Fisher Scientific[™] NanoDrop[™] technology.

98

99 Droplet Digital PCR assay

100 Primers were designed to amplify a 100-base pair (bp) fragment of the mitochondrial NADH dehydrogenase subunit
101 2 (mtDNA ND2) gene in *P. pectinata*, but not in other elasmobranchs that could co-occur with this species in U.S.
102 waters, or in other *Pristis* sawfishes. To design these primers, mtDNA ND2 sequences for *P. pectinata* (GenBank
103 accession no. KP400584.1) and 17 genetically similar exclusion species were downloaded from GenBank (Online
104 Resource 1) and aligned in CodonCode v. 6.0.2 (CodonCode Corporation, Dedham, MA, USA). Forward
105 (PpecND21F: 5'-CTGGTTCACATTGACTCTTAATTTG-3') and reverse (PpecND21R: 5'-
106 GCTACAGCTTCAGCTCTCCTTC-3') primers and an internal PrimeTime[®] double-quenched ZEN[™]/IOWA
107 Black[™] FQ probe labeled with 6-FAM (PpecND2Probe1IBQF: 5'-TACCATAGCCATCATCCATTATTATTC-
108 3') were designed to amplify DNA in only *P. pectinata* by including bp differences in the primers and the probe in
109 all exclusion species (see Online Resource 1). To confirm that the combination of the primers and probe amplified
110 the desired locus, PCRs were conducted using quantitative real-time PCR (qRT-PCR) and total genomic DNA
111 (gDNA) from four *P. pectinata* individuals. Reaction mixtures contained 1.1 µL of extracted DNA (~25 ng/µL), 1X
112 Bio-Rad[®] iTaq[™] universal probe supermix, 900 nanomolar (nM) of each primer, and 170 nM of probe, adjusted to

113 22 μ L using PCR-grade water. Cycling conditions consisted of enzyme activation at 95°C for 10 min, followed by
114 40 cycles of: 94°C for 30 s and 64°C for 2 min, followed by enzyme deactivation at 98°C for 10 min, using a ramp
115 rate of 1°C/s. The resulting amplicon from one *P. pectinata* individual was cleaned using a QIAGEN® QIAquick
116 PCR Purification Kit using the manufacturer’s protocol, with the exception that all centrifugation steps were
117 conducted at 12,000 rpm for 2 min. Forward and reverse sequences were generated using a BigDye™ Terminator
118 v3.1 Cycle Sequencing Kit (Applied Biosystems™, Foster City, CA, USA) on an Applied Biosystems™ 3730XL
119 DNA Analyzer. A consensus sequence was assembled in CodonCode v. 6.0.2 (CodonCode Corporation, Dedham,
120 MA, USA) and its identity was verified as *P. pectinata* using the NCBI BLAST search function; the generated
121 sequence was 99.3% similar to *P. pectinata* GenBank accession no. KP400584.1 (Chen et al. 2016).

122 The PCR reaction and cycling conditions were optimized for the Bio-Rad® QX200™ AutoDG™ Droplet
123 Digital™ PCR System (Instrument no. 773BR1456) by systematically adjusting seven variables (i.e., primer and
124 probe concentrations, cycle number, ramp rate, annealing temperature, denaturation time, and elongation time) to
125 produce positive results with high relative fluorescence units (RFUs) and little to no “droplet rain” (i.e., droplets, or
126 clusters of droplets, that lie between the positive and negative droplet bands on the ddPCR™ scatter plot) (see
127 Online Resource 2). Optimized ddPCR™ reaction mixtures contained 1.1 μ L of extracted DNA, 1X Bio-Rad®
128 ddPCR™ supermix for probes (no deoxyuridine triphosphate (dUTP)), 900 nanomolar (nM) of each primer, and 170
129 nM of probe, adjusted to 22 μ L using PCR-grade water. Optimal ddPCR™ cycling conditions were enzyme
130 activation at 95°C for 10 min, followed by 40 cycles of: 94°C for 30 s and 64°C for 2 min, with a final enzyme
131 deactivation step at 98°C for 10 min, using a ramp rate of 1°C/s. To ensure the assay was species-specific for *P.*
132 *pectinata* in U.S. waters, the optimized ddPCR™ reaction and cycling conditions were tested using 0.20 ng/ μ L
133 gDNA derived from fin clips from four *P. pectinata* individuals and one individual for each of 12 representative
134 exclusion species (Table 1). The target DNA fragment was amplified in all five ddPCR™ replicates for each *P.*
135 *pectinata* individual but was not amplified in any of the ddPCR™ replicates for any representative species from five
136 genetically similar ray genera and two shark genera that could co-occur with *P. pectinata*, or in other *Pristis*
137 sawfishes.

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141 Data analysis

142 Data were analyzed using three criteria for positive *P. pectinata* detections: 1) droplets fell above a manual threshold
143 (MT) defined for this assay, 2) droplets above the MT fell within the prescribed range of the positive droplet
144 population for this assay (5000–7000 RFUs; Fig. 1), and 3) the concentration of target DNA, determined using Bio-
145 Rad® QuantaSoft™ software using the Rare Event Detection (RED) setting, was at or above the Limit of Detection
146 (LoD) of the assay. Defining an assay-specific MT minimizes the likelihood of incorrectly calling artifact droplets
147 (i.e., droplets that fall above the negative band population in the absence of target DNA; see Online Resource 3) as
148 positive detections (e.g., Hunter et al. 2017). To define an appropriate MT for the *P. pectinata* eDNA assay, a No
149 Template Control (NTC) plate with no target DNA was analyzed on the ddPCR™ platform, using the described
150 reaction and cycling conditions. The highest amplitude of the artifact droplets was 2,700 RFUs; therefore, 3000
151 RFUs was chosen to adopt a more conservative approach to minimize the risk of a calling a false positive.

152 To determine the LoD of the assay, ddPCR™ reactions were performed using gDNA from three *P.*
153 *pectinata* individuals with a 6-fold series of 10X dilutions from starting concentrations of 20 ng/μL (i.e., 1:10 to
154 1:1,000,000). Target DNA was reliably detected in all replicates for all individuals up to the 1:10,000 dilutions, but
155 not in the 1:100,000 dilutions. The standard error of the 1:1,000,000 also overlapped with zero (Fig. 2a), making
156 detection at this concentration unreliable. To further refine the LoD, ddPCR™ reactions were performed on
157 subsequent 3-fold series of 2X dilutions from the 1:10,000 dilutions (Fig. 2b). The LoD of this assay was found to
158 be the 1:80,000 dilutions, corresponding to 0.25 pg of target DNA in the reaction (Fig. 2c). The standard errors of
159 the 1:80,000 dilutions did not include zero, or overlap with the 1:100,000 dilutions; so using the average number of
160 copies of target DNA/μL in the 1:80,000 dilutions and applying the lower standard error as the relaxed detection
161 threshold (see Baker et al. 2018), the LoD of the assay was determined to be 0.08 copies/μL.

162

163 **Validation of the *Pristis pectinata* eDNA assay**

164 To validate the ddPCR™ assay, positive *P. pectinata* eDNA samples were acquired via analysis of a water sample
165 from known habitat and through a tank experiment. To collect these positive water samples, a pre-cleaned ~160 L
166 tank was filled with ambient surface water from known *P. pectinata* nursery habitat in the Caloosahatchee River,
167 Florida, approximately 330 m outside of the Harbour Isles Marina. A 3 L water sample was immediately collected
168 from the tank to assess whether *P. pectinata* eDNA was present in the Caloosahatchee River water. One juvenile

169 female *P. pectinata*, measuring 786 mm stretched total length, was captured in a gill net inside the Harbour Isles
170 Marina and placed into the tank. An aerator was added to the tank and dissolved oxygen and water temperature were
171 monitored for the duration of the experiment. A 3 L water sample was collected from the tank immediately after the
172 juvenile was added (time zero) and again after 30 min. All water samples were filtered, DNA extracted, run on
173 ddPCR™, and analyzed using the methods developed in this study.

174 Applying all three criteria for a positive detection of target DNA, the ddPCR reactions containing DNA
175 extracted from ambient water from the nursery contained an average of 11.54 copies/μL (SE = 0.72) of *P. pectinata*
176 eDNA (Fig. 3). The amount of target eDNA increased to 739.4 copies/μL (SE = 38.31) immediately after the
177 juvenile was added to the tank (time zero) and then increased to 3,175.8 copies/μL (SE = 589.3) after 30 min (Fig.
178 3). At 30 min, the large quantity of target DNA isolated from the water sample oversaturated the PCR product,
179 resulting in a high standard error.

180

181 Discussion

182 The developed eDNA assay provides a rapid-assessment tool to conduct targeted surveys to investigate the
183 occurrence and infer the status of *P. pectinata* beyond their contemporary core range. This assay has been validated
184 in the Caloosahatchee River, Florida, where *P. pectinata* is the sole species of sawfish; however, because the assay
185 did not amplify DNA in other *Pristis* sawfishes, it can also be used in locations where other sawfishes have been
186 known to co-occur, at least historically, such as Texas (Brame et al. 2019). It is promising to note that there are no
187 bp differences in the primer or probe sequences designed in this study compared to a mtDNA ND2 sequence for a *P.*
188 *pectinata* collected in Mexico (GenBank accession no. MF682494.1; Diaz-Jaimes et al. 2018), indicating the
189 developed assay should amplify the target gene in this species from other, nearby waters in the Western Atlantic.
190 Use of the assay outside these waters requires careful consideration and, potentially, further testing. MtDNA genes
191 are often variable among populations within a species (Rubinoff et al. 2006); therefore, before using this assay to
192 conduct eDNA surveys in other geographic regions (e.g., Eastern Atlantic), the primers and probe should be tested
193 with *P. pectinata* tissue samples obtained from the local population. Where fresh *P. pectinata* tissue samples are not
194 available for such testing due to the possibility of local extinctions, historic rostra can be used as an alternative
195 source of DNA (Phillips et al. 2009). Finally, the primers and probe developed here were cross-tested with
196 representative species from closely related genera found in U.S. waters; testing with additional exclusion species is

197 required to ensure that the assay remains species-specific in other geographic regions, highlighting the need for local
198 fisheries knowledge (Poulakis and Grubbs 2019).

199 The use of ddPCR™ for single species detections is gaining popularity in eDNA research due to its unparalleled
200 ability to detect minute quantities of target DNA amongst high concentrations of non-target DNA and in the
201 presence of natural inhibitors found in water samples (Evans et al. 2017; Hunter et al. 2018). DdPCR™ assays
202 developed for species such as the Bull Shark, *Carcharhinus leucas* (Schweiss et al. In Press) and Killer Whale,
203 *Orcinus orca* (Baker et al. 2018), have found this platform to be capable of detecting less than 0.5 pg of target DNA
204 in a reaction. Such highly sensitive assays are especially critical in eDNA surveys targeting Critically Endangered or
205 Endangered species, where there can be substantial conservation outcomes based on the results of such surveys
206 (Hunter et al. 2018; Poulakis and Grubbs 2019). For instance, the use of ddPCR™ could reduce the risk of false
207 negatives (i.e., where target DNA is present but not detected) stemming from the use of a less sensitive PCR
208 platform such as conventional or qRT-PCR, which are unlikely to detect such minute quantities of target DNA (Doi
209 et al. 2015). Conservation and management strategies developed on the basis of such false negatives could lead to
210 slower implementation and inadequate protections along with incomplete habitat designations for threatened species,
211 ultimately hindering species recovery.

212 Using the three-criteria approach described here to define positive detections on the ddPCR™ platform
213 provides a rigorous approach to interpret the results of eDNA surveys, reducing the risk of incorrectly calling PCR
214 artifacts as positive species detections (e.g., false positives). For example, using only a MT, an artifact droplet just
215 above the threshold could be incorrectly interpreted as a positive detection. Ensuring the quantity of target DNA is
216 also within the detection capabilities of an assay allows for more robust and confident positive detections. False
217 positives can also result from contamination between eDNA samples or from exogenous DNA. Given the detection
218 capabilities of ddPCR™ assays, strict protocols to prevent contamination (see Goldberg et al. 2016; Schweiss et al.
219 In press) coupled with testing for contamination at every stage in sample processing are critical in producing reliable
220 data from eDNA surveys that may be used as a part of conservation planning. This is especially important when the
221 results of eDNA surveys could be used to prioritize research and management initiatives as well as in the allocation
222 of resources (Poulakis and Grubbs 2019).

223 With a well-designed water sampling regime, strict field and laboratory controls, and a highly sensitive
224 ddPCR™ assay, targeted species eDNA surveys provide a powerful tool to improve our knowledge of the

225 occurrence of *P. pectinata*. The eDNA tool developed here can be used to provide quantitative baseline data in non-
226 core areas from which to measure future progress towards species recovery. Recovery in *P. pectinata* populations is
227 predicted to be a slow process due to their life history characteristics and will be dependent on the mitigation of
228 anthropogenic activities (e.g., accidental fisheries mortalities; Carlson and Simpfendorfer 2015). Range re-expansion
229 during recovery is predicted to begin in locations closest to the core population(s) as a result of spillover from
230 adjacent areas, in a stepping-stone fashion (see Saura et al. 2014). There is, however, the possibility that because
231 female *P. pectinata* have been shown to exhibit philopatry (Feldheim et al. 2017), occurrence and encounter reports
232 of juvenile *P. pectinata* in non-core areas further away from SWFL (e.g. Texas, Mississippi) represent remnant *P.*
233 *pectinata* populations scattered over portions of their former range; under such a scenario, patterns of recovery could
234 be more complex and would ultimately depend on the availability of suitable habitat and the mitigation of threats
235 from anthropogenic activities (Seitz and Poulakis 2006; Poulakis et al. 2011; Norton et al. 2012; Scharer et al.
236 2017). Conducting targeted eDNA surveys for *P. pectinata* across all historically-occupied regions in U.S. waters
237 could not only aid in conservation planning and prioritizing areas for research, but could also increase our
238 understanding of patterns of recovery in a highly threatened marine species.

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253 **References:**

- 254 Akçakaya HR, Bennett EL, Brooks TM, Grace MK, Heath A, Hedges S, Hilton-Taylor C, Hoffman M, Keith DA,
255 Long B, Mallon DP (2018) Quantifying species recovery and conservation success to develop an IUCN
256 Green List of Species. *Conserv Biol* 32:1128-1138. doi: 10.1111/cobi.13112
257
- 258 Baker CS, Steel D, Nieu Kirk S, Klinck H (2018) Environmental DNA (eDNA) from the wake of the whales: droplet
259 digital PCR for detection and species identification. *Front Mar Sci* 5:133.
260 <https://doi.org/10.3389/fmars.2018.00133>
261
- 262 Bakker J, Wangensteen OS, Chapman DD, Boussarie G, Buddo D, Guttridge TL, Hertler H, Moulliot D, Vigliola L,
263 Mariani S (2017) Environmental DNA reveals tropical shark diversity in contrasting levels of
264 anthropogenic impact. *Sci Rep* 7:16886. doi:10.1038/s41598-017-17150-2
265
- 266 Brame AB, Wiley TR, Carlson JK, Fordham SV, Grubbs RD, Osborne J, Scharer RM, Bethea DM, Poulakis GR
267 (2019) Biology, ecology, and status of the Smalltooth Sawfish, *Pristis pectinata*, in the USA. *Endanger*
268 *Species Res* 39:9-23. <https://doi.org/10.3354/esr00952>
269
- 270 Boussarie G, Bakker J, Wangensteen OS, Mariani S, Bonnin L, Juhel J, Kiszka JJ, Kulbicki M, Manel S, Robbins
271 WD, Vigliola L (2018) Environmental DNA illuminates the dark diversity of sharks. *Sci Adv* 4:9661. doi:
272 10.1126/sciadv.aap9661
273
- 274 Carlson JK, Simpfendorfer CA (2015) Recovery potential of smalltooth sawfish, *Pristis pectinata*, in the United
275 States determined using population viability models. *Aquat Conserv Mar Freshw Ecosyst* 25:187-200. doi:
276 10.1002/aqc.2434
277
- 278 Carlson J, Wiley T, Smith K (2013) *Pristis pectinata* (errata version published in 2019). The IUCN Red List of
279 Threatened Species 2013: e.T18175A141791261. Accessed on 24 April 2019.
280
- 281 Carlson JK, Osborne J (2012) Relative abundance of smalltooth sawfish (*Pristis pectinata*) based on the Everglades
282 National Park Creel Survey. NOAA Technical Memorandum NMFS-SEFSC-626.
283
- 284 Chen X, Wiley T, Kyne PM, Fuetry P (2016) Complete mitochondrial genome of the Critically Endangered
285 Smalltooth Sawfish *Pristis pectinata* (Rajiformes: Pristidae). *Mitochondr DNA Part A* 27:3331-3332.
286
- 287 Diaz-Jaimes P, Bonfil R, Palacios-Barreto P, Bolano-Martinez N, Nayona-Vasquez NJ (2018) Mitochondrial
288 genome of the Critically Endangered Smalltooth Sawfish *Pristis pectinata* from Veracruz, Mexico.
289 *Conserv Genet Resour* 10:663-666. doi: 10.007/s12686-017-0896-9
290
- 291 Deiner K, Walser JC, Mächler E, Altermatt F (2015) Choice of capture and extraction methods affect detection of
292 freshwater biodiversity from environmental DNA. *Biol Conserv* 183:53-63.
293 <https://doi.org/10.1016/j.biocon.2014.11.018>
294
- 295 Doi H, Takahara T, Minamoto T, Matsuhashi S, Uchii K, Yamanaka H (2015) Droplet digital polymerase chain
296 reaction (PCR) outperforms real-time PCR in the detection of environmental DNA from an invasive fish
297 species. *Environ Sci Technol* 49:5601-5608. doi:10.1021/acs.est.5b00253
298
- 299 Dulvy NK, Davidson LN, Kyne PM, Simpfendorfer CA, Harrison LR, Carlson JK, Fordham SV (2016) Ghosts of
300 the coast: global extinction risk and conservation of sawfishes. *Aquat Conserv Mar Freshw Ecosyst* 26:134-
301 153. doi: 10.1002/aqc.2525
302
- 303 Dulvy NK, Fowler SL, Musick JA, Cavanagh RD, Kyne PM, Harrison LR, Carlson JK, Davidson LN, Fordham SV,
304 Francis MP, Pollock CM (2014) Extinction risk and conservation of the world's sharks and rays. *eLife*
305 3:e00590. doi:10.7554/eLife.00590
306
307

- 308 Evans NT, Shirey PD, Wieringa JG, Mahon AR, Lamberti GA (2017) Comparative cost and effort of fish
309 distribution detection via environmental DNA analysis and electrofishing. *Fisheries* 42:90-99.
310 <https://doi.org/10.1080/03632415.2017.1276329>
311
- 312 Feldheim KA, Fields AT, Chapman DD, Scharer RM, Poulakis GR (2017) Insights into reproduction and behavior
313 of the smalltooth sawfish *Pristis pectinata*. *Endanger Species Res* 34:463-471.
314 <https://doi.org/10.3354/esr00868>
315
- 316 Ficetola GF, Taberlet P, Coissac E (2016) How to limit false positives in environmental DNA and
317 metabarcoding? *Mol Ecol Resour* 16:604-607. <https://doi-org.lynx.lib.usm.edu/10.1111/1755-0998.12508>
318
- 319 Gargan LM, Morato T, Pham CK, Finarelli JA, Carlsson JE, Carlsson J (2017) Development of a sensitive detection
320 method to survey pelagic biodiversity using eDNA and quantitative PCR: a case study of devil ray at
321 seamounts. *Mar Biol* 164:112. doi:10.1007/s00227-017-3141-x
322
- 323 Goldberg CS, Pilliod DS, Arkle RS, Waits LP (2011) Molecular detection of vertebrates in stream water: a
324 demonstration using Rocky Mountain tailed frogs and Idaho giant salamanders. *PLoS One* 6:e22746.
325 <https://doi.org/10.1371/journal.pone.0022746>
326
- 327 Goldberg CS, Turner CR, Deiner K, Klymus KE, Thomsen PF, Murphy MA, Spear SF, McKee A, Oyler-McCance
328 SJ, Cornman RS, Laramie MB (2016) Critical considerations for the application of environmental DNA
329 methods to detect aquatic species. *Methods Ecol Evol* 7:1299-1307. [https://doi-](https://doi-org.lynx.lib.usm.edu/10.1111/2041-210X.12595)
330 [org.lynx.lib.usm.edu/10.1111/2041-210X.12595](https://doi-org.lynx.lib.usm.edu/10.1111/2041-210X.12595)
331
- 332 Hunter ME, Meigs-Friend G, Ferrante JA, Kamla AT, Dorazio RM, Diagne LK, Luna F, Lanyon
333 JM, Reid JP (2018) Surveys of environmental DNA (eDNA): a new approach to estimate occurrence in
334 vulnerable manatee populations. *Endanger Species Res* 35:101-111. <https://doi.org/10.3354/esr00880>
335
- 336 Hunter ME, Dorazio RM, Butterfield JS, Meigs-Friend G, Nico LG, Ferrante JA (2017) Detection limits of
337 quantitative and digital PCR assays and their influence in presence-absence surveys of environmental
338 DNA. *Mol Ecol Res* 17:221-229. doi: 10.1111/1755-0998.12619
339
- 340 Jerde CL, Mahon AR, Chadderton WL, Lodge DM (2011) "Sight-unseen" detection of rare aquatic species using
341 environmental DNA. *Conserv Lett* 4:150-157. doi: 10.1111/j.1755-263X.2010.00158.x
342
- 343 Lafferty KD, Benesh KC, Mahon AR, Jerde CL, Lowe CG (2018) Detecting southern California's white sharks with
344 environmental DNA. *Front Mar Sci* 5:1-6. doi: 10.3389/fmars.2018.00355
345
- 346 Lewison RL, Crowder LB, Read AJ, Freeman SA (2004) Understanding impacts of fisheries bycatch on marine
347 megafauna. *Trends Ecol Evol* 19:598-604. <https://doi.org/10.1016/j.tree.2004.09.004>
348
- 349 Mulligan CJ (2005) Isolations and analysis of DNA from archaeological, clinical, and natural history specimens.
350 *Methods Enzymol* 395:87-103. [https://doi.org/10.1016/S0076-6879\(05\)95007-6](https://doi.org/10.1016/S0076-6879(05)95007-6)
351
- 352 National Marine Fisheries Service (2003) Endangered and threatened species; final endangered status for a distinct
353 population segment of smalltooth sawfish (*Pristis pectinata*) in the United States. *Federal Register*
354 68:15674-15680.
355
- 356 National Marine Fisheries Service (2009) Recovery Plan for Smalltooth Sawfish (*Pristis pectinata*). Prepared by the
357 Smalltooth Recovery Team for the National Marine Fisheries Service. Silver Spring, Maryland.
358
- 359 National Marine Fisheries Service (2018) Smalltooth Sawfish 5-Year review: summary and evaluation. National
360 Oceanic and Atmospheric Administration, National Marine Fisheries Service, Protected Resources
361 Division, St. Petersburg, FL.
362
363

- 364 Norton SL, Wiley TR, Carlson JK, Frick AL, Poulakis GR, Simpfendorfer CA (2012) Designating critical habitat
365 for juvenile endangered smalltooth sawfish in the United States. *Mar Coast Fish* 4:473-480.
366 <https://doi.org/10.1080/19425120.2012.676606>
367
- 368 Phillips N, Chaplin J, Morgan D, Peverell S (2009) Extraction and amplification of DNA from the dried rostra of
369 sawfishes (Pristidae) for applications in conservation genetics. *Pac Conserv Biol* 15:128-134. doi:
370 10.1071/PC090128
371
- 372 Pilliod DS, Goldberg CS, Arkle RS, Waits LP (2013) Estimating occupancy and abundance of stream amphibians
373 using environmental DNA from filtered water samples. *Can J Fish Aquat Sci* 70:1123-1130. doi:
374 10.1139/cjfas-2013-0047
375
- 376 Port JA, O'Donnell JL, Romero-Maraccini OC, Leary PR, Litvin SY, Nickols KJ, Yamahara KM, Kelly RP (2016)
377 Assessing vertebrate biodiversity in a kelp forest ecosystem using environmental DNA. *Mol Ecol* 25:527-
378 541. doi: 10.1111/mec.13481
379
- 380 Poulakis GR, Grubbs RD (2019) Biology and ecology of sawfishes: global status of research and future outlook.
381 *Endanger Species Res* 39:77-90. <https://doi.org/10.3354/esr00955>.
382
- 383 Poulakis GR, Stevens PW, Timmers AA, Wiley TR, Simpfendorfer CA (2011) Abiotic affinities and spatiotemporal
384 distribution of the endangered smalltooth sawfish, *Pristis pectinata*, in a south-western Florida
385 nursery. *Aquat Conserv Mar Freshw Ecosyst* 62:1165-1177. <https://doi.org/10.1071/MF11008>
386
- 387 Rees HC, Maddison BC, Middleditch DJ, Patmore JR, Gough KC (2014) The detection of aquatic animal species
388 using environmental DNA—a review of eDNA as a survey tool in ecology. *J Appl Ecol* 51:1450-1459.
389 doi:10.1111/1365-2664.12306
390
- 391 Rubinoff D (2006) Utility of mitochondrial DNA barcodes in species conservation. *Conserv Biol* 20:1026-1033.
392 <https://doi.org/10.1111/j.1523-1739.2006.00372.x>
393
- 394 Saura S, Bodin Ö, Fortin MJ (2014) Stepping stones are crucial for species' long-distance dispersal and range
395 expansion through habitat networks. *J Appl Ecol* 51:171-182. <https://doi.org/10.1111/1365-2664.12179>
396
- 397 Seitz JC, Poulakis GR (2006) Anthropogenic effects on the smalltooth sawfish (*Pristis pectinata*) in the United
398 States. *Mar Pollut Bull* 52:1533-1540. <https://doi.org/10.1016/j.marpolbul.2006.07.016>
399
- 400 Simpfendorfer CA, Kyne PM, Noble TH, Goldsbury J, Basiita RK, Lindsay R, Shields A, Perry C, Jerry DR (2016)
401 Environmental DNA detects Critically Endangered largetooth sawfish in the wild. *Endanger Species Res*
402 30:109-116. doi:10.3354/esr00731
403
- 404 Stevens JD, Bonfil R, Dulvy NK, Walker PA (2000) The effects of fishing on sharks, rays, and chimaeras
405 (chondrichthyans), and the implications for marine ecosystems. *ICES J Mar Sci* 57:476-494.
406 doi:10.1006/jmsc.2000.0724
407
- 408 Thomsen PF, Kielgast J, Iversen LL, Møller PR, Rasmussen M, Willerslev E (2012) Detection of a diverse marine
409 fish fauna using environmental DNA from seawater samples. *PLoS One* 7:e41732.
410 <https://doi.org/10.1371/journal.pone.0041732>
411
- 412 Weltz K, Lyle JM, Ovenden J, Morgan JA, Moreno DA, Semmens JM (2017) Applications of environmental DNA
413 to detect an endangered marine skate species in the wild. *PLoS One* 12:e0178124.
414 <https://doi.org/10.1371/journal.pone.0178124>

Table 1 List of 12 exclusion species that were tested to ensure species-specificity of the primers and probe developed for the mitochondrial NADH dehydrogenase subunit 2 (mtDNA ND2) gene in *Pristis pectinata* on the Bio-Rad® QX200™ AutoDG™ Droplet Digital™ PCR System. The country of origin for each tissue sample is included

| Species | Tissue origin |
|---|----------------------|
| Green Sawfish, <i>Pristis zijsron</i> | Australia |
| Dwarf Sawfish, <i>Pristis clavata</i> | Australia |
| Largetooth Sawfish, <i>Pristis pristis</i> | Australia |
| Atlantic Guitarfish, <i>Rhinobatos lentiginosus</i> | United States |
| Atlantic Stingray, <i>Hypanus sabinus</i> | United States |
| Bluntnose Stingray, <i>Hypanus say</i> | United States |
| Cownose Ray, <i>Rhinoptera bonasus</i> | United States |
| Spotted Eagle Ray, <i>Aetobatus narinari</i> | United States |
| Clearnose Skate, <i>Raja eglanteria</i> | United States |
| Roundel Skate, <i>Raja texana</i> | United States |
| Bigeye Thresher Shark, <i>Alopias superciliosus</i> | United States |
| Atlantic Sharpnose Shark, <i>Rhizoprionodon terraenovae</i> | United States |

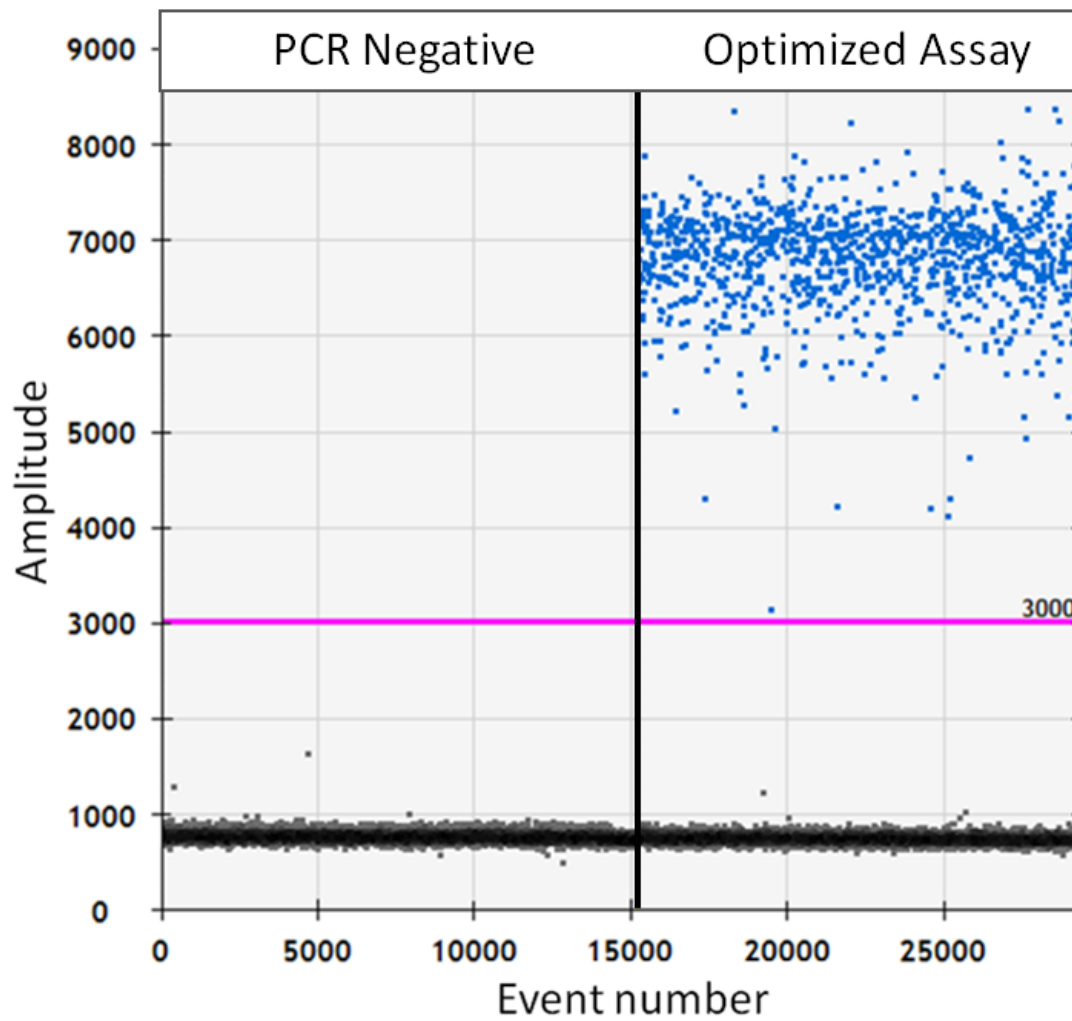


Fig. 1 Raw droplet scatter plot of ddPCRTM products using genomic DNA from one Smalltooth Sawfish, *Pristis pectinata*, individual with the optimized assay conditions with a corresponding negative control. Each droplet in each well was classified as either positive (above 3000) or negative (below 3000) for target DNA based on a manual threshold amplitude of 3000 relative fluorescence units (RFUs); detected using a Bio-Rad[®] QX200TM Droplet Reader, QuantaSoftTM software and RED analysis setting. Each well is separated by vertical lines, and is labeled to correspond with the sample it represents

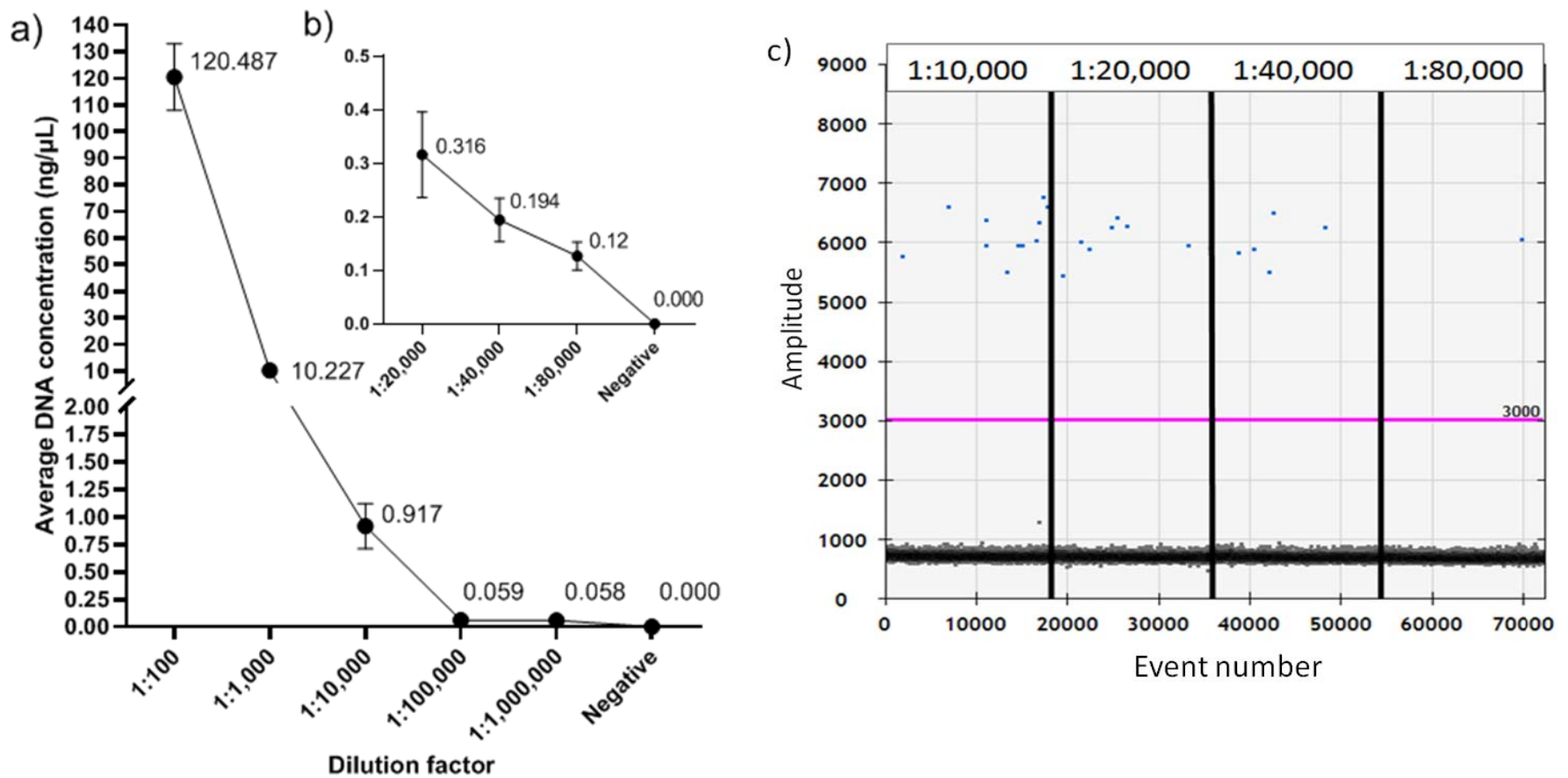


Fig. 2 Average target DNA concentrations (copy number/μL) of the Limit of Detection (LoD) dilution series, using genomic DNA from three Smalltooth Sawfish, *Pristis pectinata*, individuals with five replicates each in a) a 6-fold series of 10X dilutions from a starting concentration of 20 ng/μL, b) a 3-fold series of 2X dilutions from the 1:10,000 dilution, and c) a corresponding raw ddPCR™ scatter plot of serial dilution reactions from one replicate of one Smalltooth Sawfish, *Pristis pectinata*, individual. The Bio-Rad® QX200™ Droplet Reader and QuantaSoft™ software using the RED analysis setting was used across all samples, and each droplet in each well was classified as either positive (above 3000) or negative (below 3000) for target DNA based on a manual threshold amplitude of 3000 relative fluorescence units (RFUs)

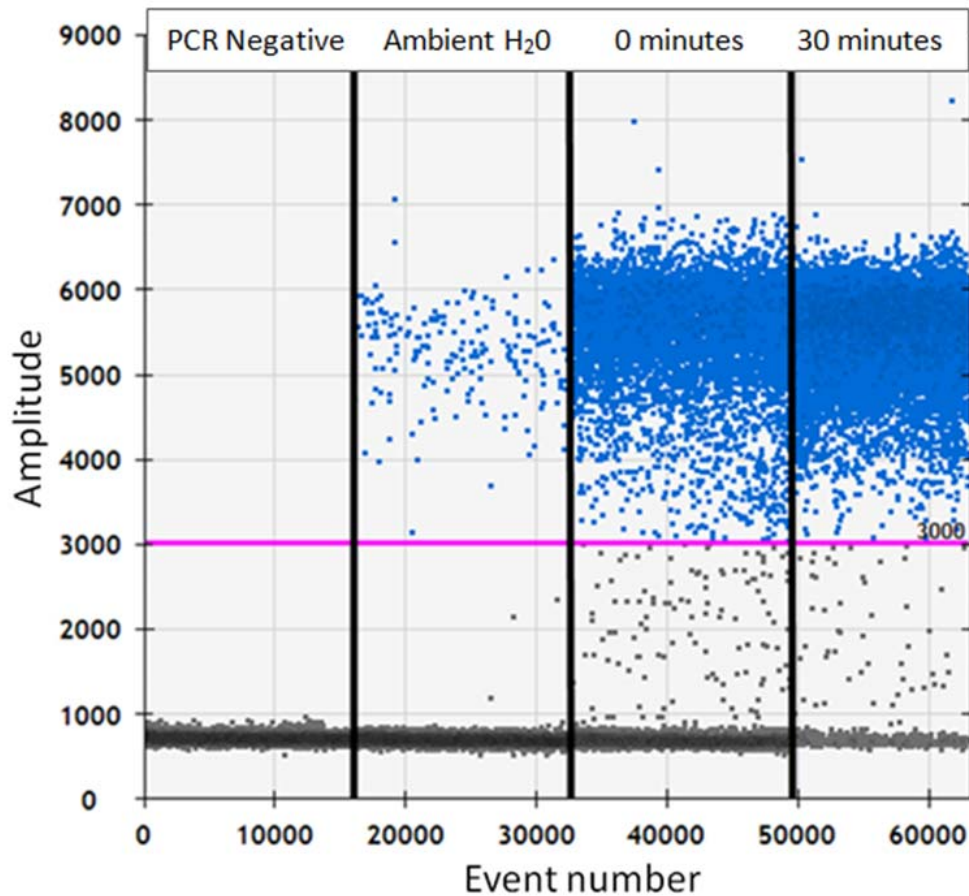


Fig. 3 Raw droplet scatter plot of ddPCRTM serial dilution products from one reaction for each of the Smalltooth Sawfish, *Pristis pectinata*, positive eDNA samples. Ambient water refers to water samples collected from the Caloosahatchee River, known *P. pectinata* nursery habitat. 0 and 30 min reactions correspond to the positive water samples collected from the *ex situ* tank containing a live *P. pectinata*. Each droplet in each well was classified as either positive (above 3000) or negative (below 3000) for target DNA based on a manual threshold amplitude of 3000 relative fluorescence units (RFUs) detected using a Bio-Rad[®] QX200TM Droplet Reader, and QuantaSoftTM software using the RED analysis setting. Each well is separated by vertical lines and is labeled to correspond with the sample or time stage it represents. Note: “Droplet rain” (i.e., droplets, or clusters of droplets, that lie between the positive and negative droplet bands on the ddPCRTM scatter plot) is seen at 30 min due to an oversaturation of target DNA