Trophic cascade driven by behavioural fine-tuning as naïve prey rapidly adjust to a novel predator

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The arrival of novel predators can trigger trophic cascades driven by shifts in prey numbers. Predators also elicit behavioural change in prey populations, via plasticity and rapid evolution, and such behavioural responses by prey may also contribute to trophic cascades. Here we document the effects of a novel predator on the behaviour and demography of a native prey species. We reveal rapid behavioural responses in the prey species (grassland melomys, a granivorous rodent) following the introduction of a novel marsupial predator (northern quoll). Within months of quolls appearing, populations of melomys exhibited reduced survival and population declines relative to control populations. Quoll-exposed populations were also significantly shyer than nearby, predator-free populations of conspecifics. This rapid but generalised response to a novel threat was replaced over the following two years with more threat-specific antipredator behaviours (i.e. predator-scent aversion). Predator-exposed populations, however, remained more neophobic than predator-free populations throughout the study. These behavioural responses manifested rapidly in changed rates of seed predation by melomys across treatments. Quoll-exposed melomys populations exhibited lower per-capita seed take rates, and rapidly developed an avoidance of seeds associated with quoll scent, with discrimination playing out over a spatial scale of tens of metres. Presumably the significant and novel predation pressure induced by the introduction of quolls drove melomys to fine-tune their behavioural response through time as the nature of the threat became clearer. These behavioural shifts could reflect individual plasticity in behaviour or may be adaptive shifts from natural selection imposed by quoll predation. Our study provides a rare insight into the rapid ecological and behavioural shifts enacted by prey to mitigate the impacts of a novel predator and shows that trophic cascades can be strongly influenced by behavioural as well as numerical responses.

INTRODUCTION

Predation is one of the most pervasive and powerful forces acting on populations. Not only does predation directly impact a population's demography (Schoener & Spiller 1996), it also imposes natural selection (Abrams 2000). Although pervasive, predation is not a constant. The pressure that predators impose on populations will vary through time and space (Lima & Dill 1990; Sih 1992) as a consequence of tightly coupled predator-prey dynamics; prey switching; or stochastic processes. The fact that predation is not constant, and that antipredator defences may be costly, suggests that flexible, plastic responses to predation pressure will often be favoured (Sih *et al.* 2000; Berger *et al.* 2001). There is, in fact, a great deal of empirical evidence that plastic responses to predation are common (e.g. Relyea 2003; Brown *et al.* 2013; Cunningham *et al.* 2019): investment in antipredator traits across morphology, life-history, and behaviour is often flexible, and dependent on the perceived risk of predation.

As well as impacting prey populations, it is increasingly apparent that predators play a powerful role in structuring communities (Estes *et al.* 2011). Some of our best evidence for this comes from the introduction of predators to naïve communities. Invasive predators can cause extinctions (Medina *et al.* 2011; Woinarski *et al.* 2015; Doherty *et al.* 2016), and alter trophic structures and ecosystem function within recipient communities (Courchamp *et al.* 2003; Simberloff *et al.* 2013). We typically think of these cascading outcomes as purely numeric effects: predators depress the size of prey populations, and the altered numbers of prey can cause cascading numerical changes down trophic levels (e.g. Ripple *et al.* 2001). These numerical effects are undeniably important, but the fact that predators can also elicit phenotypic change in prey populations – through plasticity and natural selection – means that subtler ecological effects may also manifest. Prey species living in a landscape of fear

may forage at different times, or in different places compared with the same species in a predator-free environment (Laundre *et al.* 2010). Such behavioural shifts will alter all downstream species interactions in potentially complex ways (e.g. Fortin *et al.* 2005; Suraci *et al.* 2016).

Because predator invasions are rarely intentional or anticipated, there is a scarcity of controlled empirical work on the effects of novel predators on recipient communities and the mechanisms via which these effects play out (but see Lapiedra *et al.* 2018; Pringle *et al.* 2019). Such tests are needed, however, if we are to predict invasive species impacts, advance conservation management (Sih *et al.* 2010), and improve our understanding of how communities are structured (Sax *et al.* 2007). Here, we exploit a conservation-motivated predator introduction, and show that (as well as numerical effects) rapid shifts in prey behaviour occurred, and these behavioural shifts caused measurable change in downstream trophic interactions.

Northern quolls (*Dasyurus hallucatus*) were, until recently, a common predator across northern Australia. They have declined over the last several decades, following the general decline in northern Australian mammals (Woinarski *et al.* 2015), thought to be driven by changes in grazing, fire, and predation regimes (Braithwaite & Griffiths 1994). More recently, the invasion of toxic invasive prey (cane toads, *Rhinella marina*) has resulted in dramatic, range-wide population declines in northern quolls (Shine 2010; Oakwood *et al.* 2016). Due to local extinction, northern quolls are now absent from large tracts of their former range and their ecological function as a medium-sized mammalian predator has been lost (Moore *et al.* 2019). For their conservation, northern quolls have recently been introduced to a number of offshore islands where they have never previously existed. We exploit the introduction of northern quolls to a new island to directly test the effects of

quolls as a novel predator on an island ecosystem and observe how native prey adjust to mitigate the impacts of their arrival.

In 2017, a population of 55 “toad-smart” northern quolls were introduced to a 25km^2 island off the coast of north-western Northern Territory, Australia (Kelly 2019). Prior to this introduction, Indian Island (Kabal) lacked mammalian predators, and large native reptilian predators had recently been reduced to near extinction by the invasion of cane toads. The introduction of northern quolls provided a unique opportunity to test the effects of the arrival of a novel predator in a system that has been isolated from major mammalian predators for thousands of years.

METHODS

Introduction of northern quolls

In May 2017, 55 adult northern quolls were introduced to the north-eastern tip of Indian Island, Bynoe Harbor, Northern Territory, Australia (12°37’24.60”S, 130°30’0.72”E) to field test the conservation strategy of targeted gene flow (Kelly & Phillips 2016). Quolls are a voracious, opportunistic generalist predator (< 1.5kg; Oakwood 1997), and their introduction presented a unique opportunity to monitor the behavioural and demographic impacts on grassland melomys (*Melomys burtoni*), a native mammalian granivorous prey species (mean body mass 56 g, 5.6–103.7g). Immediately prior to the introduction of quolls, we started monitoring populations of melomys in one woodland and two monsoon vine thicket plots in the vicinity of where quolls were to be released and radio tracked. After quolls were introduced and tracked it became immediately apparent that quolls were largely avoiding monsoon vine thicket sites and these sites were dropped from the on-going monitoring. For this reason, we only present data from before the introduction of quolls

from one invaded site. Most of our data compare quoll-invaded versus quoll-free sites over time, commencing within a few months of quoll arrival.

Melomys population monitoring

To determine whether the arrival of a novel predator resulted in demographic impacts (population size and survival) to native prey species, we monitored populations of melomys on Indian Island, during four trips occurring in May (site 1) and August 2017 (sites 2–7), April 2018 (sites 1–7), and May 2019 (sites 1–7). Melomys were caught from seven independent 1ha (100 m x 100 m) plots (sites 1–7) spread out across Indian Island using a standard mark-recapture trapping regime designed for a monitoring project (Begg *et al.* 1983; Kemper *et al.* 1987). Four "impact", quoll-invaded sites were established in the north of the island in the vicinity of where quolls were released and three "control", quoll-free sites were established in the south of the island. Sites in the north (quoll-invaded) and south (quoll-free) of the island were between 8.7 and 9.8km apart (Table 1) and were composed of similar habitat types. The northern and southern sections of Indian Island are divided by mangrove habitat which is inundated at high tide. We assumed quolls would be confined to the northern half of the island for at least a few years. Cage and camera trapping as well as track surveys confirmed that quolls were present at the "impact" sites and absent from the "control" sites for the duration of the study (Jolly *et al.* unpub. data).

Table 1. Pairwise distance matrix between sites on Indian Island, Northern Territory, Australia. Quolls were present at sites 1–4 and quolls were absent at sites 5–7 for the duration of the study.

Distance (m)	Site 1	Site 2	Site 3	Site 4	Site 5	Site 6	Site 7
Site 1							
Site 2	270						
Site 3	260	350					
Site 4	400	300	250				
Site 5	8760	9030	8710	9000			
Site 6	8470	8730	8450	8070	300		
Site 7	9670	9920	9590	9820	1260	1500	

Each of the seven monitoring sites consisted of 100 Elliott traps (Elliott Scientific Equipment, Upwey, Victoria) spaced at 10 m intervals in a 10 x 10 grid. Most trapping grids were open for four nights, however, the first trapping grid (site 1, May 2017) was open for six nights. After four trap nights, the majority of the melomys population had been captured at least once (Jolly *et al.* 2019). Traps were baited with peanut butter and rolled oats. These baits were replaced daily for the duration of each trapping session. Traps were checked for captures early each morning and all traps were cleared within two hours of sunrise.

Captured melomys were weighed (g) and sexed. Before release, each melomys was implanted with a microchip (Trovan Unique ID100). On successive mornings, all melomys were scanned (Trovan LID575 Handheld Reader), and any new individuals were microchipped. On the last morning of each trapping session, all melomys caught were retained for behavioural assays. Throughout the study 435 individual melomys were captured and given microchips. Of these, 191 (44%) were caught on the final night of trapping and were retained for behavioural trials. Only large, healthy juveniles (n = 12), adult males (n = 83), and adult non-visibly pregnant females (n = 96) were retained for behavioural experiments. Indian Island is remote and uninhabited by humans, so all

behavioural experiments were conducted in the field under near natural conditions (see Jolly *et al.* 2019 for detailed experimental procedures).

Modified open field tests

We employed modified open field tests (also referred to as emergence tests: see Brown & Braithwaite 2004; López *et al.* 2005; Carter *et al.* 2013; Jolly *et al.* 2019) to assess boldness in grassland melomys and whether the arrival of a novel predator resulted in behavioural shifts in invaded populations. All open field tests were conducted on the night after the last trap night (night 5) and in opaque-walled experimental arenas (540mm x 340mm x 370mm). Experimental arenas were modified plastic boxes that had an inverted Elliott trap sized hole cut in one end and were illuminated by strings of red LED lights (Jolly *et al.* 2019). Each experimental arena had natural sand as substrate, and a bait ball located both in the centre and along one wall of the arena (Jolly *et al.* 2019). After dark, Elliott traps containing a melomys were inserted into the hole in the side of each experimental arena and melomys were allowed to habituate for 10 min. At the start of each trial, Elliott trap doors were locked open – the inverted orientation of the trap prevented them from being triggered closed. Melomys were given 10 min to explore the open field arena. After 10 min, individuals were rounded back into their retreat (the Elliott trap) and a novel object (plastic bowl) was placed at the end of the arena opposite the Elliott trap (Jolly *et al.* 2019). Melomys were then given a further 10 min to explore the arena and interact with the novel object. Elliott traps remained open during the open field tests and melomys could shelter and emerge from them under their own volition. All trials were recorded using a GoPro HERO 3. A previous study in this system determined that melomys showed repeatable behaviour between trials (Jolly *et al.* 2019), therefore the data presented in this study were

188 from a single behavioural trial of each animal (*n* = 191). Once trials were complete, each
189 melomys was released at its point of capture.
190 To measure the boldness of individual melomys, we scored three behaviours
191 typically associated with boldness and neophobia in rodents (Dielenberg & McGregor 2001;
192 McGregor *et al.* 2002; Réale *et al.* 2007; Cremona *et al.* 2015): whether melomys fully
193 emerged from their Elliott trap hide and entered the open arena during the 0–10 min period
194 (scored 0 or 1, respectively); whether they fully emerged and entered the trial arena during
195 the 10–20 min period (scored 0 or 1); and whether they interacted (made contact) with the
196 novel object that was placed in the arena during the 10–20 min period (scored 0 or 1).
197 Videos were scored by a single observer who was blind to each melomys' origin and
198 identity. Because interacting with the novel object was predicated on a melomys'
199 willingness to emerge from their hide during the 10–20 min period, for analysis we
200 combined their emergence during this period and interaction with the novel object into a
201 single binary score: 0 (neophobic) = did not emerge or emerged but did not interact with
202 novel object; or 1 (not neophobic): emerged and interacted with novel object.
203
204 *Seed removal plots*
205 To assess whether the arrival of a novel predator affected the seed harvesting behaviour of
206 granivorous melomys, we established seed removal plots at each site and sampled them
207 each trapping session (night 6). After trapping and open field tests were conducted and
208 melomys had been returned to their capture location, we set up 81 seed plots at each site
209 by scraping away leaf litter with a shovel to create bare earth plots. These bare earth plots
210 were created so that they were located in the centre between four Elliott traps within the
211 10x10 trapping grid. Just before dark on the night of the seed removal experiment, we

placed a single wheat seed in the centre of each bare earth plot. These seeds were either unscented, control seeds (*n* = 40) or predator-scented seeds that had been maintained in a sealed clip-lock bag filled with freshly collected northern quoll fur (*n* = 41). The placement of predator-scented and unscented seeds was alternated so that there was a chequered arrangement of scented and unscented seeds across the site. To ensure that the predator-scent was strong enough to be detected by melomys, along with the predator-scented seeds, we also placed a few strands of quoll fur around the predator-scented seeds. Before light the next morning, we returned back to each plot and counted the number of seeds of each scent-type that were removed from the plot. Melomys are the only nocturnal granivorous animal that occurs on Indian Island, and to avoid diurnal granivorous birds from removing seeds we conducted this experiment during the night only.

Wildfire on northern Indian Island

Immediately following our monitoring and experiments in August 2017, a wildlife broke out on northern Indian Island in the vicinity of the four quoll-invaded sites and burnt through all of the sites. Because of this, our experimental design is confounded by the fact that all of our quoll-invaded sites were burnt, and all of our quoll-free sites were unburnt. Fire is a regular disturbance in this landscape (Andersen *et al.* 2005), and previous work has shown little effect of fire on abundance, survival or recruitment of grassland melomys (Griffiths & Brook 2015; Liedloff *et al.* 2018). Nonetheless, this confound exists and we proceed with caution when interpreting the effects of quolls on population size and survival of melomys.

Statistical analysis

235 During trapping sessions we identified individual melomys that were captured at each site
236 by their unique microchips. Because melomys on Indian Island have very small home ranges
237 (tending to be caught in the same or adjacent traps throughout the trapping period: Jolly *et*
238 *al.* unpub. data) and since we never observed captures of melomys marked at other sites
239 (Jolly *et al.* unpub. data), we treated each site as independent with regard to demographics
240 and behaviour (Table 1).

241 To estimate between-session survival, we analysed the mark-recapture data to
242 estimate recapture and survival rates using Cormack-Jolly-Seber models in program MARK.
243 At each site, there were three primary trapping sessions of four nights, for a total of 12 time
244 intervals in the input file. Because quolls prey on melomys, we hypothesised that survival
245 rates of melomys would be lower between trapping sessions at sites with quolls than at
246 sites without quolls. We included two groups, quoll-free (control) and quoll-invaded
247 (impact), in the input file. We ran a series of models in MARK to test the following *a priori*
248 hypotheses: (1) survival rates between sessions are lower at quoll-free sites than at quoll-
249 invaded sites; (2) survival rates are lower between sessions than within sessions, but are
250 unaffected by quolls; (3) survival is constant through time; and (4) survival varies through
251 time. All candidate models were ranked according to their AICc values and associated AIC
252 weights (Burnham & Anderson 1998).

253 To test whether the presence of quolls impacted melomys population size, we used
254 a hierarchical model in which population size was made a function of quoll
255 presence/absence, capture session, and the interaction between these factors. Population
256 size at each site during each session is estimated in this process, and we fitted this model in
257 a Bayesian framework. Our observations consisted of a capture history for each observed
258 individual over the number of nights at each site for each trapping session. We denoted the

number of individuals at site *s* during session *k* as N_{ks}. To estimate N_{ks} we used a closed population mark-recapture analysis in which each individual, *I*, was either observed, or not, according to a Bernoulli distribution:

$$O_{iks} \sim \text{Bernoulli}(d_s)$$

Our previous MARK analysis found clear evidence for variation in detection probability across sessions, but detection probabilities of melomys on Indian Island had previously been found not to vary measurably between individuals nor to change over time within a trapping session (Jolly *et al.* 2019). Thus, we made detection probability a function of session according to:

$$\text{logit}(d_s) = \mu_d + t_s$$

Where d_s denotes the expected detection probability within session s, μ_d is the expected detection probability in the first session, and t_s denotes the (categorical) effect of session on detection.

We used the "data augmentation" method (Tanner & Wong 1987; Royle *et al.* 2007; Kery & Schaub 2011) in combination with this detection probability to estimate N_{ks} for each site.session. Using this approach, the data were 'padded' by adding an arbitrary number of zero-only encounter histories of 'potential' unobserved individuals. The augmented dataset was modelled as a zero-inflated model (Royle *et al.* 2007) which changes the problem from estimating a count, to estimating a proportion. This was executed by adding a latent binary indicator variable, R_{iks}, to classify each row in the augmented data matrix as a 'real' individual or not, where $R_{iks} \sim \text{Bernoulli}(\Omega_{ks})$. The parameter Ω_{ks} was estimated from the data, and $N_{ks} = \Sigma_i R_{iks}$.

We then made Ω_{ks} (which scales with population size) a function of quoll presence/absence, q_c; session, b_k; and the interaction between the two:

$$\text{logit}(\Omega_{ks}) = \mu_p + q_c + b_k + \gamma_{ck}$$

The model was fitted using Bayesian Markov Chain Monte Carlo (MCMC) methods and minimally informative priors (Table 2) within the package JAGS (Plummer *et al.* 2017) using R (R Core Team 2019). Parameter estimates were based on 30,000 iterations with a thinning interval of 5 following a 10,000 sample burn-in. Three MCMC chains were run, and model convergence assessed by eye, and using the Gelman-Rubin diagnostic (Gelman & Rubin 1992a, 1992b).

To assess whether the introduction of quolls affected the behaviour of melomys populations, we divided the responses of melomys in open field tests into two independent response variables: the proportion that emerged during the 0-10 min period; and the proportion that emerged and interacted with the novel object during the 10-20 min period. We used generalised linear mixed-effects models with binomial errors and a logit link to test the effect of quoll presence and trapping session, with site included as a random effect, on the behavioural response variables. *P*-values were obtained by likelihood ratio tests of the full model with the effect in question against the model without the effect. This analysis was performed using R with the *lme4* software package (R Core Team 2019).

To assess whether the numerical impact of quolls on melomys affected the seed harvesting rate of invaded melomys populations, we first examined the relationship between melomys population size (estimated above) and the total number of seeds harvested in the control (unscented) seeds. Here we used a simple linear model with number of seeds harvested as a function of population size, quoll presence/absence and the interaction between these effects. To test whether there was an additional effect of quoll presence, beyond their effect on population size, we defined a new variable, Δ, as the difference in seed take between scented and unscented treatments within each site.session.

307 Here any effect of melomys density is cancelled out (because density is common to both
308 treatments within each site.session). Thus, we fitted a model in which Δ is a function of
309 quoll presence/absence, session and the interaction between these effects. These analyses
310 were performed using R (R Core Team 2019).
311
312 **RESULTS**
313 *Effect of novel predator on survival*
314 When we assessed the impact of quolls on melomys survival between trapping sessions the
315 best supported model was one in which survival rates between sessions were lower at
316 quoll-invaded sites than at quoll-free sites, and recapture rates were session-dependent
317 (Table 2). All other models were more than 4 AIC units from this best model, and so clearly
318 inferior descriptions of the data. From the best-supported model, estimates of apparent
319 survival for the intervals between the capture sessions were substantially higher at quoll-
320 free sites ($S_{2017–2018}$ = 0.368; $S_{2018–2019}$ = 0.225) than at quoll-invaded sites ($S_{2017–2018}$ = 0.207;
321 $S_{2018–2019}$ = 0.091; Fig. 1). The differing survival probability between sessions is largely
322 explained by the time difference between intervals (2017–2018 = 9 months vs. 2018–2019
323 = 13 months; Fig. 2).
324

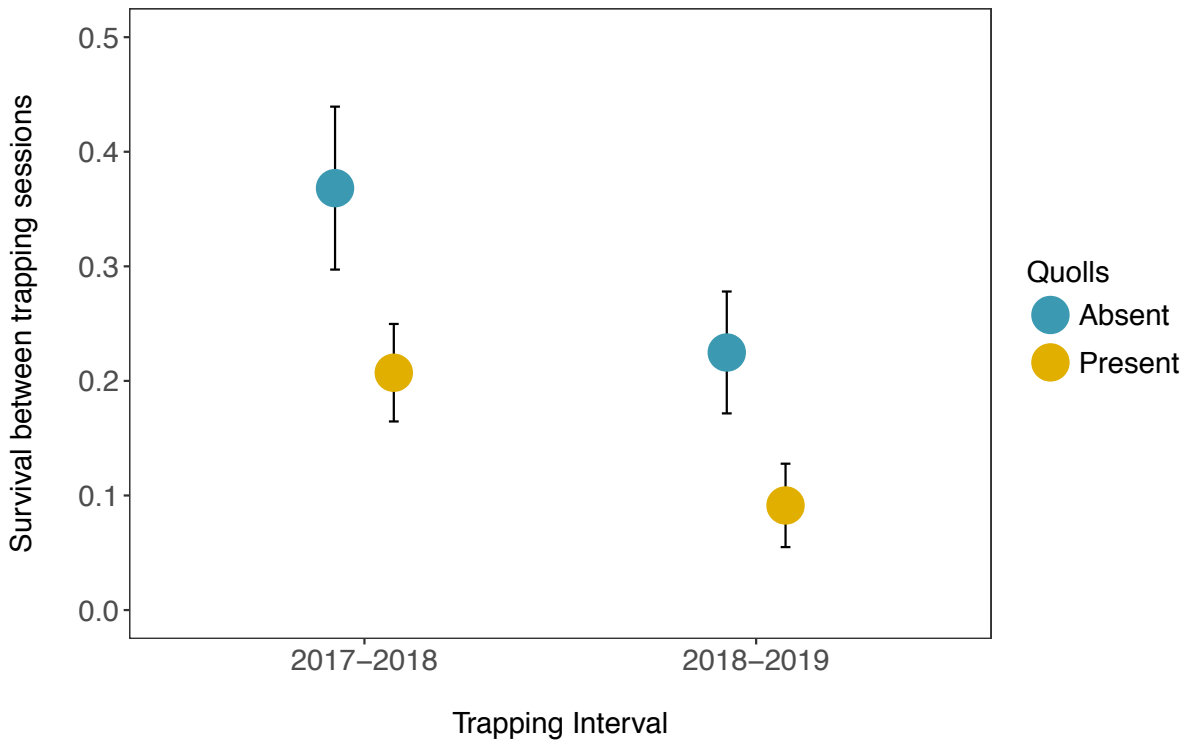

Figure 1. Between trapping session survival (± s.e.) of grassland melomys (*Melomys burtoni*) on Indian Island in quoll-invaded (*n* = 4) and quoll-free (*n* = 3) populations on Indian Island, Northern Territory, Australia.

Table 2. Results of Cormack-Jolly-Seber analyses used to compare survival (*Phi*) and recapture (*p*) probabilities of grassland melomys (*Melomys burtoni*) on Indian Island, Northern Territory, Australia. The symbols '.' and '*t*' refer to constant and time, respectively. Table shows AIC values and associated AIC weights, model likelihood, number of parameters (*N*), and model deviance. The term 'w/b' indicates that within trapping session survival rates (s_1-s_3, s_5-s_7, s_9-s_{11}) were constant and equivalent, and different to the between trapping session survival rates (s_4, s_8). The term 'group w/b' is as above, except that between trapping session survival rates differed between the two groups.

Model	AICc	Delta AICc	AICc Weights	Model Likelihood	N	Deviance
Phi (group w/b) p(t)	1688.629	0	0.92162	1	16	477.5224
Phi (w/b) p(t)	1693.562	4.933	0.07823	0.0849	14	486.6202
Phi (t) p (t)	1706.168	17.5394	0.00014	0.0002	21	484.553
Phi (g*t) p(t)	1712.896	24.2675	0	0	32	467.664
Phi (t) p (g*t)	1718.863	30.2344	0	0	32	473.631
Phi (t) p (.)	1719.862	31.2339	0	0	12	517.0641
Phi (t) p (g)	1720.751	32.1229	0	0	13	515.8843

Effect of novel predator on population size

Populations of melomys declined dramatically in quoll-invaded sites in the year following their introduction but not in quoll-free sites (Fig. 2). A strong negative interaction between the presence of quolls and trapping session in 2018 (mean = -1.194, 95% credible interval [-1.732, -0.665]) and 2019 (mean = -1.097, 95% confidence interval [-1.652, -0.551]) suggested that the presence of quolls dramatically impacted these populations of melomys (Fig. 2; Table 3).

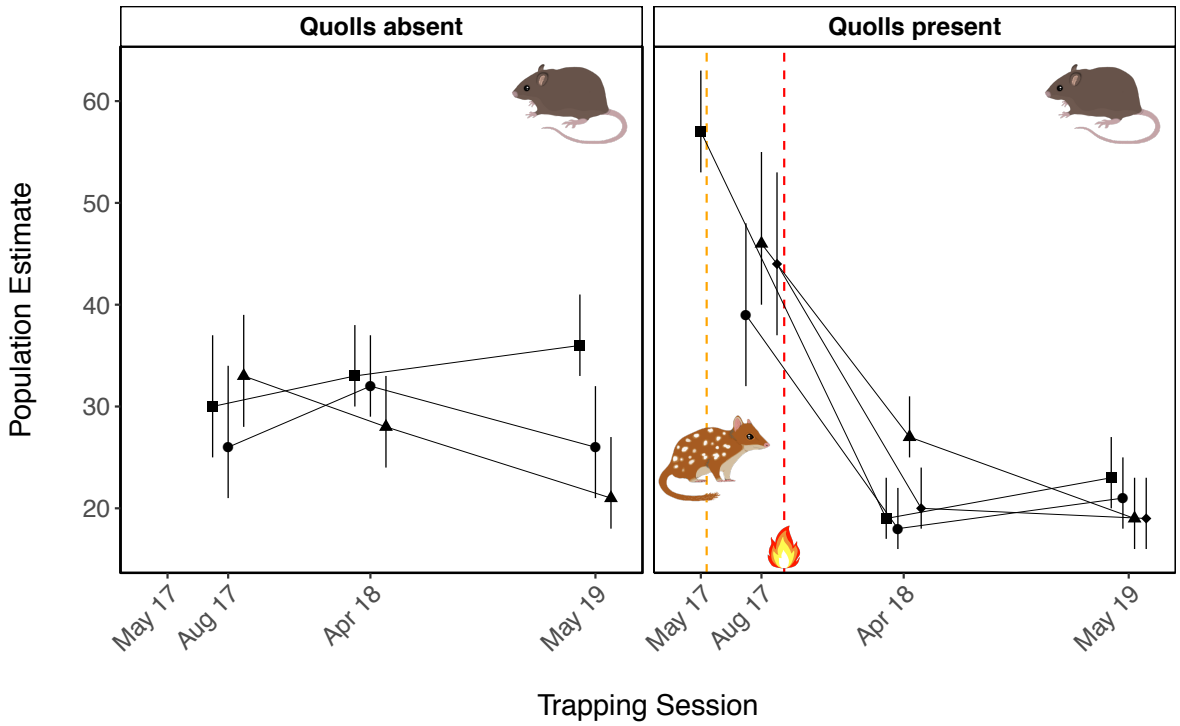

Figure 2. Posterior mean population sizes ($N_{ks} \pm$ 95% CI) for quoll-invaded and quoll-free populations of grassland melomys (*Melomys burtoni*) on Indian Island, Northern Territory, Australia. The orange dotted vertical line denotes the timing of the introduction of quolls. The red dotted vertical line denotes the timing of an unplanned fire that burnt through the quoll-invaded sites. Estimates assume closure of the population within each session and detection probability that varies across sessions.

353 **Table 3.** Model parameters and their priors including prior distributions, standard deviation, estimated
354 posterior means and their 95% credible intervals. N denotes normal probability distribution with mean and
355 standard deviation.

Model Parameters				
Name for parameter	**Parameter**	**Prior (mean, SD)**	**Posterior mean**	**95% credible intervals**
Detection:				
Intercept for detection	μ_d	N (0, 2.71)	-0.938	-1.118, -0.760
Effect of session 2 on detection	t_2	N (0, 2.71)	0.592	0.332, 0.852
Effect of session 3 on detection	t_3	N (0, 2.71)	0.458	0.179, 0.730
Population size:				
Intercept for Omega	μ_P	N (0, 2.71)	-0.906	-1.222, -0.583
Quoll Presence	r_2	N (0, 2.71)	0.698	0.314, 1.093
Trapping Session 2	b_2	N (0, 2.71)	0.061	-0.360, 0.478
Trapping Session 3	b_3	N (0, 2.71)	-0.100	-0.527, 0.339
Interaction 1 [Quoll Presence * Trapping Session 2]	$\gamma_{2,2}$	N (0, 2.71)	-1.194	-1.732, -0.665
Interaction 2 [Quoll Presence * Trapping Session 3]	$\gamma_{2,3}$	N (0, 2.71)	-1.097	-1.652, -0.551

356

357 *Effects of novel predator on prey behaviour*

358 For the proportion of melomys emerging in open field tests during the 0–10 min period,
359 there was a significant interaction between quoll presence and trapping session (χ^2 (5) =
360 4.386, P = 0.04; Fig. 3). There was no interaction between quoll presence and trapping
361 session for the proportion of melomys emerging and interacting with the novel object
362 during 10–20 min period (χ^2 (5) = 2.567, P = 0.109; Fig. 3). The model without this

interaction, however, revealed a significant effect of quoll presence on the neophobic behaviour of melomys in open field tests, with fewer melomys emerging from hiding and interacting with the novel object during the 10–20 min period of open field tests from sites where quolls were present than from sites where quolls were absent (χ^2 (5) = -4.696, P < 0.001; Fig. 3).

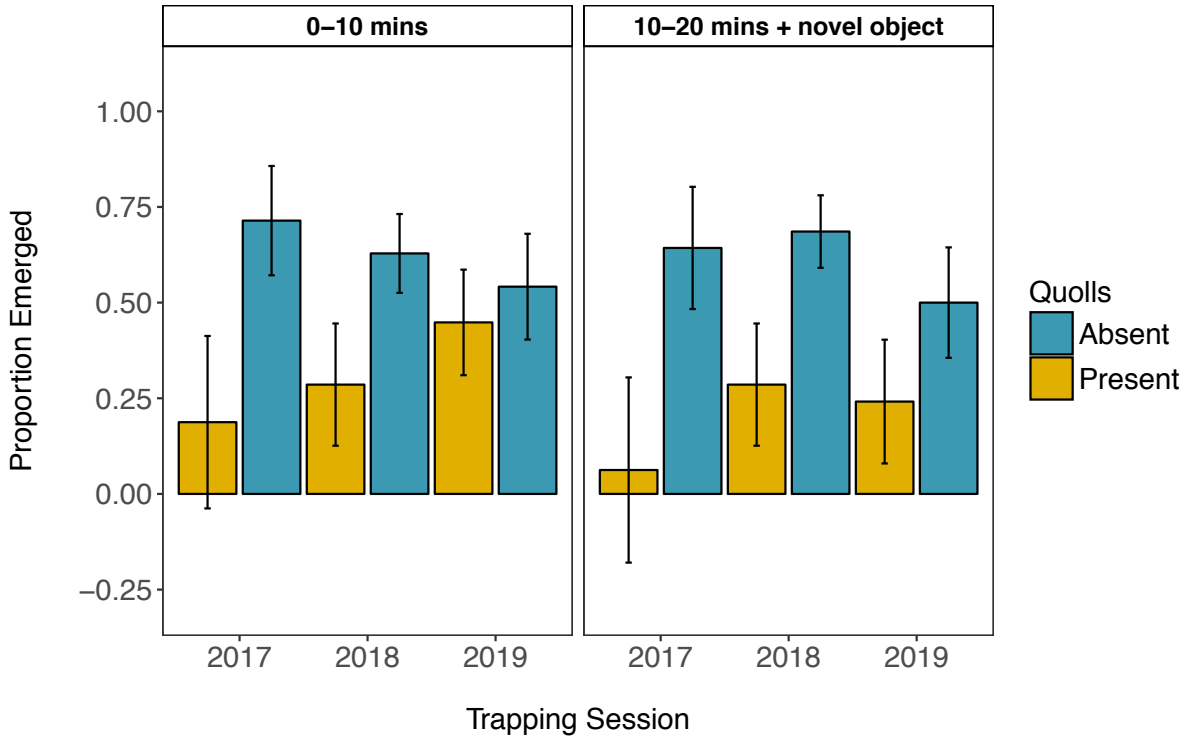

Figure 3. Mean proportion (± s.e.) of grassland melomys (*Melomys burtoni*) emerging from hiding during open field tests from quoll-invaded sites in 2017 (n = 16), 2018 (n = 28) and 2019 (n = 29), and quoll-free sites in 2017 (n = 14), 2018 (n = 35) and 2019 (n = 24) on Indian Island, Northern Territory, Australia.

Effects of novel predator on seed harvesting and predator-scent aversion

There was a very clear positive relationship between melomys density and seed take (t_{18} = 5.112, P < 0.001; Fig. 4) and a clear negative relationship between quoll presence and seed take (t_{18} = -2.344, P = 0.031; Fig. 4). There was, however, no interaction between melomys

density and quoll presence (t_{18} = -0.251, P = 0.805; Fig. 4). When we looked at the difference in seed take (Δ) between scent treatments within site.session, a striking pattern emerges, in which there is a clear interaction between the presence of quolls and session ($F_{3,17}$ = 18.61, P < 0.001; Fig. 5). Soon after the introduction of quolls to Indian Island in 2017, responses of melomys to predator-scented seeds were indistinguishable between populations with and without quolls, however, the difference in the responses of melomys diverged after quolls had been present for one (2018) and two years (2019), respectively (Fig. 5).

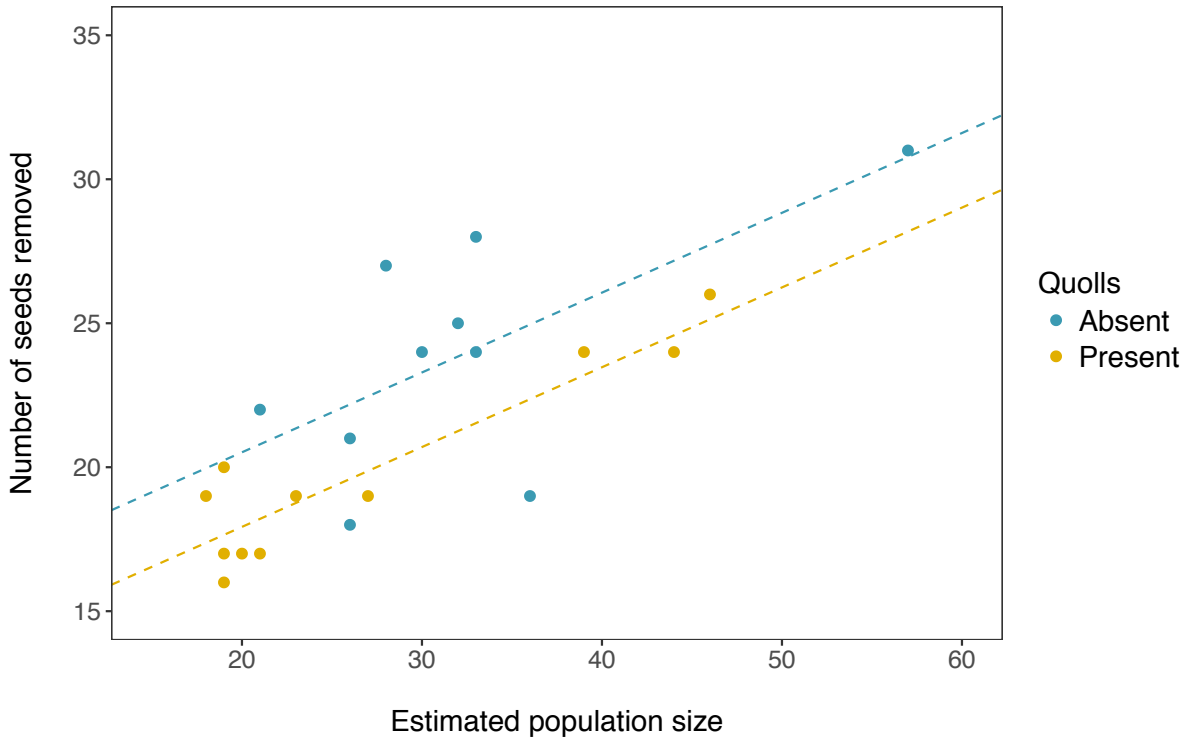

Figure 4. Effect of estimated population size on the number of control, unscented seeds removed from seed plots (n = 21) in quoll-invaded and quolls-free sites. Dotted lines denote the effect of quoll presence on seed removal rate.

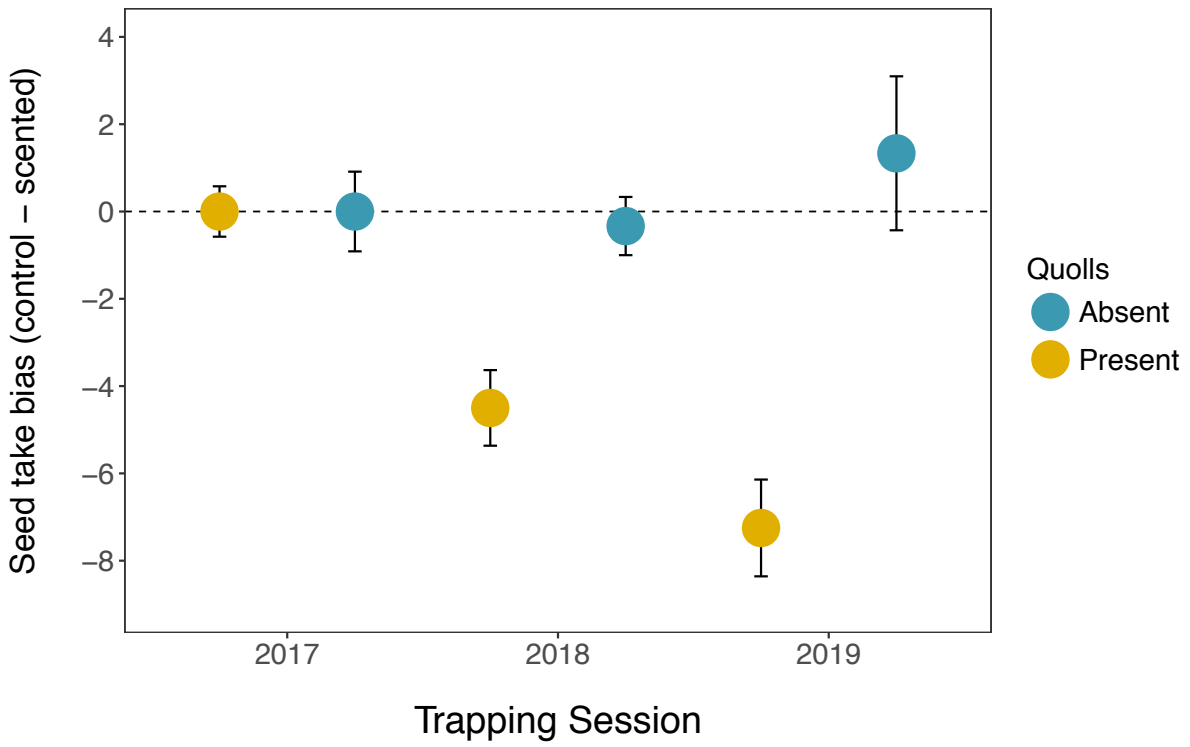

Figure 5. Mean (± s.e.) difference (Δ) between the number of control, unscented seeds and predator-scented seeds removed by melomys from quoll-invaded (*n* = 3; 2017 & *n* = 4; 2018-19) and quoll-free (*n* = 4; 2017 & *n* = 3; 2018-19) sites during each trapping session.

DISCUSSION

The introduction of northern quolls to Indian Island was associated with lowered survival (Fig. 1) and an immediate drop in population size (Fig. 2) in quoll-invaded melomys populations. This numerical effect on melomys density had an immediate impact on seed predation rates (Fig. 4), because seed take is strongly associated with the density of melomys in this system. This is a classic trophic cascade: predation suppresses herbivore density, which reduces the pressure that herbivores place on primary producers. Our study, however, also reveals an additional, subtler, cascade effect; driven by altered prey behaviour rather than by altered prey density (Fig. 3 & 5).

403 Within months of quolls appearing on the island, invaded populations of melomys
404 were significantly shyer than nearby, predator-free populations of conspecifics (Fig. 3). This
405 rapid but generalised response to a novel threat appears to have had a subtle effect on seed
406 predation rates: when we examine unscented seeds, per capita seed take is slightly lower in
407 quoll-invaded populations (Fig. 4). This generalised response appears to have been
408 supplemented over time with more threat-specific antipredator behaviours (Fig. 5).
409 Although the willingness of predator-exposed melomys to emerge from shelter (i.e.
410 boldness) converged through time with that of predator-free melomys, predator-exposed
411 melomys continued to be more neophobic than their predator-free conspecifics throughout
412 the study (Fig. 3). Meanwhile, predator-scent aversion, as evidenced by seed plots, steadily
413 increased over time (Fig. 5). Presumably the significant and novel predation pressure
414 induced by the introduction of quolls resulted in selection on behaviour and/or learning in
415 impacted rodent populations, allowing them to fine-tune their behavioural response
416 (decrease general shyness, but maintain neophobia, and respond to specific cues) as the
417 nature of the threat became clearer. These changing behavioural responses imply a
418 generalised reduction in seed take that also becomes fine-tuned over time, with high risk
419 sites (those that smell of predators) ultimately displaying substantially much lower seed
420 take than low risk sites. Thus, we see a reduction in seed take that applies to the entire
421 landscape becoming a fine-grained response, varying on a spatial scale measured in the tens
422 of metres.
423 Predation is a pervasive selective force in most natural systems, driving evolutionary
424 change in prey morphology, physiology, life history and behaviour. Unlike morphology and
425 physiology, however, the labile nature of behaviour makes it a particularly powerful trait for
426 rapid response in a changing world (Réale *et al.* 2007; Dall & Griffith 2014). Behavioural

427 comparisons of wild populations exposed to differing predation regimes provides some
428 support for the prediction that reduced boldness would be selected for under high
429 predation scenarios (Åbjörnsson *et al.* 2004; Bell 2005; Brydges *et al.* 2008) and that the
430 appearance of novel predators can result in bold individuals becoming shyer (Niemelä *et al.*
431 2012), however, the opposite pattern of response can also occur (Brown *et al.* 2005; Urban
432 2007) or behavioural phenotypes can be unrelated to predation regime (Laurila 2000;
433 Carlson & Langkilde 2014). Interestingly, a number of studies have demonstrated that
434 individuals from high-predation areas were quicker to emerge (Harris *et al.* 2010) and were
435 bolder and more aggressive (Bell & Sih 2007; Dingemanse *et al.* 2007) than predator-naïve
436 conspecifics. Although we found the opposite pattern to this immediately following the
437 arrival of a novel predator, by the second year after predator introduction we found the
438 emergence latency and boldness of melomys converging with that of predator-free
439 populations. It is plausible that with more time they may become bolder than their
440 predator-free conspecifics (by way of compensating for additional costs of acquiring food).
441 Thus, it is clear that the behavioural state of the population is dynamic, and it seems likely
442 this dynamism (and perhaps the capacity of the prey species to identify specific threats) may
443 explain some of the variation between earlier studies.
444 Although boldness may change over time, neophobia, as a generalised adaptive
445 response to predation pressure, is now well supported across a number of studies (Crane *et*
446 *al.* 2019). Individuals living under high predation risk scenarios have been shown to typically
447 display generalized neophobia (Brown *et al.* 2015; Elvidge *et al.* 2016), and neophobia can
448 increase the survival of predator-naïve individuals in initial encounters with predators
449 (Ferrari *et al.* 2015; Crane *et al.* 2018). Certainly, in our study, predator-exposed melomys

450 were significantly more neophobic than their predator-free conspecifics; an effect
451 maintained throughout the study.
452 Despite reduced survival, significant population declines, and clear behavioural
453 changes in invaded populations, it is impossible to determine with certainty from our data
454 whether changes in the behaviour of predator-invaded melomys populations are the result
455 phenotypic plasticity (learning) or natural selection. The low between trapping session
456 survival ($S_{2017–2018}$ = 0.207; $S_{2018–2019}$ = 0.091; Fig. 1) of melomys in quoll-invaded
457 populations means few individuals survive between sessions, so natural selection is a
458 possibility, and selection on these behavioural traits is potentially very strong. Although
459 behavioural changes in predator-invaded populations have been documented in a few
460 systems where predator introductions have been staged and experimentally controlled
461 (Lapiedra *et al.* 2018; Blumstein *et al.* 2019; Cunningham *et al.* 2019; Pringle *et al.* 2019),
462 elucidating whether these observed changes arise because of behavioural plasticity or
463 natural selection can be exceptionally difficult. Rapid behavioural responses of vulnerable
464 prey to recovered predators has been observed in a single prey generation, presumably
465 due to behavioural plasticity (Berger *et al.* 2001; Cunningham *et al.* 2019). Similarly,
466 behavioural adjustments to an introduced predator have been observed as a result of
467 natural selection on advantageous behavioural traits (Lapiedra *et al.* 2018). In this study,
468 although we had measures of individual behaviour, our between session recapture rates of
469 these individuals was sufficiently low that we had no longitudinal data on the behaviour of
470 individuals to test whether individuals were adjusting their behaviour or whether natural
471 selection was resulting in population-level change. It thus remains possible (and quite
472 likely) that both mechanisms were in play.

473 Although northern quolls represent a novel predator to melomys on Indian Island,
474 the two species' shared evolutionary history on the northern Australian mainland may
475 provide some explanation as to why this staged introduction resulted in rapid, finely-tuned
476 behavioural adjustment in melomys, rather than extinction. Isolation from predators can
477 rapidly result in the loss of antipredator behaviours from a prey species' behavioural
478 repertoire (Blumstein & Daniel 2005; Jolly *et al.* 2018a), dramatically increasing an
479 individual's susceptibility to predation following the introduction of either predator or prey
480 (Carthey & Banks 2014; Jolly *et al.* 2018b). But such outcomes are not inevitable: length of
481 isolation, co-evolutionary history, degree of predator novelty, density-dependent effects,
482 population size, and pre-existing predator-prey associations (Berger *et al.* 2001; Blumstein
483 2006; Banks & Dickman 2007; Sih *et al.* 2010; Carthey & Banks 2014) are all likely to be
484 hugely influential in determining whether an invaded population adjusts to the invader or
485 proceeds towards extinction. Recently, a conservation introduction of Tasmanian devils to
486 an island previously lacking them found that their possum prey rapidly adjusted their
487 foraging behaviour to accommodate this newly arrived predator (Cunningham *et al.* 2019).
488 Despite possums having lived on the island in isolation from devils since the 1950s,
489 presumably, their long evolutionary history together on mainland Tasmania had them
490 primed to respond to this predatory archetype (Sih *et al.* 2010; Carthey & Banks 2014;
491 Cunningham *et al.* 2019). This shared evolutionary history is likely responsible for both
492 possums' and melomys' ability to rapidly mount appropriate antipredator responses to the
493 introduction of these predators. The predators are novel within an individual's lifetime, but
494 the individual's ancestors have encountered them before.
495 Such considerations are of more than just theoretical interest. While native
496 Australian mammals clearly possess the ability to rapidly respond to some evolutionary

497 novel predators under certain circumstances (Carthey & Banks 2014; Blumstein *et al.* 2019),
498 many native mammals have been unable to mount rapid and/or effective enough defences
499 to thwart the invasion of evolutionarily novel predators (cats, foxes). This situation has led
500 to widespread predator-driven extinction on the Australian continent, resulting in Australia
501 leading the world in modern mammal extinctions (Woinarski *et al.* 2015).

502 Although our study documented dramatic population declines in predator-invaded
503 melomys populations, and we are assigning the causation of these declines to the
504 introduction of quolls, we need to address the unplanned, confounding factor of the fire
505 that burnt through northern Indian Island after completion of our population monitoring
506 and behavioural assays in 2017 (Fig. 2). Such fires are commonplace in the Australian wet-
507 dry tropics (Russell-Smith & Yates 2007); a regular disturbance that is often rapidly offset by
508 the annual monsoon driven wet season. Since our sites are composed of grass-free
509 woodland, the fire that burnt through them mostly burnt leaf-litter (though it reached the
510 mid-storey in other parts of the island). While this likely reduced the short-term availability
511 of food and cover for melomys, it is unlikely to directly explain the demographic effects we
512 observed (Fig. 1 & 2). A previous study investigating the effect of fire regimes on native
513 mammals in savanna woodland in Kakadu National Park, Northern Territory was unable to
514 detect an effect of fire frequency or intensity on the survival or recruitment of grassland
515 melomys, despite finding fire impacts in all other co-occurring native mammals studied
516 (Griffiths & Brook 2015). Interestingly, even in a system where fire is much more infrequent
517 and significantly more intense (e.g. mesic habitats of eastern Australia), grassland melomys
518 were found to be relatively unaffected by a wildfire that caused significant impacts to a co-
519 occurring native rodent, and any demographic impacts felt by melomys were entirely absent
520 within months of the fire (Liedloff *et al.* 2018). Additionally, the most dramatic behavioural

521 difference (boldness and neophobia) between quoll-invaded and quoll-free sites was
522 observed immediately prior to the occurrence of the fire (early August *vs.* mid-August 2017;
523 Fig 2 & 3). For the behavioural changes we observed that were potentially confounded by
524 fire, such as predator-scent aversion, we would expect to see these effects decreasing with
525 time since fire if fire was driving this response, instead we see the opposite trend (Fig. 5).
526 Finally, if food had become strongly limiting as a consequence of the fire, we would expect
527 to have observed an increase in seed take in the burned (quoll-invaded) sites, instead we
528 saw a decrease. For these reasons, we suspect the fire was unlikely to be directly
529 responsible for the demographic effects to melomys we observed, and fire cannot in any
530 way explain the response we observed to quoll-scented seeds. We, therefore, believe our
531 interpretation of these changes as being driven mostly by the addition of a novel predator
532 to the system is the most parsimonious and globally coherent interpretation of the data.
533 Fires in northern Australian savannas have, however, been shown to reduce cover for prey
534 species, facilitating and improving the hunting ability of predators (McGregor *et al.* 2014).
535 There is, therefore, a strong possibility that this fire may have facilitated the hunting
536 capabilities of quolls, increased the predation pressure and selection on behaviour imposed
537 by their arrival in this system.
538 Although our results suggest that invaded melomys populations are beginning to
539 adjust to the presence of northern quolls as a novel predator on Indian Island, there has
540 been no sign of demographic recovery from the addition of this predation pressure on the
541 island (Fig. 1 & 2). Data from our seed removal experiment clearly demonstrated that the
542 function of melomys as seed harvesters and dispersers scales with density (Fig. 4). Trophic
543 cascades resulting from the addition and loss of predators from ecosystems has been
544 observed in a number of systems globally (Ripple *et al.* 2001; Terborgh *et al.* 2001; Estes *et*

545 *al.* 2011), and the results can profoundly shape entire systems. As the only rodent and the
546 dominant granivore in this system, while melomys populations may not go extinct as a
547 result of quoll invasion, their reduced abundance and weakened ability to harvest and
548 disperse seeds may have dramatic, yet to be observed, longer-term consequences for the
549 vegetation structure and ecosystem function of Indian Island (McConkey & O'Farrill 2016).
550 Currently, grass is a rare vegetation feature on Indian island (though it is a dominant feature
551 of savanna woodlands generally), and this is quite possibly a result of the high density of
552 melomys on this (previously) predator-free island. The presence of quolls may well change
553 that, as both numerical and behaviour responses of melomys cascade down to the grass
554 community.

555 Empirical research on the effects of novel predators on recipient communities under
556 controlled conditions on a landscape-scale is exceptionally difficult and remains relatively
557 rare. The introduction of threatened predators to landscapes from which they have been
558 lost (Cunningham *et al.* 2019) or where they are entirely novel (Lapiedra *et al.* 2018),
559 however, provides a unique opportunity to observe how naïve prey can respond to novel
560 predators, and the mechanisms by which predators can structure communities. Our study
561 provides empirical support that some impacted prey populations can adjust rapidly to the
562 arrival of a novel predator via a generalised behavioural response (decreased boldness)
563 followed by development of a species-specific antipredator response (behavioural fine-
564 tuning). The arrival of the novel predator set off a trophic cascade that was driven, not only
565 by changed prey density, but also by changed prey behaviour. Thus, rapid adaptive shift may
566 allow prey populations to persist, but large-scale, system-wide changes may still follow.

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Data availability

The datasets generated and analysed during this study are available via figshare.com. https://doi.org/10.26188/5dd8fe5651e8b is currently embargoed but access can be granted on request.

586 **REFERENCES**

587 Åbjörnsson, K., Hansson, L.-A. & Brönmark, C. (2004). Responses of prey from habitats with
588 different predator regimes: local, adaptation and heritability. *Ecology*, 85, 1859–
589 1866.
590 Abrams, P.A. (2000). The evolution of predator-prey interactions: theory and evidence.
591 *Annu. Rev. Ecol. Syst.*, 31, 79–105.
592 Andersen, A.N., Cook, G.D., Corbett, L.K., Douglas, M.M., Eager, R.W., Russell-Smith, J., *et al.*
593 (2005). Fire frequency and biodiversity conservation in Australian tropical savannas:
594 implications from the Kapalga fire experiment. *Austral Ecol*, 30, 155–167.
595 Banks, P.B. & Dickman, C.R. (2007). Alien predation and the effects of multiple levels of prey
596 naiveté. *Trends in Ecology & Evolution*, 22, 229–230.
597 Begg, R., Walsh, B., Woerle, F. & King, S. (1983). Ecology of *Melomys burtoni*, the grassland
598 melomys (Rodentia: Muridae) at Cobourg Peninsula, N.T. *Wildlife Research*, 10, 259–
599 267.
600 Bell, A.M. (2005). Behavioural differences between individuals and two populations of
601 stickleback (*Gasterosteus aculeatus*): behavioural syndromes. *Journal of Evolutionary*
602 *Biology*, 18, 464–473.
603 Bell, A.M. & Sih, A. (2007). Exposure to predation generates personality in threespined
604 sticklebacks (*Gasterosteus aculeatus*). *Ecology Letters*, 10, 828–834.
605 Berger, J., Swenson, J.E. & Persson, I.-L. (2001). Recolonizing carnivores and naive prey:
606 conservation lessons from pleistocene extinctions. *Science*, 291, 1036–1039.
607 Blumstein, D.T. (2006). The multipredator hypothesis and the evolutionary persistence of
608 antipredator behavior. *Ethology*, 112, 209–217.

609 Blumstein, D.T. & Daniel, J.C. (2005). The loss of anti-predator behaviour following isolation
610 on islands. *Proceedings of the Royal Society B: Biological Sciences*, 272, 1663–1668.
611 Blumstein, D.T., Letnic, M. & Moseby, K.E. (2019). *In situ* predator conditioning of naive prey
612 prior to reintroduction. *Phil. Trans. R. Soc. B*, 374, 20180058.
613 Braithwaite, R. & Griffiths, A. (1994). Demographic variation and range contraction in the
614 northern quoll, *Dasyurus hallucatus* (Marsupialia: Dasyuridae). *Wildlife Research*, 21,
615 203.
616 Brown, C. & Braithwaite, V.A. (2004). Size matters: a test of boldness in eight populations of
617 the poeciliid *Brachyraphis episcopi*. *Animal Behaviour*, 68, 1325–1329.
618 Brown, C., Jones, F. & Braithwaite, V. (2005). *In situ* examination of boldness–shyness traits
619 in the tropical poeciliid, *Brachyraphis episcopi*. *Animal Behaviour*, 70, 1003–1009.
620 Brown, G.E., Elvidge, C.K., Ramnarine, I., Ferrari, M.C.O. & Chivers, D.P. (2015). Background
621 risk and recent experience influences retention of neophobic responses to
622 predators. *Behav Ecol Sociobiol*, 69, 737–745.
623 Brown, G.E., Ferrari, M.C.O., Elvidge, C.K., Ramnarine, I. & Chivers, D.P. (2013).
624 Phenotypically plastic neophobia: a response to variable predation risk. *Proceedings*
625 *of the Royal Society B: Biological Sciences*, 280, 20122712–20122712.
626 Brydges, N.M., Colegrave, N., Heathcote, R.J.P. & Braithwaite, V.A. (2008). Habitat stability
627 and predation pressure affect temperament behaviours in populations of three-
628 spined sticklebacks. *J Anim Ecology*, 77, 229–235.
629 Burnham, K.P. & Anderson, D.R. (1998). *Model selection and inference: a practical*
630 *information-theoretic approach*. Springer, New York.
631 Carlson, B.E. & Langkilde, T. (2014). No evidence of selection by predators on tadpole
632 boldness. *Behav*, 151, 23–45.

633 Carter, A.J., Feeney, W.E., Marshall, H.H., Cowlishaw, G. & Heinsohn, R. (2013). Animal
634 personality: what are behavioural ecologists measuring? *Biological Reviews*, 88, 465–
635 475.
636 Carthey, A.J.R. & Banks, P.B. (2014). Naïveté in novel ecological interactions: lessons from
637 theory and experimental evidence. *Biological Reviews*, 89, 932–949.
638 Courchamp, F., Chapuis, J.-L. & Pascal, M. (2003). Mammal invaders on islands: impact,
639 control and control impact. *Biological Reviews*, 78, 347–383.
640 Crane, A.L., Brown, G.E., Chivers, D.P. & Ferrari, M.C.O. (2019). An ecological framework of
641 neophobia: from cells to organisms to populations. *Biol Rev*, brv.12560.
642 Crane, A.L., Chivers, D.P. & Ferrari, M.C.O. (2018). Embryonic background risk promotes the
643 survival of tadpoles facing surface predators. *PLoS ONE*, 13, e0193939.
644 Cremona, T., Mella, V.S.A., Webb, J.K. & Crowther, M.S. (2015). Do individual differences in
645 behavior influence wild rodents more than predation risk? *Journal of Mammalogy*,
646 96, 1337–1343.
647 Cunningham, C.X., Johnson, C.N., Hollings, T., Kreger, K. & Jones, M.E. (2019). Trophic
648 rewilding establishes a landscape of fear: Tasmanian devil introduction increases
649 risk-sensitive foraging in a key prey species. *Ecography*, ecog.04635.
650 Dall, S.R.X. & Griffith, S.C. (2014). An empiricist guide to animal personality variation in
651 ecology and evolution. *Front. Ecol. Evol.*, 2, 1–7.
652 Dielenberg, R.A. & McGregor, I.S. (2001). Defensive behavior in rats towards predatory
653 odors: a review. *Neuroscience & Biobehavioral Reviews*, 25, 597–609.
654 Dingemanse, N.J., Wright, J., Kazem, A.J.N., Thomas, D.K., Hickling, R. & Dawnay, N. (2007).
655 Behavioural syndromes differ predictably between 12 populations of three-spined
656 stickleback. *J Anim Ecology*, 76, 1128–1138.

Doherty, T.S., Glen, A.S., Nimmo, D.G., Ritchie, E.G. & Dickman, C.R. (2016). Invasive predators and global biodiversity loss. *Proc Natl Acad Sci USA*, 113, 11261–11265.

Elvidge, C.K., Chuard, P.J.C. & Brown, G.E. (2016). Local predation risk shapes spatial and foraging neophobia patterns in Trinidadian guppies. *Curr Zool*, 62, 457–462.

Estes, J.A., Terborgh, J., Brashares, J.S., Power, M.E., Berger, J., Bond, W.J., *et al.* (2011). Trophic downgrading of planet Earth. *Science*, 333, 301–306.

Ferrari, M.C.O., McCormick, M.I., Meekan, M.G. & Chivers, D.P. (2015). Background level of risk and the survival of predator-naive prey: can neophobia compensate for predator naivety in juvenile coral reef fishes? *Proc. R. Soc. B*, 282, 20142197.

Fortin, D., Beyer, H.L., Boyce, M.S., Smith, D.W., Duchesne, T. & Mao, J.S. (2005). Wolves influence elk movements: behavior shapes a trophic cascade in Yellowstone National Park. *Ecology*, 86, 1320–1330.

Gelman, A. & Rubin, D.B. (1992a). Inference from iterative simulation using multiple sequences. *Statistical Science*, 7, 457–472.

Gelman, A. & Rubin, D.B. (1992b). [Practical Markov Chain Monte Carlo]: Rejoinder: replication without contrition. *Statistical Science*, 7, 503–511.

Griffiths, A.D. & Brook, B.W. (2015). Fire impacts recruitment more than survival of small-mammals in a tropical savanna. *Ecosphere*, 6, 1–22.

Harris, S., Ramnarine, I.W., Smith, H.G. & Pettersson, L.B. (2010). Picking personalities apart: estimating the influence of predation, sex and body size on boldness in the guppy *Poecilia reticulata*. *Oikos*, 119, 1711–1718.

Jolly, C.J., Webb, J.K. & Phillips, B.L. (2018a). The perils of paradise: an endangered species conserved on an island loses antipredator behaviours within 13 generations. *Biol. Lett.*, 14, 20180222.

Jolly, C.J., Kelly, E., Gillespie, G.R., Phillips, B. & Webb, J.K. (2018b). Out of the frying pan: Reintroduction of toad-smart northern quolls to southern Kakadu National Park. *Austral Ecology*, 43, 139–149.

Jolly, C.J., Webb, J.K., Gillespie, G.R., Hughes, N.K. & Phillips, B.L. (2019). Bias averted: personality may not influence trappability. *Behav Ecol Sociobiol*, 73, 129.

Kelly, E. (2019). *Targeted gene flow for conservation: northern quolls and the invasive cane toad.* PhD Thesis. University of Melbourne, Australia.

Kelly, E. & Phillips, B.L. (2016). Targeted gene flow for conservation. *Conservation Biology*, 30, 259–267.

Kemper, C.M., Kitchener, D.J., Humphreys, W.F., How, R.A., Bradley, A.J. & Schmitt, L.H. (1987). The demography and physiology of *Melomys* sp. (Rodentia: Muridae) in the Mitchell Plateau area, Kimberley, Western Australia. *Journal of Zoology*, 212, 533–562.

Kery, M. & Schaub, M. (2011). *Bayesian Population Analysis using WinBUGS: A Hierarchical Perspective*. Elsevier Science & Technology, San Diego, United States.

Lapiedra, O., Schoener, T.W., Leal, M., Losos, J.B. & Kolbe, J.J. (2018). Predator-driven natural selection on risk-taking behavior in anole lizards. *Science*, 360, 1017–1020.

Laundre, J.W., Hernandez, L. & Ripple, W.J. (2010). The landscape of fear: ecological implications of being afraid. *TOECOLJ*, 3, 1–7.

Laurila, A. (2000). Behavioural responses to predator chemical cues and local variation in antipredator performance in *Rana temporaria* tadpoles. *Oikos*, 88, 159–168.

Liedloff, A.C., Wilson, J.C. & Engeman, R.M. (2018). The effect of wildfire on population dynamics for two native small mammal species in a coastal heathland in Queensland, Australia. *Acta Oecologica*, 88, 58–64.

Lima, S.L. & Dill, L.M. (1990). Behavioral decisions made under the risk of predation: a review and prospectus. *Can. J. Zool.*, 68, 619–640.

López, P., Hawlena, D., Polo, V., Amo, L. & Martín, J. (2005). Sources of individual shy–bold variations in antipredator behaviour of male Iberian rock lizards. *Animal Behaviour*, 69, 1–9.

McConkey, K.R. & O'Farrill, G. (2016). Loss of seed dispersal before the loss of seed dispersers. *Biological Conservation*, 201, 38–49.

McGregor, H.W., Legge, S., Jones, M.E. & Johnson, C.N. (2014). Landscape management of fire and grazing regimes alters the fine-scale habitat utilisation by feral cats. *PLoS ONE*, 9, e109097.

McGregor, I.S., Schrama, L., Ambermoon, P. & Dielenberg, R.A. (2002). Not all 'predator odours' are equal: cat odour but not 2,4,5 trimethylthiazoline (TMT; fox odour) elicits specific defensive behaviours in rats. *Behavioural Brain Research*, 129, 1–16.

Medina, F.M., Bonnaud, E., Vidal, E., Tershy, B.R., Zavaleta, E.S., Josh Donlan, C., *et al.* (2011). A global review of the impacts of invasive cats on island endangered vertebrates. *Glob. Change Biol.*, 17, 3503–3510.

Moore, H.A., Dunlop, J.A., Valentine, L.E., Woinarski, J.C.Z., Ritchie, E.G., Watson, D.M., *et al.* (2019). Topographic ruggedness and rainfall mediate geographic range contraction of a threatened marsupial predator. *Divers Distrib*, ddi.12982.

Niemelä, P.T., DiRienzo, N. & Hedrick, A.V. (2012). Predator-induced changes in the boldness of naïve field crickets, *Gryllus integer*, depends on behavioural type. *Animal Behaviour*, 84, 129–135.

Oakwood, M. (1997). The Ecology of the Northern Quoll, *Dasyurus hallucatus*. PhD Thesis. Australian National University, Australia.

Oakwood, M., Woinarski, J.C.Z. & Burnett, S. (2016). *Dasyurus hallucatus* (northern quoll). The IUCN Red List of Threatened Species 2016: e.T6295A21947321. http://dx.doi.org/10.2305/IUCN.UK.2016-2.RLTS.T6295A21947321.en. Downloaded on 31 October 2019.

Plummer, M., Stukalov, A. & Denwood, M. (2017). rjags: Bayesian graphical models using MCMC. *URL http://CRAN. R-project. org/package= rjags. R package version, 2, 0-4.*

Pringle, R.M., Kartzinel, T.R., Palmer, T.M., Thurman, T.J., Fox-Dobbs, K., Xu, C.C.Y., *et al.* (2019). Predator-induced collapse of niche structure and species coexistence. *Nature*, 570, 58–64.

R Core Team. (2019). *R: A Language and Environment for Statistical Computing.* R Foundation for Statistical Computing, Vienna.

Réale, D., Reader, S.M., Sol, D., McDougall, P.T. & Dingemanse, N.J. (2007). Integrating animal temperament within ecology and evolution. *Biological Reviews*, 82, 291–318.

Relyea, R.A. (2003). Predators come and predators go: the reversiblity of predator-induced traits. *Ecology*, 84, 1840–1848.

Ripple, W.J., Larsen, E.J., Renkin, R.A. & Smith, D.W. (2001). Trophic cascades among wolves, elk and aspen on Yellowstone National Park's northern range. *Biological Conservation*, 102, 227–234.

Royle, J.A., Dorazio, R.M. & Link, W.A. (2007). Analysis of multinomial models with unknown index using data augmentation. *Journal of Computational and Graphical Statistics*, 16, 67–85.

Russell-Smith, J. & Yates, C.P. (2007). Australian savanna fire regimes: context, scales, patchiness. *Fire Ecology*, 3, 48–63.

752 Sax, D., Stachowicz, J., Brown, J., Bruno, J., Dawson, M., Gaines, S., *et al.* (2007). Ecological
753 and evolutionary insights from species invasions. *Trends in Ecology & Evolution*, 22,
754 465–471.
755 Schoener, T.W. & Spiller, D.A. (1996). Devastation of prey diversity by experimentally
756 introduced predators in the field. *Nature*, 381, 691–694.
757 Shine, R. (2010). The ecological impact of invasive cane toads (*Bufo marinus*) in Australia.
758 *The Quarterly Review of Biology*, 85, 253–291.
759 Sih, A. (1992). Prey uncertainty and the balancing of antipredator and feeding needs. *The*
760 *American Naturalist*, 139, 1052–1069.
761 Sih, A., Bolnick, D.I., Luttbeg, B., Orrock, J.L., Peacor, S.D., Pintor, L.M., *et al.* (2010).
762 Predator-prey naïveté, antipredator behavior, and the ecology of predator invasions.
763 *Oikos*, 119, 610–621.
764 Sih, A., Ziemba, R. & Harding, K.C. (2000). New insights on how temporal variation in
765 predation risk shapes prey behavior. *Trends in Ecology & Evolution*, 15, 3–4.
766 Simberloff, D., Martin, J.-L., Genovesi, P., Maris, V., Wardle, D.A., Aronson, J., *et al.* (2013).
767 Impacts of biological invasions: what's what and the way forward. *Trends in Ecology*
768 *& Evolution*, 28, 58–66.
769 Suraci, J.P., Clinchy, M., Dill, L.M., Roberts, D. & Zanette, L.Y. (2016). Fear of large carnivores
770 causes a trophic cascade. *Nature Communications*, 7, 10698.
771 Tanner, M.A. & Wong, W.H. (1987). The calculation of posterior distributions by data
772 augmentation. *Journal of the American Statistical Association*, 82, 528–540.
773 Terborgh, J., Lopez, L., Nunez, P., Rao, M., Shahabuddin, G., Orihuela, G., *et al.* (2001).
774 Ecological meltdown in predator-free forest fragments. *Science*, 294, 1923–1926.

Urban, M.C. (2007). Risky prey behavior evolves in risky habitats. *Proceedings of the National Academy of Sciences*, 104, 14377–14382.

Woinarski, J.C.Z., Burbidge, A.A. & Harrison, P.L. (2015). Ongoing unraveling of a continental fauna: decline and extinction of Australian mammals since European settlement. *Proceedings of the National Academy of Sciences*, 112, 4531–4540.