

1 **The cold and the drought: effects of extreme weather events on Stem**
2 **Carbon dynamic in a Mediterranean beech forest**

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30 **Summary**

31 The effects of short-term extreme events on tree functioning are still rather elusive.
32 Considering beech one of the most sensible species to late frost and water shortage, we
33 investigate the intra-annual C dynamics in stems under such conditions. Wood formation and
34 stem CO₂ efflux were monitored in a Mediterranean beech forest for three years (2015–2017)
35 which were characterized by a late frost (2016) and a summer drought (2017). The late frost
36 event reduced radial growth and consequently the amount of carbon fixed in the stem biomass
37 by 80%. Stem carbon efflux in 2016 was reduced by 25%, which can be attributed to reduced
38 effluxes due to growth respiration. Consequently, the amount of carbon emitted was higher
39 than the carbon fixed in stems, causing a net emission of 0.468 Mg C ha⁻¹ yr⁻¹. Counter to
40 our expectations, we found no effects of the 2017 summer drought on radial growth and stem
41 carbon efflux.

42 Even if forests are potentially susceptible to all weather extremes, which are becoming more
43 frequent in Mediterranean basin, the study evidenced the negative effect of spring late frost
44 but the high resilience of beech that resumed the growth in 2017, and became again C sink.

45

46 **Keywords:** *Fagus sylvatica* L. (beech), late frost, extreme weather event, wood formation,
47 stem carbon efflux, resilience, drought, growth.

48 **Introduction**

49 Even small changes in the mean or variance of a climate variable cause disproportionately
50 large changes in the frequency of extreme weather events, recognized as major drivers of
51 current and future ecosystem dynamics (Frank *et al.*, 2015). In the near future, Mediterranean
52 region is predicted to be the most vulnerable to global change between the European regions
53 (Schröter *et al.*, 2005). Changes in temperature and precipitation regimes may increase
54 drought risk (Schröter *et al.*, 2005), which can negatively affect physiological performance
55 (Rezaie *et al.*, 2018), growth and competitive strength (Peuke *et al.*, 2002) of common beech,
56 one of the most important and widespread broadleaved trees in Europe.

57 Increasing spring temperatures can trigger earlier leaf unfolding (Gordo & Sanz, 2010; Allevato *et al.*,
58 2019), which in turn results in higher risk that young leaves are exposed to spring frost, especially at
59 higher elevation (Vitasse *et al.*, 2018). Temperatures below -4°C can kill the developing new
60 shoots and leaves, thus reducing the photosynthetic area and ultimately the trees' growth. In
61 addition, depending on the intensity of damage, the formation of new leaves requires a high
62 amount of reserves (Dittmar *et al.*, 2006; D'Andrea *et al.*, 2019).

63 Tree stems play an important role in the carbon balance of forest ecosystems (Yang *et al.*,
64 2016). Part of the carbon (C) fixed by photosynthesis is allocated to stem, while some is
65 respired by stems and emitted into the atmosphere. Radial growth – an often used proxy for
66 the overall allocation of C to the stem (Bascietto *et al.*, 2004; Cuny *et al.*, 2015; Chan *et al.*,
67 2018a) – is largely related to the process of wood formation, which can be divided into five
68 (main) developmental phases: *i*) cambial cell division; *ii*) cell enlargement; *iii*) secondary wall
69 deposition and *iv*) cell wall thickening (lignifications); while *v*) in the case of vessels and
70 fibres, also genetically-programmed cell death or apoptosis (Prislan *et al.*, 2018). The whole
71 process is sensitive to many factors as leaf phenology (Michelot *et al.*, 2012), temperature
72 (Begum *et al.*, 2007), drought (Linares *et al.*, 2009), tree-size and social status (Rathgeber *et*
73 *al.*, 2011), and tree vigour (Gričar *et al.*, 2009).

74 As also described by Damesin (2003), the only current-year stem respiration may represent
75 the 8% of the overall above ground respiration and 1.4% of the annual carbon assimilation. A
76 recent global estimate shows that the only stem respiration from boreal to tropical forests
77 combined was 6.7 (\pm 1.1) Pg C yr⁻¹, which corresponds to 84% of the mean annual
78 anthropogenic emission (Yang *et al.*, 2016). Measurements of actual stem respiration (RS) are
79 difficult if not impossible (Teskey *et al.*, 2008), and the most commonly measured proxy,
80 namely stem CO₂ efflux (ES), is likely to underestimate local respiration (Trumbore *et al.*,
81 2013). Still, previous studies reported a strong correlation between RS and ES, with ES
82 ranging between 82–94% and 86–91% of RS in *Populus deltoides* (Saveyn *et al.*, 2008) and
83 *Dacrydium cupressinum* Lamb stems (Bowman *et al.*, 2005), respectively. Another study
84 using O₂ uptake as an alternative proxy for actual respiration and comparing it to traditionally
85 measured ES, showed that ES on average underestimated RS by about 41% (CO₂ was not
86 emitted locally at the point of measurement) (Hilman *et al.*, 2019). ES and RS are different
87 because part of the CO₂ produced by respiration is not released directly through the bark in
88 atmosphere, but it is dissolved in xylem sap and is carried upward by the transpiration stream
89 (Bloemen *et al.*, 2014). In addition, ES is affected by CO₂ deriving from root respiration
90 which is carried upward into the stem (Bloemen *et al.*, 2013). Moreover, part of respired CO₂
91 can be fixed in the xylem storage pools (Ubierna *et al.*, 2009)

92 ES proved to be related to stem temperature, growth rates, distribution and turnover of living
93 cells (Collalti *et al.*, 2019) nitrogen concentration (Ceschia *et al.*, 2002), to tree social class
94 (Guidolotti *et al.*, 2013) and varies seasonally due to growth (Gruber *et al.*, 2009). ES is
95 affected by growth respiration, which provides the energy for synthesizing new tissues; and
96 by maintenance respiration, which maintains existing living cells (Ceschia *et al.*, 2002).

97 Separating ES into these components allows for further investigation of stem carbon
98 budgeting and tissue costs (Chan *et al.*, 2018b). Not much is known on the interaction
99 between wood formation (xylogenesis) and ES and a deeper investigation of this link is
100 crucial, especially in the context of climate change and extreme weather events, which may
101 greatly modify the contribution of these processes to C cycle. Nevertheless, despite the crucial
102 role of extreme events and the rising attention to their prospected increasing role in future
103 climate scenarios, information on the effect of short-term extreme events on tree functioning
104 is still rather elusive (Carrer *et al.*, 2016; Gazol *et al.*, 2019).

105 In this context, we monitored xylogenesis together with ES in a Mediterranean beech forest
106 from 2015 to 2017 – a period characterized by a spring late frost (2016) and a summer
107 drought (2017) – with the objective to unravel the intra-annual C dynamics in stems under
108 these two extreme weather events. We hypothesized that: 1) cambial activity and radial
109 growth may cess soon after leaf death due to 2016 spring late frost; 2) second leaf re-
110 sprouting may start at the expense of stem growth; 3) 2017 summer drought negatively impact
111 stem biomass production and effluxes; and that 4) such extreme weather events would alter
112 the stems C dynamic at both tree and stand scale.

113

114 **Material and methods**

115 ***Study site***

116 The measurements were carried out between 2015 and 2017 on a long-term monitored beech
117 stand (*Fagus sylvatica* L.) located at Selva Piana (41°50'58" N, 13°35'17" E, 1,560 m
118 elevation) close to Collelongo (Abruzzi Region, Italy) in the Central Apennine. Site
119 information regarding forest structure, climate and soil are all in depth described in previous
120 works (Chiti *et al.*, 2010; Guidolotti *et al.*, 2013; Collalti *et al.*, 2016; Rezaie *et al.*, 2018).

121 In the night between 25th and 26th of April 2016 (Day Of Year, DOY 115), a spring late frost
122 occurred in Central and South Italy, causing leaf damage in many beech stands (Bascietto *et*
123 *al.*, 2018; Greco *et al.*, 2018; Nolè *et al.*, 2018; Allevato *et al.*, 2019). In the Selva Piana site
124 the air temperature reached – 6°C at canopy level, destroying the whole-stand canopy and
125 leaving the trees without leaves for almost two months. The 2017 growing season was
126 characterized by a prolonged summer drought, starting in June and getting more severe over
127 the next two months. In 2017, annual precipitation was 950 mm, with only 54 mm of
128 precipitation from June to August. The daily maximum average temperature was 23.9 °C in

129 August 2017, ~ 2°C warmer than the long-term average (1950–2013) (Error! Reference source
130 not found.).

131 *Tree selection, wood formation dynamic and xylem phenology*

132 Sampling was performed on five trees (Rezaie *et al.*, 2018), that were selected for their
133 similarity with site tree ring chronology, as done in other studies on wood formation and stem
134 CO₂ efflux (Ceschia *et al.*, 2002; Damesin *et al.*, 2002; Gruber *et al.*, 2009; Delpierre *et al.*,
135 2019). Micro-cores collection, and cross-sections preparation followed the standard
136 methodology described in Prislan *et al.* (2013). On each photographed cross-section, the
137 number of cambium cells was counted, and the widths of the developing xylem were
138 measured along three radial directions. The dynamics of xylem formation were analyzed by
139 fitting Gompertz function to xylem increments (Prislan *et al.*, 2018; Rathgeber *et al.*, 2018),
140 corrected for the previous tree ring width (Oladi *et al.*, 2011), as follows:

141

$$142 \quad y = \alpha \exp \left[-e^{(\beta - kt)} \right] \quad (1)$$

143

144 where y is the cumulative ring width (μm) at time t (day of the year), α is the final asymptotic
145 size representing the annual potential growth, β is the x-axis placement parameter, and k is the
146 rate of change parameter.

147 For each tree and monitoring year the following phenological xylem formation phases were
148 recorded: *i*) cambium *reactivation*, *ii*) beginning of cell *enlargement* period (bE); *iii*)
149 beginning of cell wall *thickening* (bW); *iv*) beginning of cell *maturation* (bM); *v*) *cessation* of
150 cell enlargement phase (cE); and *vi*) *cessation* of cell wall thickening and *lignification* phase
151 (cW). The date of cambium reactivation was assessed as the average between dates when an
152 increase of cambium cells was observed (i.e. from 3-4 to 6-7 cells in a radial row) (Čufar *et al.*
153 *et al.*, 2008; Deslauriers *et al.*, 2008). Phases of xylem growth and ring formation were
154 computed using logistic regressions, spanning from the 50% probability that phenophase have
155 started or ended (Rathgeber *et al.*, 2018). Based on phenological phases, the duration of key
156 wood formation phases were calculated: *i*) the overall duration of the enlargement period (dE
157 = cE – bE); *ii*) the duration of the wall-thickening period (dW = cW – bW); and *iii*) the total
158 duration of wood formation (i.e. the duration of xylogenesis) (dX = cW – bE). Data were
159 analysed using the CAVIAR (v2.10-0) package (Rathgeber *et al.*, 2018) built for R statistical
160 software (R Development Core Team, 2018).

161 Starting from the detailed time-resolved data from tree micro-cores, the annual C fixed in the
162 stem (SG_t) was estimated for each sampled tree as follows:

163

$$164 \quad SG_t = \frac{0.46 \times (BS_t - BS_{t-1})}{\Delta t} \quad (2)$$

165

166 Where SG_t is the amount of C fixed in stem per year expressed in $Mg \text{ C yr}^{-1}$, 0.46 is the
167 carbon content of the woody tissues (Scarascia-Mugnozza *et al.*, 2000), BS_t and BS_{t-1} are the
168 stem biomass in Mg of Dry Matter (DW) of at the beginning and at the end of each sampling
169 year, Δt is the time variation (one year).

170 The site-specific allometric equation for beech used for BS was that proposed by Masci
171 (2002):

172

$$173 \quad BS = \frac{283.734 \times DBH^{2.134}}{10^6} \quad (3)$$

174

175 Where BS is in Mg DW, and DBH is the diameter (in cm) at 1.30 m ($R^2 = 0.96$, p -value <
176 0.01).

177 ***Stem CO₂ efflux (ES)***

178 Two PVC collars (10 cm diameter and 5 cm high, one facing north and one south) were fixed
179 on each tree with flexible plastic ties and sealed leak tight with Terostat (Henkel KgaA,
180 Germany). When present, bark mosses and lichens were removed. Stem CO₂ efflux was
181 measured with a portable IRGA (EGM 4, PP-System, Hitchin, UK), equipped with a closed-
182 dynamic chamber (SRC-1, PP-System, Hitchin, UK), that was tightened to the collars. Each
183 measurement consisted of a 120 seconds loop, where CO₂ concentration inside the chamber
184 was measured every 5 seconds. During measurements the CO₂ concentration typically
185 increased by 10 to 50 $\mu\text{mol mol}^{-1}$.

186 Stem CO₂ efflux (ES) was calculated as:

187

$$188 \quad ES = K_{CO_2} \div V_{mol} \times \frac{V_{cuv}}{A} \quad (4)$$

189

190 ES is stem CO₂ efflux per surface area ($\mu\text{mol m}^{-2} \text{ s}^{-1}$), K_{CO_2} ($\mu\text{mol mol}^{-1} \text{ s}^{-1}$) is the slope of
191 the regression between CO₂ concentration and time during measurements, V_{mol} , the molar
192 volume, is the volume occupied by one mole of CO₂ ($\text{m}^3 \text{ mol}^{-1}$), at the air pressure (measured

193 by built-in sensor of the EGM-4) and air temperature (T_{air} , °C) at measurement time, A is the
194 exposed lateral area of stem (m^2), and V_{cuv} is the sum of SRC-1 and collar volumes (m^3).
195 Temperature was used to calculate the relationship among ES and T_{air} , using the well-known
196 Arrhenius exponential function:

197

$$198 \quad ES = a \times e^{T_{air} \times b} \quad (5)$$

199

200 ES overall temperature sensitivity for a 10 °C increase (Q_{10}) was calculated according to
201 Gruber *et al.* (2009) as:

202

$$203 \quad Q_{10} = 10^{K_T \times 10} \quad (6)$$

204

205 where K_T is the regression slope taken from linear regression of log10 of ES versus T_{air} .

206 From the wood formation and xylem phenology analysis described above, we identified wood
207 (w) and non wood (nw) formation periods for each tree, so it was possible to separate the
208 measured ES in two groups, ES_w and ES_{nw} . Thus, according to Eq. 5, we calculated per each
209 group the specific CO_2 efflux at a base air temperature of 15°C (ES_{15w} and ES_{15nw}) and the
210 specific Q_{10} (Q_{10w} and Q_{10nw}).

211 During the non-wood formation period, ES_{nw} is constituted only by the effluxes derived by
212 the maintenance respiration (ES_b , $\mu mol m^{-2} s^{-1}$) that was calculated as:

213

$$214 \quad ES_{nw} \equiv ES_b = ES_{15nw} \times Q_{10nw}^{\frac{(T_{air}-15)}{10}} \quad (7)$$

215

216 During the wood formation period ES_w ($\mu mol m^{-2} s^{-1}$), which is affected by both maintenance
217 and growth respiration, was calculated as:

218

$$219 \quad ES_w = ES_{15w} \times Q_{10w}^{\frac{(T_{air}-15)}{10}} \quad (8)$$

220

221

222 We assumed that ES_b and its relationship with air temperature was also valid during wood
223 formation period, although this approach is not accounting for the acclimation of maintenance
224 respiration to temperature during warmer periods. Nevertheless, there are contrasting
225 hypotheses on this process (Carey *et al.*, 1997; Stockfors & Linder, 1998). Under this

226 assumption, we calculated the stem CO₂ efflux due to growth respiration, ESg (μmol m⁻² s⁻¹),
227 as:

228

$$229 \quad ESg = ES_w - ESb \quad (9)$$

230

231 The daily C effluxes of the whole stem were obtained by integrating, over the entire stem
232 area, the effluxes through equation 7, 8, 9, using the half-hourly T_{air} values measured at the
233 site. The stem area was calculated as follow:

234

$$235 \quad LA = 0.464 \times DBH - 2.083 \quad (10)$$

236

237 where LA is the stem lateral area (m²) and DBH is the stem diameter (in cm) at 1.30 m (R² =
238 0.828, *p*-value < 0.01, for more details on equation see additional material). Using this
239 approach, we considered the measurement at 1.30 m representative of the whole stem, even if
240 contrasting effect of height on stem CO₂ effluxes are reported (Damesin *et al.*, 2002;
241 Katayama *et al.*, 2019).

242 Annual values of each C fluxes of the five sampled trees (TES, TESb, TESg, see Table 1 for
243 definitions) were obtained summing up the daily values.

244 ***From tree to stand level***

245 Annual values of each fluxes per tree at stand scale (ASG, AESb, AESw, AESg) were
246 estimated using a probabilistic sampling scheme that yields unbiased estimators for several
247 variables (Bascietto *et al.*, 2004, 2012).

248 DBH elevated to the power of 2.5 (DBH^{2.5}) (Bascietto *et al.*, 2012) was used as proxy
249 variable to estimate stem C fluxes. The sampling probability at the stand level (*P*_{*t_j*}) of the *j*_{th}
250 tree was calculated as

$$251 \quad P_{t_j} = \frac{A_{t_j}}{\sum_{j=1}^q A_{t_j}} \quad (11)$$

252 where *q* is the number of trees in the stand, and *A_{t_j}* is the proxy variable (DBH^{2.5}) of the *j*_{th}
253 tree. Every *j*_{th} tree yields an unbiased estimator of stem fluxes of the stand:

$$254 \quad \widehat{V}_{S_j} = \frac{W_{C_j}}{P_{t_j}} \quad (12)$$

255 Where *W_{C_j}* is the stem flux (e.g. TES, TESb, TESg, and SG) of the *j*_{th} tree. Consequently, it is
256 possible to calculate the unbiased mean and the variance of the estimator, as follow:

257
$$\widehat{V}_S = \frac{1}{n} \times \sum_{j=1}^n \widehat{V}_{S_j} \quad (13)$$

258
$$\sigma^2(\widehat{V}_S) = \frac{\sum_{j=1}^n (\widehat{V}_{S_j} - \widehat{V}_S)^2}{n \times (n-1)} \quad (14)$$

259

260 Where n is the number of sampled trees.

261 ***Meteorological and phenological data***

262 Half-hourly air temperature and precipitation were downloaded by ERA5 database of the
263 European Centre for Medium-Range Weather Forecasts (ECMWF)
264 (<https://www.ecmwf.int/en/forecasts/datasets/archive-datasets/reanalysis-datasets/era5>), and
265 downscaled to experimental site using relationships built using previously measured values
266 for temperature (°C) and precipitation (mm). Comparison against data measured at the site
267 confirm the reliability of this choice that allows to use a continuous data series, without gaps.
268 Leaf phenology was monitored using the MODIS Leaf Area Index product (LAI,
269 MOD15A2H, <https://modis.gsfc.nasa.gov/>) with 8-day temporal resolution and 500 meter
270 spatial resolution. The date of onset of photosynthetic activity (green up) and the date at
271 which plant green leaf area peaks its annual maximum (maturity) were assessed through the
272 rate of change in the curvature of the fitted logistic models (Zhang *et al.*, 2003).

273 ***Statistical data analysis***

274 Descriptive parameters of the growth and xylem phenology were tested using One Way
275 Repeated Measures Analysis of Variance, considering years as factor, followed by post-hoc
276 (Holm-Sidak method). Exponential equation was used to evaluate the relationship between ES
277 and T_{air} . Differences among ES parameters (Q_{10} and ES_{15}) were tested using Two Way
278 Repeated Measures ANOVA (Two Factor Repetition), using year and period (non-wood
279 formation, wood formation) as factors. Multiple comparisons were performed by Holm-Sidak
280 method. Linear regressions were used to assess the relationship between ES_g and SG . We
281 tested data normality and constant variance using Shapiro-Wilk test and the Spearman rank
282 correlation between the absolute values of the residuals and the observed value of the
283 dependent variable, respectively.

284 **Results**

285 ***Wood formation dynamic***

286 The date of onset of photosynthetic activity, based on leaf area index (LAI) dynamics,
287 differed among the study years, occurring at DOY 118, 95, and 127 in 2015, 2016 and 2017,

288 respectively. In all three years, cambium activation occurred after leaf unfolding at DOY 123
289 ± 4 , 118 ± 8 , 138 ± 6 in 2015, 2016 and 2017, respectively (Fig. 1). In 2016, cambium cell
290 production continued also after the late frost event, but at considerably lower rates.

291 Different intra-annual growth patterns were observed during the three study years, especially
292 in the year of the late frost (2016, Fig. 2, Table 1). In 2016, the maximal growth rate (rx) (F =
293 8.469, p -value = 0.014) was lower and was reached 3 weeks earlier (tx) (F = 22.667, p -value
294 < 0.001) than in the other two years. The different intra-annual growth patterns resulted also
295 in significantly narrower tree rings in 2016 ($230.12 \pm 1.54 \mu\text{m}$) (F = 13.272, p -value < 0.01)
296 than in 2015 ($1312.17 \pm 196.15 \mu\text{m}$) and 2017 ($1234.80 \pm 269.32 \mu\text{m}$).

297 In the study years, differences were also observed for the beginning, cessation, and duration
298 of wood formation phases (Fig. S2). The beginning of the enlargement phase occurred earliest
299 in 2016 and latest in 2017 (F = 34.789, p -value < 0.001). In contrast, the cessation of this
300 phase was observed latest in 2015 (F = 17.155, p -value < 0.01). Consequently, the duration of
301 the enlargement phase was longer in 2015 (110 ± 22 days) than in 2016 (82 ± 4 days) and
302 2017 (78 ± 4 days) (F = 8.025, p -value = 0.01). The beginning of wall thickening phase did
303 not differ among the years (F = 4.188, p -value = 0.06). The cessation of this phase occurred
304 latest in 2015 (F = 69.167, p -value < 0.001). Thus, the duration of wall thickening phase was
305 shorter in 2016 (57 ± 5 days) than 2015 (99 ± 6 days) and 2017 (75 ± 4 days) (F = 26.561, p -
306 value < 0.001). We observed also a delay in the beginning of the maturation phase in 2016 (at
307 DOY 200 ± 4) with respect to 2015 (at DOY 178 ± 4) and 2017 (at DOY 176 ± 2) (F =
308 11.650, p -value < 0.01). The overall duration of wood formation was longer in 2015 (128 ± 5
309 days) than in 2016 (98 ± 8 days) and 2017 (97 ± 8 days) (F = 12.561, p -value < 0.001).

310 ***Stem CO₂ efflux (ES)***

311 During the monitoring period (April 2015 – November 2017), the measured ES ranged
312 between $0.16 \pm 0.03 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (December 2015) and $3.01 \pm 0.40 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$
313 (August 2017) (Fig. 3). Mean ES measured in 2016 ($0.68 \pm 0.19 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) was
314 lower (F = 24.476, p -value < 0.01) than in 2015 ($1.11 \pm 0.40 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and 2017
315 ($1.29 \pm 0.30 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$).

316 In each years, ES was strongly related to air temperature through the standard exponential
317 function (Fig. 4). The relation was confirmed at the different aggregation level of
318 measurements (whole year, wood formation and non-wood formation periods; see also Table
319 S2 in Supporting Information).

320 Q_{10} shown to be 2.71 ± 0.15 , 2.11 ± 0.18 , and 2.68 ± 0.15 , in 2015, 2016, and 2017,
321 respectively. The Q_{10} parameter was not strongly affected by the sampling year (p -value =
322 0.059), although the values in 2016 were 22% lower than the other two years. Wood
323 formation affected the Q_{10} parameter ($F = 31.563$, p -value < 0.01) with Q_{10w} and Q_{10nw}
324 calculated to be 3.06 ± 0.15 and 1.93 ± 0.14 ($t = 5.571$, p -value < 0.01), respectively. This
325 difference was confirmed for all of the sampled years.

326 Also ES_{15} was affected by the different conditions of monitoring years ($F = 7.094$, p -value =
327 0.01) with mean values of 2016 ($0.63 \pm 0.07 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) lower than 2015 (0.93 ± 0.12
328 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and 2017 ($0.82 \pm 0.12 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$). Similarly to Q_{10} , wood
329 formation period affected also ES_{15} , with ES_{15w} ($0.84 \pm 0.22 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) higher than
330 ES_{15nw} ($0.73 \pm 0.02 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, $F = 7.094$, p -value = 0.01). Furthermore, during the
331 wood formation period, ES_{15w} was higher in 2015 ($1.03 \pm 0.07 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and 2017
332 ($0.90 \pm 0.07 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) than in 2016 ($0.60 \pm 0.09 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$). No differences
333 among years were found for ES during the non-wood formation periods.

334 Annual ES for individual trees ranged between $112 \text{ g C m}^{-2} \text{ yr}^{-1}$ in 2016 (tree 4) and 349 g C
335 $\text{m}^{-2} \text{ yr}^{-1}$ in 2017 (tree 2). Average total ES for all sampled trees was lower in 2016 (182 ± 25
336 $\text{g C m}^{-2} \text{ yr}^{-1}$, $F = 12.007$, p -value < 0.01) than in 2015 ($258 \pm 27 \text{ g C m}^{-2} \text{ yr}^{-1}$) and 2017 (233
337 $\pm 35 \text{ g C m}^{-2} \text{ yr}^{-1}$).

338 The estimated contribution of maintenance respiration to ES for individual trees ranged
339 between $112 \text{ g C m}^{-2} \text{ yr}^{-1}$ in 2016 (tree 4) and $284 \text{ g C m}^{-2} \text{ yr}^{-1}$ in 2017 (tree 2), and was
340 lower, on average, in 2016 ($169 \pm 21 \text{ g C m}^{-2} \text{ yr}^{-1}$) than in 2015 ($211 \pm 18 \text{ g C m}^{-2} \text{ yr}^{-1}$) ($q =$
341 5.104 , p -value = 0.017).

342 Likewise, the estimated contribution of wood formation to ES for individual trees varied
343 between 0 in 2016 (tree 4) and $70 \text{ g C m}^{-2} \text{ yr}^{-1}$ in 2015 (tree 2) and was significantly lower (F
344 = 8.144, p -value = 0.012) on average in 2016 ($14 \pm 5 \text{ g C m}^{-2} \text{ yr}^{-1}$) than in 2015 ($48 \pm 9 \text{ g C}$
345 $\text{m}^{-2} \text{ yr}^{-1}$) and 2017 ($39 \pm 8 \text{ g C m}^{-2} \text{ yr}^{-1}$). In relative terms, contribution to ES by wood
346 formation was estimated to be $18 \pm 2\%$, $9 \pm 3\%$, and $16 \pm 3\%$ in 2015, 2016, and 2017, with
347 the remaining CO_2 efflux originating from maintenance respiration.

348 ***Radial growth and stem C effluxes***

349 During the study period, annual average measured ES and tree ring widths were closely
350 related (Fig. 5). Seasonal patterns of ES were similar in the three experimental years, but with
351 different amplitudes (Fig. 6). Moreover, ES_b , the stem C effluxes affected by maintenance
352 respiration, showed a similar pattern among the three years. We observed a time-lag between

353 the time of maximum growth rate (t_x) and maximum ES values of 23 ± 2 days, 31 ± 2 days,
354 and 29 ± 1 days in 2015, 2016, and 2017, respectively. Differences between years were not
355 significant ($F = 3.317$, p -value = 0.07).

356 ***From tree to stand level***

357 Annual stand-level stem C emission (AES) was lower in 2016 (0.7 ± 0.01 Mg C ha⁻¹ yr⁻¹)
358 than 2015 (0.9 ± 0.1 Mg C ha⁻¹ yr⁻¹) and 2017 (0.9 ± 0.1 Mg C ha⁻¹ yr⁻¹) (Table 3). During
359 the experimental years, annual stem C effluxes due to maintenance respiration (AESb) were
360 similar; in contrast AES due to growth respiration (AESg) was lower in 2016 (0.05 ± 0.01 Mg
361 C ha⁻¹ yr⁻¹) than in 2015 (0.18 ± 0.03 Mg C ha⁻¹ yr⁻¹) and 2017 (0.14 ± 0.02 Mg C ha⁻¹ yr⁻¹).
362 The contribution of AESg to the annual stem effluxes was 26%, 2%, and 24% in 2015, 2016,
363 and 2017, respectively.

364 The amount of carbon fixed in stem biomass (SG) was lower in 2016 (0.2 ± 0.05 Mg C ha⁻¹
365 yr⁻¹) compared to 2015 (1.1 ± 0.1 Mg C ha⁻¹ yr⁻¹) and 2017 (1.0 ± 0.2 Mg C ha⁻¹ yr⁻¹). The
366 difference between SG and AES was positive in 2015 and 2017, but negative in 2016. At the
367 studied beech forest, the mean C construction cost of wood, assessed as the slope of the
368 relationship between AESg and SG at tree level ($R^2 = 0.849$, p -value < 0.01, see supporting
369 information Fig.S3), was 0.2 g C emitted per g C fixed. This parameter was the mean between
370 years being 0.15 ± 0.01 for 2015, 0.25 ± 0.06 for 2016, and 0.13 ± 0.06 for 2017, respectively.

371 **Discussion**

372 ***Cambial activity and radial growth are not entirely linked to leaf phenology***

373 To the best of our knowledge, for the first time the effects of a spring late frost and a
374 subsequently close summer drought on the wood formation of beech were described.

375 In all three years, cambium reactivation and wood formation occurred within 1-3 weeks after
376 leaf development, confirming the tight dependence of the radial growth on leaf phenology and
377 photosynthesis in diffuse-porous species (Čufar *et al.*, 2008; Michelot *et al.*, 2012).
378 Nevertheless, in diffuse-porous trees, stem conductivity to water occurs in several outermost
379 growth rings and is not limited to the youngest formed xylem, as in ring-porous species
380 (Schume *et al.*, 2004). Hence, in beech allocation to current year wood is not as decisive as in
381 ring-porous species, and newly formed photosynthates at the beginning of the season are
382 rather used for other crucial processes, such as foliage and fine root growth. At the same site,
383 Matteucci (1998) analysed in parallel net ecosystem exchange (NEE) and carbon allocation to
384 foliage and wood growth (dendrometers), finding that the latter started approximately 15 days
385 after photosynthesis exceeded respiration (i.e. NEE was negative). Until then, net absorbed

386 carbon was allocated mostly to foliage growth. This can be related to C allocation hierarchy
387 that identifies newly developing leaves as the main C sink as at the beginning of the growing
388 season (Capioli *et al.*, 2013; Collalti *et al.*, 2018; Merganičová *et al.*, 2019, and references
389 therein). Interestingly, and counter to our expectations, in 2016 the cambium remained active
390 at low rates even after complete canopy defoliation. D'Andrea *et al.* (2019) already showed
391 that in the absence of photosynthesis old C reserves fuelled leaves production and likely other
392 physiological activities of the tree. In the present work we found that also cambium activity,
393 besides a second leaf production, behaves as an additional C sink, giving new insight on tree
394 internal C regulation and the pivotal role of C reserve.

395 After the second re-sprouting, the cambium cell production decreased and became non-
396 productive, although the environmental conditions were still favourable for radial growth.
397 Stem radial growth of beech in Selva Piana was greatly affected by extreme late spring frost
398 event in 2016 because of the premature cessation of cambial cell production and the lower
399 growth rate during the active period, which resulted in 82% narrower annual xylem
400 increments if compared to 2015 and 2017. This can be related to a somewhat hypothesised,
401 genetically controlled, form of hierarchy in C allocation (composed by old C reserve and
402 recently fixed photosynthetates) that identifies newly developing leaves as the main C sink
403 rather to radial growth (Capioli *et al.*, 2013; Collalti *et al.*, 2018; Merganičová *et al.*, 2019).
404 Some new insights on the genetic control (rather than the environmental ones) on cambial
405 growth have only emerged recently (see Zhang *et al.* 2019; Greb, 2019) but a clear picture
406 still missing. In beech, previously reported growth reduction, as a consequence of late frost,
407 ranged from 48 to 83% in beech, with the maximum occurring at the northern fringe of the
408 Alps (Dittmar *et al.*, 2006). Radial growth rates had fully recovered in 2017 with no visible
409 long-term effects of the late spring frost event in 2016, showing the high resilience to late
410 frost of beech (Dittmar *et al.*, 2006; Principe *et al.*, 2017). However, as shown in D'Andrea *et*
411 *al.* (2019), beech trees during 2016 were able to compensate for reserve lost, completely
412 refilling the pool at the same level as before the frost event. Hence, it were not need to
413 prioritize reserve recharge over stem biomass production the subsequent year.

414 Surprisingly, the 2017 summer drought shown no effect on the stem radial growth, we can
415 only hypothesis that because it occurred when trees already completed the primary of the
416 processes, as already demonstrated in the Mediterranean region for other tree species (Forner
417 *et al.*, 2018).

418 ***Effluxes from stem are not entirely synchronised to radial growth***

419 Mean annual values of Q_{10} ranged between 2.11 (2016) to 2.71 (2015) and were similar to the
420 values estimated at the same site for co-dominant (2.59) and dominant (2.34) beech trees
421 (Guidolotti *et al.*, 2013). Q_{10nw} , and Q_{10w} estimated in this study are very comparable with the
422 dataset of various coniferous and broadleaf tree species reported in Damesin *et al.* (2002).
423 Similar intra-annual variability of Q_{10} were observed in many studies on different species
424 with higher Q_{10} during the growing period (Paembonan *et al.*, 1992; Carey *et al.*, 1997;
425 Stockfors & Linder, 1998; Gruber *et al.*, 2009). However, other studies found stable Q_{10}
426 throughout the year (Ceschia *et al.*, 2002; Damesin *et al.*, 2002).

427 As reported in other studies, ES15, stemCO₂ efflux at an air temperature of 15°C, was
428 sensitive to wood formation processes, showing an increase during the growing period
429 (Ceschia *et al.*, 2002; Damesin *et al.*, 2002).

430 Maximal xylem production and maximum ES were not synchronized while a constant delay
431 of about a month was observed, as in young beech forest where the peak of ES occurred *c.* 27
432 days after the maximum stem growth rate (Ceschia *et al.*, 2002). Furthermore, our results
433 confirmed that at the peak ES occurred when xylem cells were still in the phase of wall
434 thickening and lignification, as previously hypothesized (Ceschia *et al.*, 2002). Moreover,
435 when maximum ES was observed, it is very likely that trees were already refilling the stem C
436 reserves pool (Scartazza *et al.*, 2013).

437 ***Only spring frost affects negatively stem C fluxes***

438 The amount of C fixed by stem biomass formation in 2015 and 2017 ranged from 1.47 to 1.65
439 Mg C ha⁻¹ yr⁻¹, thus being at the low end of values reported for a beech forest in Germany
440 from 1.69 to 2.41 Mg C ha⁻¹ yr⁻¹ (Mund *et al.*, 2010). Such differences can be explained by
441 differences in environmental conditions. In 2016, however, we measured only 0.194 ± 0.05
442 Mg C ha⁻¹ yr⁻¹, i.e. only about 20% of the fixation during the two reference years,
443 emphasizing how exceptionally negative this year was.

444 Annual stem CO₂ efflux (AES) is known to be highly variable in temperate forests (Yang *et al.*
445 *et al.*, 2016). Our data ranges from 0.662 to 0.938 Mg C ha⁻¹ yr⁻¹ and thus is lower than the 1.65
446 to 2.25 Mg C ha⁻¹ yr⁻¹ reported for a younger beech forest (Damesin *et al.*, 2002).

447 Annual stem CO₂ efflux (AES) is known to be highly variable in temperate forests (Yang *et al.*
448 *et al.*, 2016). Our data ranges from 0.66 to 0.94 Mg C ha⁻¹ yr⁻¹ and thus is lower than the 1.65 to
449 2.25 Mg C ha⁻¹ yr⁻¹ reported for a younger beech forest (Damesin *et al.*, 2002).

450 An earlier estimate for AES at the study site was $0.63 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$ (Guidolotti *et al.*, 2013)
451 for 2007, another year characterized by a summer drought. The stem C efflux of the drought
452 year presented in this study ($2017, 0.85 \text{ Mg C ha}^{-1} \text{ yr}^{-1}$) was about 25% higher than 2007, this
453 could be due to an increase of stem biomass (*c.* 15% lower in 2007 than in 2017, see also
454 Collalti *et al.*, 2019) and to different measurement tools. In 2015, the contribution of AESg to
455 the annual stem effluxes (about 25%) was lower than that measured in a young beech forest
456 (Ceschia *et al.*, 2002), evidencing the importance of forest developmental stage in
457 determining wood formation and growth respiration. The construction cost ($0.23 \text{ g C fixed per}$
458 g C emitted) is in the lower end of the range reported for boreal tree species ($0.25 - 0.76 \text{ g C}$
459 $\text{g}^{-1} \text{ C}$ (Lavigne & Ryan, 1997).

460 While the late frost event in 2016 reduced both wood growth and stem CO₂ efflux with
461 respect to those measured in the other two years, the percentage reduction of growth (80%)
462 was much larger than the reduction of ES (25%). Hence, it seems that in beech the
463 contribution of growth respiration on total stem CO₂ fluxes in mature beech trees is lower
464 than that of maintenance respiration. In the year of late frost, the strong reduction of fixed
465 growth C and the contemporary lower reduction of stem CO₂ efflux strongly affected the
466 overall stem carbon balance, transforming the stems from a sink into a source of carbon. We
467 estimated the amount of carbon released into the atmosphere by stems in 2016 equal to the
468 amount of C fixed during the other two experimental years.

469 Differently, the summer drought did not have any significant effect on stem growth and
470 neither on effluxes of C related to growth respiration. The pressure of extreme events on
471 carbon cycle can become also evident at large scale, as shown in the case of the heat wave
472 occurred in the summer of 2003, when the cumulative European carbon sequestration of the
473 previous five years was released in few months (Ciais *et al.*, 2005). Forest ecosystems are
474 potentially susceptible to all climate extremes (Reichstein *et al.*, 2013), although their impact
475 on different functional processes may differ.

476 In conclusion, this study highlights the sensitivity of beech in the Mediterranean region to leaf
477 damage as a result of post-leaf out frost. Expecting leaf development to start earlier due to
478 global warming, is increasing the likelihood that spring frost may damage leaves. We
479 demonstrated that stem growth was significantly reduced due to the prolonged absence of
480 photosynthesizing leaves, despite the fact that the cambium was kept active during the leafless
481 period by using old reserves. However, the loss in growth was not completely compensated
482 for after re-growth of leaves, but rather the cambium activity ceased shortly thereafter.
483 Consequently, trees fixed less C in stem biomass, with a reduction of the stem carbon efflux

484 due to growth respiration, which was mirrored also on the annual stem C effluxes. The
485 reduction of C fixation was higher than the reduction C effluxes of stem, hence, this
486 component of forest ecosystem became an active source of C, releasing in the atmosphere an
487 amount of carbon equal to the sum of the net gain of the other two study years. It can be
488 inferred that this occurred also in other beech stands similarly affected by spring frost in the
489 Centre and South of Italy.

490

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498 **Author contribution**

499 E.D'A., N.R., G.M. contributed to the design of the research. Fieldwork was carried out by
500 E.D'A., N.R.; laboratory analysis for wood formation N.R., J.M., P.P.; data analysis was done
501 by E.D'A.; data interpretation by E.D'A., N.R., P.P., J.M., A.C., G.M., J.K.. The manuscript
502 was written by E.D'A and N.R. with major contributions by P.P., J.M., A.C., G.M., J.K..

503

504 **References**

- 505 Allevato E, Saulino L, Cesarano G, Chirico GB, D'Urso G, Falanga Bolognesi S, Rita A,
506 Rossi S, Saracino A, Bonanomi G. 2019. Canopy damage by spring frost in European
507 beech along the Apennines: effect of latitude, altitude and aspect. *Remote Sensing of*
508 *Environment* 225: 431–440.
- 509
- 510 Bascietto M, Bajocco S, Mazzenga F, Matteucci G. 2018. Assessing spring frost effects on
511 beech forests in Central Apennines from remotely-sensed data. *Agricultural and Forest*
512 *Meteorology* 248: 240–250.
- 513
- 514 Bascietto M, Cherubini P, Scarascia-Mugnozza G. 2004. Tree rings from a European beech
515 forest chronosequence are useful for detecting growth trends and carbon sequestration.
516 *Canadian Journal of Forest Research* 34: 481–492.
- 517
- 518 Bascietto M, De Cinti B, Matteucci G, Cescatti A. 2012. Biometric assessment of
519 aboveground carbon pools and fluxes in three European forests by Randomized Branch
520 Sampling. *Forest Ecology and Management* 267: 172–181.
- 521
- 522 Begum S, Nakaba S, Oribe Y, Kubo T, Funada R. 2007. Induction of Cambial Reactivation

- 523 by Localized Heating in a Deciduous Hardwood Hybrid Poplar (*Populus sieboldii* 3 P .
524 *grandidentata*). *Annals of botany*, 439–447.
- 525
- 526 Bloemen J, Agneessens L, Van Meulebroek L, Aubrey DP, McGuire MA, Teskey RO, Steppe
527 K. 2014. Stem girdling affects the quantity of CO₂ transported in xylem as well as CO₂
528 efflux from soil. *New Phytologist* 201: 897–907.
- 529 Bloemen J, McGuire MA, Aubrey DP, Teskey RO, Steppe K. 2013. Transport of root-
530 respired CO₂ via the transpiration stream affects aboveground carbon assimilation and CO₂
531 efflux in trees. *New Phytologist* 197: 555–565.
- 532
- 533 Bowman WP, Barbour MM, Turnbull MH, Tissue DT, Whitehead D, Griffin KL, Bowman
534 WP. 2005. Sap flow rates and sapwood density are critical factors in within- and between-
535 tree variation in CO₂ efflux from stems of mature *Dacrydium cupressinum* trees. *New*
536 *Phytologist* 167: 815–828.
- 537
- 538 Campioli M, Verbeeck H, Van den Bossche J, Wu J, Ibrom A, D’Andrea E, Matteucci G,
539 Samson R, Steppe K, Granier A. 2013. Can decision rules simulate carbon allocation for
540 years with contrasting and extreme weather conditions? A case study for three temperate
541 beech forests. *Ecological Modelling* 263.
- 542
- 543 Carey E V., Callaway RM, DeLucia EH. 1997. Stem respiration of ponderosa pines grown in
544 contrasting climates: implications for global climate change. *Oecologia* 111: 19–25.
- 545
- 546 Carrer M, Brunetti M, Castagneri D. 2016. The Imprint of Extreme Climate Events in
547 Century-Long Time Series of Wood Anatomical Traits in High-Elevation Conifers.
548 *Frontiers in Plant Science* 7: 1–12.
- 549
- 550 Ceschia É, Damesin C, Lebaube S, Pontailier JY, Dufrière É. 2002. Spatial and seasonal
551 variations in stem respiration of beech trees (*Fagus sylvatica*). *Annals of Forest Science*
552 59: 801–812.
- 553
- 554 Chan T, Berninger F, Kolari P, Nikinmaa E, Hölttä T. 2018a. Linking stem growth respiration
555 to the seasonal course of stem growth and GPP of Scots pine. *Tree Physiology* 38: 1356–
556 1370.
- 557
- 558 Chan T, Berninger F, Kolari P, Nikinmaa E, Hölttä T. 2018b. Linking stem growth respiration
559 to the seasonal course of stem growth and GPP of Scots pine. *Tree Physiology*: 1356–1370.
- 560
- 561 Chiti T, Papale D, Smith P, Dalmonech D, Matteucci G, Yeluripati J, Rodeghiero M,
562 Valentini R. 2010. Predicting changes in soil organic carbon in mediterranean and alpine
563 forests during the Kyoto Protocol commitment periods using the CENTURY model. *Soil*
564 *Use and Management* 26: 475–484.
- 565
- 566 Collalti A, Marconi S, Ibrom A, Trotta C, Anav A, *et al.* 2016. Validation of 3D-CMCC
567 Forest Ecosystem Model (v.5.1) against eddy covariance data for 10 European forest sites.
568 *Geoscientific Model Development* 9: 479–504.
- 569
- 570 Collalti A, Tjoelker MG, Hoch G, Mäkelä A, Guidolotti G, *et al.* 2019. Plant respiration:
571 Controlled by photosynthesis or biomass? *Global Change Biology*,
572 <https://doi.org/10.1111/gcb.14857>.

- 573
574 Collalti A, Trotta C, Keenan TF, Ibrom A, Bond-lamberty B, *et al.* 2018. Thinning Can
575 Reduce Losses in Carbon Use Efficiency and Carbon Stocks in Managed Forests Under
576 Warmer Climate. *Journal of Advances in Modeling Earth Systems* 10: 2427–2452.
577
578 Čufar K, Prislan P, De Luis M, Gričar J. 2008. Tree-ring variation, wood formation and
579 phenology of beech (*Fagus sylvatica*) from a representative site in Slovenia, SE Central
580 Europe. *Trees - Structure and Function* 22: 749–758.
581
582 Cuny HE, Rathgeber CBK, Frank D, Fonti P, Mäkinen H, *et al.* 2015. Woody biomass
583 production lags stem-girth increase by over one month in coniferous forests. *Nature Plants*
584 1: 15160.
585
586 D’Andrea E, Rezaie N, Battistelli A, Gravichkova O, Kuhlmann I, *et al.* 2019. Winter’s bite:
587 Beech trees survive complete defoliation due to spring late frost damage by mobilizing old
588 C reserves. *New Phytologist*, 224: 625–631
589
590 Damesin C, Ceschia E, Le Goff N, Ottorini J-M, Dufrêne E. 2002. Stem and branch
591 respiration of beech: from tree measurements to estimations at the stand level. *New*
592 *Phytologist* 153: 159–172.
593
594 Deslauriers A, Rossi S, Anfodillo T, Saracino A. 2008. Cambial phenology, wood formation
595 and temperature thresholds in two contrasting years at high altitude in southern Italy. *Tree*
596 *physiology* 28: 863–871.
597
598 Dittmar C, Fricke W, Elling W. 2006. Impact of late frost events on radial growth of common
599 beech (*Fagus sylvatica* L.) in Southern Germany. *European Journal of Forest Research*
600 125: 249–259.
601
602 Forner A, Valladares F, Bonal D, Granier A, Grossiord C. 2018. Extreme droughts affecting
603 Mediterranean tree species’ growth and water-use efficiency: the importance of timing.
604 *Tree physiology*.
605
606 Frank D, Reichstein M, Bahn M, Thonicke K, Frank D, Mahecha MD, Smith P, van der
607 Velde M, Vicca S, Babst F, *et al.* 2015. Effects of climate extremes on the terrestrial
608 carbon cycle: Concepts, processes and potential future impacts. *Global Change Biology*
609 21: 2861–2880.
610
611 Gazol A, Camarero JJ, Colangelo M, de Luis M, Martínez del Castillo E, Serra-Maluquer X.
612 2019. Summer drought and spring frost, but not their interaction, constrain European beech
613 and Silver fir growth in their southern distribution limits. *Agricultural and Forest*
614 *Meteorology* 278: 107695.
615
616 Gordo O, Sanz JJ. 2010. Impact of climate change on plant phenology in Mediterranean
617 ecosystems. *Global Change Biology* 16: 1082–1106.
618
619 Greco S, Infusino M, De Donato C, Coluzzi R, Imbrenda V, Lanfredi M, Simoniello T,
620 Scalercio S. 2018. Late spring frost in mediterranean beech forests: Extended crown
621 dieback and short-term effects on moth communities. *Forests* 9: 1–18.
622

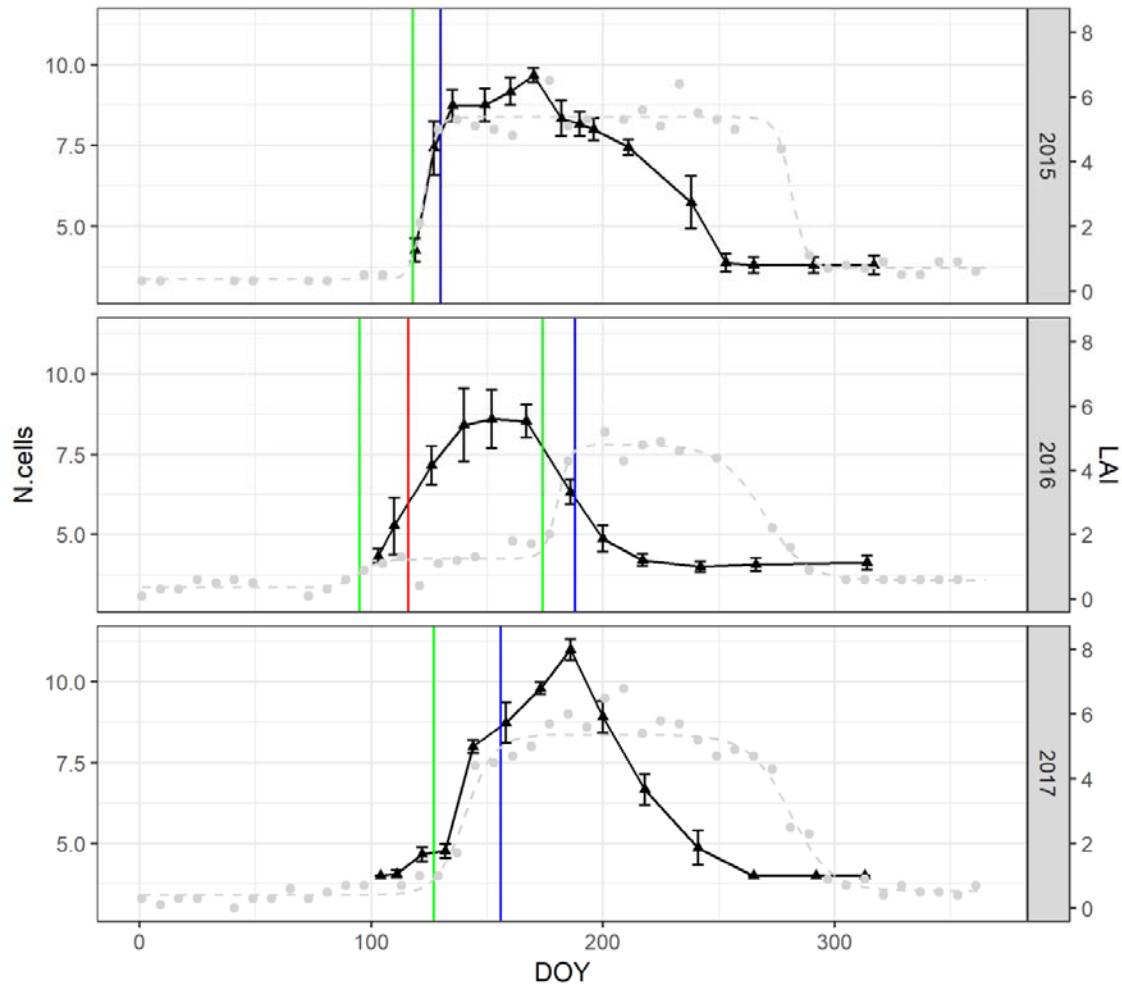
- 623 Gričar J, Krže L, Čufar K. 2009. Number of cells in xylem, phloem and dormant cambium in
624 silver fir (ABIES ALBA), in trees of different vitality. *IAWA Journal*, 30: 121–133.
625
- 626 Gruber A, Wieser G, Oberhuber W. 2009. Intra-annual dynamics of stem CO₂ efflux in
627 relation to cambial activity and xylem development in *Pinus cembra*. *Tree Physiology* 29:
628 641–649.
- 629 Guidolotti G, Rey A, D’Andrea E, Matteucci G, De Angelis P. 2013. Effect of environmental
630 variables and stand structure on ecosystem respiration components in a Mediterranean
631 beech forest. *Tree Physiology* 33: 960–972.
632
- 633 Hilman B, Muhr J, Trumbore SE, Kunert N, Carbone MS, *et al.* 2019. Comparison of CO₂
634 and O₂ fluxes demonstrate retention of respired CO₂ in tree stems from a range of tree
635 species. *Biogeosciences* 16: 177–191.
636
- 637 Lavigne MB, Ryan MG. 1997. Growth and maintenance respiration rates of aspen , black
638 spruce and jack pine stems at northern and southern BOREAS sites. *Tree Physiology*: 543–
639 552.
640
- 641 Linares JC, Camarero JJ, Carreira JA. 2009. Plastic responses of *Abies pinsapo* xylogenesis to
642 drought and competition. *Tree Physiology* 29: 1525–1536.
643
- 644 Merganičová K, Merganič J, Lehtonen A, Vacchiano G, *et al.* 2019. Forest carbon allocation
645 modelling under climate change. *Tree Physiology* doi: 10.1093/treephys/tpz105
646
- 647 Michelot A, Simard S, Rathgeber C, Dufrêne E, Damesin C. 2012. Comparing the intra-
648 annual wood formation of three European species (*Fagus sylvatica*, *Quercus petraea* and
649 *Pinus sylvestris*) as related to leaf phenology and non-structural carbohydrate dynamics.
650 *Tree Physiology* 32: 1033–1045.
651
- 652 Mund M, Kutsch WL, Wirth C, Kahl T, Knohl A, Skomarkova M V, Schulze E-D. 2010. The
653 influence of climate and fructification on the inter-annual variability of stem growth and
654 net primary productivity in an old-growth, mixed beech forest. *Tree physiology* 30: 689–
655 704.
656
- 657 Nolè A, Rita A, Ferrara AMS, Borghetti M. 2018. Effects of a large-scale late spring frost on
658 a beech (*Fagus sylvatica* L.) dominated Mediterranean mountain forest derived from the
659 spatio-temporal variations of NDVI. *Annals of Forest Science* 75: 83.
660
- 661 Oladi R, Pourtahmasi K, Eckstein D, Bräuning A. 2011. Seasonal dynamics of wood
662 formation in Oriental beech (*Fagus orientalis* Lipsky) along an altitudinal gradient in the
663 Hyrcanian forest, Iran. *Trees - Structure and Function* 25: 425–433.
664
- 665 Paembonan SA, Hagihara A, Hozumi K. 1992. Long-Term respiration in relation to growth
666 and maintenance processes of the aboveground parts of a hinoki forest tree. *Tree*
667 *Physiology* 10: 21–31.
668
- 669 Peuke AD, Schraml C, Hartung W, Rennenberg H. 2002. Identification of drought-sensitive
670 beech ecotypes by physiological parameters. *New Phytologist* 154: 373–387.
671
- 672 Principe A, Struwe T, Wilmking M, Kreyling J. 2017. Low resistance but high resilience in

- 673 growth of a major deciduous forest tree (*Fagus sylvatica* L.) in response to late spring frost
674 in southern Germany. *Trees - Structure and Function* 31: 743–751.
675
- 676 Prislán P, Čufar K, De Luis M, Gričar J. 2018. Precipitation is not limiting for xylem
677 formation dynamics and vessel development in European beech from two temperate forest
678 sites. *Tree Physiology* 38: 186–197.
- 679 Prislán P, Gričar J, De Luis M, Smith KT, Cufar K. 2013. Phenological variation in xylem
680 and phloem formation in *Fagus sylvatica* from two contrasting sites. *Agricultural and
681 Forest Meteorology*: 142–151.
682
- 683 R Development Core Team. 2018. R: A Language and Environment for Statistical
684 Computing.
685
- 686 Rathgeber CBK, Rossi S, Bontemps J-D. 2011. Cambial activity related to tree size in a
687 mature silver-fir plantation. *Annals of Botany* 108: 429–438.
688
- 689 Rathgeber CBK, Santenoise P, Cuny HE. 2018. CAVIAR: An R package for checking,
690 displaying and processing wood-formation-monitoring data. *Tree Physiology* 38: 1246–
691 1260.
692
- 693 Rezaie N, D’Andrea E, Bräuning A, Matteucci G, Bombi P, Lauteri M. 2018. Do atmospheric
694 CO₂ concentration increase, climate and forest management affect iWUE of common
695 beech? Evidences from carbon isotope analyses in tree rings. *Tree Physiology* 1975: 1110–
696 1126.
697
- 698 Saveyn A, Steppe K, Mc Guire MA, Lemeur R, Teskey RO. 2008. Stem respiration and
699 carbon dioxide efflux of young *Populus deltoides* trees in relation to temperature and
700 xylem carbon dioxide concentration. *Oecologia* 154: 637–649.
701
- 702 Scarascia-Mugnozza G, Bauer GA, Persson H, Matteucci G, Masci A. 2000. Tree Biomass,
703 Growth and Nutrient Pools. In: Schulze E-D, ed. Carbon and Nitrogen Cycling in
704 European Forest Ecosystems. Springer Verlag, 49–62.
705
- 706 Scartazza A, Moscatello S, Matteucci G, Battistelli A, Brugnoli E. 2013. Seasonal and inter-
707 annual dynamics of growth, non-structural carbohydrates and C stable isotopes in a
708 Mediterranean beech forest. *Tree physiology* 33: 730–42.
709
- 710 Schröter D, Cramer W, Leemans R, Prentice IC, Araújo MB, Arnell NW, Bondeau A,
711 Bugmann H, Carter TR, Gracia CA, et al. 2005. Ecosystem service supply and
712 vulnerability to global change in Europe. *Science (New York, N.Y.)* 310: 1333–7.
713
- 714 Schume H, Grabner M, Eckmüllner O. 2004. The influence of an altered groundwater regime
715 on vessel properties of hybrid poplar. *Trees* 18: 184–194.
716
- 717 Stockfors J, Linder S. 1998. Effect of nitrogen on the seasonal course of growth and
718 maintenance respiration in stems of Norway spruce trees. *Tree Physiology* 18: 155–166.
719
- 720 Teskey RO, Saveyn A, Steppe K, McGuire MA. 2008. Origin, fate and significance of CO₂ in
721 tree stems. *New Phytologist* 177: 17–32.
722

- 723 Trumbore SE, Angert A, Kunert N, Muhr J, Chambers JQ. 2013. What's the flux? Unraveling
724 how CO₂ fluxes from trees reflect underlying physiological processes. *New Phytologist*
725 197: 353–355.
726
- 727 Ubierna N, Kumar AS, Cernusak LA, Pangle RE, Gag PJ, Marshall JD. 2009. Storage and
728 transpiration have negligible effects on δ¹³C of stem CO₂ efflux in large conifer trees.
729 *Tree Physiology* 29: 1563–1574.
730
- 731 Vitasse Y, Schneider L, Rixen C, Christen D, Rebetez M. 2018. Increase in the risk of
732 exposure of forest and fruit trees to spring frosts at higher elevations in Switzerland over
733 the last four decades. *Agricultural and Forest Meteorology* 248: 60–69.
734
- 735 Yang J, He Y, Aubrey DP, Zhuang Q, Teskey RO. 2016. Global patterns and predictors of
736 stem CO₂ efflux in forest ecosystems. *Global Change Biology* 22: 1433–1444.
737
- 738 Zhang X, Friedl M a., Schaaf CB, Strahler AH, Hodges JCF, Gao F, Reed BC, Huete A.
739 2003. Monitoring vegetation phenology using MODIS. *Remote Sensing of Environment*
740 84: 471–475.
741
742
743
744
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748

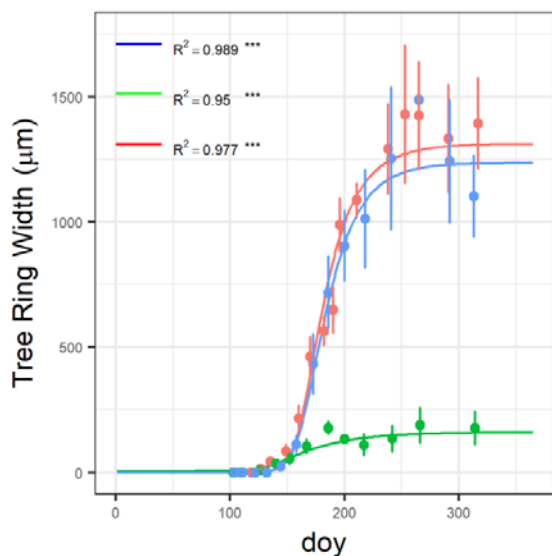
749 **Figure**

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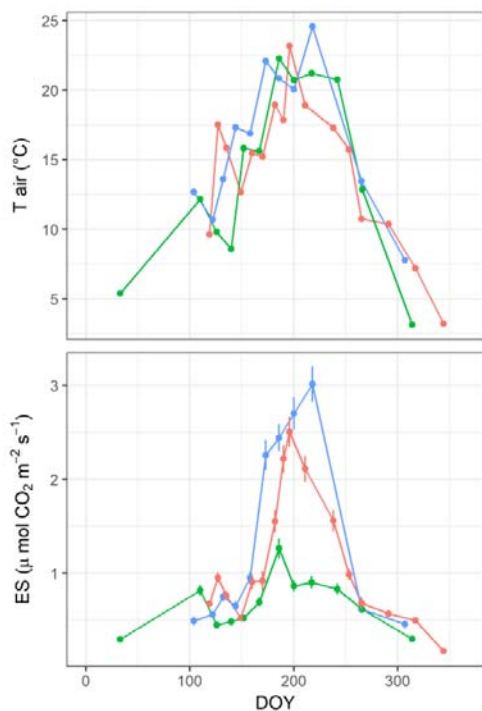
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752 **Fig. 1: Number of cambium cells (N. cells vs. Leaf area index (LAI, m^2m^{-2})).** Grey points and
753 **dashed lines are the MODIS-LAI values and modelled intra-annual dynamic of Selva Piana**
754 **beech forest, respectively. Green and blue vertical lines represent the green up and maturity**
755 **phase of leaf phenology, respectively. Red vertical line represents the late frost of 25th April**
756 **2016. Black triangles are the average number of cambial cells of five beech trees. Bars are the**
757 **standard error.**



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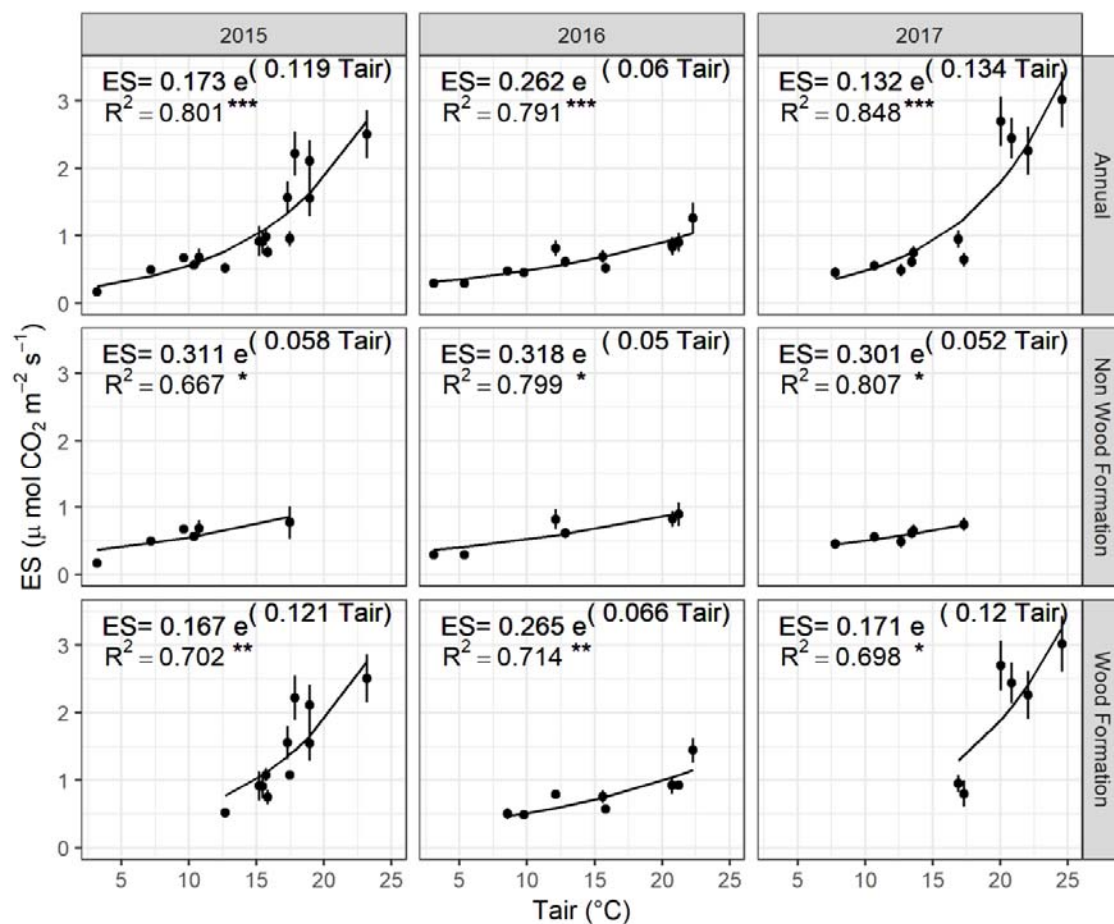
759 **Fig. 2: Intra-annual dynamics of xylem formation (µm) in 2015 (red dots and solid line), 2016**
760 **(green dots and solid line) and 2017 (blue dots and solid line). Gompertz functions were fitted to**
761 **the total xylem increment comprised of enlarging, wall thickening and mature cells. Each point**
762 **is the mean of the 5 sampled *Fagus sylvatica* trees and bars are standard errors. *** *p*-value <**
763 **0.001**



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765 **Fig. 3: Top panel: T_{air} (°C) at the measuring time in 2015 (Red), 2016 (Green), and 2017 (Blue).**
766 **Bottom panel: Measured stem CO₂ effluxes (µmol CO₂ m⁻² s⁻¹) in 2015 (Red), 2016 (Green), and**
767 **2017 (Blue). Each point is the mean of 5 *Fagus sylvatica* trees. Bars are standard errors.**

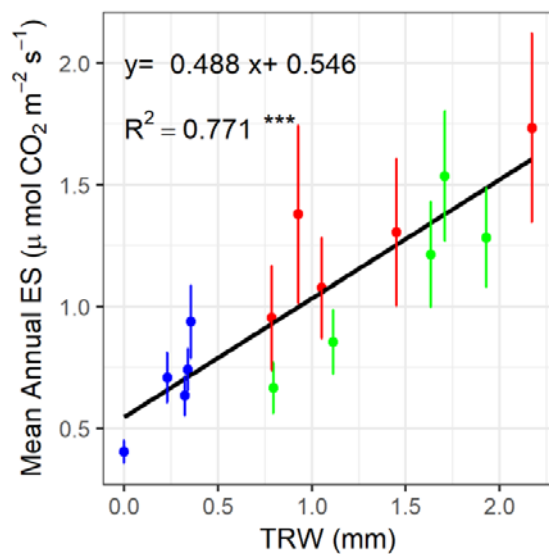
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770 **Fig. 4: Relationship between ES ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and air temperature (T_{air} , $^{\circ}\text{C}$). Annual,**
 771 **considering the whole measurements for each year, each point is the mean of the five sampled**
 772 **trees. Each point represents the mean of those trees, during non-wood and wood formation**
 773 **periods, at a given sampling date. Bars are the standard error. *** p -value < 0.001, ** p -value <**
 774 **0.01, * p -value < 0.05.**

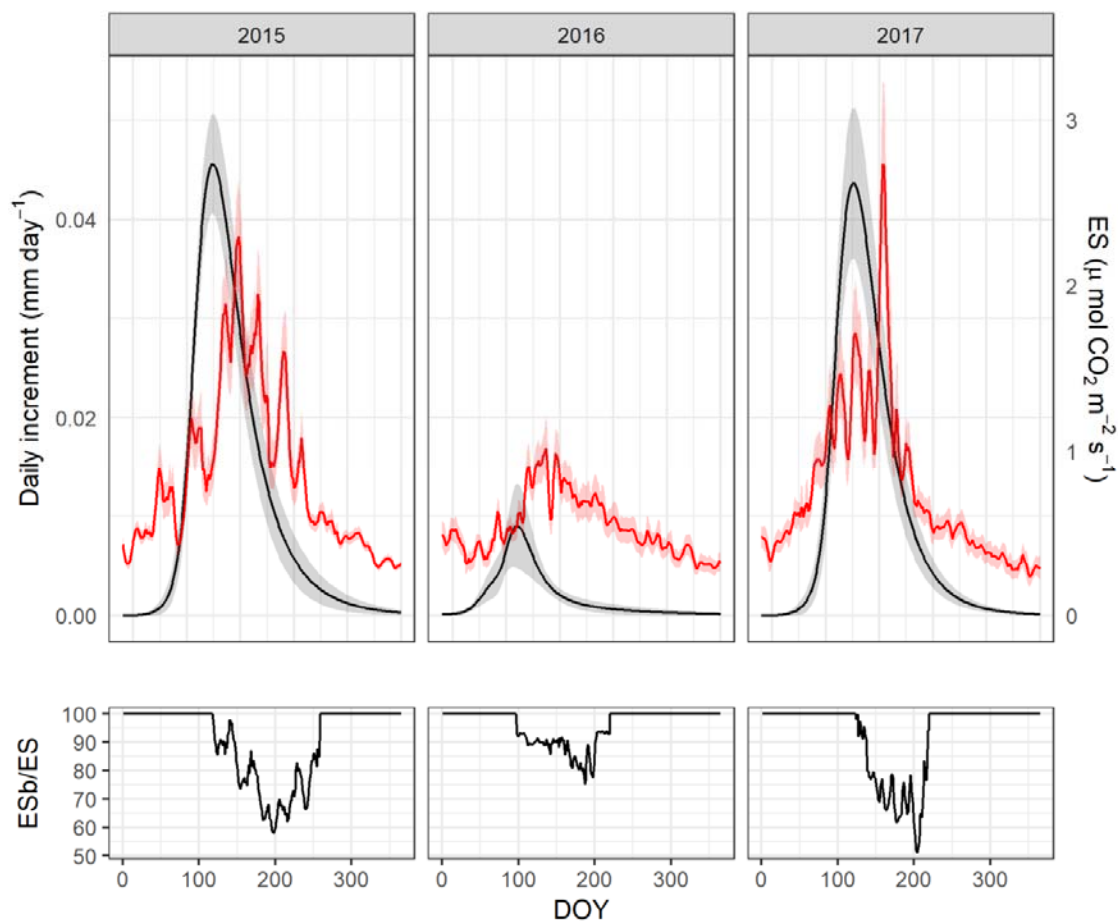
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777 **Fig. 5: Relationships between the ring widths (mm) and the mean annual ES ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$)**
778 **measured in 2015 (Green), 2016 (Blue), and 2017 (Red). Each point represents a sampled *Fagus***
779 ***sylvatica* tree per year. Bars are the standard error. *** p -value < 0.001.**

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782 **Fig. 6: Daily increment and stem Carbon effluxes of *Fagus sylvatica*. Black line is the daily**
783 **increment during 2015, 2016 and 2017, respectively. Red line represents ES, the daily C effluxes.**
784 **The below panel, ESb/ES shows the contribution of maintenance respiration to daily stem C**
785 **efflux.**

786 **Tables**

787 **Table 1: List of terms used in the text**

Terms	Definition	Spatial Scale
ES	stem CO ₂ efflux per surface area (μmol m ⁻² s ⁻¹)	Local
ES _{15w}	specific CO ₂ efflux at an air temperature of 15°C during the wood formation period (μmol m ⁻² s ⁻¹)	Local
ES _{15nw}	specific CO ₂ efflux at an air temperature of 15°C during the non-wood formation period (μmol m ⁻² s ⁻¹)	Local
Q _{10w}	ES temperature sensitivity for a 10 °C increase during the wood formation period	Local
Q _{10nw}	ES temperature sensitivity for a 10 °C increase during the non-wood formation period	Local
ESw	stem CO ₂ efflux per surface area (μmol m ⁻² s ⁻¹) during wood formation	Local
ESb	stem CO ₂ efflux per surface area (μmol m ⁻² s ⁻¹) due to maintenance respiration	Local
ESg	stem CO ₂ efflux per surface area (μmol m ⁻² s ⁻¹) due to growth respiration	Local
AES	annual stem C efflux (Mg C ha ⁻¹ yr ⁻¹)	Stand
AESb	annual stem C efflux due to maintenance respiration (Mg C ha ⁻¹ yr ⁻¹)	Stand
AESg	annual stem C efflux due to growth respiration (Mg C ha ⁻¹ yr ⁻¹)	Stand
SG	annual C fixed in stem biomass (Mg C ha ⁻¹ yr ⁻¹)	Stand

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794 **Table 2: Parameters describing the intra annual radial growth derived from Gompertz function**
 795 **for the total xylem increment comprised of enlarging, wall thickening and mature cells. α is the**
 796 **upper asymptote, representing the final ring width at the end of the growing season; t_x is the**
 797 **DOY at which the daily increment is maximal (Gompertz curve inflection point); r_x is the**
 798 **maximal daily increment ($\mu\text{m day}^{-1}$). Different letters represent significant differences among**
 799 **the monitored years.**
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Total xylem increment				
Parameter	Year	Mean (\pm S.E.)	F	<i>p</i> -value
α	2015	1312.17 (\pm 196.15) a		
α	2016	230.12 (\pm 31.54) b	13.722	< 0.01
α	2017	1234.80 (\pm 269.32) a		
r_x	2015	25.77 (\pm 2.95) a		
r_x	2016	6.09 (\pm 2.00) b	8.469	0.014
r_x	2017	22.71 (\pm 3.51) a		
t_x	2015	174 (\pm 2.19) a		
t_x	2016	157 (\pm 5.65) b	22.667	< 0.001
t_x	2017	174.80 (\pm 1.24) a		

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809 **Table 3: Annual C stem fluxes. AES is the annual stem C efflux assessed using specific**
 810 **parameters for wood formation (Q_{10w} and ES_{15w}) and non-wood formation periods (Q_{10nw} and**
 811 **ES_{15nw}); AESb is the annual stem C efflux due to maintenance respiration; AESg is the annual**
 812 **stem C efflux due to growth respiration; SG is the annual amount of C fixed in the stem**
 813 **biomass; different letters represent significant differences p -value < 0.05**
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Year	Flux type	Mean (\pm S.E.) Mg C ha ⁻¹ yr ⁻¹	F	p -value
2015	AES	0.938 (\pm 0.084) a		
2016	AES	0.662 (\pm 0.008) b	13.933	0.002
2017	AES	0.854 (\pm 0.104) a		
2015	AESb	0.760 (\pm 0.053) a		
2016	AESb	0.615 (\pm 0.072) b	6.270	0.023
2017	AESb	0.715 (\pm 0.090) a/b		
2015	AESg	0.178 (\pm 0.033) a		
2016	AESg	0.049 (\pm 0.017) b	8.310	0.009
2017	AESg	0.139 (\pm 0.027) a		
2015	SG	1.151 (\pm 0.151) a		
2016	SG	0.194 (\pm 0.050) b	18.620	< 0.001
2017	SG	1.031 (\pm 0.183) a		
2015	SG-AES	0.213 (\pm 0.087) a		
2016	SG-AES	-0.468 (\pm 0.055) b	16.838	0.001
2017	SG-AES	0.177 (\pm 0.105) a		

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