1 Divergent population recovery trajectories following sex-biased perturbation in life history traits
2

3 Daisuke Goto[1,2,*], Martin J. Hamel[3], Jeremy J. Hammen[4], Matthew L. Rugg[5], Mark A. Pegg[3],
4 and Valery E. Forbes[2,6]
5
[1]Institute of Marine Research/Havforskningsinstituttet, Postboks 1870, Nordnes, 5817 Bergen,
Norway
[2]School of Biological Sciences, Manter Hall, University of Nebraska-Lincoln, Lincoln,
Nebraska, 68588 USA
[3]School of Natural Resources, Hardin Hall, 3310 Holdrege Street, University of Nebraska-
Lincoln, Lincoln, Nebraska 68583-0973 USA
[4]U.S. Fish and Wildlife Conservation Office, 101 Park Deville Drive, Suite A, Columbia,
Missouri 65203 USA
[5]Montana Fish, Wildlife & Parks, 907 North Kendrick Avenue, Glendive, Montana 59330 USA
[6]Department of Ecology, Evolution and Behavior, University of Minnesota, 123 Snyder Hall,
1475 Gortner Ave, St. Paul, Minnesota 55108 USA
*Corresponding author Email: daisuke.goto2@gmail.com

Abstract
19
Uncertainty in risks posed by emerging stressors such as synthetic hormones impedes the
management and conservation of threatened and endangered long-lived vertebrate populations.
Synthetic hormones often induce sex-biased perturbations in exposed animals by disrupting
gonad development and early life-history stage transitions, threatening population persistence.
We use a spatially-explicit biophysical model to evaluate how sex-biased perturbations in life
history traits of individuals (maternal investment in egg production and sex determination)
modulate year class strength and population recovery trajectories of a long-lived, migratory fish–
shovelnose sturgeon (*Scaphirhynchus platorynchus*)–under spatially and temporally dynamic
synthetic androgen exposure. Simulations revealed that reduced efficiency of maternal energetic
investment in gonad development prolonged maturation time, increased the probability of
skipped spawning, and, in turn, gradually shrunk spawner biomass, ultimately weakening year
class strength and extending recovery time at higher exposure. Demographic perturbation in
individuals emerged as vastly different population projection and recovery trajectories: the
responses to skewed sex allocation were delayed at the population level by more than two
decades, followed by an abrupt shift in year class strength and population size, ultimately
reaching an alternate state without any sign of recovery. When energetics and demographics
were both perturbed, they acted antagonistically, allowing the population to gradually recover
without reaching a stable state. The emergent patterns in long-term population projections
suggest the complexity of cumulative effects of sex-biased perturbation in life history traits of
individuals exposed to low-level (often below detection) synthetic hormones, which can be
further obscured by lagged responses and environmental stochasticity. Our study illustrates that
the recovery from skewed sex allocation can pose greater conservation challenges than
diminished maternal investment in egg production–highlighting the importance of population
monitoring for restoring depleted long-lived populations such as *Scaphirhynchus* sturgeon in
ever-changing large rivers.
45
46   *Keywords:* early life history; endocrine disruptors; life history traits; endangered species; agent-
47   based modeling; spatially explicit models; conservation

1. INTRODUCTION

Humans activities have uniquely shaped life history strategies of wild populations of animals by imposing unnatural stress for centuries (Crain, Kroeker & Halpern 2008; Hendry, Farrugia & Kinnison 2008). The symptoms of human-induced stressors, in turn, manifest in life history traits varyingly among individuals (Marr *et al.* 2006) over time and space (Hope 2005; Maxwell *et al.* 2013). Sexual dimorphism, for example, is prevalent in natural populations (Fairbairn 1997), often leading to sex-biased pressure (Coltman *et al.* 2003). As stress becomes intensified, these sources of intraspecific variation in fitness may emerge as a directional shift in demographic rates (Wedekind 2017), diminishing reproductive output (Knell & Martínez-Ruiz 2017). Because of sex-dependent differential investment in reproduction, sex-biased stresses may pose a unpredictable threat to persistence of wild populations (Coltman *et al.* 2003).

One emerging stressor with such effects is synthetic chemicals, which have been increasingly detected in aquatic systems around the globe (Bernhardt, Rosi & Gessner 2017). A rising demand for human food production has increased agricultural use of synthetic chemicals (Tilman 1999; Tilman *et al.* 2001), posing a greater risk to freshwater vertebrates (Bernhardt, Rosi & Gessner 2017). Synthetic hormones–often used as growth-promoters for farmed animals–are especially potent, impairing fish development and reproduction by disrupting energy allocation and maturation (del Carmen Alvarez & Fuiman 2005). These indirect processes thus stunt somatic and gonadal growth (Arcand Hoy & Benson 1998). In pre-spawning females, even low-dose (often below detection) exposure to synthetic hormones can disrupt maturation processes such as vitellogenin production–a critical step in egg production (Arcand Hoy & Benson 1998; Nash *et al.* 2004). Because maternal investment in egg production incur greater energetic costs than somatic growth, females may more likely experience reproductive impairment than males from chemical perturbation in gonad development (Arukwe & Goksøyr 1998). Further, synthetic hormones can irreversibly perturb sex determination, skewing sex allocation in offspring (Mills & Chichester 2005). Modified demographic traits such as skewed sex ratio have been reported in wild animal populations with sex-biased mortality (McTavish, Stech & Stay 1998). However, it remains unclear how skewed sex allocation in early life stages resulting from exposure to synthetic hormones influence population persistence and recovery in nature (Ankley *et al.* 2008).

We here present a simulation study on how long-term sex-biased perturbations in life history traits can emerge and modify recovery potential of a migratory fish exposed to low-level synthetic androgens in a large river. Migratory animals are exposed to spatially and temporally dynamic flow-dependent exposure regimes as the chemicals move throughout networks of streams and rivers (O'Brien *et al.* 2016); variable exposure and subsequent symptoms among exposed individuals can obscure a causal link, if any, to population-level responses, delaying management actions (Peterson *et al.* 2003). We test varying exposure and effect scenarios on the Missouri River shovelnose sturgeon (*Scaphirhynchus platorynchus*) as an illustrative case study. Population sizes of *Scaphirhynchus* sturgeon have declined in many of their endemic habitats, and some are facing potential extirpation owing to extensive habitat modification (e.g., impoundments) (Bramblett & White 2001). Further, some of the Missouri River populations have shown signs of unsuccessful spawning (DeLonay *et al.* 2009a) and year class failure owing to low larval fish survival (Hrabik *et al.* 2007)–a threat to population persistence (DeLonay *et al.* 2009a). We use a spatially-explicit individual-based biophysical model to evaluate how modified energetics (reduced maternal investment in egg production) and demographics (skewed sex

93   allocation)–two commonly reported symptoms of synthetic androgens–influence year class
94   success, and population persistence and recovery of sturgeon.
95
96   2. MATERIALS AND METHODS
97
2.1. Study system
The study system is the lower 162 km of the Platte River (Nebraska, USA), one of the largest
tributaries of the Missouri River basin (Fig. 1a and S1), providing spawning and nursing habitats
essential for threatened and endategered fish populations such as *Scaphirhynchus* sturgeon
(Peters & Parham 2008). Due to its proximity to wastewater treatment plants and agricultural
fields, however, the lower Platte River is also one of the impaired water bodies in the region
(NDEQ 2004). Fish that inhabit this river and its tributaries such as *Scaphirhynchus* sturgeon are
exposed to elevated concentrations of synthetic hormones widely used at concentrated cattle
feedlots near the lower Platte River watershed (Huntzinger & Ellis 1993) and exhibit symptoms
of endocrine disruption (e.g., intersexuality and atresia, DeLonay *et al.* 2009a).
108
2.2. Sturgeon biophysical model
We use a spatially explicit individual-based biophysical model of shovelnsoe sturgoen
previously developed for the Lower Platte River popuatlion; the model is fully described in Goto
et al. (2015) and Goto *et al.* (2018). The model is designed to uncover ecological mechanisms
underlying early life-history processes modified by natural and human-induced stressors, and
their contributions to population persistence under spatially and temporally dynamic habitat
conditions. The model primarily hinges on energetics principles to understand how maternal and
environmental influences modulate year class variability through energy acquisition and
allocation (Houde 1989). In this study, we evaluated sex-biased effects of synthetic androgen
perturbation in maternal investment in gonad development and first-year sex allocation (Fig. 1b
and S2). The following briefly describes the essetial features of the model relevant to this study.
The model river consists of 162 equally divided longitudinal rectangular segments (i.e., 1 rkm
each). Model inputs include water temperature (°C), mean gauge height (depth, m), channel
width (m), discharge rate ($m^3 s^{-1}$), and photoperiod (hrs). Water temperature, gauge height,
channel width, and discharge rate (1995–2011) were derived from the USGS National Water
Information System (http://waterdata.usgs.gov/nwis). Further details on derivation of physical
and hydrological input data are described in Goto *et al.* (2018).
The sturgeon model keeps track of state variables of the entire life cycle (Fig. 1b). Early life
stages (fertilized eggs, yolk-sac larvae, and feeding larvae, and age-0 juveniles) are represented
as super-individuals (Scheffer *et al.* 1995), whereas age 1+ juveniles and adults are represented
as individuals. Each super-individual is a collection of individuals ("cohorts") from the same
mother. We initialized simulations with 60 000 age 1+ individuals on January 1st. The initial
location of each sturgeon was randomly assigned to a cell with non-extreme water velocity (≤0.6
$m s^{-1}$) and depth (≥0.5 m). The initial age structure follows a Poisson distribution with mean age
of 7, which is truncated at a maximum age of 16, and the initial sex ratio is 1:1. Sex was
randomly assigned to each fish, assuming a binomial distribution. For age 1+ sturgeon, initial
lengths-at-age were computed using the von Bertalanffy growth function (Ricker 1975)
parameterized for the Platte River population (Goto *et al.* 2018). Once spawned, fertilized eggs
were initialized as super-individuals; the number of super-individuals is the number of mothers,
and the number of individuals in each super-individual is the total number of eggs spawned by

each mother. After settlement, feeding larvae were initialized with individual-level processes
(e.g., growth) and traits (e.g., body size).
The model simulates the population and updates state variables at discrete daily time steps over
years and generations. The habitat conditions (including prey production) are first updated, and
then the model evaluates actions by sturgeon in the following order; (i) *spawning,* (ii) *hatching*
and larval development, (iii) *movement,* (iv) *foraging,* (v) *growth*, and (vi) *mortality* (Fig. 1b).
Density dependence is incorporated in directly foraging and indirectly in growth. Sturgeon's
abilities to sense habitat conditions and prey density, capture prey, decide when to spawn, and
survive are partially stochastic because of inherent uncertainty and randomness.
The model consists of one benthic prey submodel and six sturgeon submodels (Goto *et al.*
2018). Benthic prey production is computed using a mass-balance model described by Rashleigh
& Grossman (2005) with modifications described in Goto *et al.* (2018). The early stages of larval
development are computed using temperature-dependent nonlinear functions (Goto *et al.* 2018).
Foraging is computed using a modified version of the Beddington–DeAngelis functional
response model (Beddington 1975; DeAngelis, Goldstein & O'neill 1975). Growth is simulated
by the bioenergetics model (Hanson *et al.* 1997) modified so that the assimilated energy is
allocated to storage (reversible) and structure (irreversible) (Gurney *et al.* 2003). Sturgeon
mature based on sex-specific probability, which is length-dependent functions parameterized for
the Platte River population (Goto *et al.* 2018). Reproductive adults further allocate energy to
gonad incrementally between spawning events. Adults initiate the reproductive cycle only when
their physiological condition is above a threshold based on relative storage mass. The amount of
daily energy allocated to gonad is further adjusted when the physiological condition is below the
threshold (Goto *et al.* 2018). Spawning is simulated as a stochastic event and occurs only when
spawning cues–water temperature, river discharge rate, daylight hours, and gonad development
(gonadosomatic index or GSI), concurrently meet the thresholds (Delonay *et al.* 2016). Water
temperature suitable for the female *Scaphirhynchus* sturgeon spawning ranges is set at 12.6–
24.9°C (DeLonay *et al.* 2009b). A minimum threshold for river discharge rate was set at 141.6
$m^3 s^{-1}$, which is expected to reach roughly 50% connectivity (migratory pathway connectedness)
in the Lower Platte River (Peters & Parham 2008) and directly influences movement and
subsequently habitat conditions for other individual-level actions. A minimum GSI to initiate
spawning is set at 0.18 for females and 0.03 for males (Wildhaber *et al.* 2007a). After spawning,
fish lose all gonad mass, and their physiological condition and total mass are updated; fish do not
re-enter the reproductive cycle until the physiological condition meets the threshold. When either
temperature or discharge rate does not meet the thresholds, fish skip spawning. We assume that
when the physiological condition does not meet the threshold to remain in the reproductive
cycle, gonadal energy is reabsorbed and reallocated to storage. Movement is tracked in the
longitudinal direction only. The planktonic life stage fish (eggs and yolk-sac larvae) move by
passive drift with water currents (Braaten *et al.* 2008). Feeding larvae, juveniles, and adults move
in the direction with probability of attaining higher habitat quality using the empirical models
developed for the Lower Platte River population (Peters & Parham 2008). Mortality results from
thermal stress (eggs and yolk-sac larvae), predation (eggs, larvae, and juveniles), starvation, and
recreational fishing (age 3+). We assume that a constant proportion of eggs and yolk-sac larvae
die when water temperature rises above 24 °C (Kappenman, Webb & Greenwood 2013).
Realized predation, starvation, and fishing mortality are a stochastic event. Starvation mortality
depends on relative storage mass (Goto *et al.* 2018).
184

2.3. Simulation scenarios for sex-biased purtubations by synthetic androgens
We ran 57-year (27-year exposure + 30-year recovery) simulations for the following four
scenarios to assess the population performance under two types of perturbations in life history
traits of individuals induced by low-level synthetic androgens, energetics and demorgraphics;
1) No sex-biased purtubation–baseline;
2) Synthetic androgens perturb maternal energetic investment in egg production by imposing
reduction in vitellogenin production rate;
3) Synthetic androgens perturb sex allocation by imposing skewness in sex ratio at recruitment;
and
4) Synthetic androgens perturb both maternal energetic investment and sex allocation. We
assumed no direct mortality caused by synthetic androgen exposure. Further, we assumed that: a)
synthetic androgens are released between May 1st and October 31st (a six-month growing
season); b) androgens affect only fish located below the confluence with the Elkhorn River (~51
rkm from Plattsmouth, NE, Fig 1a), where a number of cattle feedlots are located; and c) because
the androgen concentration varies with stream flow rate, the realized exposure level for
individual fish is spatially and temporally dynamic, and is thus adusted based on flow rate in a
grid cell in which the given fish is located.
Energetic perturbation. We explicitly evaluated the mechanism underlying synthetic androgen-
induced reduction in egg production by simulating the dose-dependent reduction of vitellogenin
production in individual fish based on an exponential decay function (Ankley *et al.* 2003).
Vitellogenin (VTG) concentration (mg ml^{-1}) is converted to relative gonad mass (GSI) using an
empirically derived linear function based on field survey data on the Missouri River shovelnose
sturgeon reported by Wildhaber *et al.* (2007b) (Fig. S2) as GSI = 0.984VTG + 35.103. Reduction
in daily energy allocation to gonad (g d^{-1}) in exposed females was then computed based on
somatic mass on the given day.
We tested the scenarios of 15%, 25%, and 35% reduction in daily vitellogenin production in
fish that are located at the Elkhorn River–Platte River confluence; these reduction rates in
vitelogenin production approximately corresspond to the responses in fish exposed to
environmentally realistic, low level sysnthetic androgens (e.g., <~15 ng l^{-1} 17β-trenbolone, a
synthetic androgen commonly used in cattle feedlots, Ankley *et al.* 2003). For fish that are
located in cells below the confluence, the reduction rates were proportionally adjusted by the
difference in flow rate (e.g., if flow rates at the confluence and in the cell where a fish is located
are 100 and 120 m^3 s^{-1}, respectively, the realized reduction rate for this fish is 12.5% under the
15% reduction scenario). Because the model sturgeon cannot detect low-level synthetic
androgens in water, we assumed that their movement and energy acquisition (foraging) vary
independently of synthetic androgens.
Demographic perturbation. We implemented synthetic androgen-induced skewness in sex
allocation by proportionally reducing the probability of being female. We tested the scenarios of
20%, 60%, and 80% reduction in probability of being females for post-hatch fish that remain in
the cell at the Elkhorn River–Platte River confluence; these rates approximately corresspond to
the responses in fish exposed to environmentally realistic concentrations (e.g., <~15 ng l^{-1} 17β-
trenbolone, Morthorst, Holbech & Bjerregaard 2010). We assumed that the probability of being
female incrementally declines by remaining in the exposure area each day (i.e., 0.11%, 0.33%,
and 0.44%, respectively). For fish in cells below the confluence, the probability was
proportionally adjuted by the difference in flow rates (as above).

For all the scenarios, we used hydrological and temperature inputs empirically derived from
the lower Platte River field surveys during 1995–2011 (Goto *et al.* 2018), from which we
randomly selected each year; we assumed no systematic change or shift in hydrological and
thermal conditions in this study. Further, we assumed no change in turbidity, channel
geomorphology, and photoperiod. We ran ten simulations for each exposure scenario and
computed summary statistics and % deviations in demographic and life-history traits from the
baseline simulations.
237
238   3. RESULTS
239
3.1. Baseline sturgeon population dynamics
Under the baseline conditions, the sturgeon population remained stable over 57 years (Fig. 2
and S3). The population maintained, on average, ~32% of adults being female, ~15.8 metric ton
spawner biomass, and ~7100 recruits (Fig. S3, Table S1 and S2). Adult females allocated, on
average, ~9.4 % of total mass to gonad while they were in the reproductive cycle; ~22% of
reproductive females eventually spawned, producing, on average, 23 million eggs (or ~33000
eggs spawner^{-1}) annually (Table S1 and S2). Most spawning and subsequent larval settlement
occurred below the confluence of the Elkhorn River and the Platte River (112 rkm) in late April
to early May (Table S1 and S2). Although the recruit numbers highly varied among years, a non-
linear (hump-shaped) density-dependent relationship emerged between spawners and recruits
(Fig. S3); on average, 0.11% of first-year fish survived (Table S1 and S2).
251
3.2. Recovery from reduced vitellogenin production efficiency
Simulated reduction in vitellogenin production rate gradually lowered reproductive capacities
and, in turn, biomass production of the sturgeon population exposed to synthetic androgens (Fig.
2 and 3). The proportion of energy allocated to gonad mass (10%) or storage mass (24%) in
reproductive females under exposure did not differ from those under baseline (Table S1).
However, the number of adult females in the reproductive cycle declined by 7.5–34% during
exposure (Fig. 4a,b). Further, spawner biomass and recruit numbers declined especially when
exposed to moderate and high levels of the perturbation (6.8–25% and 23–38%, respectively,
Fig. 3b, c), which were linearly correlated with relative changes in reproductive female
abundance (Fig. 3e). Lowered reproductive capacities and year class resulted in 8.6–43%
reduction (relative to baseline) in population biomass during exposure (Fig. 2b and 3d).
Similar relationships remained during recovery; reproductive females, spawners, and recruits
of the population with moderate exposures returned to the baseline levels with population size
fully recovered (i.e., mean population size returns to the baseline level) within five years (Fig. 2b
and 3), whereas the reproductive females, spawners, and recruits of the population with high
exposure continued to decline, ultimately reducing mean population biomass by up to 83%
before showing a sign of recovery after 15 years (~year 42, Fig. 2b).
269
3.3. Recovery from skewed recruit sex ratio
Sturgeon spawner biomass abruptly declined by up to 67% by ~year 22 under all scenarios of
simulated skewed sex ratios during the first year of life (Fig. 2c). Skewed sex ratios in recruits
were reflected in adults within ~6 years with relative female numbers declining, on average, by
6.3–25% during exposure (Fig. 3g). Subsequently, the number of adult females entering the
reproductive cycle and spawner biomass both started declining by ~year 8. During exposure,

mean reproductive female abundance declined by 19–21%; mean spawner biomass declined by
up to 29% under high exposure (Fig. 3h and 3i). In contrast, mean recruit number declined under
low and moderate exposure (by 15% and 23%, respectively), but differed little from the baseline
population under high exposure (Fig. 3j).
During recovery, mean sex ratios under all the exposure scenarios remained slightly (by 9.3–
16%) skewed toward males with high interannual variation (Fig. 3g). However, only the low-
exposure population fully recovered, whereas the moderate and high exposure populations
continued to show declining reproductive female abundance, spawner biomass, and recruit
number (Fig. 3). Despite large year class in early recovery years (year 31–33), more than 85% of
adults were males under high exposure and did not contribute to spawner biomass (Fig. 3). The
moderate and high exposure spawner biomass ultimately reached a stable state with less than
20% of the baseline population (Fig. 2c).
288
3.4. Recovery from combined effects of reduced vitellogenin production and skewed sex ratio
Concurrently simulating reduction in vitellogenin production and skewed sex ratios modulated
the sturgeon population dynamics additively under low exposure, but antagonistically under
moderate and high exposure (Fig. 2d). Spawner biomass declined within 5 years and remained
relatively stable and then declined further during the last five years (Fig. 2d). During exposure,
adult sex ratio differed little under low exposure, but skewness toward males increased by 11%
under medium and high exposure (Fig. 4a). Further, reproductive female abundance, spawner
biomass, and recruit number declined by 31%, 23%, and 31% (respectively) under low and
moderate exposure, and declined by 46%, 29%, and 47% (respectively) under high exposure
(Fig. 4b-d).
During recovery, adult sex ratios remained skewed toward males (by ~11%) under all the
exposure scenarios (Fig. 4a). Mean reproductive female number, spawner biomass, and recruit
number (except for reproductive females with low exposure) recovered after low and moderate
exposures but remained low under high exposure (Fig. 4b-d). Spawner biomass under low and
moderate exposure recovered within ~25 years of recovery, whereas the biomass of the
population under high exposure reached a stable state in ~22 years at ~44% lower than the
baseline population (Fig. 2d).
306
307   4. DISCUSSION
308
Simulations revealed that energetic (gonad maturation) and demographic (sex allocation)
perturbations in life history traits of long-lived, migratory freshwater vertebrates such as
Scaphirhynchus sturgeon can emerge as vastly different response and recovery patterns in
population projection in a spatially and temporally dynamic environment. We find that lowered
vitellogenin production rate prolongs maturation and, in turn, reduces spawner biomass,
ultimately weakening year class strength. Further, the energetic perturbation operates as
exposure-dependent processes–the higher the exposure level, the more severe the impacts,
thereby resulting in longer recovery time. In contrast, skewed sex allocation emerges as delayed
(by more than two decades) population-level responses, followed by an abrupt shift in year class
strength and population trajectory under higher exposures, ultimately reaching an alternative
(stable) state without any sign of recovery. When energetics and demographics are concurrently
perturbed, however, they act antagonistically as the exposure level increases, allowing it to
gradually recover without reaching a stable state. These contrasting emergent patterns at the

population level highlight the complexity of cumulative effects of stochastic perturbations in life
history traits by long-term, low-level exposure to stressors, posing challenges to conservation
efforts to restore the depleted vertebrate population.
Delayed gonad maturation in energetically perturbed females prolongs the spawning cycle of
our model sturgeon population instead of reducing fecundity. Sex-biased differences in energy
allocation to gonadal maturation often determine life-history strategies such as foraging and mate
selection, shaping sexual dimorphism in animal populations (Huse 1998; Rennie *et al.* 2008). For
a capital spawner such as sturgeon, slowed female gonad maturation delays spawning timing.
Although sturgeon are iteroparous, because of infrequent spawning and long maturation time,
delayed spawning events can reduce lifetime reproductive output (our simulations show that
some individuals spawned only once). Inefficient energy use triggered by synthetic hormones in
animals with long maturation time may thus threaten population persistence. Our simulations,
however, also suggest that the energetic perturbation triggers plastic responses. Although
reproductive and spawning females decline by up to 25%, the population eventually recovers,
indicating the reproductive capacity diminished during androgen exposure can be restored if
energy allocation to gonad resumes.
Delayed spawning events may also increase the probability of year class failures owing to
environmental stochasticity. We find that exposed (energetically challenged) female sturgeon are
more likely to abort the reproductive cycle as they divert energy to meet metabolic needs when
poor environmental conditions persist. Although energetic tradeoffs between migration and
fecundity are beyond the scope of our study, animals with long spawning migration such as
sturgeon likely skip spawning or produce offspring with lower survival rate owing to reduced
energy reserves (Glebe & Leggett 1981). Sufficient energy reserves in eggs are vital to larval
development and survival; maternal investment in egg production determines the nutritional
status of post-hatching larvae (Kjesbu *et al.* 1996). Our model assumes that individual eggs do
not vary in energy content; however, energetic tradeoffs in maternal investment in egg
production and offspring survival can arise and vary greatly among taxa (Jönsson 1997; McBride
et al. 2015) and thus need further investigation for *Scaphirhynchus* sturgeon.
Our study suggests that the recovery from demographic perturbation can pose greater
conservation challenges than energetic perturbation. We find that a long-lived animal population
such as sturgeon experiencing skewed sex allocation can remain in a small population size for
long time. Although skewed sex ratio has been widely reported for a variety of vertebrate taxa
(Donald 2007), causes of the skewness vary; it can result from phenotypic sex-biased differences
in physiology, dispersal, and (natural or human-induced) mortality (Weimerskirch, Lallemand &
Martin 2005; Pen *et al.* 2010). In our baseline model population, adult sex ratios are skewed
toward males (~1:2.2), primarily resulting from 1) a longer maturation time for females than
males and 2) more energy investment for reproduction in females, resulting in higher probability
of starvation in poor years. Simulations, however, show that further deviation (by 13–23%) in
adult sex ratio skewness result in long-lasting threats to year class success and population
viability. Consequences of skewed adult sex ratio may not immediately emerge for long-lived
species (Holmes & York 2003; Lawson *et al.* 2010). Once the skewness reaches a tipping point
(Scheffer *et al.* 2009), however, it often becomes difficult for a population to recover (Scheffer *et*
al. 2001) as it may stabilize in an alternative state facilitated by a positive feedback (Van Nes &
Scheffer 2007; Dai *et al.* 2012). In conservation planning for long-lived migratory species such
as sturgeon, thus, to prevent the shifts and minimize recovery cost, it is critical to identify early

warnings that signal impending demographic shifts; skewness in adult sex ratio can be a sensitive metric for assessing viability of stressed populations (Ancona *et al.* 2017).

Long-term, low-level exposure to stressors such as synthetic hormones often obscures causative relationships (Windsor, Ormerod & Tyler 2018). Establishing ecologically relevant indicators such as deviation in adult sex ratio and vitellogenin production rate would guide the risk assessment of synthetic hormones and habitat conservation planning for depleted populations. Further, given potential time lags across biological hierarchy, continued biological monitoring of individual-level traits would help detect early signals that may indicate population-level changes. Although our study was designed to evaluate consequences of low-level exposure, our findings indicate larger implications of among-individual variation in sex-biased perturbations for population persistence and recovery of threatened and endangered species. Population models widely used in risk assessments often assume no individual-level variation in each size or age group. These models may not only gloss over individual variation in fertility and age at maturity that may contribute to population viability, but also under- or over- estimate effects of synthetic hormones and other stressors (Jobling & Tyler 2003).

Large-river migratory populations such as *Scaphirhynchus* sturgeon likely experience multiple synthetic hormones over agriculturally rich landscapes in regions such as the Midwestern United States (Welcomme 1995). Further confounding factors in conservation and management efforts can arise in hydrologically altered systems such as the Platte River. Precipitation can affect the amount of agrichemicals released into rivers. Assessing the likelihood of ecological risks of agrichemicals for long-lived species requires the development of trait-based predictive models using anticipated scenarios. With expected trends in climate variability, which can have considerable effects on river hydrology and water chemistry, these forecasting tools are critical for fisheries and water resource management.

AUTHORS' CONTRIBUTIONS
DG, VEF, and MAP conceived the ideas; MJH, JJH and MLR collected and managed the data; DG led the writing of the manuscript. All authors contributed to revisions and gave final approval for publication.

ACKNOWLEDGEMENTS
We greatly appreciate comments by Maxime Vaugeois (University of Minnesota) and anonymous reviewers on the earlier version of this manuscript. This project was partially funded by University of Nebraska–Lincoln and the Nebraska Game and Parks Commission (Project F-180-R).

REFERENCES
Ancona, S., Dénes, F.V., Krüger, O., Székely, T. & Beissinger, S.R. (2017) Estimating adult sex ratios in nature. *Philosophical Transactions of the Royal Society B: Biological Sciences,* **372,** 20160313.
Ankley, G.T., Jensen, K.M., Makynen, E.A., Kahl, M.D., Korte, J.J., Hornung, M.W., Henry, T.R., Denny, J.S., Leino, R.L. & Wilson, V.S. (2003) Effects of the androgenic growth promoter 17-β-trenbolone on fecundity and reproductive endocrinology of the fathead minnow. *Environmental Toxicology and Chemistry,* **22,** 1350-1360.

Ankley, G.T., Miller, D.H., Jensen, K.M., Villeneuve, D.L. & Martinovic, D. (2008)
Relationship of plasma sex steroid concentrations in female fathead minnows to
reproductive success and population status. *Aquatic Toxicology,* **88,** 69-74.
Arcand Hoy, L.D. & Benson, W.H. (1998) Fish reproduction: an ecologically relevant indicator
of endocrine disruption. *Environmental Toxicology and Chemistry,* **17,** 49-57.
Arukwe, A. & Goksøyr, A. (1998) Xenobiotics, xenoestrogens and reproduction disturbances in
fish. *Sarsia,* **83,** 225-241.
Beddington, J. (1975) Mutual interference between parasites or predators and its effect on
searching efficiency. *Journal of Animal Ecology*, 331-340.
Bernhardt, E.S., Rosi, E.J. & Gessner, M.O. (2017) Synthetic chemicals as agents of global
change. *Frontiers in Ecology and the Environment,* **15,** 84-90.
Braaten, P.J., Fuller, D.B., Holte, L.D., Lott, R.D., Viste, W., Brandt, T.F. & Legare, R.G. (2008)
Drift dynamics of larval pallid sturgeon and shovelnose sturgeon in a natural side channel
of the upper Missouri River, Montana. *North American Journal of Fisheries*
Management, **28,** 808-826.
Bramblett, R.G. & White, R.G. (2001) Habitat use and movements of pallid and shovelnose
sturgeon in the Yellowstone and Missouri rivers in Montana and North Dakota.
Transactions of the American Fisheries Society, **130,** 1006-1025.
Coltman, D.W., O'Donoghue, P., Jorgenson, J.T., Hogg, J.T., Strobeck, C. & Festa-Bianchet, M.
(2003) Undesirable evolutionary consequences of trophy hunting. *Nature,* **426,** 655-658.
Crain, C.M., Kroeker, K. & Halpern, B.S. (2008) Interactive and cumulative effects of multiple
human stressors in marine systems. *Ecology Letters,* **11,** 1304-1315.
Dai, L., Vorselen, D., Korolev, K.S. & Gore, J. (2012) Generic indicators for loss of resilience
before a tipping point leading to population collapse. *science,* **336,** 1175-1177.
DeAngelis, D., Goldstein, R. & O'neill, R. (1975) A model for tropic interaction. *Ecology*, 881-
892.
del Carmen Alvarez, M. & Fuiman, L.A. (2005) Environmental levels of atrazine and its
degradation products impair survival skills and growth of red drum larvae. *Aquatic*
Toxicology, **74,** 229-241.
Delonay, A.J., Chojnacki, K.A., Jacobson, R.B., Albers, J.L., Braaten, P.J., Bulliner, E.A.,
Elliott, C.M., Erwin, S.O., Fuller, D.B., Haas, J.D., Ladd, H.L.A., Mestl, G.E., Papoulias,
D.M. & Wildhaber, M.L. (2016) Ecological requirements for pallid sturgeon
reproduction and recruitment in the Missouri River—A synthesis of science, 2005 to
2012. *Scientific Investigations Report*, pp. 242. Reston, VA.
DeLonay, A.J., Jacobson, R.B., Papoulias, D.M., Simpkins, D.G., Wildhaber, M.L., Reuter, J.M.,
Bonnot, T.W., Chojnacki, K.A., Korschgen, C.E. & Mestl, G.E. (2009a) Ecological
requirements for pallid sturgeon reproduction and recruitment in the Lower Missouri
River: a research synthesis 2005-08.
DeLonay, A.J., Jacobson, R.B., Papoulias, D.M., Simpkins, D.G., Wildhaber, M.L., Reuter, J.M.,
Bonnot, T.W., Chojnacki, K.A., Korschgen, C.E., Mestl, G.E. & Mac, M.J. (2009b)
Ecological requirements for pallid sturgeon reproduction and recruitment in the Lower
Missouri River: A research synthesis 2005–08. U.S. Geological Survey Scientific
Investigations Report 2009–5201.
Donald, P.F. (2007) Adult sex ratios in wild bird populations. *Ibis,* **149,** 671-692.

Fairbairn, D.J. (1997) Allometry for sexual size dimorphism: pattern and process in the
coevolution of body size in males and females. *Annual review of Ecology and*
Systematics, **28,** 659-687.
Glebe, B. & Leggett, W. (1981) Temporal, intra-population differences in energy allocation and
use by American shad (*Alosa sapidissima*) during the spawning migration. *Canadian*
Journal of Fisheries and Aquatic Sciences, **38,** 795-805.
Goto, D., Hamel, M.J., Hammen, J.J., Rugg, M.L., Pegg, M.A. & Forbes, V.E. (2015)
Spatiotemporal variation in flow-dependent recruitment of long-lived riverine fish:
Model development and evaluation. *Ecological modelling,* **296,** 79-92.
Goto, D., Hamel, M.J., Pegg, M.A., Hammen, J.J., Rugg, M.L. & Forbes, V.E. (2018) Spatially
dynamic maternal control of migratory fish recruitment pulses triggered by shifting
seasonal cues. *Marine and Freshwater Research,* **69,** 942-961.
Gurney, W.S.C., Jones, W., Veitch, A.R. & Nisbet, R.M. (2003) Resource allocation,
hyperphagia, and compensatory growth in juveniles. *Ecology,* **84,** 2777-2787.
Hanson, P.C., Johnson, T.B., Schindler, D.E. & Kitchell, J.F. (1997) Fish bioenergetics 3.0.
University of Wisconsin Sea Grant Institute, Madison, WI.
Hendry, A.P., Farrugia, T.J. & Kinnison, M.T. (2008) Human influences on rates of phenotypic
change in wild animal populations. *Molecular ecology,* **17,** 20-29.
Holmes, E. & York, A. (2003) Using age structure to detect impacts on threatened populations: a
case study with Steller sea lions. *Conservation Biology,* **17,** 1794-1806.
Hope, B.K. (2005) Performing spatially and temporally explicit ecological exposure assessments
involving multiple stressors. *Human and Ecological Risk Assessment,* **11,** 539-565.
Houde, E.D. (1989) Comparative growth, mortality, and energetics of marine fish larvae:
temperature and implied latitudinal effects. *Fishery Bulletin,* **87,** 471-495.
Hrabik, R., Herzog, D., Ostendorf, D. & Petersen, M. (2007) Larvae provide first evidence of
successful reproduction by pallid sturgeon, Scaphirhynchus albus, in the Mississippi
River. *Journal of Applied Ichthyology,* **23,** 436-443.
Huntzinger, T.L. & Ellis, M.J. (1993) Central Nebraska river basins, Nebraska. *JAWRA Journal*
of the American Water Resources Association, **29,** 533-574.
Huse, G. (1998) Sex-specific life history strategies in capelin (*Mallotus villosus*)? *Canadian*
Journal of Fisheries and Aquatic Sciences, **55,** 631-638.
Jobling, S. & Tyler, C.R. (2003) Topic 4.3: Endocrine disruption in wild freshwater fish. *Pure*
and applied chemistry, **75,** 2219-2234.
Jönsson, K.I. (1997) Capital and income breeding as alternative tactics of resource use in
reproduction. *Oikos***,** 57-66.
Kappenman, K.M., Webb, M.A.H. & Greenwood, M. (2013) The effect of temperature on
embryo survival and development in pallid sturgeon *Scaphirhynchus albus* (Forbes &
Richardson 1905) and shovelnose sturgeon *S. platorynchus* (Rafinesque, 1820). *Journal*
of Applied Ichthyology, **29,** 1193-1203.
Kjesbu, O.S., Solemdal, P., Bratland, P. & Fonn, M. (1996) Variation in annual egg production
in individual captive Atlantic cod (*Gadus morhua*). *Canadian Journal of Fisheries and*
Aquatic Sciences, **53,** 610-620.
Knell, R.J. & Martínez-Ruiz, C. (2017) Selective harvest focused on sexual signal traits can lead
to extinction under directional environmental change. *Proceedings of the Royal Society*
B: Biological Sciences, **284,** 20171788.

Lawson, D.M., Regan, H.M., Zedler, P.H. & Franklin, J. (2010) Cumulative effects of land use, altered fire regime and climate change on persistence of Ceanothus verrucosus, a rare, fire-dependent plant species. *Global Change Biology,* **16,** 2518-2529.

Marr, A., Arcese, P., Hochachka, W., Reid, J.M. & Keller, L. (2006) Interactive effects of environmental stress and inbreeding on reproductive traits in a wild bird population. *Journal of Animal Ecology,* **75,** 1406-1415.

Maxwell, S.M., Hazen, E.L., Bograd, S.J., Halpern, B.S., Breed, G.A., Nickel, B., Teutschel, N.M., Crowder, L.B., Benson, S. & Dutton, P.H. (2013) Cumulative human impacts on marine predators. *Nature communications,* **4,** 2688.

McBride, R.S., Somarakis, S., Fitzhugh, G.R., Albert, A., Yaragina, N.A., Wuenschel, M.J., Alonso-Fernández, A. & Basilone, G. (2015) Energy acquisition and allocation to egg production in relation to fish reproductive strategies. *Fish and Fisheries,* **16,** 23-57.

McTavish, K., Stech, H. & Stay, F. (1998) A modeling framework for exploring the population level effects of endocrine disruptors. *Environmental Toxicology and Chemistry,* **17,** 58-67.

Mills, L.J. & Chichester, C. (2005) Review of evidence: Are endocrine-disrupting chemicals in the aquatic environment impacting fish populations? *Science of the total environment,* **343,** 1-34.

Morthorst, J.E., Holbech, H. & Bjerregaard, P. (2010) Trenbolone causes irreversible masculinization of zebrafish at environmentally relevant concentrations. *Aquatic Toxicology,* **98,** 336-343.

Nash, J.P., Kime, D.E., Van der Ven, L.T.M., Wester, P.W., Brion, F., Maack, G., Stahlschmidt-Allner, P. & Tyler, C.R. (2004) Long-term exposure to environmental concentrations of the pharmaceutical ethynylestradiol causes reproductive failure in fish. *Environmental Health Perspectives,* **112,** 1725.

NDEQ (2004) 2004 surface water quality integrated report. Water Quality Division. pp. 118 pp.

O'Brien, D., Lewis, S., Davis, A., Gallen, C., Smith, R., Turner, R., Warne, M., Turner, S., Caswell, S. & Mueller, J.F. (2016) Spatial and temporal variability in pesticide exposure downstream of a heavily irrigated cropping area: application of different monitoring techniques. *Journal of Agricultural and Food Chemistry,* **64,** 3975-3989.

Pen, I., Uller, T., Feldmeyer, B., Harts, A., While, G.M. & Wapstra, E. (2010) Climate-driven population divergence in sex-determining systems. *Nature,* **468,** 436.

Peters, E.J. & Parham, J.E. (2008) Ecology and management of sturgeon in the lower Platte River, Nebraska. *Nebraska Technical Series*.

Peterson, C.H., Rice, S.D., Short, J.W., Esler, D., Bodkin, J.L., Ballachey, B.E. & Irons, D.B. (2003) Long-term ecosystem response to the Exxon Valdez oil spill. *Science,* **302,** 2082.

Rashleigh, B. & Grossman, G.D. (2005) An individual-based simulation model for mottled sculpin (*Cottus bairdi*) in a southern Appalachian stream. *Ecological Modelling,* **187,** 247-258.

Rennie, M.D., Purchase, C.F., Lester, N., Collins, N.C., Shuter, B.J. & Abrams, P.A. (2008) Lazy males? Bioenergetic differences in energy acquisition and metabolism help to explain sexual size dimorphism in percids. *Journal of Animal Ecology,* **77,** 916-926.

Ricker, W.E. (1975) *Computation and interpretation of biological statistics for fish populations*. Fisheries Research Board of Canada, Ottawa, Ontario.

Scheffer, M., Bascompte, J., Brock, W.A., Brovkin, V., Carpenter, S.R., Dakos, V., Held, H.,
Van Nes, E.H., Rietkerk, M. & Sugihara, G.J.N. (2009) Early-warning signals for critical
transitions. **461,** 53.
Scheffer, M., Baveco, J., DeAngelis, D., Rose, K. & Van Nes, E. (1995) Super-individuals a
simple solution for modelling large populations on an individual basis. *Ecological*
modelling, **80,** 161-170.
Scheffer, M., Carpenter, S., Foley, J.A., Folke, C. & Walker, B. (2001) Catastrophic shifts in
ecosystems. *Nature,* **413,** 591.
Tilman, D. (1999) Global environmental impacts of agricultural expansion: the need for
sustainable and efficient practices. *Proceedings of the National Academy of Sciences,* **96,**
5995-6000.
Tilman, D., Fargione, J., Wolff, B., D'antonio, C., Dobson, A., Howarth, R., Schindler, D.,
Schlesinger, W.H., Simberloff, D. & Swackhamer, D. (2001) Forecasting agriculturally
driven global environmental change. *science,* **292,** 281-284.
Van Nes, E.H. & Scheffer, M. (2007) Slow recovery from perturbations as a generic indicator of
a nearby catastrophic shift. *The American Naturalist,* **169,** 738-747.
Wedekind, C. (2017) Demographic and genetic consequences of disturbed sex determination.
Philosophical Transactions of the Royal Society B: Biological Sciences, **372,** 20160326.
Weimerskirch, H., Lallemand, J. & Martin, J. (2005) Population sex ratio variation in a
monogamous long-lived bird, the wandering albatross. *Journal of Animal Ecology,* **74,**
285-291.
Welcomme, R.L. (1995) Relationships between fisheries and the integrity of river systems.
Regulated Rivers: Research & Management, **11,** 121-136.
Wildhaber, M., Papoulias, D., DeLonay, A., Tillitt, D., Bryan, J. & Annis, M. (2007a) Physical
and hormonal examination of Missouri River shovelnose sturgeon reproductive stage: a
reference guide. *Journal of Applied Ichthyology,* **23,** 382-401.
Wildhaber, M.L., Papoulias, D.M., DeLonay, A.J., Tillitt, D.E., Bryan, J.L. & Annis, M.L.
(2007b) Physical and hormonal examination of Missouri River shovelnose sturgeon
reproductive stage: a reference guide. *Journal of Applied Ichthyology,* **23,** 382-401.
Windsor, F.M., Ormerod, S.J. & Tyler, C.R. (2018) Endocrine disruption in aquatic systems: up-
scaling research to address ecological consequences. *Biological Reviews,* **93,** 626-641.
575

Figures

(a)

(b)

Fig. 1. Study system and model structure. a) a map of the study system, Lower Platte River (Nebraska–NE, USA) and b) a schematic of the spatially explicit individual-based biophysical model for shovelnose sturgeon.

Fig. 2. Shovelnose sturgeon population projections during exposure and recovery under synthetic androgen scenarios – a) a baseline scenario, b) an energetic perturbation (reduced maternal investment in gonad development) scenario with three exposure levels (shown as biomass relative to the baseline scenario); c) a demographic perturbation (pre-recruit sex ratio skewness) scenario with three exposure levels; and d) a combined scenario of energetic and demographic perturbations with three exposure levels. Red solid, green dashed, and blue dotted lines indicate low, medium and high androgen exposure.

Fig. 3. Changes (%, compared to the baseline scenario) in life history traits of shovelnose sturgeon populations– under energetic perturbation scenarios: a) proportion of mature females in the reproductive cycle, b) reproductive female number, c) spawner biomass, d) recruit number, e) relationship between reproductive female number and spawner biomass or recruit number during exposure, f) relationship between reproductive female number and spawner biomass or recruit number during recovery; and demographic perturbation scenarios: g) adult sex ratio, h) reproductive female number, i) spawner biomass, j) recruit number, k) relationship between adult sex ratio and reproductive female number, spawner biomass, or recruit number during exposure, and l) relationship between adult sex ratio and reproductive female number, spawner biomass, or recruit number during recovery. In, a–d and g–j, red solid lines in box plots indicate mean values and black solid lines show 25th, 50th, and 75th percentile (bottom to top). In e, f, k, and l, data points show mean values from low, moderate, and high androgen exposure levels.

Fig. 4. Changes (%, compared to the baseline scenario) in life history traits of shovelnose sturgeon populations under combined scenario of energetic and demographic perturbations – a) proportion of mature females in the reproductive cycle, b) adult sex ratio, c) reproductive female number, d) spawner biomass, e) recruit number, f) relationship between reproductive female number and spawner biomass or recruit number during exposure, g) relationship between reproductive female number and spawner biomass or recruit number during recovery, h) relationship between adult sex ratio and reproductive female number, spawner biomass, or recruit number during exposure, and i) relationship between adult sex ratio and reproductive female number, spawner biomass, or recruit number during recovery. In, a–e, red solid lines in box plots show mean values and black solid lines indicate 25th, 50th, and 75th percentile (bottom to top). In f–i, data points show mean values from low, moderate, and high androgen exposure levels.