

1 **Title: ‘Tree demographic performance of European tree**
2 **species at their hot and cold edges.’**

3 **Running title: ‘Demographic performance at climatic**
4 **edges.’**

5 **Georges Kunstler¹, Arnaud Guyennon¹, Sophia Ratcliffe^{2,3}, Nadja R uger^{4,5}, Paloma Ruiz-**
6 **Benito^{6,7}, Dylan Z. Childs⁸, Jonas Dahlgren⁹, Aleksi Lehtonen¹⁰, Wilfried Thuiller¹¹, Chris-**
7 **tian Wirth^{2,4,12}, Miguel A. Zavala⁷, Roberto Salguero-Gomez¹³.**

8 ¹Univ. Grenoble Alpes, Irstea, UR LESSEM, 38000 Grenoble, France.; ²Department of Systematic Botany
9 and Functional Biodiversity, University of Leipzig, Johannisallee 21-23, 04103 Leipzig, Germany.; ³NBN
10 Trust, 14-18 St. Mary’s Gate, Lace Market, Nottingham NG1 1PF, UK.; ⁴German Centre for Integrative Bio-
11 diversity Research (iDiv) Halle-Jena-Leipzig, Deutscher Platz 5e, 04103 Leipzig, Germany.; ⁵Smithsonian
12 Tropical Research Institute, Apartado 0843-03092, Balboa, Anc n, Panama.; ⁶Departamento de Biologia y
13 Geologia, Fisica y Quimica Inorganica, Escuela Superior de Ciencias Experimentales y Tecnologia, Univer-
14 sidad Rey Juan Carlos, Tulipan, 28933, Mostoles (Madrid), Spain.; ⁷Grupo de Ecologia y Restauracion Fore-
15 stal, Departamento de Ciencias de la Vida, Universidad de Alcal , Edificio de Ciencias, Campus Universi-
16 tario, 28805 Alcal  de Henares, Madrid, Spain.; ⁸Department of Animal & Plant Sciences, The University of
17 Sheffield, Sheffield, UK.; ⁹Swedish University of Agricultural Sciences, Ume , 90183 Sweden.; ¹⁰Natural Re-
18 sources Institute Finland (Luke), Latokartanonkaari 9 FI-00790 Helsinki Finland.; ¹¹Univ. Grenoble Alpes,
19 CNRS, Univ. Savoie Mont Blanc, CNRS, LECA, Laboratoire d’Ecologie Alpine, F-38000 Grenoble, France.;
20 ¹²Max-Planck-Institute for Biogeochemistry, Hans-Kn llstr. 10, 07745 Jena, Germany.¹³Department of Zo-
21 ology, University of Oxford, 11a Mansfield Rd OX1 3SZ, Oxford, UK.

22 Species range limits are thought to result from a decline in demographic performance because
23 of unsuitable climate at the edges. However, recent studies reporting contradictory patterns in
24 tree species demographic performance at their edges cast doubt on our ability to predict climate
25 change impacts on species ranges. Here we parameterised integral projection models with climate
26 and competition effects for 27 tree species using forest inventory data from over 90,000 plots across

27 Europe. Then, we predicted growth, survival, lifespan, and passage time – the time to grow to a
28 large size – at the hot and cold edges and compared them to the range centre. We found that while
29 growth and passage time of European tree species are constrained at their cold edge, survival
30 and lifespan are constrained at their hot edge. Our study shows a more complicated picture than
31 previously thought with demographic responses that differ between hot and cold edges.

32 *Keywords:* “demography, IPM, passage time, vital rate, range edge”

33

34 **Introduction**

35 Increasing concerns have emerged regarding potential major redistributions of plant species ranges
36 in the coming decades due to climate change (Zimmermann *et al.* 2013). Indeed, some studies
37 have confirmed that these range shifts are already underway (Chen *et al.* 2011). Range shifts are
38 directly related to changes in demographic rates and population dynamics. Demographic theories
39 propose that if populations are at equilibrium, mean population growth rate (λ) will drop at the
40 species’ range edge ($\lambda < 1$) due to alterations in one or more vital rates contributing to λ (Case *et*
41 *al.* 2005; Holt & Keitt 2005). This prediction is grounded in a long-standing hypothesis in biogeog-
42 raphy, known as the ‘abundance-centre hypothesis’ (hereafter ACH, Brown 1984), which proposes
43 that demographic performance decline at the range edge results in a decrease in abundance. This
44 decline in demographic performance can arise directly because of abiotic constraints (*e.g.* frost)
45 or indirectly because of changes in biotic constraints (*e.g.* competition) (Hargreaves *et al.* 2014;
46 Pironon *et al.* 2017).

47 The number of studies that have directly tested the ACH with field data on population growth
48 rates, and the vital rates that contribute to them, at the centre and the edge of species range is sur-
49 prisingly limited. Nonetheless, these studies showed only weak or contradictory support for the
50 ACH. Transplant experiments have shown that population growth rate or some vital rates tend
51 to decline beyond the edge but not necessarily right at the edge (Hargreaves *et al.* 2014; Lee-Yaw
52 *et al.* 2016). For long-lived organisms such as trees, their generation time rules out transplant ex-
53 periments that would cover their full life cycle. Rather, researchers have to rely on models based

54 on natural population monitoring data (see Purves 2009). Only a few studies have used this ap-
55 proach over the full species range and they found no clear evidence of a decrease in demographic
56 performance at the edge (Purves 2009; Thuiller *et al.* 2014; Csergo *et al.* 2017). Some studies also
57 tested the role of biotic interactions and found only weak evidence for the idea that biotic interac-
58 tions constrain the demographic performance at the edge and that this effect is stronger for edges
59 in productive environments than in unproductive environments (Cahill *et al.* 2014; Hargreaves *et*
60 *al.* 2014; Louthan *et al.* 2015).

61 Our limited understanding of the demographic underpinnings of species ranges in long-lived
62 organisms represents a major handicap to forecast climate change impacts on trees. Yet, trees play
63 a crucial role in the biosphere by sheltering a significant proportion of biodiversity and carbon
64 stocks and contributing to the livelihoods of local populations (van der Plas *et al.* 2018). Existing
65 studies cover only a limited number of species and rather restricted areas of the species ranges,
66 making it difficult to explain what drives the large variability in the demographic performance
67 observed between edges and species. Several characteristics of the edges and species are likely to
68 explain this variability. Firstly, the decline in demographic performance is likely to vary depend-
69 ing on the type of biophysical constraints at the edge (Gaston 2009). For instance, demographic
70 constraints could differ between drought- and cold-limited edges because tolerance to different
71 abiotic stress requires different adaptative strategies (Niinemets & Valladares 2006). Secondly, the
72 decline in demographic performance is likely to vary depending on the vital rate considered (Gas-
73 ton 2009; Hargreaves *et al.* 2014). Finally, the constraints on the demographic performance at the
74 edge are likely to vary with species' physiological strategy (Anderegg *et al.* 2019). These physio-
75 logical differences can be captured by species' climatic optimum and by functional traits related
76 to species physiological climate response, such as wood (Chave *et al.* 2009) or leaf characteristics
77 (Wright *et al.* 2017).

78 Here, we explore these questions at the continental scale for 27 European tree species using for-
79 est inventory data documenting survival and growth of more than 1 million adult trees. First, we
80 fitted survival and growth models to capture climate and competition impacts on these vital rates.
81 Second, we built size-structured population models using integral projection models (IPM) (Ellner
82 *et al.* 2016) to evaluate demographic performance based on four metrics: two vital rates - growth
83 and survival, and two life trajectory metrics - mean lifespan and passage time (time to grow from

84 small to large size). We then used these models to compare demographic performance at the hot
85 or cold edges with the performance at the range centre for each species. Using these metrics we
86 tested the following hypotheses: (1) demographic performance is reduced at the edge compared
87 to the centre for one or more metric; (2) the decline in demographic performance is stronger with
88 competition than without because competition restricts the demographic performance at the edge;
89 (3) different metrics are reduced at the hot and the cold edge due to different biophysical con-
90 straints operating at these edges; demographic performance at the edges depends on (4) species
91 climatic optimum and (5) functional traits related to species' climatic response (wood density, leaf
92 economic spectrum traits, leaf size, and xylem vulnerability to embolism).

93 **Materials and Methods**

94 *Forest inventory*

95 We used the European forest inventory (NFI) data compiled in the FunDivEUROPE project (Baeten
96 *et al.* 2013; Ratcliffe *et al.* 2015). The data covers 91,528 plots and more than 1 million trees in Spain,
97 France, Germany, Sweden and Finland. NFIs record information on individual trees in each plot,
98 including species identity, diameter at breast height (dbh), and status (alive, dead, harvested or
99 recruited). Plot design varies between countries but generally plots are circular with variable radii
100 depending on tree size (see Supplementary Materials). The minimum dbh of trees included in the
101 dataset was 10 cm. Plots were remeasured over time allowing estimations of individual growth
102 and survival. Only the French NFI is based on a single measurement but includes measurement
103 of radial growth with cores and estimation of time since death allowing to estimate these vital
104 rates. We selected species with > 2,000 individuals and > 500 plots, to ensure a good coverage of
105 their range, growth, and survival. We excluded exotic species for which the distribution is mainly
106 controlled by plantation operations. For the demographic analyses, we also excluded all plots
107 with records of harvesting operations or disturbances between the two surveys, which would
108 otherwise influence our estimation of local competition.

109 *Climate variables*

110 We used two bioclimatic variables known to control tree demography (Kunstler *et al.* 2011): (1)
111 the sum of degree days above 5.5 °C (*sgdd*), and (2) the water availability index (*wai*). *sgdd* is the
112 cumulative day-by-day sum of the number of degrees > 5.5 °C and is related to the mean annual
113 temperature and the length of the growing season. It was extracted from E-OBS, a high resolution
114 (1 km²) downscaled climate data-set (Moreno & Hasenauer 2016) for the years between the two
115 surveys plus two years before the first survey. *wai* was computed using precipitation (*P*, extracted
116 from E-OBS) and potential evapotranspiration (*PET*) from the Climatic Research Unit (Harris *et*
117 *al.* 2014) data-set, as $(P - PET)/PET$ and is related to the water availability.

118 *Integral projection models*

119 An IPM predicts the size distribution, $n(z', t + 1)$, of a population at time $t + 1$ from its size dis-
120 tribution at t , $n(z, t)$ (with z the size at t and z' the size at $t + 1$), based on the following equation
121 (Easterling *et al.* 2000; Ellner *et al.* 2016):

$$n(z', t + 1) = \int_L^U K(z', z)n(z, t)dz \quad (1)$$

122 The kernel $K(z', z)$ can be split into the survival and growth kernel ($P(z', z)$) and the fecundity
123 kernel ($F(z', z)$), as follow $K(z', z) = P(z', z) + F(z', z)$. $P(z', z)$ is defined as $P(z', z) = s(z)G(z', z)$
124 and represents the probability that an individual of size z survives between t and $t + 1$ and reaches
125 the size z' . The size of the individuals z can range between L and U . NFI data do not provide
126 direct information on tree fecundity, thus our models describe the fate of a cohort (a cohort IPM
127 for individuals with dbh ≥ 10 cm) by focusing only on $P(z', z)$ (in Supplementary Materials, we
128 provide a sensitivity analysis of tree population growth to fecundity)

129 For each of the 27 species, we fitted growth and survival functions depending on tree size, the
130 two climatic variables (*sgdd* and *wai*) and local competition estimated as the sum of basal area of
131 competitors (following Kunstler *et al.* 2011). The shape of the response curve for these variables
132 and the type of interaction between climate and size and climate and competition can have a large
133 impact on vital rates predictions. To account for such uncertainties, we re-sampled 100 times 70%
134 of the data to fit the model and select the best type of response curve and interactions based on

135 the Akaike information criteria (*i.e.*, lowest AIC) (Burnham & Anderson 2002). Because there were
136 fewer plots in extreme climatic conditions, we resampled the data with a higher probability of
137 sampling plots in extreme climatic conditions for the given species. Then we used the remaining
138 30% of the data to evaluate the goodness of fit of the growth and survival models. Goodness of fit
139 and response curves of growth and survival models are presented in the Supplementary Materials
140 (Figures 4 to 13).

141 *Growth model*

142 After preliminary exploration, we selected two alternative shapes of the climatic response curves:
143 asymptotic or quadratic polynomial corresponding to the equations 2 and 3. These choices corre-
144 spond to two alternative biological models: (i) either all species have their optimum at high water
145 availability and sum of degree days; or (ii) species have bell-shaped climate response curves with
146 different optima along the climatic variables:

$$\log(G_{i,p}) = a_0 + a_{0,p} + a_1 D_i + a_2 \log(D_i) + a_3 BA_i + a_4 \frac{1}{sgdd_p} + a_5 \frac{1}{wai_p} + \varepsilon_i \quad (2)$$

$$\log(G_{i,p}) = a_{0,c} + a_{0,p} + a_1 D_i + a_2 \log(D_i) + a_3 BA_i + a_4 sgdd_p + a_5 sgdd_p^2 + a_6 wai_p + a_7 wai_p^2 + \varepsilon_i \quad (3)$$

147 Where $G_{i,p}$ is the annual diameter growth of tree i in plot p , D_i is the dbh of tree i , BA_i is the
148 sum of basal area of competitors of tree i per ha, $sgdd_p$ is the sum of growing degree days, wai_p
149 is the water aridity index, a_0 to a_7 are estimated parameters, and $a_{0,p}$ is a normal random plot
150 effect accounting for unexplained variation at the plot level. The intercept $a_{0,c}$ is country specific
151 to account for differences in protocol between NFIs and ε_i is the unexplained tree level variability
152 following a normal distribution. We fitted the models in R-cran separately for each species (R Core
153 Team 2019) using the 'lmer' function ("lme4" package, Bates *et al.* 2015). We also tested models
154 with interactions between the climatic variables - $1/sgdd_p$ and $1/wai_p$ for model (2) and $sgdd_p$ and

155 wai_p for model (3)) - and size (D_i and $\log(D_i)$) and the climatic variables and competition.

156 *Survival model*

157 Survival models were fitted with a generalised linear model with a binomial error. The predictors
158 and interactions explored were the same as in the growth model. To account for variable survey
159 times between plots we used the complementary log-log link with an offset representing the num-
160 ber of years between the two surveys (y_p) (Morris *et al.* 2013). We fitted the model in R-cran using
161 the 'glm' function. We did not include a random plot intercept because in most plots no individ-
162 uals died between the surveys, making the estimation of the random plot effect challenging.

163 *Tree harvesting*

164 Although we excluded plots with evidence of harvesting between the two surveys to fit the sur-
165 vival functions, most European forests are subject to management, which has a strong impact on
166 population dynamics (Schelhaas *et al.* 2018). Harvesting of dying or damaged trees is probably
167 resulting in an underestimation of the natural mortality rate. To make sensible predictions with
168 our IPMs it was thus necessary to incorporate a harvesting rate to prevent an overestimate of tree
169 lifespan. We set a mean harvesting rate, estimated across all species and inventories, as 0.5 % per
170 year. We did not model size and climate dependence of the harvesting rate, as we focused on
171 climatic and not anthropogenic constraints on tree demography.

172 *Prediction of demographic metrics at the edges and centre of species range*

173 *Species distribution*

174 To identify the edge of a species range, a simple representation of its distribution is necessary.
175 Across Europe, there is a strong correlation between $sgdd$ and wai , and so we described species
176 ranges along a single climatic axis corresponding to the first axis (PC1) of the PCA of $sgdd$ and wai
177 (Supplementary Materials, Figure 3). Species showed a clear segregation along this climatic axis
178 in Europe (Figure 1). Based on the coordinates on PC1 of the plots where the species was present,
179 we identified the centre of the range as their median value, the hot and dry edge (hereafter hot
180 edge) and the cold and wet edge (hereafter cold edge), respectively, as their 5% and 95% quantiles.

181 To evaluate which species' edges corresponded to an actual limit in the species distribution and
182 not just to limits in the coverage of the data, we fitted species distribution models with BIOMOD2
183 (Thuiller *et al.* 2009). To do so we used presence/absence data covering all Europe (Mauri *et*
184 *al.* 2017) (see Supplementary Materials). For comparison of the demographic performance at the
185 edge *vs.* the centre of the distribution, we retained only the edges with a at least 10% drop in the
186 probability of presence of the species predicted by the SDM (Figure 1).

187 *Demographic metrics*

188 To evaluate how individual performance varied between the range centre and the edges, we de-
189 rived four metrics representing key dimensions of population performance. The first two metrics
190 were related to individual vital rates, and were defined by the growth and survival of 15 cm dbh
191 individuals (focusing on small individual because of their strong effect on population dynamics,
192 Grubb 1977). The last two metrics were related to individual lifetime performance integrating the
193 vital rates in the IPM, and were defined by the mean lifespan of a 10 cm dbh individual and the
194 passage time of a 10 cm dbh individual to 60 cm. The details of the numerical methods used to
195 compute lifespan and passage time from the IPM are provided in the Supplementary Materials.
196 Model diagnostics showed that our numerical approach was not sensitive to the number of size
197 bins retained for the IPM (*i.e.* # bins > 800 , see Figure 14 in Supplementary Materials).

198 We predicted the four demographic metrics at the centre and the hot and cold edges of the
199 species range using their positions on the climatic axis (each position on PC1 is associated with
200 a unique combination of *sggd* and *wai*). We integrated uncertainty into our estimate by deriving
201 a prediction for each of the 100 re-sampled growth and survival models (see above). Because
202 competitive interactions may also be important in controlling species demography at the edge of
203 the range (Louthan *et al.* 2015), we made these predictions either without competition (by setting
204 *BA* to 0) or with a high level of competition (by setting *BA* to $30m^2ha^{-1}$, corresponding to a closed
205 forest).

206 *Analysis of the relative demographic performance at the edges*

207 For each demographic metric (*m*) we computed the relative difference in the metric at the edge
208 (hot or cold) *vs.* the centre as: $\Omega_{edge}^m = (m_{edge} - m_{centre}) / m_{centre}$. Firstly, for each metric, we tested

209 whether species demographic performance declines at the edge compared to the centre (hypothesis
210 1) by fitting a mixed model to m as function of the range position type (edge *vs.* centre) and
211 a random species effect (with the function *lmer* in *lme4*). Secondly, we tested whether the effects
212 were different without or with competition (hypothesis 2). We ran this analysis separately for hot
213 and cold edges to see how demographic responses differ between them (hypothesis 3). Thirdly,
214 we explored whether Ω_{edge}^m was dependent on the climatic optimum of the species (the median
215 position on PC1) (hypothesis 4) by fitting Phylogenetic generalised least squares (PGLS) regres-
216 sion using a phylogeny extracted from Zanne *et al.* (2014). We accounted for the uncertainty in the
217 demographic response by including a weight proportional to the inverse of the variance of Ω_{edge}^m .
218 The PGLS regression with maximum likelihood estimation of Pagel's lambda (a measure of the
219 phylogenetic signal ranging between 0 and 1) did not always converged. In those cases we fitted
220 a PGLS model with a Brownian model (Pagel's lambda set at 1). We retained only the regressions
221 that were both significant (after a Bonferroni correction to account for multiple comparisons) and
222 had a non-negligible magnitude of the effect (Camp *et al.* 2008). The magnitude of the effect was
223 considered negligible when the confidence interval of the effect size intercepted the interval -0.10
224 and 0.10 (Camp *et al.* 2008). Effect sizes were computed as the standardised slope (Schielzeth
225 2010).

226 Finally, we explored the effect of four functional traits (see below) that are known to influence
227 tree response to climate (hypothesis 5), by testing the link between each trait individually and
228 Ω_{edge}^m with the same PGLS regression approach. We selected: (i) wood density, which is related
229 to drought and temperature response (Chave *et al.* 2009; Stahl *et al.* 2014); (ii) the leaf economic
230 spectrum (LES) because species at the conservative end of the spectrum are thought to be more
231 tolerant to extreme climate (Reich 2014); (iii) leaf size, which is related to plant response to water
232 stress and frost (Wright *et al.* 2017); and (iv) xylem vulnerability to embolism measured by the
233 water potential leading to 50% loss of xylem conductivity, Ψ_{50} , a strong predictor of drought-
234 induced mortality (Anderegg *et al.* 2016). LES is based on the covariance of specific leaf area, leaf
235 lifespan, and leaf nitrogen per mass (Wright *et al.* 2004). We used leaf nitrogen per mass (N_{mass}), as
236 it was the LES trait with the best coverage across our species. Trait data were sourced from open
237 databases (Wright *et al.* 2004, 2017; Chave *et al.* 2009; Choat *et al.* 2012; Maire *et al.* 2015).

238 Results

239 *Demographic responses differ between edge types and metrics*

240 Across the 27 species we found evidence of a significant decrease in growth and increase in pas-
241 sage time (longer time needed to grow from 10 to 60 cm) at the cold edge in comparison with the
242 centre of the distribution but no effect at the hot edge (Figure 2). In contrast, at the hot edge, we
243 found evidence of a significant decrease in both tree survival and lifespan (Figure 2). This is con-
244 sistent with the hypothesis that at least one metric will decline in performance at the edge, and that
245 different metrics are affected depending on the edge type. In contrast, we found that lifespan was
246 significantly longer at the cold edge than at the centre of the distribution (Figure 2). Generally,
247 these patterns were unaffected by local competition (Figure 3), however, it is important to note
248 that the increase in lifespan at the cold edge became non-significant at high levels of competition,
249 which may show that competition constrains performance at the edge for some species.

250 Behind the overall demographic response at the edge, there were large variations between
251 species. For each metric and edge type we found species showing a decrease and species showing
252 an increase in performance (Supplementary Materials; Figures 16 to 23).

253 *Demographic responses vary with species climatic optimum*

254 Growth response at the hot and cold edges was related to the climatic optimum of the species;
255 species with climatic optimum in hot climates were more constrained at their hot edge while
256 species with climatic optimum in cold climates were more constrained at their cold edge. This
257 result is depicted in Figure 4 by a positive relationship between Ω_{edge}^m for growth at the hot edge
258 and the median climate of the species and a negative relationship Ω_{edge}^m for the growth at the cold
259 edge. The same pattern is visible for passage time, but in the opposite direction, because passage
260 time is longer when growth is slower (Figure 4). The responses of Ω_{edge}^m for survival and lifespan
261 were much weaker or null. We found a negative relationship for survival at the hot edge, which
262 was largely related to a few extreme species, and no effect for lifespan (Figure 4).

263 *Weak links between demographic response and species traits*

264 N_{mass} had the strongest relationship with Ω_{edge}^m of all the traits we tested. At the hot edge, species
265 with high N_{mass} experienced a stronger decrease in their survival and lifespan than species with a
266 low N_{mass} (Figure 5). In contrast, at the cold edge, species with low N_{mass} experienced a stronger
267 decrease in their survival and lifespan than species with high N_{mass} (Figure 5). In addition, species
268 with high N_{mass} had less limitation of their growth at the hot edge than species with low N_{mass}
269 (Figure 5).

270 Relationships between Ω_{edge}^m and wood density, leaf size and xylem vulnerability to embolism
271 were generally weak (Supplementary Materials, Figures 25 to 27). Species with small leaf area
272 had better survival, growth, and passage time at the cold edge than large leafed species (Supple-
273 mentary Materials, Figure 27) and species with high Ψ_{50} experienced a stronger decrease in their
274 growth at the hot edge than species with low Ψ_{50} . However, the pattern was driven by a few
275 species (Supplementary Materials, Figure 26).

276 **Discussion**

277 Our analysis based on pan-European forest inventory data and integral projection models of 27
278 species, found weak support for the ACH prediction that demographic performance is lower at
279 the edge than at the centre of the species range. Instead, decline in demographic performance
280 was strikingly different between the cold and the hot edges. At cold and wet edges, growth
281 and passage time were constrained, whereas at hot and dry edges, survival and lifespan were
282 constrained. Beyond these general patterns, we found important variability between species in
283 their demographic performance at the edge, which was partially explained by species' climatic
284 optimum and traits.

285 *Different demographic responses at the hot and the cold edge*

286 We found mixed support for the ACH, not all the demographic metrics were limited at the edges
287 and patterns were variable between species. This is consistent with observational studies that
288 found limited evidence of a relationship between species demography and their distribution. For
289 instance, both Thuiller *et al.* (2014) and Csergo *et al.* (2017) found limited correlation between

290 plants demographic performance and probability of presence. In addition, Purves (2009) reported
291 mixed evidence of a decrease in demographic performance at the south and north edges of North
292 American tree species.

293 Growth and passage time were constrained at the cold edge in comparison with the centre
294 of the species distribution. This is consistent with studies on North American tree species, that
295 found a decrease in growth at the cold edge in adult trees (Purves 2009) and juveniles (Ettinger &
296 HilleRisLambers 2017; Putnam & Reich 2017). In contrast with the ACH, we found a tendency for
297 a slightly faster growth at the hot edge than at the centre, which has also been reported in North
298 American trees (Purves 2009; Ettinger & HilleRisLambers 2017; Putnam & Reich 2017).

299 At the hot and dry edge, tree survival and lifespan were lower than at the centre of the range.
300 In contrast, Purves (2009) found no such decrease in survival at the hot edge of eastern North
301 American species. This difference could be explained by the fact that the hot edge of most Eu-
302 ropean species corresponds to both a hot and a dry climate, whereas in eastern America the hot
303 edge is less constrained by drought (Zhu *et al.* 2014). We found that lifespan was longer at the
304 cold edge than at the centre of the distribution. This is in contradiction with the classical view that
305 survival is constrained in cold climates and the results of Purves (2009). Given that tree growth
306 rate is constrained at the cold edge, this longer lifespan could be explained by a tradeoff between
307 tree growth rate and tree longevity (both inter- and intra-specific, see Black *et al.* 2008; Di Fil-
308 ippo *et al.* 2015) and the observation that survival rate correlates negatively with site productivity
309 (Stephenson *et al.* 2011).

310 *Lack of competition effect*

311 Numerous studies have proposed that competitive interactions could be crucial in setting demo-
312 graphic limits, particularly at the productive edge (see Hargreaves *et al.* 2014; HilleRisLambers *et*
313 *al.* 2013; Louthan *et al.* 2015; Alexander *et al.* 2016; Ettinger & HilleRisLambers 2017; Jump *et al.*
314 2017). In our analyses, we explored the effect of competition by comparing the demographic met-
315 rics estimated without competition or with a high level of competition. Despite the strong effects
316 of competition on both growth and survival and interactions between competition and climate,
317 competition did not influence the demographic response at the edges. The only evidence of com-
318 petition constraints at the edge that we found was that without competition; lifespan increased at

319 the cold edge only in the absence of competition.

320 Three main reasons could explain the lack of competitive effect on the demographic response
321 at the edge in our study. Firstly, we did not differentiate between intra- and inter-specific compe-
322 tition, whereas inter-specific competition might have the strongest impact at the edge (Alexander
323 *et al.* 2016). Secondly, we did not analyse competitive effect on population growth rate, it was
324 thus not possible to evaluate whether competitive exclusion could be at play at the edges (Ches-
325 son 2018). Thirdly, properly estimating competitive effect with observational data is notoriously
326 difficult (Tuck *et al.* 2018).

327 *Strong effect of species median climate on growth response at the edge*

328 We found that the hotter the centre of the species range the greater was the constraints on growth
329 and passage time at its hot edge. The same pattern was found with the cold edge and the species
330 climate optimum proximity to cold extreme. This is in agreement with the general pattern of
331 vegetation productivity in Europe, which is at its maximum in temperate climates where both
332 drought and cold stress are limited (Jung *et al.* 2007).

333 *Weak traits effect on species demographic response at the edge*

334 Part of the variation in the demographic response at the edge was related to N_{mass} , a key dimension
335 of the leaf economic spectrum. An important difficulty in the interpretation of these results is
336 that our understanding of the link between leaf economic traits and climate is limited. Multiple
337 mechanisms, some of them contradictory, have been proposed to explain the link between leaf
338 N and climate. For instance, it is generally considered that species with low N_{mass} have a more
339 conservative strategy of resource use and perform better in stressful conditions than species with
340 high N_{mass} (Reich 2014). In agreement with this finding, we found that species with low N_{mass}
341 had a better survival and lifespan at the hot edge. In contrast, high leaf N has been linked with
342 photosynthesis tolerance to drought and low temperature because of higher enzyme activities
343 (Wright *et al.* 2003; Reich & Oleksyn 2004). Consistent with this mechanism, we found that species
344 with high N_{mass} had a better survival and lifespan at the cold edge and a better growth at the hot
345 edge.

346 We found limited relationships between leaf size or xylem vulnerability to embolism and de-

347 mographic response, which was surprising as the mechanisms related to climate response are
348 better understood for these traits. Smaller leaves were related to a longer lifespan at the hot edge
349 and a better survival, growth and passage time at the cold edge. This in agreement with Wright *et*
350 *al.* (2017) who proposed that large leaves are disadvantaged in hot and dry climates because their
351 transpiration rate during the day is too high and are disadvantaged in cold climate because they
352 have greater risks of reaching critical low temperatures during the night. Anderegg *et al.* (2019)
353 also reported weak link between traits and drought related mortality at the edge, with only an
354 effect for xylem vulnerability to embolism. The effect was, however, that drought adapted species
355 experienced higher drought mortality at the edge.

356 *On the challenge of connecting population dynamics and species ranges*

357 Based on our results, it is difficult to conclude if there is a clear decrease in population growth rate
358 at the edges. A key limitation is that our analysis did not include the regeneration phase, which
359 is thought to be a bottleneck in tree population dynamics (Grubb 1977). In the Supplementary
360 materials, we provide an evaluation of the importance of this phase for tree population growth
361 rate with an elasticity analysis of matrix population models extracted from the COMPADRE Plant
362 matrix database (Salguero-Gómez *et al.* 2015). The analysis shows that the regeneration and adult
363 phases were equally important (see Figure 29 in Supplementary Materials). Our analysis thus
364 captures an important part of a tree's life cycles. However, we can not rule out the possibility that
365 the regeneration phase has a disproportional importance for the dynamics at the edge, as several
366 studies have shown that this phase is extremely sensitive to climate [Clark *et al.* (2014); Defosse-
367 2016]. Integrating regeneration in IPMs is challenging because we have much less data on this
368 stage (Needham *et al.* 2018).

369 It is also important to keep in mind that species ranges are not necessarily related to the mean
370 population growth rate but rather to its temporal variability and the population resilience because
371 they control extinction risk (Holt *et al.* 2005). Another explanation is that suitable habitats where
372 population growth rates are unaffected might exist up to the edge due to the presence of suitable
373 microsites (Cavin & Jump 2017). In this case, the species ranges arise because the fraction of
374 suitable habitats available to the metapopulation decrease (Holt & Keitt 2000). Finally, tree species
375 distributions might not be in equilibrium with the current climate. This could be because species

376 are either still in the process of recolonising from their ice age refugia (Svenning & Skov 2004)
377 or already affected by climate change. Such disequilibrium should however be visible by better
378 performance at the cold edge (Talluto *et al.* 2017) and we found no evidence of this in our results.

379 *Synthesis*

380 Our field study shows that trees' demographic responses at range edges are more complex than
381 predicted by the ACH. Here, the patterns of demographic response of the 27 European tree species
382 varied between their hot and cold edges. We only found strong evidence of demographic limits
383 for edges occurring in extreme conditions (hot edges of hot-distributed species and cold edges
384 of cold-distributed species). Our findings open an important perspective, as they show that one
385 should not expect the same demographic response at the hot *vs.* the cold edge and that we need
386 to refine predictions of climate change impacts in function of the species and edge characteristics.

387 *Acknowledgments*

388 This paper is a joint effort of the working group sAPROPOS - 'Analysis of PROjections of POP-
389 ulationS', kindly supported by sDiv (Synthesis Centre of the German Centre for Integrative Bio-
390 diversity Research - iDiv), funded by the German Research Foundation (FZT 118). GK and AG
391 received support from the REFORCE - EU FP7 ERA-NET Sumforest 2016 through the call "Sus-
392 tainable forests for the society of the future", with the ANR as national funding agency (grant
393 ANR-16-SUMF-0002). The NFI data synthesis was conducted within the FunDivEUROPE project
394 funded by the European Union's Seventh Programme (FP7/2007–2013) under grant agreement
395 No. 265171. We thank Gerald Kandler (Forest Research Institute Baden-Wurtemberg) for his help
396 to format the German data. We thank the MAGRAMA, the Johann Heinrich von Thunen-Institut,
397 the Natural Resources Institute Finland (LUKE), the Swedish University of Agricultural Sciences,
398 and the French Forest Inventory (IGN) for making NFI data available. We are grateful to the
399 Glopnet, the global wood density, the global leaf size, and the global xylem embolism vulnerabil-
400 ity data bases for making their data publicly available. We are grateful to all the participants of
401 the sAPROPOS working group for their stimulating discussion. We are grateful to Fabian Roger
402 for his help to build the species phylogeny.

403 **FIGURES**

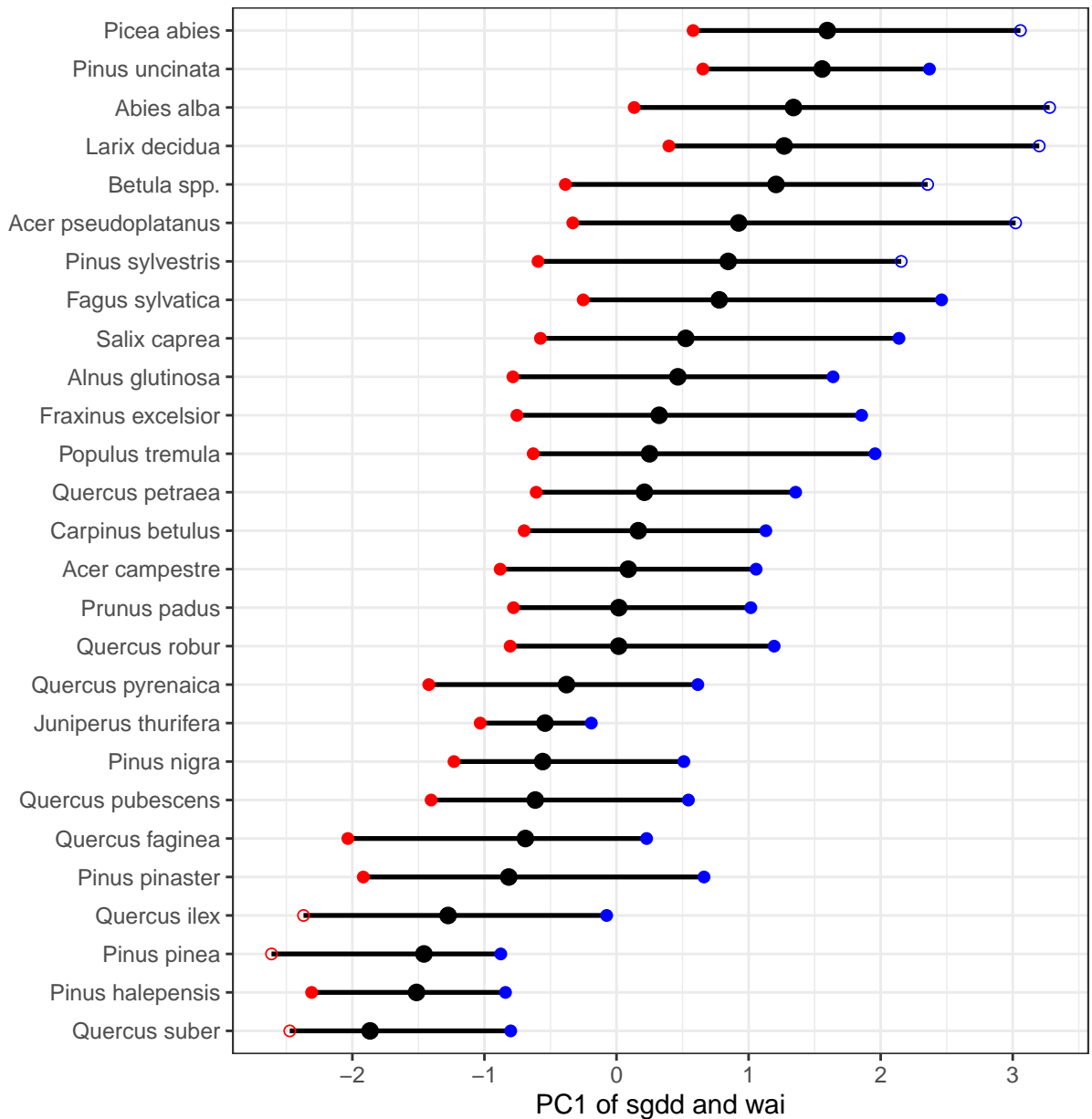


Figure 1: Species distribution along the axis one of the PCA of the two climatic variables sgdd and wai. The centre of the species distribution along this axis is represented by a black circle and the hot and dry edge and the cold and wet edge by red and blue circle respectively. Filled circles represent edges with a clear drop of the probability of presence that were selected for the analysis.

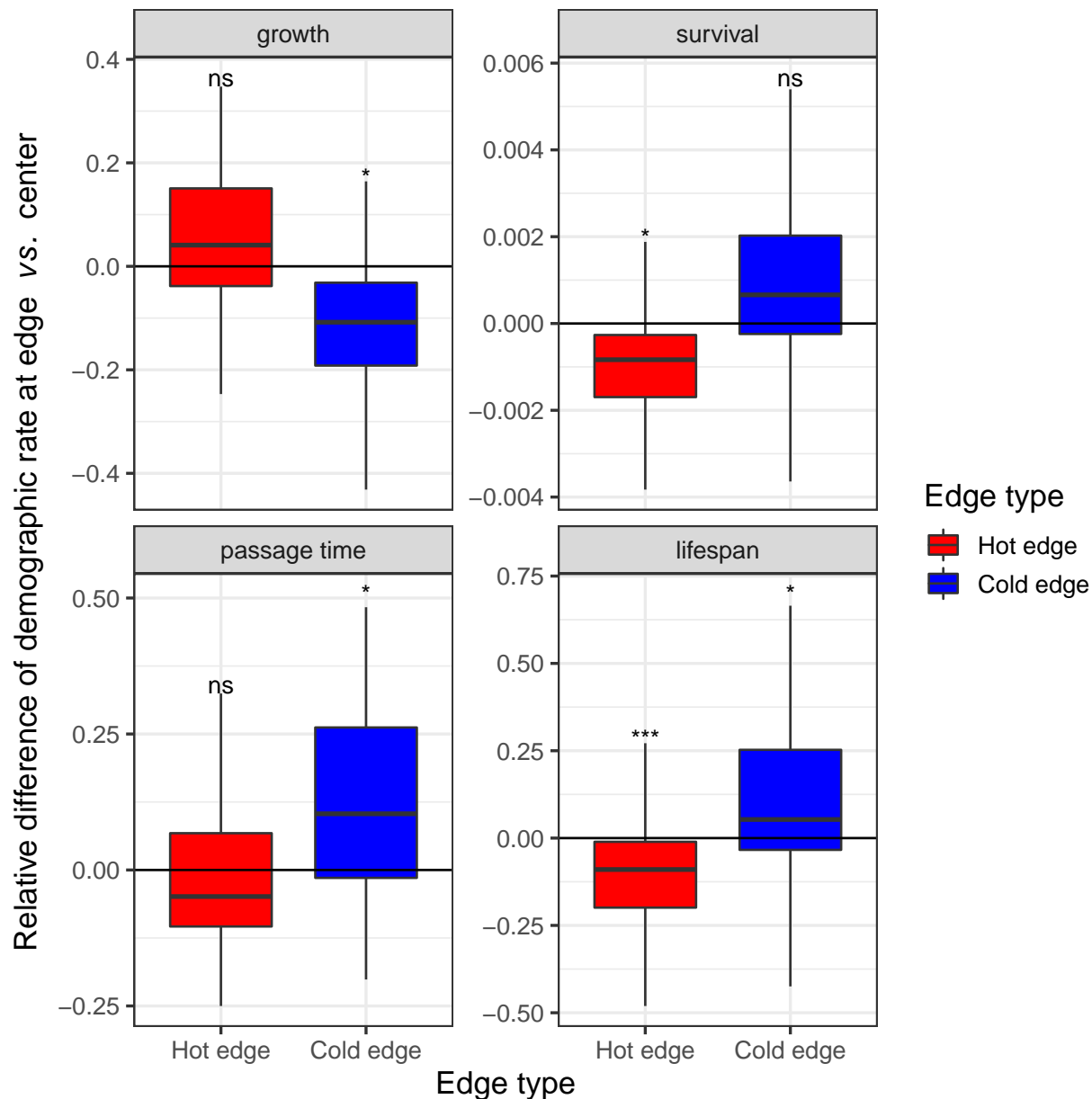


Figure 2: **Differences in the demographic metrics at edge vs. the centre of the distribution** (/Users/kunstler/Dropbox/sAPROPOS/sAPROPOS_ForInv/figures/Demo_diff_resample_overall_sp_best2.pdf). The box-plot of the relative difference between the edge and the centre computed over the 100 data resampling and the 27 species is represented for the four demographic metrics (annual diameter growth and survival for an individual 15cm in diameter, passage time from 10cm in diameter to 60cm in diameter and lifespan of tree 15cm in diameter) and the two edge types (hot in red, cold in blue). The p value of the test for the difference in each demographic metric and edge type is presented at the top of the box-plot (ns : non significant, * : p value < 0.05, ** : p value < 0.01, *** : p value < 0.001). The p value was computed with a mixed model with species as a random effect (see Methods for details).

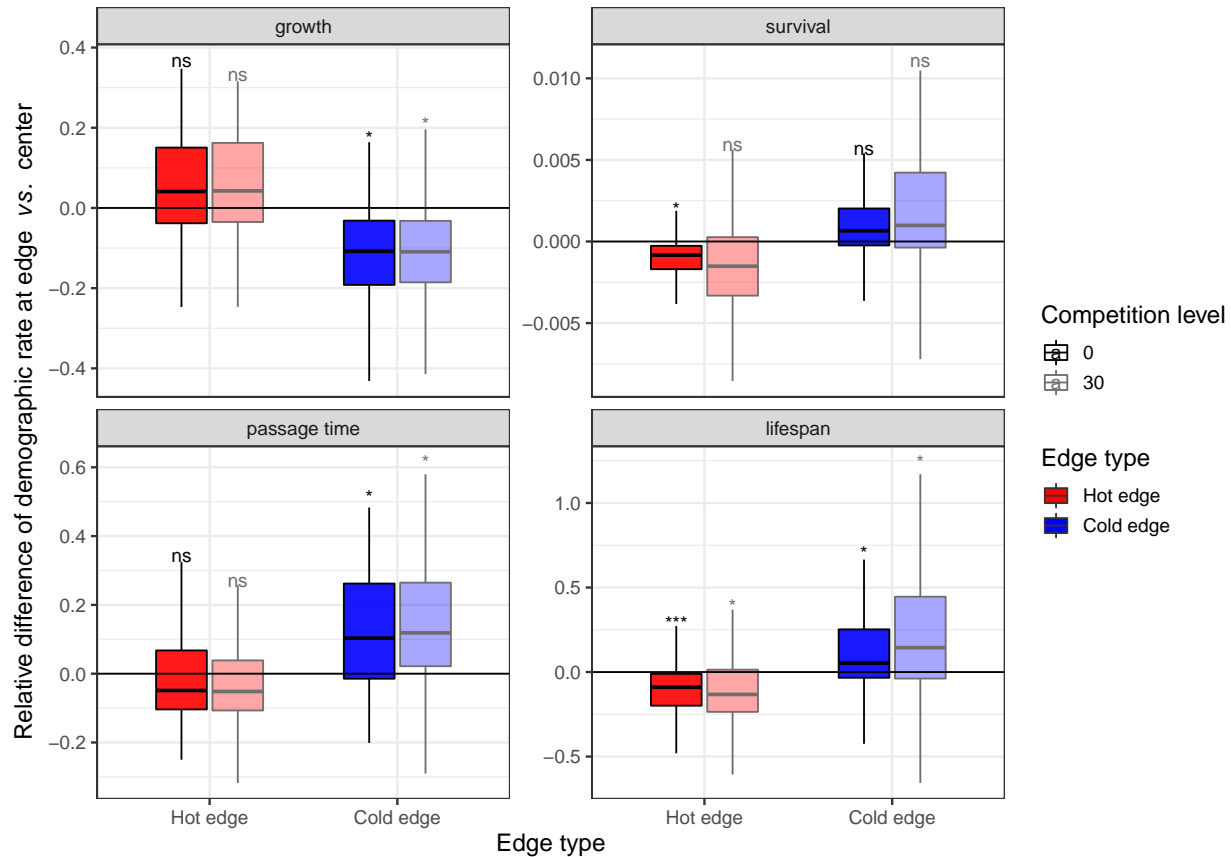


Figure 3: **Differences in the demographic metrics at edge vs. the centre of the distribution** (/Users/kunstler/Dropbox/sAPROPOS/sAPROPOS_ForInv/figures/Demo_diff_resample_overall_sp_BATOT without and with a high level of competition. The box-plot of the relative difference between the edge and the centre computed over the 100 data resampling and the 27 species is represented for the four demographic metrics (annual diameter growth and survival for an individual 15cm in diameter, passage time from 10cm in diameter to 60cm in diameter and lifespan of tree 15cm in diameter), the two edge types (hot in red, cold in blue), and the two levels of competition (without competition: basal area of competitors, $BA = 0$, no transparency, with a high level of competition: basal area of competitors, $BA = 30m^2 ha^{-1}$ color transparency). The p value of the test for the difference in each demographic metric and edge type is presented at the top of the box-plot (ns : non significant, * : p -value < 0.05 , ** : p -value < 0.01 , *** : p -value < 0.001). The p -value was computed with a mixed model with species as a random effect (see Methods for details)

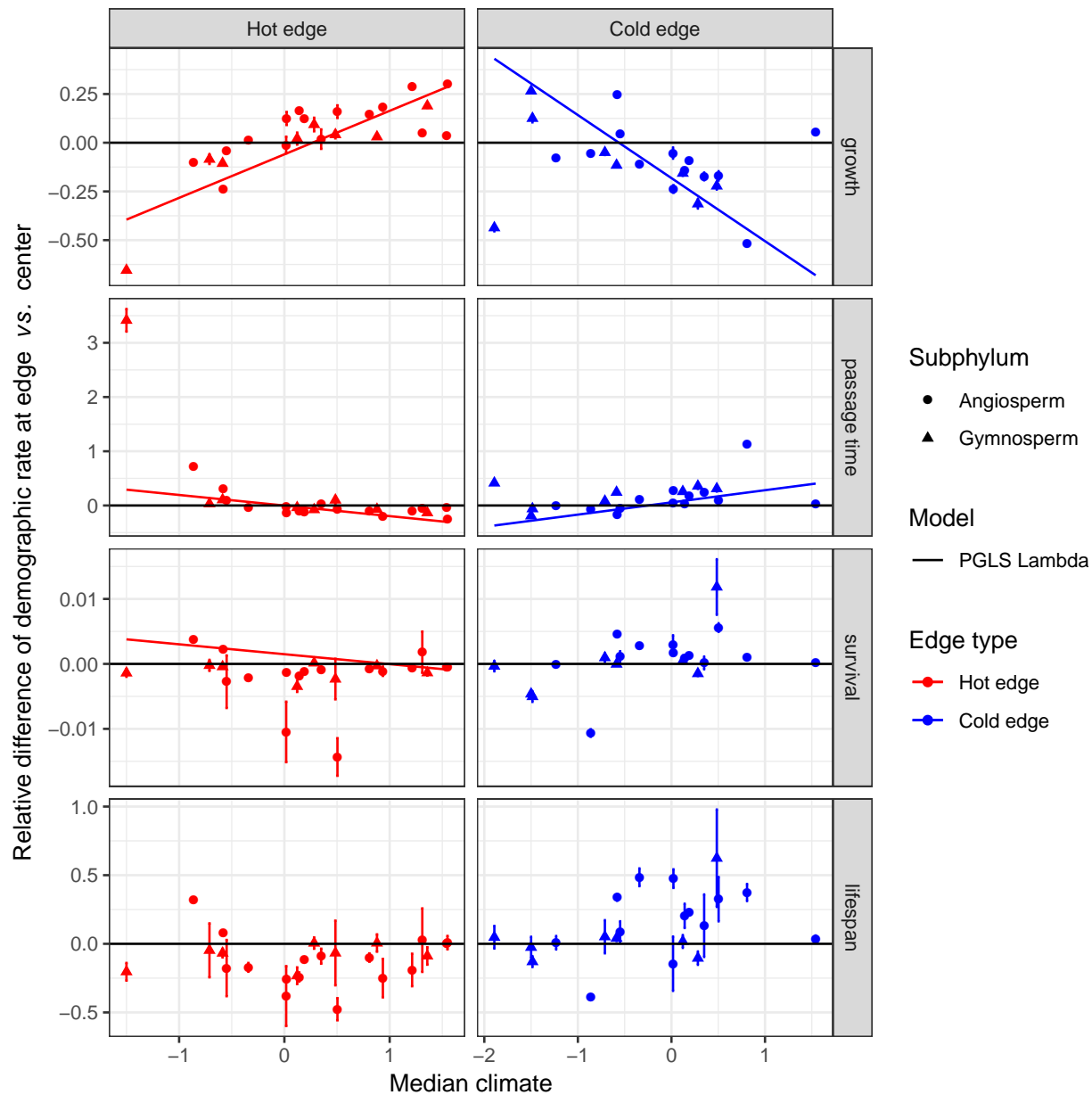


Figure 4: **Changes in demographic responses at the edge** ([/Users/kunstler/Dropbox/sAPROPOS/sAPROPOS_ForInv/figures/Demo_diff_resample_PC1_lm_overall_sp](#)) **in function of species climatic optimum.** Species demographic response at the edge - measured as the relative differences of the demographic metrics at edge *vs.* the centre of the distribution - in function of the median position of the species on the first axis of the climate PCA. For each species the mean (point) and the 95% quantiles (error bar) of the demographic response over the 100 data resampling is represented for both the hot (red) and the cold (blue) edges. Phylogenetic generalised least squares (PGLS Lambda) regressions are represented only for significant relationship with a non negligible magnitude of the effect. Gymnosperm and angiosperm species are represented with different symbols.

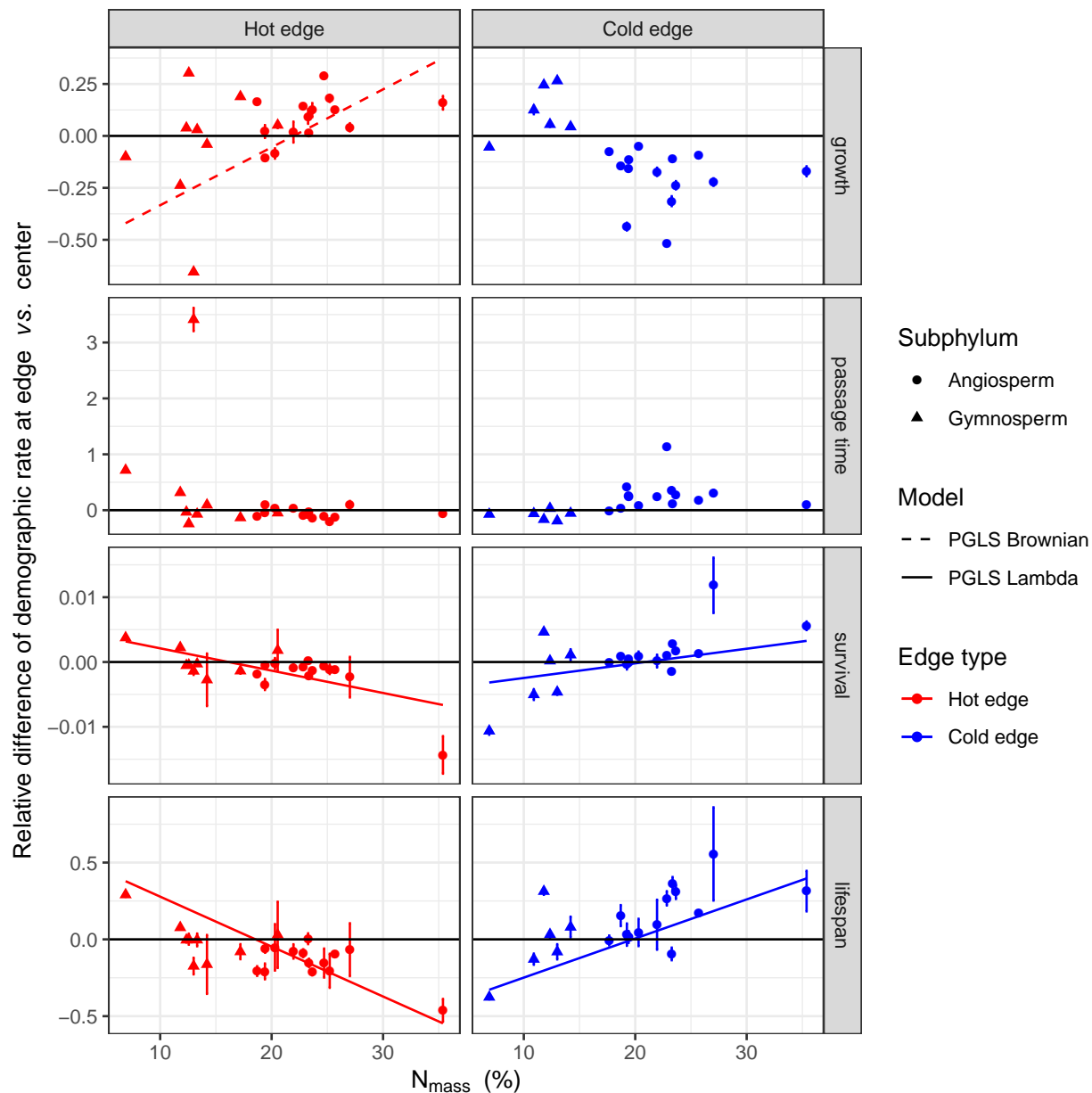


Figure 5: **Changes in demographic responses at the edge** (/Users/kunstler/Dropbox/sAPROPOS/sAPROPOS_ForInv/figures/Demo_diff_resample_Nmass_lm_overall) **in function of species leaf N per mass.** Species demographic response at the edge - measured as the relative differences in the demographic metric at edge *vs.* the centre of the distribution - as a function of species leaf nitrogen per mass. For each species the mean (point) and 95% quantiles (error bar) of the demographic response over the 100 data resampling is represented for both the hot (red) and the cold (blue) edges. Phylogenetic generalised least squares (PGLS) regressions are represented only for significant relationship with a non negligible magnitude of the effect (see details in caption of Figure 4).

404 REFERENCES

405 Alexander, J.M., Diez, J.M., Hart, S.P. & Levine, J.M. (2016). When Climate Reshuffles Competitors:
406 A Call for Experimental Macroecology. *Trends Ecol. Evol.*, 31, 831–841.

407 Anderegg, W.R., Klein, T., Bartlett, M., Sack, L., Pellegrini, A.F. & Choat, B. *et al.* (2016). Meta-
408 analysis reveals that hydraulic traits explain cross-species patterns of drought-induced tree mor-
409 tality across the globe. *P. Natl. Acad. Sci. USA.*, 113, 5024–5029.

410 Anderegg, W.R.L., Anderegg, L.D.L., Kerr, K.L. & Trugman, A.T. (2019). Widespread drought-
411 induced tree mortality at dry range edges indicates that climate stress exceeds species' compen-
412 sating mechanisms. *Glob. Change Biol.*, gcb.14771.

413 Baeten, L., Verheyen, K., Wirth, C., Bruelheide, H., Bussotti, F. & Finér, L. *et al.* (2013). A novel
414 comparative research platform designed to determine the functional significance of tree species
415 diversity in European forests. *Perspect. Plant Ecol.*, 15, 281–291.

416 Bates, D., Mächler, M., Bolker, B. & Walker, S. (2015). Fitting linear mixed-effects models using
417 lme4. *J. Stat. Softw.*, 67, 1–48.

418 Black, B.A., Colbert, J.J. & Pederson, N. (2008). Relationships between radial growth rates and
419 lifespan within North American tree species. *Écoscience*, 15, 349–357.

420 Brown, J.H. (1984). On the relationship between abundance and distribution of species. *Am.*
421 *Nat.*, 124, 255–279.

422 Burnham, K.P. & Anderson, D.R. (2002). *Model selection and multimodel inference: A practical*
423 *information-theoretic approach*. Springer-Verlag, New-York.

424 Cahill, A.E., Aiello-Lammens, M.E., Caitlin Fisher-Reid, M., Hua, X., Karanewsky, C.J. & Ryu,
425 H.Y. *et al.* (2014). Causes of warm-edge range limits: Systematic review, proximate factors and
426 implications for climate change. *J. Biogeogr.*, 41, 429–442.

427 Camp, R.J., Seavy, N.E., Gorresen, P.M. & Reynolds, M.H. (2008). A statistical test to show
428 negligible trend: comment. *Ecology*, 89, 1469–1472.

429 Case, T.J., Holt, R.D., McPeck, M.A. & Keitt, T.H. (2005). The community context of species'
430 borders: Ecological and evolutionary perspectives. *Oikos*, 108, 28–46.

431 Cavin, L. & Jump, A.S. (2017). Highest drought sensitivity and lowest resistance to growth
432 suppression are found in the range core of the tree *Fagus Sylvatica* L. Not the equatorial range

433 edge. *Glob. Change Biol.*, 23, 362–379.

434 Chave, J., Coomes, D.A., Jansen, S., Lewis, S.L., Swenson, N.G. & Zanne, A.E. (2009). Towards
435 a worldwide wood economics spectrum. *Ecol. Lett.*, 12, 351–66.

436 Chen, I.-C., Hill, J.K., Ohlemüller, R., Roy, D.B. & Thomas, C.D. (2011). Rapid range shifts of
437 species associated with high levels of climate warming. *Science*, 333, 1024–6.

438 Chesson, P. (2018). Updates on mechanisms of maintenance of species diversity. *J. Ecol.*, 106,
439 1773–1794.

440 Choat, B., Jansen, S., Brodribb, T.J., Cochard, H., Delzon, S. & Bhaskar, R. *et al.* (2012). Global
441 convergence in the vulnerability of forests to drought. *Nature*, 491, 752–755.

442 Clark, J.S., Bell, D.M., Kwit, M.C. & Zhu, K. (2014). Competition-interaction landscapes for the
443 joint response of forests to climate change. *Glob. Change Biol.*, 20, 1979–1991.

444 Csergo, A.M., Salguero-Gomez, R., Broennimann, O., Coutts, S.R., Guisan, A. & Angert, A.L.
445 *et al.* (2017). Less favourable climates constrain demographic strategies in plants. *Ecol. Lett.*, 20,
446 969–980.

447 Di Filippo, A., Pederson, N., Baliva, M., Brunetti, M., Dinella, A. & Kitamura, K. *et al.* (2015).
448 The longevity of broadleaf deciduous trees in Northern Hemisphere temperate forests: Insights
449 from tree-ring series. *Front. Ecol. Evol.*, 3, 46.

450 Easterling, M.R., Ellner, S.P. & Dixon, P.M. (2000). Size-specific sensitivity: Applying a new
451 structured population model. *Ecology*, 81, 694–708.

452 Ellner, S.P., Childs, D.Z. & Rees, M. (2016). Data-driven modelling of structured populations.
453 *Springer International Publishing*.

454 Ettinger, A. & HilleRisLambers, J. (2017). Competition and facilitation may lead to asymmetric
455 range shift dynamics with climate change. *Glob. Change Biol.*, 23, 3921–3933.

456 Gaston, K.J. (2009). Geographic range limits: Achieving synthesis. *P. R. Soc. B.*, 276, 1395–1406.

457 Grubb, P.J. (1977). The maintenance of species-richness in plant communities: The importance
458 of the regeneration niche. *Biological Review*, 52, 107–145.

459 Hargreaves, A.L., Samis, K.E. & Eckert, C.G. (2014). Are Species' Range Limits Simply Niche
460 Limits Writ Large? A Review of Transplant Experiments beyond the Range. *Am. Nat.*, 183, 157–
461 173.

462 Harris, I., Jones, P.D., Osborn, T.J. & Lister, D.H. (2014). Updated high-resolution grids of

463 monthly climatic observations - the CRU TS3.10 Dataset: updated high-resolution grids of monthly
464 climatic observations. *Int. J. Climatol.*, 34, 623–642.

465 HilleRisLambers, J., Harsch, M.A., Ettinger, A.K., Ford, K.R. & Theobald, E.J. (2013). How will
466 biotic interactions influence climate change-induced range shifts?: Biotic interactions and range
467 shifts. *Ann. Ny. Acad. Sci.*, 1297, 112–125.

468 Holt, R.D. & Keitt, T.H. (2000). Alternative causes for range limits: A metapopulation perspec-
469 tive. *Ecol. Lett.*, 3, 41–47.

470 Holt, R.D. & Keitt, T.H. (2005). Species' borders: A unifying theme in ecology. *Oikos*, 108, 3–6.

471 Holt, R.D., Keitt, T.H., Lewis, M.A., Maurer, B.A. & Taper, M.L. (2005). Theoretical models of
472 species' borders: Single species approaches. *Oikos*, 108, 18–27.

473 Jump, A.S., Ruiz-Benito, P., Greenwood, S., Allen, C.D., Kitzberger, T. & Fensham, R. *et al.*
474 (2017). Structural overshoot of tree growth with climate variability and the global spectrum of
475 drought-induced forest dieback. *Glob. Change Biol.*, 23, 3742–3757.

476 Jung, M., Vetter, M., Herold, M., Churkina, G., Reichstein, M. & Zaehle, S. *et al.* (2007). Un-
477 certainties of modeling gross primary productivity over Europe: A systematic study on the effects
478 of using different drivers and terrestrial biosphere models: effects on simulated GPP. *Global Bio-*
479 *geochem. Cy.*, 21, GB4021.

480 Kunstler, G., Albert, C.H., Courbaud, B., Lavergne, S., Thuiller, W. & Vieilledent, G. *et al.* (2011).
481 Effects of competition on tree radial-growth vary in importance but not in intensity along climatic
482 gradients. *J. Ecol.*, 99, 300–312.

483 Lee-Yaw, J.A., Kharouba, H.M., Bontrager, M., Mahony, C., Csergo, A.M. & Noreen, A.M.E. *et*
484 *al.* (2016). A synthesis of transplant experiments and ecological niche models suggests that range
485 limits are often niche limits. *Ecol. Lett.*, 19, 710–722.

486 Louthan, A.M., Doak, D.F. & Angert, A.L. (2015). Where and When do Species Interactions Set
487 Range Limits? *Trends Ecol. Evol.*, 30, 780–792.

488 Maire, V., Wright, I.J., Prentice, I.C., Batjes, N.H., Bhaskar, R. & Bodegom, P.M. van *et al.* (2015).
489 Global effects of soil and climate on leaf photosynthetic traits and rates: Effects of soil and climate
490 on photosynthetic traits. *Global Ecol. Biogeogr.*, 24, 706–717.

491 Mauri, A., Strona, G. & San-Miguel-Ayanz, J. (2017). EU-Forest, a high-resolution tree occur-
492 rence dataset for Europe. *Sci. Data*, 4, 160123.

- 493 Moreno, A. & Hasenauer, H. (2016). Spatial downscaling of European climate data: Spatial
494 Downscaling of European Climate Data. *Int. J. Climatol.*, 36, 1444–1458.
- 495 Morris, W.K., Vesk, P.A. & McCarthy, M.A. (2013). Profiting from pilot studies: Analysing
496 mortality using Bayesian models with informative priors. *Basic Appl. Ecol.*, 14, 81–89.
- 497 Needham, J., Merow, C., Chang-Yang, C.-H., Caswell, H. & McMahon, S.M. (2018). Inferring
498 forest fate from demographic data: From vital rates to population dynamic models. *P. R. Soc. B.*,
499 285, 20172050.
- 500 Niinemets, Ü. & Valladares, F. (2006). Tolerance to shade, drought, and waterlogging of tem-
501 perate Northern Hemisphere trees and shrubs. *Ecol. Monogr.*, 76, 521–547.
- 502 Pironon, S., Papuga, G., Villellas, J., Angert, A.L., García, M.B. & Thompson, J.D. (2017). Ge-
503 ographic variation in genetic and demographic performance: New insights from an old biogeo-
504 graphical paradigm: The centre-periphery hypothesis. *Biol. Rev.*, 92, 1877–1909.
- 505 Purves, D.W. (2009). The demography of range boundaries versus range cores in eastern US
506 tree species. *P. R. Soc. B.*, 276, 1477–1484.
- 507 Putnam, R.C. & Reich, P.B. (2017). Climate and competition affect growth and survival of
508 transplanted sugar maple seedlings along a 1700-km gradient. *Ecol. Monogr.*, 87, 130–157.
- 509 Ratcliffe, S., Liebergesell, M., Ruiz-Benito, P., Madrigal Gonzalez, J., Munoz Castaneda, J.M. &
510 Kändler, G. *et al.* (2015). Modes of functional biodiversity control on tree productivity across the
511 European continent. *Global Ecol. Biogeogr.*, 25, 251–262.
- 512 R Core Team. (2019). *R: A language and environment for statistical computing*. R Foundation for
513 Statistical Computing, Vienna, Austria.
- 514 Reich, P.B. (2014). The world-wide “fast-slow” plant economics spectrum: A traits manifesto.
515 *J. Ecol.*, 102, 275–301.
- 516 Reich, P.B. & Oleksyn, J. (2004). Global patterns of plant leaf N and P in relation to temperature
517 and latitude. *P. Natl. Acad. Sci. USA.*, 101, 11001–11006.
- 518 Salguero-Gómez, R., Jones, O.R., Archer, C.R., Buckley, Y.M., Che-Castaldo, J. & Caswell, H. *et*
519 *al.* (2015). The compadrePlant Matrix Database: An open online repository for plant demography.
520 *J. Ecol.*, 103, 202–218.
- 521 Schelhaas, M.-J., Fridman, J., Hengeveld, G.M., Henttonen, H.M., Lehtonen, A. & Kies, U. *et al.*
522 (2018). Actual European forest management by region, tree species and owner based on 714,000

- 523 re-measured trees in national forest inventories. *PLOS ONE*, 13, e0207151.
- 524 Schielzeth, H. (2010). Simple means to improve the interpretability of regression coefficients:
525 Interpretation of regression coefficients. *Methods Ecol. Evol.*, 1, 103–113.
- 526 Stahl, U., Reu, B. & Wirth, C. (2014). Predicting species' range limits from functional traits for
527 the tree flora of North America. *P. Natl. Acad. Sci. USA.*, 111, 13739–13744.
- 528 Stephenson, N.L., Van Mantgem, P.J., Bunn, A.G., Bruner, H., Harmon, M.E. & O'Connell,
529 K.B. *et al.* (2011). Causes and implications of the correlation between forest productivity and tree
530 mortality rates. *Ecol. Monogr.*, 81, 527–555.
- 531 Svenning, J.-C. & Skov, F. (2004). Limited filling of the potential range in European tree species.
532 *Ecol. Lett.*, 7, 565–573.
- 533 Talluto, M.V., Boulangeat, I., Vissault, S., Thuiller, W. & Gravel, D. (2017). Extinction debt and
534 colonization credit delay range shifts of eastern North American trees. *Nat. Ecol. Evol.*, 1, 0182.
- 535 Thuiller, W., Lafourcade, B., Engler, R. & Araújo, M.B. (2009). BIOMOD - a platform for en-
536 semble forecasting of species distributions. *Ecography*, 32, 369–373.
- 537 Thuiller, W., Münkemüller, T., Schiffrers, K.H., Georges, D., Dullinger, S. & Eckhart, V.M. *et al.*
538 (2014). Does probability of occurrence relate to population dynamics? *Ecography*, 37, 1155–1166.
- 539 Tuck, S.L., Porter, J., Rees, M. & Turnbull, L.A. (2018). Strong responses from weakly interact-
540 ing species. *Ecol. Lett.*, 21, 1845–1852.
- 541 van der Plas, F., Ratcliffe, S., Ruiz-Benito, P., Scherer-Lorenzen, M., Verheyen, K. & Wirth, C.
542 *et al.* (2018). Continental mapping of forest ecosystem functions reveals a high but unrealised
543 potential for forest multifunctionality. *Ecol. Lett.*, 21, 31–42.
- 544 Wright, I.J., Dong, N., Maire, V., Prentice, I.C., Westoby, M. & Díaz, S. *et al.* (2017). Global
545 climatic drivers of leaf size. *Science*, 357, 917–921.
- 546 Wright, I.J., Reich, P.B. & Westoby, M. (2003). Least-cost input mixtures of water and nitrogen
547 for photosynthesis. *Am. Nat.*, 161, 98–111.
- 548 Wright, I.J., Reich, P.B., Westoby, M., Ackerly, D.D., Baruch, Z. & Bongers, F. *et al.* (2004). The
549 worldwide leaf economics spectrum. *Nature*, 428, 821–7.
- 550 Zanne, A.E., Tank, D.C., Cornwell, W.K., Eastman, J.M., Smith, S.A. & FitzJohn, R.G. *et al.*
551 (2014). Three keys to the radiation of angiosperms into freezing environments. *Nature*, 506, 89–92.
- 552 Zhu, K., Woodall, C.W., Ghosh, S., Gelfand, A.E. & Clark, J.S. (2014). Dual impacts of climate

553 change: Forest migration and turnover through life history. *Glob. Change Biol.*, 20, 251–264.

554 Zimmermann, N.E., Jandl, R., Hanewinkel, M., Kunstler, G., Klling, C. & Gasparini, P. *et al.*

555 (2013). Potential Future Ranges of Tree Species in the Alps. In: *Management Strategies to Adapt*

556 *Alpine Space Forests to Climate Change Risks* (eds. Cerbu, G., Hanewinkel, Marc, Gerosa Giacomo &

557 Jandl, Robert). InTech.