Biochar-induced Root Exudates Rewire the Rhizosphere Microbiome and Its

Functionality

Hanyue Yang[1], Amina Furrkukh Mughal[2], Yaqi You[1,2,*]

[1]Department of Environmental Resources Engineering, SUNY College of Environmental

Science and Forestry

[2]Graduate Program in Environmental Science, SUNY College of Environmental Science and

Forestry

*Correspondence:

Yaqi You

402 Baker Lab, 1 Forestry Dr, Syracuse, NY 13210

E-mail: yyou@esf.edu

Phone: +1-315-470-6765

ABSTRACT

Biochar amendment has multifaceted benefits on the soil-plant system; yet the underlying mechanisms remain poorly understood. Integrating microfluidics and multi-omics and using wheat as a model plant, we demonstrate that biochar induces differential root exudation which rewires the rhizosphere microbiome and its functionality. Complex molecules in biochar-induced root exudates, particularly plant secondary metabolites and phytohormones, evoke a plant-beneficial rhizosphere microbiome centered by diverse plant growth promoting rhizobacteria (PGPR). Functionally, the restructured microbiome exhibits shifted nitrogen metabolism, characterized by enhanced nitrogen fixation and complete denitrification with reduced N_2O emission potential. Additionally, biochar suppresses methane production through orchestrated microbial community functional shifts, including stimulated methionine salvage that deprives the methanogenesis precursor methanethiol; suppressed biosynthesis of coenzyme M, a cofactor required for the methane-forming step; and an enrichment of methanotrophs and thus enhanced methane oxidation. The restructured microbiome also contributes to the transformation of phytohormones and biosynthesis of redox-active quinone compounds, having profound impact on the soil-plant system. Together, our work provides new insights into biochar's multifaceted roles in reprogramming the rhizosphere and highlights the promise of engineering the rhizosphere through reshaping root-microbe interactions using designed biochar or root exudate cocktails. Given the fact that fire-derived black carbon shares similar chemistry with biochar and that increased frequency of wildfire could cause more black carbon deposition, this research also spotlights the overlooked impact of fire-derived black carbon on belowground carbon and nitrogen dynamics through intricate rhizosphere interactions. New mechanisms unveiled in this research have implications for both sustainable agriculture and general soil biogeochemistry.

KEYWORDS: *biochar, multi-omics, root exudates, secondary metabolites, phytohormone, rhizosphere microbiome, plant growth promoting rhizobacteria, soil metabolome, nitrogen cycling, methane cycling*

INTRODUCTION

Sustainable soil management is vital for safeguarding food production, environmental health, and human well-being[1]. Biochar has been widely recognized as a multifunctional soil enhancer, capable of improving soil health, boosting crop productivity and stress resistance, reducing nutrient leaching, and mitigating greenhouse gas (GHG) emissions[2, 3]. These benefits are often attributed to biochar-induced changes in soil physicochemical properties, such as increased nutrient availability, enhanced water-holding capacity, reduced acidity or toxic substances[4, 5]. Emerging evidence, however, started to unveil more profound influence of biochar on the soil-plant system.

Soil is a dynamic living system, comprising diverse and complex microbial communities that drive numerous biogeochemical reactions[6]. Increasing evidence suggests that biochar amendment can alter microbial abundance, diversity, community structure, and functional compositions[5, 7]. However, few studies have investigated biochar's effects on the rhizosphere—a narrow, dynamic zone that serves as a critical interface between plant roots and soil microorganisms and a key component of the plant holobiont[2, 8, 9]. Our recent study found that biochar can modulate root endometabolites and rhizosphere microbiome[10]. But not all root endometabolites are secreted into the soil; root exudates that are actively secreted directly influence rhizosphere biogeochemistry and thus require further investigation.

Plants use root exudates to recruit plant-growth promoting rhizobacteria (PGPR) and thereby gain growth advantages such as enhanced resistance to diseases and stresses[11, 12]. Whether biochar induces root exudates favoring PGPR is still unknown. Further, nitrogen (N) is a key limiting nutrient for plant productivity in most terrestrial ecosystems, especially agroecosystems[13]. While biochar is reported to affect microbially mediated N cycling, including N fixation, nitrification, and denitrification, in bulk soil[14-16], its effect on rhizosphere N cycling is unclear. Specifically, biochar-induced root exudates may play an important yet understudied role therein[17, 18]. Unleashing biochar's potential in sustainable agriculture requires a mechanistic understanding of biochar's modulating effects on root exudation and the rhizosphere microbiome.

This study aimed to use wheat, a fertilizer-intensive cereal crop, as a model plant to

elucidate biochar-induced chemical and biological shifts in the rhizosphere, with a focus on

molecular mechanisms of biochar-induced rhizosphere reprogramming. Three specific objectives

were: (1) identify biochar-induced root exudates, (2) characterize biochar-induced, root exudate-

mediated rhizosphere microbiome restructuring, and (3) determine biochar's modulation on

rhizosphere biogeochemistry, especially N and methane cycling. We conducted controlled

experiments using laboratory-produced, well-characterized wheat biochar and a microfluidics

device called EcoFAB (ecosystem fabrication), a modular growth system developed for studying

the rhizosphere throughout plant development stages[19, 20]. We analyzed root exudates and the

soil metabolome using untargeted metabolomics, the rhizosphere microbiome and its function

using next-generation sequencing and real-time quantitative PCR (qPCR), and root exudate-

microbe associations using a combination of network analysis and machine learning. Our work

provides new mechanistic insights into biochar's multifaceted roles in reprogramming the

rhizosphere and the potential of engineering the rhizosphere through reshaping root-microbe

interactions.

MATERIALS AND METHODS

Biochar production and characterization

Biochar was produced by slow pyrolysis of wheat straw (350 °C for 30 min) with no oxygen

in a laboratory reactor, the same as our prior work[10]. We chose low pyrolysis temperature due to

the fact that biochar produced at lower pyrolysis temperatures tends to promote soil enzyme

activity more than biochar produced at higher pyrolysis temperatures[2]. Biochar was characterized

using scanning electron microscopy (SEM) for microscopic structure, energy-dispersive X-ray

spectroscopy (EDS) for surface elemental composition, Fourier-transform infrared spectroscopy

(FTIR) for molecular composition and functional groups, Boehm titration for three oxygen-

containing functional groups (acidic carboxyl, lactone, and phenolic group) on the surface, and X-

ray diffraction (XRD) for crystal structure, in addition to conventional analysis of pH, major and

trace elements, electrical conductivity, and cation exchange capacity (**Table S1**). Details are in

the Supporting Information (SI).

EcoFAB fabrication

We used EcoFAB devices, a fabricated ecosystem, for wheat plant cultivation, rhizosphere

observation, and sample collection[20]. This setting enables continuous investigation of the

rhizosphere under highly controlled conditions. We downloaded the original design files

(https://eco-fab.org/device-design/) and customized the design to accommodate wheat seedlings.

Acrylonitrile butadiene styrene (ABS) mold parts were fabricated using a MakerBot Replicator 2X

3D printer, acrylic mold parts were fabricated on a lathe machine, and the assembled mold was

used to cast polydimethylsiloxane (PDMS) layers (**Fig S1**)[21]. Our prototype EcoFAB device

consisted of a PDMS oval chamber (97.6 × 64 × 3.2 mm) adhering to a large rectangular coverslip

(102 × 76 × 1.2 mm) (Ted Pella) in an airtight way (**Fig S1**). Each device was enclosed in a

transparent cover box (**Fig 2A; Fig S1**) to maintain a sterile environment for plant experiment.

Before experiments, the EcoFAB chamber was rinsed with sterile Milli-Q water; the device surface

and the cover box were sterilized using 70% ethanol for 30 minutes and then 100% ethanol for 5

min. More details about EcoFAB fabrication and device assembly are in the SI.

Plant experiment and sample collection

In this study, we applied wheat biochar at a rate of 0.25% (w/w) based on cumulative

evidence that lower rates promote plant growth while higher rates inhibit growth and our prior

observations that wheat biochar at 0.25% rate yields unique rhizosphere alterations[10]. To prepare

soil slurries, 100 g of agricultural soil (**Table S1**) was added to 1 L sterile Milli-Q water, with or

without biochar. Soil slurries were shaken for 16 hours to reach thorough mixing, and slowly

injected into each EcoFAB chamber with a sterile syringe. Wheat (*Triticum aestivum* L.) seeds

were germinated and seedlings of similar sizes were transferred aseptically into sterilized EcoFAB

devices containing soil slurries with or without biochar (one plant per device; four biological

replicates for each treatment). Plants were cultivated under full spectrum LED light simulating

natural sunlight (12 hour light, 12 hour dark) and controlled humidity (80%) at 25 °C for 21 days,

and the development of root systems was monitored (**Fig S1**). Milli-Q water was added to EcoFAB

devices every three days to maintain humidity. To ensure reproducibility, the experiment was

conducted twice (8 EcoFAB devices in each independent round).

After 21 days, individual EcoFAB devices were dissembled, and remaining soil slurries

and plant seedlings were collected. Soil slurries were immediately stored at -80°C until metabolite

extraction, while seedlings were immediately processed to collect rhizosphere soils like in our

prior studies[10, 22]. After rhizosphere soil removal, individual plants with clean root surfaces were

transferred to sterile 15 mL centrifuge tubes containing 15 mL Milli-Q water to allow for exudation

for 24 hours like in our prior study[23]. An aliquot of the collected exudated solution (1.5 mL) was

transferred to a 2.0 mL sterile microcentrifuge tube, dried using a vacuum concentrator (Savant

SpeedVac SPD300DDA, Thermo Scientific), and 1.5 mL of methanol solution (80% v/v in Milli-Q

water) was added to the tube, followed by vortexing for 15 minutes and sonication bath for 5

minutes at room temperature. The resulting extract was filtered through a 0.22 μm PVDF filter

(Millipore) and stored at -80 °C until metabolomics analysis. Similar procedures were used to

extract soil exometabolites (detailed in the SI).

Untargeted metabolomics

Root exudates and soil slurries were analyzed using liquid chromatography-tandem mass

spectrometry (LC-MS/MS) performed on a Vanquish LC system in conjunction with a Q Exactive

HF Orbitrap LC-MS system[24] coupled with an electrospray ionization (ESI) source (Thermo

Fisher). The MS has a resolving power of up to 240,000 FWHM (Full Width at Half Maximum) at

500 m/z and a mass range of 20-1200 m/z. Extract aliquots of 10 μL were injected into a Hypersil

GOLD C18 column (1.9 μm, 2.1×150 mm) (Thermo Fisher) operated at 30 °C. LC mobile phases

consisted of solvent A (0.1% formic acid in Milli-Q water) and solvent B (0.1% formic acid in LC/MS

grade acetonitrile), and the flow rate was maintained at 300 μL/min. MS was operated in both

positive and negative ion mode (ESI+ and ESI-). More LC-MS details are in the SI.

Spectral data was processed as before[10]. Raw data was converted to the generic mzXML

format and processed using three platforms (MetaboAnalyst 6.0, XCMS, MZmine3)[25-27], and the

resulting feature tables were merged without redundancy and further processed in MetaboAnalyst

6.0 for functional analysis. Compounds were converted to the International Union of Pure and

Applied Chemistry (IUPAC) names and classified to the "Class" level of the purely structure-based

chemical taxonomy (ChemOnt) using ClassyFire[28]. Principal coordinates analysis (PCoA) was

performed based on Jaccard distance using the "vegan" package[29]. Significant fold change (FC)

was determined by the Mann-Whitney U test with Benjamini-Hochberg correction ($|log_2FC| > 1$

and $p < 0.05$). For network analysis, pairwise Pearson correlation with false discovery rate (FDR)

control was calculated using the "WGCNA" package[30]; only significant correlations ($r > 0.9$ and

adjusted $p < 0.01$) were retained. Network topology was assessed with the "igraph" package[31].

Node degree distributions were fitted to a power-law model, and nodes with degrees exceeding

the 95th percentile were identified as highly connected hub nodes[10].

Next generation sequencing, qPCR, and biofilm assay

Metagenomic DNA was extracted from rhizosphere soils using the ZymoBIOMICS

DNA/RNA Miniprep Kit following the manufacturer's protocol (Zymo Research). A PowerLyzer

Homogenizer was used to maximize extraction efficiency. The resulting DNA was assessed on a

Qubit 4 fluorometer with the 1× dsDNA High Sensitivity Assay Kit (Thermo Fisher) and by

electrophoresis. 16S rRNA gene sequencing and data processing were performed using QIIME2,

DADA2, and the "phyloseq" package as before[10, 22]. The "edgeR" (Empirical Analysis of Digital

Gene Expression in R) package was used to identify differential taxa between samples.

PICRUSt2 was used to infer metagenomic functional contents[32]; PICRUSt2 predictions were

subjected to differential abundance analysis using the R package "ggpicrust2"[33]. Microbial

community growth rate, i.e., median predicted minimum doubling time (mPMDT), was predicted

using the R package "gRodon"[34, 35]. Microbial network analysis was constructed using SparCC[36].

Community assembly mechanisms were assessed using the null model approach. We calculated

the β-nearest taxon index (βNTI) using the R package "iCAMP" to estimate the relative importance

of deterministic versus stochastic assembly processes[37], and calculated the Raup-Crick metric

using Bray-Curtis dissimilarities (RC_{BC}) to estimate the relative importance of selection, drift,

dispersal limitation, and homogenizing dispersion[38]. More details on sequencing and data

processing are in the SI.

While PICRUSt2 inference provides a more comprehensive snapshot of community

functions, inference accuracy varies. Therefore, we used qPCR to quantify the abundance of

eubacterial 16S rRNA gene (proxy of total bacteria) and key genes involved in nitrogen (*nifH*, *nxrA*,

nirK, *nirS*, *nosZ* Clade I) and methane (*mcrA*, *pmoA*) cycling using established primers (**Table

S2**). qPCR was performed on a CFX Duet Real-Time PCR System (BioRad) with SsoAdvanced

universal SYBR green supermix (Bio-Rad), 2 μL of DNA template (2-fold diluted to mitigate soil

matrix inhibition effect), and primers (0.48 μM each) in a 25 μL reaction volume. The thermal

program consisted of an initial denaturation at 95°C for 2 min, followed by 35 (16S, *mcrA*, *nxrA*)

or 40 (*nifH*, *nirK, nirS*, *nosZ, pmoA*) cycles of 10 sec at 95 °C, 20 sec at the annealing temperature

(**Table S2**), and 30 sec at 72 °C. A melting curve analysis was conducted after each run to verify

reaction specificity. Standard curves were generated using a cloned PCR product for 16S rRNA

gene and synthetic DNA (gBlocks, Integrated DNA Technologies) for other genes (**Fig S2**),

following the MIQE guidelines[39]. More details of qPCR and data processing are in the SI.

To test whether biochar-induced root exudates promote biofilm formation compared to root

exudates from control plants, we conducted a biofilm assay using a model soil bacterium

Pseudomonas putida KT2440[40]. The results were normalized against dissolved organic carbon

(DOC) content in root exudates to exclude the effect of varying C content (details in the SI).

Multi-omics integration

We used two approaches to identify associations between root exudates and rhizosphere

microbes: (1) OmicsNet 2.0 which utilizes multiple databases to create multi-omics networks, and

(2) mmvec (microbe-metabolite vectors), a neural network approach that predicts an entire

metabolite profile from a single microbe sequence[41, 42]. OmicsNet 2.0 calculates metabolite-

microbe correlations based on microbial metabolic capacity and pertinent proteins that are

predicted using high-quality genome-scale metabolic models. The mmvec approach estimates

the conditional probability of observing a metabolite given that a microbe is present to identify the

most likely metabolite-microbe interactions. Compared to approaches based on correlation (e.g.,

Spearman's) or network (e.g., SparCC), mmvec can identify previously undiscovered patterns

and is less prone to false discovery (positive or negative) as it provides consistency between

absolute and relative abundances of compositional data. However, mmvec does not calculate the

statistical significance or confidence interval of each metabolite-microbe pair. Here we combined

the two approaches to maximize the discovery of true metabolite-microbe interactions. The highly

co-occurring 10 root exudates and 12 microbial genera, identified by both approaches, were

visualized using a heatmap. Similar approaches were used to identify and visualize associations

between rhizosphere microbes and soil exometabolites. More details are in the SI.

General statistics and data visualization

Bioinformatics, statistics, and data visualization were performed in R (4.2.1) and Python

(3.7) with various R packages and Python modules. Significant differences were determined using

one-way analysis of variance (ANOVA) or Welch's *t* test, nonparametric statistics (Wilcoxon or

Kruskal-Wallis), and permutational multivariate analysis of variance (PERMANOVA), depending

on the variable. Network visualization used Gephi (0.10.1) and Cytoscape (3.9.1). More details

are in the SI.

RESULTS AND DISCUSSION

1. Biochar significantly promotes wheat plant growth

SEM indicated the wheat biochar used here had a porous morphology (**Fig 1A**). FTIR

spectra showed the biochar contained aromatic compounds (particularly those with benzene rings

and quinone groups), C–O bonds from esters or ethers, nitriles, silicates and/or carbonates, and

possibly C–N or P–O (**Fig 1B; Fig S3**) (more details in the SI)[43-45]. EDS analysis revealed the

surface of the biochar predominantly contained C (62.36%), O (24.52%), Si (6.13%), K (3.06%),

and Ca (1.04%) (averaged atomic%, n = 29) (**Fig 1C**). Consistent with the EDS results, XRD

spectrum indicated the presence of amorphous C, graphite, and various mineral crystals

containing Si, K, and Ca (**Fig 1D**) (more details in the SI)[44, 46]. These results reflected the biochar's

feedstock (wheat straw) and the production condition (slow pyrolysis at 350 °C).

In the 21-day EcoFAB experiment (**Fig 2A**), application of 0.25% (w/w) wheat biochar

yielded 1.59 times more fresh weight of shoots (Welch t test: $p < 0.001$) (**Fig S4**). Fresh weight of

roots had no significant difference (Welch t test: $p = 0.413$), but biochar-treated plants had more

lateral roots (**Fig S4**). These results are consistent with our prior 15-day pot study[10] and suggest

biochar may promote wheat growth more significantly in larger systems.

2. Biochar induces differential root exudation

Biochar induced differential root exudation (**Table S3**), as shown by PCoA of Jaccard

distance of 312 exudate compounds detected across samples from two independent rounds of

EcoFAB experiments (PERMANOVA: $p = 0.145$ for round x treatment) (**Fig S5**). This is due to

systematic regulation of root metabolism by biochar, as evidenced by increased average degree

and centrality of root exudate co-occurrence network (**Fig S6**). Specifically, biochar significantly

regulates carbohydrate metabolism, glycan biosynthesis, the metabolism of terpenoids and

polyketides, and nucleotide metabolism, all of which are highly involved in plant growth and

development, signaling, and stress defense[47-49].

Among the 312 exudate compounds, six compounds were only identified in biochar-

induced root exudates, including two flavanonols, one flavanone and its direct precursor, a

phenolic ester, and dUMP (**Table S3**). Further, biochar significantly affected 39 compounds

($|Log_2FC| > 1$, p-$adjusted < 0.05$) from 14 ChemOnt classes: 10 organooxygen compounds, 6

flavonoids, 3 prenol lipids, 3 fatty acyls, 2 from each of four classes (organonitrogen compounds,

steroids and steroid derivatives, imidazopyrimidines, carboxylic acids and derivatives), 1 from

each of five classes (5'-deoxyribonucleosides, pyrimidine nucleotides, furanoid lignans, linear 1,3-

diarylpropanoids, organic phosphoric acids and derivatives), and 4 unclassed chemicals (**Fig 2B**).

Three compounds having regulatory roles in plant defense were upregulated, namely beta-1,4-

poly-N-acetyl-D-glucosamine (chitin) (2.2 fold), (22S)-22-hydroxycampest-4-en-3-one (a direct

precursor of brassinosteroids (BRs)) (2.6 fold), and 2,4-dihydroxy-1,4-benzoxazin-3-one-

glucoside (DIBOA-Glc, a benzoxazinoid (BX) compound) (5.2 fold). Chitin enhances plant

immune responses by activating defense-related genes as an elicitor while chitin decomposition

by-products also inhibit pathogen[50]. BRs are phytohormones indispensable for root growth and

BR biosynthesis is largely restricted to the root elongation zone[51]. DIBOA-Glc is the glycosylation

product of DIBOA, the most abundant BX in cereal crops, and the precursor of defense-active

compound DIMBOA-Glc. BX glucosides are stored in vacuoles and released over time to the

rhizosphere where their derivatives can suppress weeds and hostile microbes while enriching

beneficial microbes[52-54]. Our results indicate that biochar promotes plant growth by upregulating

root exudates that have antimicrobial and antioxidant activities. Further, BRs and BXs are

secondary metabolites known for their signaling functions, and increasing evidence reveals they

also mediate plant-microbe interactions such that shifts in their biosynthesis can reshape the

rhizosphere microbiome [52, 54, 55]. Here, by upregulating biosynthesis of plant signaling metabolites,

biochar may modulate the rhizosphere microbiome, which indeed occurred as discussed later.

Thirty-six root exudates were downregulated, including six flavonoids. Flavonoids are essential in

plant fitness and rhizosphere microbiome assembly[56]. Downregulation of flavonoid biosynthesis

could reflect either reduced stress due to scavenging of reactive oxygen species (ROS) or BR-

induced inhibition of flavonoid biosynthesis[57]. It may also reflect shifted flavonoid biosynthesis as

several flavonoids, including 2 flavanonols (dihydrokaempferol, (+)-2,3-dihydrofisetin), one

flavanone and its direct precursor (eriodictyol and eriodictyol chalcone), and a phenolic ester (4-

coumaroylshikimate) were only detected in biochar-treated root exudates. Overall, our results

provide chemical insights into widely observed biochar benefits on plant productivity[58].

A majority (81.4%) of the 312 exudate compounds were previously detected inside of root

tissues (**Fig S7**; also see the SI text)[10]. But we also identified 58 exudate compounds that were

not detected within root tissues, including 8 fatty acyls, 6 organooxygen compounds, 5 pyrimidine

nucleotides, 5 purine nucleotides, 4 carboxylic acids and derivatives, 3 flavonoids, 1 isoflavonoid,

3 prenol lipids, among others, such as DIBOA-Glc and gibberellin A29-catabolite (GA29-catabolite,

a deactivated catabolite from GA20) (**Fig S7; Table S4**). These 58 actively exudated compounds

likely have critical roles in the rhizosphere: Fatty acyls and prenol lipids, organooxygen

compounds, carboxylic acids and derivatives, and flavonoids and isoflavonoids are involved in

rhizosphere signaling and able to mediate plant root-microbe interactions[56, 59, 60]. Among the

detected fatty acyls, linolenic acids are plant defense signaling molecules[61]; 2'-deoxymugineic

acid is an iron (III) phytosiderophore, influencing rhizosphere microbiome composition through

iron competition[9, 62]. Five of the six detected organooxygen compounds were inositol phosphates,

a major pool of organic phosphorus in soil and signaling molecules mediating microbial root

colonization[63]. These results, along with our prior pot study[10], unveil biochar-induced metabolic

alterations in roots and subsequent exudation of molecules with profound modulating effects on

plant root-microbe interactions.

3. Biochar-induced root exudates drive the rhizosphere microbiome restructuring

3.1. Biochar rewires the rhizosphere microbiome and microbial interactions

In the native wheat rhizosphere, the top 10 phyla and top 40 genera comprised 60% and

35% of all prokaryotes, respectively, reflecting the core rhizosphere microbiome of plant,

especially cereals[9]. Biochar had limited influence on rhizosphere microbial alpha diversity but

considerably altered microbiome composition (PERMANOVA test: p = 0.001 for round × treatment)

(**Fig 3A**), promoting 12 genera while suppressing 9 genera (**Fig S8-S9; Table S5**) (more details

in the SI).

Remarkably, biochar enhanced microbial interactions, evidenced by greater microbial

network modularity and centrality, and a large number of new keystone genera emerged (60 vs.

53 in the treated vs. native rhizosphere) (**Fig 3C; Fig S10; Table S6-S8**). Twelve common

keystone bacterial genera were observed in the native and treated rhizosphere (detailed in the

SI), including nine PGPR from Actinobacteria, Alphaproteobacteria and Gammaproteobacteria:

Bradyrhizobium and *Microvirga* are N-fixing bacteria[64]; *Paracoccus* is capable of metabolizing

nitrogen, nitrate, and ammonia[65]; *Lysobacter* and *Stenotrophomonas* can suppress soil-borne

pathogens by producing lytic or antimicrobial substances[66, 67]; *Shinella* can degrade polycyclic

aromatic hydrocarbons[68]; *Cellvibrio* has an extensive repertoire of carbohydrate-active enzymes

for degrading polysaccharides (cellulose, dextran, xylan, chitin, and starch), thus enhancing

organic matter turnover[69, 70]; *Rhodococcus* contributes to nutrient cycling[71] and influences plant

stress response by metabolizing the phytohormone abscisic acid (ABA)[72]; *Solirubrobacter* can

produce the phytohormone indole-3-acetic acid (IAA) and improve nutrient bioavailability through

assimilatory reduction of sulphate and utilizing a wide range of carbohydrates[73] (more in the SI).

Besides these common keystone genera, the native rhizosphere microbiome had classic

rhizosphere keystone genera, such as those able to degrade plant cell wall polysaccharides

(details in the SI), whereas the biochar-treated rhizosphere had additional diverse PGPR as

keystone genera (**Fig 3C**). One example was *Chryseobacterium* (with greatest betweenness and

closeness centrality) that can enhance plant nutrient acquisition, degrade aromatic compounds,

produce antioxidant enzymes, synthesize IAA, and lower plant stress levels by reducing the

immediate precursor of the stress phytohormone ethylene using 1-aminocyclopropane-1-

carboxylate   (ACC)   deaminase[74].   Another   example   was   methylotrophic

Methylobacterium/*Methylorubrum* (connector hub) that can synthesize developmental

phytohormones (auxins, gibberellins, cytokinins) and ACC deaminase and fix nitrogen[75]. Other

keystone PGPR genera in the biochar-treated rhizosphere included N-fixers *Allorhizobium-*

Neorhizobium-Pararhizobium-Rhizobium and *Azospirillaceae*, denitrifiers *Magnetospirillum*[76, 77],

Nocardia able to produce IAA[78, 79], Planctomycetes *OM190* (module hub) able to produce diverse

secondary metabolites including antimicrobial compounds[80], and *Solibacillus* that can produce

IAA and plant singling molecule hydrogen cyanide and secrete siderophores to solubilize or

chelate minerals[81, 82].

These results reveal a new microbial mechanism for the widely-reported but poorly-

understood plant-promoting effect of biochar: Biochar elicits diverse PGPR, including nitrogen

fixers, denitrifiers, and phytohormone producers, to emerge as microbiome keystone taxa.

Phytohormones like auxins, gibberellins, and cytokinins regulate plant growth and development

(e.g., cell division and differentiation, root branching), immunity, and abiotic/biotic stress

responses[83, 84]; therefore, the biochar-treated rhizosphere microbiome may grant the plant with

multiple benefits. Indeed, while we did not track phytohormone dynamics, we observed higher

shoot mass and more lateral roots of biochar-treated plants (**Fig S4**), a phenomenon likely

attributed to the enrichment of phytohormone-producing PGPR and the aforementioned

upregulation of plant BRs biosynthesis. Given the significance of keystone taxa in overall

microbiome structure and functioning[85], future research should directly measure microbial

processing of phytohormones in the biochar-treated rhizosphere.

3.2. Rhizosphere microbiome rewiring is linked to biochar-induced root exudates

Plants use root exudates to recruit and tune the rhizosphere microbiome[9]. Since biochar

induced differential exudation, particularly polyphenols and lipids, we hypothesized that shifts in

root exudate chemistry restructured the rhizosphere microbiome, selectively enriching or

suppressing specific bacterial taxa, through metabolic coupling of biochar-induced root exudates

with microbial substrate-use preferences as suggested by recent evidence from laboratory

experiments and genome-informed trait-based modeling[86-88]. In line with this hypothesis, the

biochar-rewired rhizosphere microbial community had a slower growth rate than the native

microbiome (**Fig 3D**), suggesting a switch from higher growth and lower substrate utilization

efficiency to lower growth and higher substrate utilization efficiency[86, 87]. To validate this genome-

informed, community-level prediction, we tested biofilm formation of a model soil bacterium *P.*

putida KT2440 under exposure to normal or biochar-induced root exudates (**Fig S11A**). While

root exudates from control and biochar-treated plants had similar levels of DOC (**Fig S11C**) (*p* =

0.607), biochar-induced root exudates resulted in less *P. putida* biofilm within 48 hours (**Fig S11B**)

(*p* = 0.004), consistent with our hypothesis that biochar-induced root exudates favor slow-growing

microbes capable of utilizing complex molecules. It should be noted that this shift did not

significantly change the total bacteria number. Instead, biochar insignificantly increased the

number of rhizosphere bacteria as shown by qPCR results (**Fig S12**). Analysis of community

assembly mechanisms found that dispersal limitation dominated rhizosphere community turnover

(63.11% of all assembly mechanisms) (**Fig S13A-B**), suggesting low dispersal rates in the

rhizosphere niche where weak or unstructured selective pressures from substrate utilization

enable high compositional turnover[89]. Together, computational prediction, biofilm assay, and

community assembly analysis support our hypothesis that biochar rewired the rhizosphere

microbiome through metabolic coupling of biochar-induced root exudates with microbial

substrate-use preferences.

Our hypothesis is further supported by the coincidence of biochar-enriched root exudate

composition, including chitin and DIBOA-Glc, and the enrichment of bacteria having chitinolytic

activity and BXs-enhanced growth (detailed in the SI). This motivated us to explore specific

microbe-exudate compound associations. Combining the co-occurrence network (OmicsNet 2.0)

and neural network (mmvec) approaches, we identified 12 bacterial genera, including 11 PGPR

(*Bacillus*,   *Blastococcus*,   *Bradyrhizobium*,   *Brevundimonas*,   *Mesorhizobium*,   *Nevskia*,

Paenibacillus, *Pantoea*, *Pseudomonas*, *Rhodococcus*, *Roseomonas*)[90, 91], with highest

associations with biochar-induced root exudates (mainly plant secondary metabolites) (**Fig 4; Fig**

S13C). Extremely high co-occurrence probability may manifest when a compound is a preferred

substrate to certain bacteria and multiple lines of evidence support the metabolic coupling nature

of the observed microbe-exudate compound associations[54, 86-88].

Four flavonols, including three quercetin derivatives and one kaempferol derivative, had

varying co-occurrence with the 12 bacterial genera: Isoquercitrin (quercetin 3-O-glucoside) had

very high co-occurrence with all the 12 genera, particularly *Pantoea*; quercetin 3-O-rhamnoside-

7-O-glucoside and kaempferol-3-O-rhamnoside had similar co-occurrence patterns with several

microbes, opposite to quercitrin (quercetin 3-O-rhamnoside). Quercetin and its derivatives can

stimulate PGPR and bacteria like *Pantoea* harbor quercetin dioxygenases for degrading

flavonols[56, 92]. This aligns with our observation of very high co-occurrence probabilities between

Pantoea and two quercetin derivatives (quercitrin > isoquercitrin). Three gibberellins (GAs),

GA12-aldehyde (early precursor of bioactive GAs), GA44 open lactone (key intermediate in

bioactive GA1 synthesis), and GA34-catabolite (deactivated catabolite from bioactive GA4), had

varying co-occurrence with the 12 bacterial genera, with GA44 open lactone and GA34-catabolite

showing opposite patterns. GA phytohormones are crucial for many aspects of plant growth and

development. Appropriate levels of bioactive GA are tightly regulated in plants and also modulated

by PGPR such as *Bacillus*, *Bradyrhizobium*, and *Mesorhizobium* that possess GA metabolic

pathways[93-95]. Here *Bacillus* co-occurred with GA34-catabolite, while *Bradyrhizobium* and

Mesorhizobium had very high co-occurrence probability with GA44 open lactone and GA34-

catabolite, respectively, suggesting their involvement in GA metabolism. (S)-Abscisic acid, the

naturally occurring and bioactive form of phytohormone ABA, and its direct catabolite 8'-

hydroxyabscisic acid had high co-occurrence with *Blastococcus*, *Bradyrhizobium*, and

Roseomonas, suggesting ABA metabolism by these bacteria. Indeed, PGPR like *Bradyrhizobium*

can regulate ABA levels in the rhizosphere through biosynthesis and catabolism of ABA [52, 94, 96, 97].

Two glycosides of phytohormone cytokinins, cis-zeatin-O-glucoside and dihydrozeatin-O-

glucoside, had opposite patterns, with cis-zeatin-O-glucoside showing very high co-occurrence

with several PGPR (*Mesorhizobium* > *Rhodococcus* > *Pseudomonas* > *Bacillus*). It is known that

Rhodococcus and *Pseudomonas* can produce cytokinins[94, 98, 99].

Taken together, our results reveal biochar-induced root exudates as a key driver reshaping

the rhizosphere microbiome, likely through metabolic coupling of root exudates with microbial

substrate-use preferences. In particular, secondary metabolites such as flavonoids (flavonols)

and developmental phytohormones (GAs, ABA, cytokinins) evoke diverse PGPR capable of

utilizing these secondary metabolites. Our study thus provides a new chemical mechanism for

the widely-reported but poorly-understood plant-promoting effect of biochar. Further research

employing metagenomics and metatranscriptomics is required to ascertain our observations and

elaborate underlying metabolic mechanisms.

4. Biochar-treated rhizosphere microbiome has shifted functions

4.1. Shifted N and methane cycling

Many studies have reported biochar's impact on bulk soil N and methane cycling[2, 3]. We

speculate such effects, likely due to changes in microbial functional composition and metabolic

pathway, can also occur in the rhizosphere. qPCR measurements found increased abundance of

nifH (N fixation), *nxrA* (nitrification, NO_2^- to NO_3^-), *nirS* (denitrification, NO_2^- to NO) and *nosZ*

(denitrification, N_2O to N_2), and reduced abundance of *nirK* (denitrification, NO_2^- to NO) in biochar-

treated rhizosphere (*p*-adjusted > 0.05) (**Fig 5B**). Increases in N fixation and nitrification genes

due to biochar treatment were also detected by community-level function prediction using

PICRUSt2 (**Fig S14**) (detailed in the SI). These trends are consistent with the emergence of N-

fixing bacteria (e.g., *Allorhizobium-Neorhizobium-Pararhizobium-Rhizobium*, *Azospirillaceae*) as

keystone taxa and the enrichment of ammonia-oxidizing bacteria (e.g., *Nitrosomonas*) in the

biochar-treated rhizosphere (**Fig 3C**; **Table S5**). Interestingly, biochar-induced increase in *nxrA*

gene coincided with a drastic enrichment of *Roseomonas*, a genus known to harbor *nxrA*[100] and

showing high co-occurrence with biochar-induced-phytohormone ABA (a terpenoid) and its direct

catabolite 8'-hydroxyabscisic acid (**Fig 4**). In complete denitrification, NO_2^- is transformed to NO

by nitrite reductases (*nirK* and *nirS*), while NO is further reduced to N_2O by nitric oxide reductases

(*nor*) and N_2O reduced to N_2 by nitrous oxide reductase (*nosZ*)[101]. Changes in the relative

abundance of these genes are associated with the production and consumption of N_2O and thus

the emission of this strong GHG. Specifically, the (*nirK*+*nirS*)/*nosZ* gene ratio has been used to

estimate N_2O emission potential[102]. Our qPCR results indicate a reduced (*nirK*+*nirS*)/*nosZ* ratio

in biochar-treated rhizosphere, indicating a promotion of complete denitrification and a reduction

of net N_2O emissions, likely due to biochar's role in facilitating electron transfer to denitrifying

microbes[103]. qPCR results also showed a reduction in the (*nirK*/*nirS*) ratio, suggesting a shift from

nirK-type nitrifiers, which often lack the *nosZ* gene, to *nirS*-type nitrifiers[104]. This observation is

consistent with the emergence of *Magnetospirillum* as keystone taxa and the enrichment of

Sporomusa in the biochar-treated rhizosphere (**Fig 3C**; **Table S5**). *Magnetospirillum* harbors a

complete denitrification pathway, able to reduce NO_3^- to N_2[105]. It should be noted that *Sporomusa*

species contain a recently-characterized novel lactonase-type N_2O reductase that cannot be

captured by existing primers[106]. Overall, these qPCR results align with prior observations from the

bulk soil that biochar promotes nitrogen fixation and nitrification while mitigating N_2O emissions[2],

and unveil microbiome restructuring as the underlying mechanism.

Neither qPCR or PICRUSt2 detected *mcrA* encoding the methyl-coenzyme M reductase

complex in any sample, reflecting the low representation of methanogens in the ectorhizosphere;

meanwhile, function prediction suggested that biochar suppressed coenzyme M biosynthesis I

required for methanogenesis (**Fig 5A**)[107]. Both qPCR and PICRUSt2 found increasing *pmoA*

(encoding membrane-bound particulate methane monooxygenase) in the biochar-treated

rhizosphere, indicating stimulated methane oxidation (**Fig 5B; Fig S15**)[108]. Consistently, two

exometabolites, isopentenyl pyrophosphate and dimethylallylpyrophosphate, were only present

in the biochar-treated rhizosphere (**Fig 5C; Table S12**). These two prenol lipids are five-carbon

building blocks for all isoprenoids, including bacteriohopanepolyols that are crucial structural

components of methanotroph membranes. Further, PICRUSt2 results showed biochar enriched

the methionine salvage pathway by >1000 fold (**Fig 5A**). In a sulfur-adequate environment like

here (**Table S1**), stimulated methionine salvage pulls 5-methylthioribulose-1-phosphate (MTRu-

1-P) towards L-methionine biosynthesis, limiting bioavailable methanethiol for archaeal

methanogenesis. Indeed, MTRu-1-P and its upstream precursor S-adenosylmethioninamine were

detected in root exudates, while MTRu-1-P downstream product 3-methylthiopropionic acid and

S-adenosylmethioninamine were detected in the soil metabolome (**Table S3; Table S12**). In

biochar-treated rhizosphere, microbial cysteine and methionine metabolism was significantly

associated with these root exudates and soil exometabolites (*p-adjusted* < 0.05) (**Table S9; Table**

S13), and four PGPR (*Roseomonas* > *Nevskia* > *Blastococcus* > *Bacillus*) had high co-occurrence

with 3-methylthiopropionic acid (**Fig S18**). Collectively, our results from qPCR, function prediction,

and untargeted metabolomics suggest that in the rhizosphere, biochar inhibits methane emission

through orchestrated microbiome functional shifts that are more complicated than previously

understood: stimulated methionine salvage that deprives the methanogenesis precursor

methanethiol; suppressed biosynthesis of coenzyme M, a cofactor required for the methane-

forming step; and enhanced methane oxidation.

Collectively, our results reveal new mechanisms of biochar's inhibition of two strong GHGs

in the rhizosphere, indicate the pivotal role of biochar-induced changes in root exudate chemistry

in restructuring the rhizosphere microbiome and its functionality, and identify PGPR genera whose

response to biochar-induced root exudates likely contributes to the observed shifts in N and

methane cycling.

4.2. Broader changes in rhizosphere microbiome functions

A total of 506 exometabolites were detected in rhizosphere soils, including 180

compounds detected in root exudates collected by hydroponic secretion and 326 compounds only

found in the soil metabolome (**Fig S16; Table S12**). The 326 exometabolites consisted of

compounds secreted by roots in contact with rhizosphere microbes and microbial metabolites.

For example, DIBOA-Glc was only present in hydroponic secretion (**Fig 2B**), while its active form

DIBOA was only present in soil exometabolites and had high co-occurrence probability with 7

PGPR (**Fig S18**), likely reflecting BXs being used as recruitment signals for specific beneficial

microbes[52-54]. Various GAs were present in both root exudates and soil metabolites, but an early

precursor of bioactive GAs (GA12-aldehyde) was only present in root exudates, whereas two GA

catabolites (GA8, GA8-catabolite) and one intermediate (GA24) were only present in the soil

metabolome (**Table S12**), with GA24 showing high co-occurrence probability with numerous

PGPR, especially *Bradyrhizobium* known to possess GA metabolic pathways (**Fig S18**)[93-95].

Biochar significantly affected the soil metabolome chemistry (PERMANOVA: $p < 0.05$ for

treatment) (**Fig S17**), reflected in 64 most discriminative compounds ($|Log_2FC| > 1$, *p-adjusted* <

0.05) from 14 ChemOnt classes (**Fig 5C**) (detailed in the SI). Remarkably, several plant defense

compounds, including lotaustralin and methylthioalkylmalates (derived from methionine and

important precursors of sulfur-containing glucosinolates), were only present in biochar-treated

rhizosphere (**Fig 5C; Table S12**). Further, IAA and an indole derivative (5-

hydroxyindoleacetaldehyde) were present in the control rhizosphere but absent in biochar-treated

rhizosphere where 5-hydroxyindoleacetic acid, a direct oxidation product of 5-

hydroxyindoleacetaldehyde was detected, suggesting microbial metabolism of IAA and consistent

with aforementioned enrichment of IAA-producing PGPR (**Fig 5C; Table S3; Table S12**).

After mapping soil exometabolites to predicted microbial pathways (**Table S13**) and

comparing root exudate-microbe co-occurrence with microbe-exometabolite co-occurrence

(**Table S9**), we found biochar likely promoted microbial synthesis of phylloquinone and

menaquinol (more in the SI). Specifically, the menaquinol percussor 2-succinyl-5-enolpyruvyl-6-

hydroxy-3-cyclohexene-1-carboxylic acid was only present in root exudates, and its immediate

metabolites (1R,6R)-6-hydroxy-2-succinylcyclohexa-2,4-diene-1-carboxylate and 2-

succinylbenzoate were only present in the soil metabolome (**Table S3; Table S12**). Biochar

significantly reduced these two intermediates of menaquinone biosynthesis (**Fig 5C**).

Coincidentally, according to function prediction, biochar increased all but one menaquinone

biosynthesis gene (**Fig S19**). Menaquinol, the bioactive reduced form of menaquinone, was

detected in the soil metabolome, which can serve as an extracellular electron donor in the

rhizosphere environment. This may explain biochar-promoted complete denitrification observed

in this study and biochar-facilitated electron transfer as proposed by others[103].

Overall, these results from integrating multi-omics data indicate biochar's profound

modulating effects on the rhizosphere microbiome functionality, ranging from phytohormone

transformation attributed to biochar-enriched PGPR to biosynthesis of redox-active quinone

compounds. These changes have significant implications for plant development and soil

biogeochemistry. Future research should employ metagenomics and metatranscriptomics to

validate our observations and elaborate underlying microbiology mechanisms.

5. Environmental Implications and Significance

The research community is striving to develop sustainable technologies to support crop production while reducing GHG emissions and reliance on agrochemicals. The rhizosphere microbiome plays an essential role in plant growth and general soil biogeochemistry, and rhizosphere engineering has attracted increasing interest as a promising approach for sustainable agriculture. While biochar is a promising multifunctional soil conditioner, its effect on the rhizosphere is unknown. This study is among the first that elaborates biochar-induced root exudates as a key driver that rewires the rhizosphere microbiome and its functionality. By integrating multi-omics, we reveal new microbial and chemical mechanisms behind the widely reported but poorly understood biochar effects on the soil-plant system. Complex molecules in biochar-induced root exudates, including a wide range of plant secondary metabolites and phytohormones, evoke a plant-beneficial rhizosphere microbiome centered by diverse PGPR. The restructured microbiome is enhanced in N fixation, complete denitrification, methionine salvage, and methane oxidation, thus suppressing N_2O and CH_4 emissions. The restructured microbiome also contributes to the transformation of phytohormones and biosynthesis of redox-active quinone compounds, with critical implications for the soil-plant system. Future research should apply integrated metagenomics/metatranscriptomics, leverage larger-scale and field experiments, and investigate other plant models to validate and extend our findings. Future research should also directly demonstrate biochemical connections between the root exudate compounds and PGPR identified in this study. Nevertheless, this research provides new insights into biochar's multifaceted roles in reprogramming the rhizosphere and highlights the promise of engineering the rhizosphere through reshaping root-microbe interactions using designed biochar or root exudate cocktails. Further, given the fact that fire-derived black carbon shares similar chemistry with biochar and that increased frequency of wildfire could cause more black carbon deposition, this research spotlights the overlooked impact of fire-derived black carbon on belowground carbon and nitrogen dynamics through intricate rhizosphere interactions, which itself

requires in-depth investigation.

SUPPORTING INFORMATION

Detailed material and methods, additional results and discussion, figures (S1-S19), and

tables (S1-S13) are shown in SUPPORTING INFORMATION (PDF; XLSX).

ACKNOWLEDGEMENT

This work was supported by the State University of New York College of Environmental

Science and Forestry (SUNY ESF) through startup funds to Y. You. It was partially supported by

the   US   Department   of   Agriculture   through   a   McIntireStennis   Capacity   Grant

(NI23MSCFRXXXG070). H. Yang is grateful to the ESF Alumni Association. We would like to

posthumously acknowledge Kevin Guerin (ESF Analytical and Technical Services) for great

assistance with fabricating acrylic mold parts on a lathe machine. We thank Dr. Alex Artyukhin for

training H. Yang on the Q Exactive HF Orbitrap mass spectrometer. We thank Olivia Abry and

Hayden Bregg (Department of Biology, ESF) for assistance with the EcoFAB experiment, and Dr.

Leanne Powers (Department of Chemistry, ESF) for assistance with DOC measurement. We are

grateful to Syracuse University Research Computing for providing high-performance computing

resources.

Fig 1. Physicochemical properties of the wheat biochar used in this study. (A) An representative

SEM image at 100´ magnification. White scale bar indicates 100 mm. (B) FTIR spectrum in the

range 400-2000 cm^{-1} with annotations of characteristic absorbance peaks. Full spectrum (400-

4000 cm^{-1}) is in **Fig S3**. (C) EDS-derived elemental distribution on biochar surface. Pie chart

indicates atomic percentages averaged from 29 subsamples. (D) XRD spectrum with

characteristics diffraction angles highlighted.

585 **Fig 2.** (A) Seedlings grown in EcoFAB devices. (B) Thirty-nine most discriminative root exudates

586 in biochar-treated wheat roots (|Log$_2$FC| > 1, *p-adjusted* < 0.05). Individual compounds are

587 grouped into ChemOnt classes.

Fig 3. (A) PCoA plot of the native and biochar-treated rhizosphere microbiomes (experiment round and treatment along the first and second coordinate, respectively) (PERMANOVA test: $p <$ 0.05 for round x treatment). (B) Relative abundance of the top 10 microbial phyla. (C) Hub networks of keystone genera in the native and biochar-treated rhizosphere ($p < 0.05$, 1000 bootstrapping). Node colors and sizes indicate phyla and betweenness values, respectively. Purple and blue edges indicate positive and negative correlations. (D) Biochar-rewired rhizosphere microbiome had slightly slower growth rate (i.e., larger median predicted minimum doubling time (mPMDT)) than the native rhizosphere microbiome. Prediction was made using gRodon based on enriched taxa identified using DESeq2 ($|Log_2FC| > 1$, *p-adjusted* < 0.05).

Fig 4. Associations between exudate compounds (column) and bacterial genera (row) inferred by the mmvec neural network. Colors indicate the estimated log conditional probability (i.e., association strength) of observing a root exudate compound given the presence of a microbial genus. Only those genera identified by both the OmicsNet 2.0 and mmvec approaches are shown here to ensure true positivity. Root exudates are clustered based on their associations on microbial genera.

Fig 5. (A) Microbial pathways significantly regulated by biochar ($|Log_2FC| > 1$, *p-adjusted* < 0.05).

Pathways were inferred using PICRUSt2. (B) qPCR-measured abundance of key nitrogen- and

methane-cycling genes. Boxplots represent the median and interquartile range (IQR) of gene

abundance between groups. (C) Rhizosphere soil exometabolites were significantly influenced

by biochar ($|Log_2FC| > 1$, *p-adjusted* < 0.05). Individual compounds are grouped into ChemOnt

classes.

REFERENCES

(1) Matson, P. A.; Parton, W. J.; Power, A. G.; Swift, M. J. Agricultural intensification and ecosystem properties. *Science* **1997**, *277* (5325), 504-509.

(2) Kerner, P.; Struhs, E.; Mirkouei, A.; Aho, K.; Lohse, K. A.; Dungan, R. S.; You, Y. Q. Microbial responses to biochar soil amendment and influential factors: a three-Level meta-analysis. *Environ. Sci. Technol.* **2023**, *57* (48), 19838-19848.

(3) Das, S. K.; Ghosh, G. K.; Avasthe, R. Biochar application for environmental management and toxic pollutant remediation. *Biomass Convers. Biorefin.* **2020**, 555-566.

(4) Hersh, B.; Mirkouei, A.; Sessions, J.; Rezaie, B.; You, Y. A review and future directions on enhancing sustainability benefits across food-energy-water systems: the potential role of biochar-derived products. *AIMS Environ. Sci.* **2019**, *6* (5), 379-416.

(5) Chen, L. J.; Jiang, Y. J.; Liang, C.; Luo, Y.; Xu, Q. S.; Han, C.; Zhao, Q. G.; Sun, B. Competitive interaction with keystone taxa induced negative priming under biochar amendments. *Microbiome* **2019**, *7*, 77.

(6) Vandenkoornhuyse, P.; Quaiser, A.; Duhamel, M.; Le Van, A.; Dufresne, A. The importance of the microbiome of the plant holobiont. *New Phytol.* **2015**, *206* (4), 1196-1120.

(7) Brtnicky, M.; Datta, R.; Holatko, J.; Bielska, L.; Gusiatin, Z. M.; Kucerik, J.; Hammerschmiedt, T.; Danish, S.; Radziemska, M.; Mravcova, L.; et al. A critical review of the possible adverse effects of biochar in the soil environment. *Sci. Total Environ.* **2021**, *796* (18), 148756.

(8) Mahmud, K.; Missaoui, A.; Lee, K. D.; Ghimire, B.; Presley, H. W.; Makaju, S. Rhizosphere microbiome manipulation for sustainable crop production. *Curr. Plant Biol.* **2021**, *27* (17), 100210.

(9) Trivedi, P.; Leach, J. E.; Tringe, S. G.; Sa, T.; Singh, B. K. Plant–microbiome interactions: from community assembly to plant health. *Nat. Rev. Microbiol.* **2020**, *18* (11), 607-621.

(10) Yang, H.; Kerner, P.; Liang, X.; Struhs, E.; Mirkouei, A.; You, Y. Biochar amendment modulates root metabolome and rhizosphere microbiome of wheat. *Biochar* **2025**, *7* (1), 46.

(11) Liu, Y.; Zhang, H.; Wang, J.; Gao, W.; Sun, X.; Xiong, Q.; Shu, X.; Miao, Y.; Shen, Q.; Xun, W. Nonpathogenic Pseudomonas syringae derivatives and its metabolites trigger the plant "cry for help" response to assemble disease suppressing and growth promoting rhizomicrobiome. *Nat. Commun.* **2024**, *15* (1), 1907.

(12) Luo, J.; Gu, S.; Guo, X.; Liu, Y.; Tao, Q.; Zhao, H.-P.; Liang, Y.; Banerjee, S.; Li, T. Core microbiota in the rhizosphere of heavy metal accumulators and its contribution to plant performance. *Environ. Sci. Technol.* **2022**, *56* (18), 12975-12987.

(13) Klimasmith, I. M.; Kent, A. D. Micromanaging the nitrogen cycle in agroecosystems. *Trends Microbiol.* **2022**, *30* (11), 1045-1055.

(14) Xiao, Z.; Rasmann, S.; Yue, L.; Lian, F.; Zou, H.; Wang, Z. The effect of biochar amendment on N-cycling genes in soils: A meta-analysis. *Sci. Total Environ.* **2019**, *696*, 133984.

(15) Liu, X.; Shi, Y.; Zhang, Q.; Li, G. Effects of biochar on nitrification and denitrification-mediated N_2O emissions and the associated microbial community in an agricultural soil. *Environ. Sci. Pollut. Res.* **2021**, *28*, 6649-6663.

(16) Lyu, H.; Zhang, H.; Chu, M. W.; Zhang, C. F.; Tang, J. C.; Chang, S. X.; Masek, O.; Ok, Y. S. Biochar affects greenhouse gas emissions in various environments: A critical review. *Land*

Degradation & Development **2022**, *33* (17), 3327-3342.

(17) Coskun, D.; Britto, D. T.; Shi, W. M.; Kronzucker, H. J. How plant root exudates shape the
nitrogen cycle. *Trends Plant Sci.* **2017**, *22* (8), 661-673.

(18) Lei, X.; Shen, Y. T.; Zhao, J. N.; Huang, J. J.; Wang, H.; Yu, Y.; Xiao, C. W. Root exudates
mediate the processes of soil organic carbon input and efflux. *Plants-Basel* **2023**, *12* (3), 630.

(19) Sasse, J.; Kant, J.; Cole, B. J.; Klein, A. P.; Arsova, B.; Schlaepfer, P.; Gao, J.; Lewald, K.;
Zhalnina, K.; Kosina, S. Multilab EcoFAB study shows highly reproducible physiology and
depletion of soil metabolites by a model grass. *New Phytol.* **2019**, *222* (2), 1149-1160.

(20) Zengler, K.; Hofmockel, K.; Baliga, N. S.; Behie, S. W.; Bernstein, H. C.; Brown, J. B.; Dinneny,
J. R.; Floge, S. A.; Forry, S. P.; Hess, M.; et al. EcoFABs: advancing microbiome science through
standardized fabricated ecosystems. *Nat. Methods* **2019**, *16* (7), 567-571.

(21) Gao, J.; Sasse, J.; Lewald, K. M.; Zhalnina, K.; Cornmesser, L. T.; Duncombe, T. A.; Yoshikuni,
Y.; Vogel, J. P.; Firestone, M. K.; Northen, T. R. Ecosystem fabrication (EcoFAB) protocols for the
construction of laboratory ecosystems designed to study plant-microbe interactions. *JoVE*
(Journal of Visualized Experiments) **2018**, (134), e57170.

(22) You, Y.; Kerner, P.; Shanmugam, S.; Khodakovskaya, M. V. Emerging investigator series:
differential effects of carbon nanotubes and graphene on the tomato rhizosphere microbiome.
Environ. Sci.: Nano **2023**, *10* (6), 1570-1584.

(23) Yang, H.; Wang, C.; Chen, F.; Yue, L.; Cao, X.; Li, J.; Zhao, X.; Wu, F.; Wang, Z.; Xing, B.
Foliar carbon dot amendment modulates carbohydrate metabolism, rhizospheric properties and
drought tolerance in maize seedling. *Sci. Total Environ.* **2022**, *809*, 151105.

(24) Scheltema, R. A.; Hauschild, J.-P.; Lange, O.; Hornburg, D.; Denisov, E.; Damoc, E.; Kuehn,
A.; Makarov, A.; Mann, M. The Q Exactive HF, a Benchtop mass spectrometer with a pre-filter,
high-performance quadrupole and an ultra-high-field Orbitrap analyzer. *Mol. Cell. Proteomics*
2014, *13* (12), 3698-3708.

(25) Pang, Z.; Zhou, G.; Ewald, J.; Chang, L.; Hacariz, O.; Basu, N.; Xia, J. Using MetaboAnalyst
5.0 for LC–HRMS spectra processing, multi-omics integration and covariate adjustment of global
metabolomics data. *Nat. Protoc.* **2022**, *17* (8), 1735-1761.

(26) Schmid, R.; Heuckeroth, S.; Korf, A.; Smirnov, A.; Myers, O.; Dyrlund, T. S.; Bushuiev, R.;
Murray, K. J.; Hoffmann, N.; Lu, M.; et al. Integrative analysis of multimodal mass spectrometry
data in MZmine 3. *Nat. Biotechnol.* **2023**, *41* (4), 447-449.

(27) Tautenhahn, R.; Patti, G. J.; Rinehart, D.; Siuzdak, G. XCMS online: a web-based platform
to process untargeted metabolomic data. *Anal. Chem.* **2012**, *84* (11), 5035-5039.

(28) Djoumbou Feunang, Y.; Eisner, R.; Knox, C.; Chepelev, L.; Hastings, J.; Owen, G.; Fahy, E.;
Steinbeck, C.; Subramanian, S.; Bolton, E. ClassyFire: automated chemical classification with a
comprehensive, computable taxonomy. *J. Cheminf.* **2016**, *8* (1), 61.

(29) Oksanen, J.; Blanchet, F.; Kindt, R.; Legendre, P.; Minchin, P.; O'hara, R.; Simpson, G.;
Solymos, P.; Stevens, M.; Wagner, H. Package 'vegan'. Community ecology package, version 2.
The comprehensive R network (CRAN). **2013**.

(30) Langfelder, P.; Horvath, S. WGCNA: an R package for weighted correlation network analysis.
BMC Bioinformatics **2008**, *9*, 559.

(31) Csardi, G.; Nepusz, T. The igraph software package for complex network research. *InterJournal, complex systems* **2006**, *1695* (5), 1-9.

(32) Douglas, G. M.; Maffei, V. J.; Zaneveld, J. R.; Yurgel, S. N.; Brown, J. R.; Taylor, C. M.; Huttenhower, C.; Langille, M. G. PICRUSt2 for prediction of metagenome functions. *Nat. Biotechnol.* **2020**, *38* (6), 685-688.

(33) Yang, C.; Mai, J.; Cao, X.; Burberry, A.; Cominelli, F.; Zhang, L. ggpicrust2: an R package for PICRUSt2 predicted functional profile analysis and visualization. *Bioinformatics* **2023**, *39* (8), btad470.

(34) López, J. L.; Fourie, A.; Poppeliers, S. W. M.; Pappas, N.; Sánchez-Gil, J. J.; de Jonge, R.; Dutilh, B. E. Growth rate is a dominant factor predicting the rhizosphere effect. *Isme J.* **2023**, *17* (9), 1396-1405.

(35) Weissman, J. L.; Hou, S.; Fuhrman, J. A. Estimating maximal microbial growth rates from cultures, metagenomes, and single cells via codon usage patterns. *Proc. Natl. Acad. Sci. U.S.A.* **2021**, *118* (12), e2016810118.

(36) Friedman, J.; Alm, E. J. Inferring correlation networks from genomic survey data. **2012,** *PLoS Comput Biol* 8(9): e1002687.

(37) Ning, D.; Wang, Y.; Fan, Y.; Wang, J.; Van Nostrand, J. D.; Wu, L.; Zhang, P.; Curtis, D. J.; Tian, R.; Lui, L.; et al. Environmental stress mediates groundwater microbial community assembly. *Nat. Microbiol.* **2024**, *9* (2), 490-501.

(38) Stegen, J. C.; Lin, X. J.; Fredrickson, J. K.; Chen, X. Y.; Kennedy, D. W.; Murray, C. J.; Rockhold, M. L.; Konopka, A. Quantifying community assembly processes and identifying features that impose them. *Isme J.* **2013**, *7* (11), 2069-2079.

(39) Taylor, S. C.; Mrkusich, E. M. The state of RT-quantitative PCR: firsthand observations of implementation of minimum information for the publication of quantitative real-time PCR experiments (MIQE). *J. Mol. Microbiol. Biotechnol.* **2014**, *24* (1), 46-52.

(40) Burmølle, M.; Hansen, L. H.; Sørensen, S. J. Establishment and early succession of a multispecies biofilm composed of soil bacteria. *Microb. Ecol.* **2007**, *54*, 352-362.

(41) Morton, J. T.; Aksenov, A. A.; Nothias, L. F.; Foulds, J. R.; Quinn, R. A.; Badri, M. H.; Swenson, T. L.; Van Goethem, M. W.; Northen, T. R.; Vazquez-Baeza, Y.; et al. Learning representations of microbe–metabolite interactions. *Nat. Methods* **2019**, *16* (12), 1306-1314.

(42) Zhou, G.; Pang, Z.; Lu, Y.; Ewald, J.; Xia, J. OmicsNet 2.0: a web-based platform for multi-omics integration and network visual analytics. *Nucleic Acids Res.* **2022**, *50* (W1), W527-W533.

(43) Liu, L.; Liu, G.; Zhou, J.; Wang, J.; Jin, R.; Wang, A. Improved bioreduction of nitrobenzene by black carbon/biochar derived from crop residues. *RSC advances* **2016**, *6* (87), 84388-84396.

(44) Yao, S.; Li, X.; Cheng, H.; Zhang, C.; Bian, Y.; Jiang, X.; Song, Y. Resource utilization of a typical vegetable waste as biochars in removing phthalate acid esters from water: a sorption case study. *Bioresour. Technol.* **2019**, *293*, 122081.

(45) Zhou, Q.; Jiang, X.; Li, X.; Jia, C. Q.; Jiang, W. Preparation of high-yield N-doped biochar from nitrogen-containing phosphate and its effective adsorption for toluene. *RSC Advances* **2018**, *8* (53), 30171-30179.

(46) Waqas, M.; Aburiazaiza, A. S.; Miandad, R.; Rehan, M.; Barakat, M. A.; Nizami, A. S.

Development of biochar as fuel and catalyst in energy recovery technologies. *J. Clean. Prod.* **2018**, *188*, 477-488..

(47) Joseph, S.; Cowie, A. L.; Van Zwieten, L.; Bolan, N.; Budai, A.; Buss, W.; Cayuela, M. L.; Graber, E. R.; Ippolito, J. A.; Kuzyakov, Y. How biochar works, and when it doesn't: A review of mechanisms controlling soil and plant responses to biochar. *Gcb Bioenergy* **2021**, *13* (11), 1731-1764.

(48) Ali, I.; Ullah, S.; Iqbal, A.; Quan, Z.; Liang, H.; Ahmad, S.; Muhammad, I.; Amanullah; Imran; Guo, Z.; et al. Combined application of biochar and nitrogen fertilizer promotes the activity of starch metabolism enzymes and the expression of related genes in rice in a dual cropping system. *BMC Plant Biol.* **2021**, *21* (1), 600.

(49) Kolton, M.; Graber, E. R.; Tsehansky, L.; Elad, Y.; Cytryn, E. Biochar-stimulated plant performance is strongly linked to microbial diversity and metabolic potential in the rhizosphere. *New Phytol.* **2017**, *213* (3), 1393-1404.

(50) Shamshina, J. L.; Kelly, A.; Oldham, T.; Rogers, R. D. Agricultural uses of chitin polymers. *Environ. Chem. Lett.* **2020**, *18* (1), 53-60.

(51) Vukašinović, N.; Wang, Y.; Vanhoutte, I.; Fendrych, M.; Guo, B.; Kvasnica, M.; Jiroutová, P.; Oklestkova, J.; Strnad, M.; Russinova, E. Local brassinosteroid biosynthesis enables optimal root growth. *Nat. Plants* **2021**, *7* (5), 619-632.

(52) Yu, J.; Tu, X.; Huang, A. C. Functions and biosynthesis of plant signaling metabolites mediating plant–microbe interactions. *Nat. Prod. Rep.* **2022**, *39* (7), 1393-1422.

(53) Cotton, T. A.; Pétriacq, P.; Cameron, D. D.; Meselmani, M. A.; Schwarzenbacher, R.; Rolfe, S. A.; Ton, J. Metabolic regulation of the maize rhizobiome by benzoxazinoids. *ISME J.* **2019**, *13* (7), 1647-1658.

(54) Hu, L.; Robert, C. A.; Cadot, S.; Zhang, X.; Ye, M.; Li, B.; Manzo, D.; Chervet, N.; Steinger, T.; Van Der Heijden, M. G. Root exudate metabolites drive plant-soil feedbacks on growth and defense by shaping the rhizosphere microbiota. *Nat. Commun.* **2018**, *9* (1), 2738.

(55) Kudjordjie, E. N.; Sapkota, R.; Steffensen, S. K.; Fomsgaard, I. S.; Nicolaisen, M. Maize synthesized benzoxazinoids affect the host associated microbiome. *Microbiome* **2019**, *7* (1), 59.

(56) Wang, L.; Chen, M.; Lam, P.-Y.; Dini-Andreote, F.; Dai, L.; Wei, Z. Multifaceted roles of flavonoids mediating plant-microbe interactions. *Microbiome* **2022**, *10* (1), 233.

(57) Liang, T.; Shi, C.; Peng, Y.; Tan, H.; Xin, P.; Yang, Y.; Wang, F.; Li, X.; Chu, J.; Huang, J. Brassinosteroid-activated BRI1-EMS-SUPPRESSOR 1 inhibits flavonoid biosynthesis and coordinates growth and UV-B stress responses in plants. *Plant cell* **2020**, *32* (10), 3224-3239.

(58) Biederman, L. A.; Harpole, W. S. Biochar and its effects on plant productivity and nutrient cycling: a meta-analysis. *GCB bioenergy* **2013**, *5* (2), 202-214.

(59) Li, M.; Song, Z.; Li, Z.; Qiao, R.; Zhang, P.; Ding, C.; Xie, J.; Chen, Y.; Guo, H. Populus root exudates are associated with rhizosphere microbial communities and symbiotic patterns. *Front. Microbiol.* **2022**, *13*, 1042944.

(60) Macabuhay, A.; Arsova, B.; Walker, R.; Johnson, A.; Watt, M.; Roessner, U. Modulators or facilitators? Roles of lipids in plant root–microbe interactions. *Trends Plant Sci.* **2022**, *27* (2), 180-190.

(61) He, M.; Ding, N. Z. Plant unsaturated fatty acids: multiple roles in stress response. *Front.*
Plant Sci. **2020**, *11*. 562785.
(62) Gu, S.; Wei, Z.; Shao, Z.; Friman, V.-P.; Cao, K.; Yang, T.; Kramer, J.; Wang, X.; Li, M.; Mei,
X. Competition for iron drives phytopathogen control by natural rhizosphere microbiomes. *Nature*
Microbiology **2020**, *5* (8), 1002-1010.
(63) O'Banion, B. S.; Jones, P.; Demetros, A. A.; Kelley, B. R.; Knoor, L. H.; Wagner, A. S.; Chen,
J.-G.; Muchero, W.; Reynolds, T. B.; Jacobson, D. Plant myo-inositol transport influences bacterial
colonization phenotypes. *Curr. Biol.* **2023**, *33* (15), 3111-3124.
(64) Missbah El Idrissi, M.; Bouhnik, O.; ElFaik, S.; Alami, S.; Lamin, H.; Bedmar, E. J.;
Abdelmoumen, H. Characterization of bradyrhizobium spp. nodulating lupinus cosentinii and L.
luteus microsymbionts in Morocco. *Front. Agron.* **2021**, *3*, 661295.
(65) Hollensteiner, J.; Schneider, D.; Poehlein, A.; Brinkhoff, T.; Daniel, R. Pan-genome analysis
of six Paracoccus type strain genomes reveal lifestyle traits. *Plos one* **2023**, *18* (12), e0287947.
(66) de Bruijn, I.; Cheng, X.; de Jager, V.; Expósito, R. G.; Watrous, J.; Patel, N.; Postma, J.;
Dorrestein, P. C.; Kobayashi, D.; Raaijmakers, J. M. Comparative genomics and metabolic
profiling of the genus Lysobacter. *BMC Genomics* **2015**, *16* (1), 991.
(67) Kumar, A.; Rithesh, L.; Kumar, V.; Raghuvanshi, N.; Chaudhary, K.; Abhineet; Pandey, A. K.
Stenotrophomonas in diversified cropping systems: friend or foe? *Front. Microbiol.* **2023**, *14*,
1214680.
(68) Wang, Z.; Hu, H.; Zhang, Z.; Xu, Y.; Xu, P.; Tang, H. IA multiple PAHs-degrading Shinella sp.
strain and its potential bioremediation in wastewater. *Sci Total Environ* **2023**, *879*, 162974.
(69) Wu, Y.-R.; He, J. Characterization of a xylanase-producing Cellvibrio mixtus strain J3-8 and
its genome analysis. *Sci. Rep.* **2015**, *5* (1), 10521.
(70) Forsberg, Z.; Nelson, C. E.; Dalhus, B.; Mekasha, S.; Loose, J. S.; Crouch, L. I.; Røhr Å, K.;
Gardner, J. G.; Eijsink, V. G.; Vaaje-Kolstad, G. Structural and functional analysis of a lytic
polysaccharide monooxygenase important for efficient utilization of chitin in cellvibrio japonicus.
J Biol Chem **2016**, *291* (14), 7300-7312.
(71) Ma, S.; Huang, S. W.; Tian, Y.; Lu, X. Y. Heterotrophic ammonium assimilation: An important
driving force for aerobic denitrification of Rhodococcus erythropolis strain Y10. *Chemosphere*
2022, *291*..
(72) Yuzikhin, O. S.; Gogoleva, N. E.; Shaposhnikov, A. I.; Konnova, T. A.; Osipova, E. V.; Syrova,
D. S.; Ermakova, E. A.; Shevchenko, V. P.; Nagaev, I. Y.; Shevchenko, K. V.; et al. Rhizosphere
bacterium Rhodococcus sp. P1Y metabolizes abscisic acid to form dehydrovomifoliol.
Biomolecules **2021**, *11* (3), 345.
(73) Jara-Servin, A.; Mejia, G.; Romero, M. F.; Peimbert, M.; Alcaraz, L. D. Unravelling the
genomic and environmental diversity of the ubiquitous Solirubrobacter. *Environ. Microbiol.* **2024**,
26 (8), e16685.
(74) Jung, H.; Lee, D.; Lee, S.; Kong, H. J.; Park, J.; Seo, Y.-S. Comparative genomic analysis of
Chryseobacterium species: deep insights into plant-growth-promoting and halotolerant capacities.
Microb. Genom. **2023**, *9* (10), 001108.
(75) Zhang, C.; Zhou, D.-F.; Wang, M.-Y.; Song, Y.-Z.; Zhang, C.; Zhang, M.-M.; Sun, J.; Yao, L.;

Mo, X.-H.; Ma, Z.-X. Phosphoribosylpyrophosphate synthetase as a metabolic valve advances Methylobacterium/Methylorubrum phyllosphere colonization and plant growth. *Nat. Commun.* **2024**, *15* (1), 5969.

(76) You, Y.; Aho, K.; Lohse, K. A.; Schwabedissen, S. G.; Ledbetter, R. N.; Magnuson, T. S. Biological soil crust bacterial communities vary along climatic and shrub cover gradients within a sagebrush steppe ecosystem. *Front. Microbiol.* **2021**, *12*, 569791.

(77) Lu, Y.; Rosencrantz, D.; Liesack, W.; Conrad, R. Structure and activity of bacterial community inhabiting rice roots and the rhizosphere. *Environ. Microbiol.* **2006**, *8* (8), 1351-1360.

(78) Shutsrirung, A.; Chromkaew, Y.; Pathom-Aree, W.; Choonluchanon, S.; Boonkerd, N. Diversity of endophytic actinomycetes in mandarin grown in northern Thailand, their phytohormone production potential and plant growth promoting activity. *Soil Sci. Plant Nutr.* **2013**, *59* (3), 322-330.

(79) Alotaibi, F.; St-Arnaud, M.; Hijri, M. In-depth characterization of plant growth promotion potentials of selected alkanes-degrading plant growth-promoting bacterial isolates. *Front. Microbiol.* **2022**, *13*, 863702.

(80) Ludington, W. B.; Seher, T. D.; Applegate, O.; Li, X.; Kliegman, J. I.; Langelier, C.; Atwill, E. R.; Harter, T.; DeRisi, J. L. Assessing biosynthetic potential of agricultural groundwater through metagenomic sequencing: A diverse anammox community dominates nitrate-rich groundwater. *PLoS One* **2017**, *12* (4), e0174930.

(81) Radhakrishnan, N.; Krishnasamy, C. Isolation and characterization of salt-stress-tolerant rhizosphere soil bacteria and their effects on plant growth-promoting properties. *Sci. Rep.* **2024**, *14* (1), 24909.

(82) Kaur, M.; Karnwal, A. Screening of plant growth-promoting attributes bearing endogenous bacteria from abiotic stress resisting high altitude plants. *J. Agric. Food Res.* **2023**, *11*, 100489.

(83) Venturi, V.; Keel, C. Signaling in the rhizosphere. *Trends Plant Sci.* **2016**, *21* (3), 187-198.

(84) Tzipilevich, E.; Russ, D.; Dangl, J. L.; Benfey, P. N. Plant immune system activation is necessary for efficient root colonization by auxin-secreting beneficial bacteria. *Cell host & microbe* **2021**, *29* (10), 1507-1520. e1504.

(85) Banerjee, S.; Schlaeppi, K.; van der Heijden, M. G. A. Keystone taxa as drivers of microbiome structure and functioning. *Nat. Rev. Microbiol.* **2018**, *16* (9), 567-576.

(86) Marschmann, G. L.; Tang, J.; Zhalnina, K.; Karaoz, U.; Cho, H.; Le, B.; Pett-Ridge, J.; Brodie, E. L. Predictions of rhizosphere microbiome dynamics with a genome-informed and trait-based energy budget model. *Nat. Microbiol.* **2024**, *9* (2), 421-433.

(87) Zhalnina, K.; Louie, K. B.; Hao, Z.; Mansoori, N.; Da Rocha, U. N.; Shi, S.; Cho, H.; Karaoz, U.; Loqué, D.; Bowen, B. P. Dynamic root exudate chemistry and microbial substrate preferences drive patterns in rhizosphere microbial community assembly. *Nat. Microbiol.* **2018**, *3* (4), 470-480.

(88) Thoenen, L.; Giroud, C.; Kreuzer, M.; Waelchli, J.; Gfeller, V.; Deslandes-Hérold, G.; Mateo, P.; Robert, C. A. M.; Ahrens, C. H.; Rubio-Somoza, I.; et al. Bacterial tolerance to host-exuded specialized metabolites structures the maize root microbiome. *Proc. Natl. Acad. Sci. U.S.A.* **2023**, *120* (44), e2310134120.

(89) Freire-Zapata, V.; Holland-Moritz, H.; Cronin, D. R.; Aroney, S.; Smith, D. A.; Wilson, R. M.;

Ernakovich, J. G.; Woodcroft, B. J.; Bagby, S. C.; 13, E. F. T. H. S. B. M. R. Microbiome–metabolite linkages drive greenhouse gas dynamics over a permafrost thaw gradient. *Nat. Microbiol.* **2024**, *9* (11), 2892-2908.

(90) Orozco-Mosqueda, M. D.; Santoyo, G.; Glick, B. R. Recent advances in the bacterial phytohormone modulation of plant growth. *Plants-Basel* **2023**, *12* (3). 606.

(91) Lin, H.; Chang, J. J.; Liang, T. N.; Yang, C.; Wei, Y. Y.; Peng, W.; Xu, A. K.; Duan, C. Q.; Zhao, Y. G. Effects of norfloxacin on the interaction between duckweed and its growth-promoting bacterial assemblages. *J. Hazard. Mater.* **2025**, *492*. 138249.

(92) Nascimento, F. X.; Hernandez, A. G.; Glick, B. R.; Rossi, M. J. The extreme plant-growth-promoting properties of Pantoea phytobeneficialis MSR2 revealed by functional and genomic analysis. *Environ. Microbiol.* **2020**, *22* (4), 1341-1355.

(93) Nett, R. S.; Montanares, M.; Marcassa, A.; Lu, X.; Nagel, R.; Charles, T. C.; Hedden, P.; Rojas, M. C.; Peters, R. J. Elucidation of gibberellin biosynthesis in bacteria reveals convergent evolution. *Nat. Chem. Biol.* **2017**, *13* (1), 69-74.

(94) Hartmann, M.; Six, J. Soil structure and microbiome functions in agroecosystems. *Nat. Rev. Earth Environ.* **2023**, *4* (1), 4-18.

(95) Nett, R. S.; Contreras, T.; Peters, R. J. Characterization of CYP115 as a gibberellin 3-oxidase indicates that certain rhizobia can produce bioactive gibberellin A4. *ACS Chem. Biol.* **2017**, *12* (4), 912-917.

(96) Pang, Z.; Chen, J.; Wang, T.; Gao, C.; Li, Z.; Guo, L.; Xu, J.; Cheng, Y. Linking plant secondary metabolites and plant microbiomes: a review. *Front. Plant Sci.* **2021**, *12*, 621276.

(97) Boiero, L.; Perrig, D.; Masciarelli, O.; Penna, C.; Cassán, F.; Luna, V. Phytohormone production by three strains of Bradyrhizobium japonicum and possible physiological and technological implications. *Appl. Microbiol. Biotechnol.* **2007**, *74* (4), 874-880.

(98) Eichmann, R.; Richards, L.; Schäfer, P. Hormones as go-betweens in plant microbiome assembly. *Plant J.* **2021**, *105* (2), 518-541.

(99) Pertry, I.; Václavíková, K.; Depuydt, S.; Galuszka, P.; Spíchal, L.; Temmerman, W.; Stes, E.; Schmülling, T.; Kakimoto, T.; Van Montagu, M. C. Identification of Rhodococcus fascians cytokinins and their modus operandi to reshape the plant. *Proc. Natl. Acad. Sci. U.S.A.* **2009**, *106* (3), 929-934.

(100) Yin, L.-Z.; Li, J.-L.; Fang, B.-Z.; Liu, Z.-T.; Wang, P.; Dong, L.; Duan, L.; Luo, X.-Q.; Li, S.-H.; Li, W.-J. Roseomonas ponticola sp. nov., a novel bacterium isolated from Pearl River estuary. *Int. J. Syst. Evol. Microbiol.* **2021**, *71* (10), 004994.

(101) Kuypers, M. M.; Marchant, H. K.; Kartal, B. The microbial nitrogen-cycling network. *Nat. Rev. Microbiol.* **2018**, *16* (5), 263-276.

(102) Zhang, X.; Qiu, Y.; Gilliam, F. S.; Gillespie, C. J.; Tu, C.; Reberg-Horton, S. C.; Hu, S. Arbuscular mycorrhizae shift community composition of N-cycling microbes and suppress soil N_2O emission. *Environ. Sci. Technol.* **2022**, *56* (18), 13461-13472.

(103) Cayuela, M. L.; Sánchez-Monedero, M. A.; Roig, A.; Hanley, K.; Enders, A.; Lehmann, J. Biochar and denitrification in soils: when, how much and why does biochar reduce N_2O emissions? *Sci. Rep.* **2013**, *3* (1), 1732.

(104) Shan, J.; Sanford, R. A.; Chee-Sanford, J.; Ooi, S. K.; Löffler, F. E.; Konstantinidis, K. T.;
Yang, W. H. Beyond denitrification: the role of microbial diversity in controlling nitrous oxide
reduction and soil nitrous oxide emissions. *Glob. Change Biol.* **2021**, *27* (12), 2669-2683.
(105) Li, Y.; Katzmann, E.; Borg, S.; Schüler, D. The periplasmic nitrate reductase Nap is required
for anaerobic growth and involved in redox control of magnetite biomineralization in
Magnetospirillum gryphiswaldense. *J. Bacteriol.* **2012**, *194* (18), 4847-4856.
(106) He, G.; Wang, W.; Chen, G.; Xie, Y.; Parks, J. M.; Davin, M. E.; Hettich, R. L.; Konstantinidis,
K. T.; Löffler, F. E. A novel bacterial protein family that catalyses nitrous oxide reduction. *Nature*
2025, *646*, 152–160.
(107) Garcia, P. S.; Gribaldo, S.; Borrel, G. Diversity and evolution of methane-related pathways
in archaea. *Annu. Rev. Microbiol.* **2022**, *76*, 727-755.
(108) Nazaries, L.; Murrell, J. C.; Millard, P.; Baggs, L.; Singh, B. K. Methane, microbes and
models: fundamental understanding of the soil methane cycle for future predictions. *Environ.*
Microbiol. **2013**, *15* (9), 2395-2417.