

1 The need to include phylogeny in trait-based analyses of
2 community composition

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22

23 **Summary**

24 1. A growing number of studies incorporate functional trait information to analyse patterns and
25 processes of community assembly. These studies of trait-environment relationships generally
26 ignore phylogenetic relationships among species. When functional traits and the residual
27 variation in species distributions among communities have phylogenetic signal, however,
28 analyses ignoring phylogenetic relationships can decrease estimation accuracy and power, inflate
29 type I error rates, and lead to potentially false conclusions.

30 2. Using simulations, we compared estimation accuracy, statistical power, and type I error rates
31 of linear mixed models (LMM) and phylogenetic linear mixed models (PLMM) designed to test
32 for trait-environment interactions in the distribution of species abundances among sites. We
33 considered the consequences of both phylogenetic signal in traits and phylogenetic signal in the
34 residual variation of species distributions generated by an unmeasured (latent) trait with
35 phylogenetic signal.

36 3. When there was phylogenetic signal in the residual variation of species among sites, PLMM
37 provided better estimates (closer to the true value) and greater statistical power for testing
38 whether the trait-environment interaction regression coefficient differed from zero. LMM had
39 unacceptably high type I error rates when there was phylogenetic signal in both traits and the
40 residual variation in species distributions. When there was no phylogenetic signal in the residual
41 variation in species distributions, LMM and PLMM had similar performances.

42 4. LMMs that ignore phylogenetic relationships can lead to poor statistical tests of trait-
43 environment relationships when there is phylogenetic signal in the residual variation of species
44 distributions among sites, such as caused by unmeasured traits. Therefore, phylogenies and

45 PLMMs should be used when studying how functional traits affect species abundances among
46 communities in response to environmental gradients.

47 **Introduction**

48 Species composition and abundance in ecological communities depend in part on both the
49 environmental conditions at a site and the traits expressed by species that allow them to live
50 under these environmental conditions. Typically, environmental conditions at a site allow only a
51 subset of species from the regional species pool to reach high abundances, with different
52 functional traits favouring species in different sites. Therefore, both environmental conditions
53 and functional traits play an important role in explaining species abundances in communities. To
54 better understand community assembly, we need to study the statistical interaction between
55 environmental conditions at a site and the functional traits of species that live there (McGill et al.
56 2006; Westoby & Wright, 2006).

57
58 Common statistical approaches to analyse how traits mediate species responses to environmental
59 variables have used either ordination with permutation tests (the fourth-corner problem and RLQ
60 analysis, Legendre, Galzin & Harmelin-Vivien 1997; Dray & Legendre, 2008) or an indirect
61 two-step approach. The fourth-corner problem links three data matrix tables: a site \times species
62 incidence/abundance matrix (L), a site \times environmental variables matrix (R), and a species \times
63 traits matrix (Q). The traits \times environmental variables matrix (R'LQ) is the fourth matrix (thus
64 explaining the etymology of the approach). While this approach provides a good qualitative
65 overview of how traits and environmental variables are associated, it does not give information
66 about species-specific variation in responses to environmental variables, and it is difficult to use

67 for prediction. The second, two-step approach first fits species-specific regressions of abundance
68 against environmental variables; the resulting regression coefficients are then regressed against
69 traits (e.g., Soudzilovskaia et al. 2013). This approach, while informative at the species level,
70 does not incorporate all community data in a single analysis and has low statistical power (Jamil
71 et al. 2013).

72
73 The interactions between traits and environmental variables can also be directly tested with
74 model-based methods (Bolker et al. 2009; Jamil et al. 2013; Brown et al. 2014; Warton et al.
75 2014, Ovaskaine, De Knecht & Delgado 2016). Statistically, the interaction between traits and
76 environmental variables can be estimated as the trait-environment interaction coefficient in
77 generalized linear models (GLMs, Brown et al. 2014), linear mixed models (LMMs, Ovaskaine,
78 De Knecht & Delgado 2016), or generalized linear mixed models (GLMMs, Pollock, Morris &
79 Veski 2012; Jamil et al. 2013). These model-based methods allow model selection and prediction,
80 and are often more flexible, powerful, and informative than fourth-corner and two-step
81 approaches (Ives & Helmus 2011; Jackson et al. 2012; Brown et al. 2014; Warton et al. 2014).

82
83 Most analyses of trait-environment interactions ignore phylogenetic relationships among species,
84 despite the large literature on phylogenetic analyses in comparative studies (Felsenstein 1985;
85 Harvey & Pagel 1991; Paradis 2012; Garamszegi 2014) and the relevance of phylogeny to many
86 areas of ecology (Webb et al. 2002; Cavender-Bares et al. 2009). This can lead to statistical
87 problems because functional traits often exhibit a phylogenetic pattern in which closely related
88 species share similar trait values (i.e., phylogenetic signal, Blomberg, Garland & Ives 2003). If
89 there are multiple traits that affect species abundance or incidence (or other characteristic of

90 interest), then the unmeasured traits with phylogenetic signal may generate covariance in the
91 unexplained, residual variation after accounting for measured traits. This covariance in the
92 residual variation will reflect the phylogeny, and will affect model estimation and hypothesis
93 testing of regression coefficients (e.g., Felsenstein 1985, Martins & Hansen 1997; Garland,
94 Bennett & Rezende 2005; Revell 2010).

95
96 Here, we investigate the need to incorporate phylogenetic covariance among species in
97 regressions for trait-environment interactions. We considered a regression problem in which
98 there is a causal but unmeasured (latent) trait that introduces unexplained variability in species
99 abundance, and phylogenetic covariance in the unexplained variation if the unmeasured trait has
100 phylogenetic signal. This gives four possible cases (Revell 2010): the pairwise combinations of
101 whether or not there is phylogenetic signal in the measured trait in the regression, and whether or
102 not there is phylogenetic signal in the residual variation. We then compared the accuracy, type I
103 error rates, and statistical power of linear mixed models (LMMs) and phylogenetic linear mixed
104 models (PLMMs, Ives & Helmus 2011) in estimating the trait-environment interaction
105 coefficient. We show that when there is phylogenetic signal in the residual variation (latent trait),
106 PLMM outperformed LMM, with LMM performing particularly poorly when there is also
107 phylogenetic signal in the measured trait.

108

109 **Materials and methods**

110 We simulated data to test the importance of accounting for phylogenetic relationships when
111 studying how functional traits interact with environmental variables to affect species abundances.
112 All simulations and calculations were performed with R (R Core Team, 2015).

113 Simulations

114 We simulated the abundance Y of species j ($j = 1, \dots, n$) at site s ($s = 1, \dots, m$) that depends on
115 two site environmental variables (env1 and env2) and two species functional traits
116 (trait1 and trait2) using the model

$$118 \quad Y_i = \alpha + \beta_1 \text{env1}_{\text{site}[i]} + \beta_2 \text{env2}_{\text{site}[i]} + \beta_3 \text{trait1}_{\text{spp}[i]} + \beta_4 \text{trait2}_{\text{spp}[i]} + \\ 119 \quad \beta_5 (\text{env1}_{\text{site}[i]} \times \text{trait1}_{\text{spp}[i]}) + \beta_6 (\text{env1}_{\text{site}[i]} \times \text{trait2}_{\text{spp}[i]}) + e_i. \quad (1)$$

120
121 Functions $\text{spp}[i]$ and $\text{site}[i]$ map the observation i to the identity of the species and site,
122 respectively (Gelman & Hill, 2007, p251-252), so i takes values from 1 to nm . We assume both
123 environmental variable env1 and functional trait trait1 are measured. Env1 (e.g., soil fertility,
124 canopy cover) affects the abundance of all species among sites ($\beta_1 \neq 0$), and trait1 (e.g., nutrient
125 absorption capacity, specific leaf area) determines in part the overall abundance of species ($\beta_3 \neq$
126 0). Furthermore, there is an interaction between env1 and trait1 ($\beta_5 \neq 0$) implying that trait1
127 affects the performance of species along the environmental gradient env1.

128
129 To introduce unexplained variation and phylogenetic signal, we treated env2 and trait2 as
130 unmeasured (latent) variables. Like env1, env2 has a direct effect on species abundances ($\beta_2 \neq 0$).
131 Like trait1, trait2 determines in part species abundances ($\beta_4 \neq 0$) and has an interactive effect
132 with env1 ($\beta_6 \neq 0$). As we are mainly interested in the trait \times environment interactions for the
133 measured data (env1 and trait1), we did not include the interactions between env2 and trait1 or
134 trait2. Our goal is to investigate the interaction between env1 and trait1 which is given by β_5 .
135 Consequently, we set all parameters in equation 1 other than β_5 to be 1. Finally, we simulated e_i

136 as a normal random variable that is independent among species and sites. In this way, we treated
137 the abundance of species Y as log-transformed values from count data. We did not simulate
138 abundance as raw count data because log-transformation of count data usually does not affect the
139 significance tests for regression coefficients when low count values (<5) are uncommon (Ives
140 2015; Warton et al. 2016).

141
142 We simulated the phylogeny as a uniform birth-death process with birth rate = 1 and death rate =
143 0 using the `sim.bdtree` function of the `geiger` R package (Harmon et al. 2008). The
144 phylogeny gives the expected phylogenetic covariances among species under Brownian motion
145 evolution (Grafen 1989; Martins & Hansen 1997) that can be used to construct a matrix \mathbf{C} , and
146 when there is no phylogenetic signal the (zero) covariance structure is given by the identity
147 matrix \mathbf{I} . Because functional traits may or may not have phylogenetic signal, we simulated four
148 scenarios for the two functional traits: trait1 with phylogenetic signal but not trait2 (trait1: \mathbf{C} ;
149 trait2: \mathbf{I}); trait2 with phylogenetic signal but not trait1 (trait1: \mathbf{I} ; trait2: \mathbf{C}); both traits with
150 phylogenetic signal (trait1: \mathbf{C} ; trait2: \mathbf{C}); and neither trait with phylogenetic signal (trait1: \mathbf{I} ;
151 trait2: \mathbf{I}). Functional traits without phylogenetic signal were simulated as $N(0, 1)$ normal random
152 variables; functional traits with phylogenetic signal were simulated using the `fastBM` function of
153 the `phytools` R package (Revell 2012). We simulated `env1` as a uniform distribution ranging
154 from -1 and 1 to generate a strong environmental gradient. Variable `env2` and residuals e_i were
155 simulated as $N(0, 1)$ normal random variables.

156
157 We conducted simulations with 30 sites. To study type I error rates (false positives that
158 incorrectly reject the true null hypothesis), we set $\beta_5 = 0$ and varied the number of species (20,

159 30, 40, 50, 60, 70, 80). To study statistical power, we varied the value of β_5 (0, 0.25, 0.5, 0.75, 1)
160 and fixed the number of species at 50. For each case we performed 1000 simulations.

161

162 **Model fitting**

163 We fit both LMM and PLMM to the simulated datasets with R package `pez` (Pearse et al. 2015).

164 The LMM has the form

165

$$166 Y_i = \alpha + a_{\text{spp}[i]} + b_{\text{site}[i]} + (\beta_1 + c_{\text{spp}[i]}) \text{env1}_{\text{site}[i]} + \beta_3 \text{trait1}_{\text{spp}[i]} + \beta_5 \text{env1}_{\text{site}[i]} \times \text{trait1}_{\text{spp}[i]} + e_i$$

$$167 a \sim \text{Gaussian}(\mathbf{0}, \sigma_a^2 \mathbf{I}_n)$$

$$168 b \sim \text{Gaussian}(\mathbf{0}, \sigma_b^2 \mathbf{I}_m)$$

$$169 c \sim \text{Gaussian}(\mathbf{0}, \sigma_c^2 \mathbf{I}_n) \tag{2}$$

170

171 Here, we use the convention of multilevel models (Gelman & Hill, 2007), with fixed and random
172 effects given by Greek and Latin letters, respectively. The fixed effects β_1 , β_3 , and β_5 correspond
173 to the same coefficients in the simulation model (equation 1). Random effect $a_{\text{spp}[i]}$ allows
174 different species to have different overall abundance to capture effects of the term $\beta_4 \text{trait2}_{\text{spp}[i]}$ in
175 equation 1. Random effect $b_{\text{site}[i]}$ allows different sites to have different overall abundance across
176 all species within that site to capture effects of the term $\beta_2 \text{env2}_{\text{site}[i]}$ in equation 1. Finally,
177 random effect $c_{\text{spp}[i]}$ allows different species to have different responses to env1 to capture effects
178 of the term $\beta_6 \text{env1}_{\text{site}[i]} \times \text{trait2}_{\text{spp}[i]}$ in equation 1.

179

180 The PLMM includes all terms of equation 2, plus phylogenetic version of random terms $a_{\text{spp}[i]}$

181 and $c_{\text{spp}[i]}$.

182

$$183 \quad Y_i = \alpha + (a_{\text{spp}[i]} + a_{\text{spp}[i]}^p) + b_{\text{site}[i]} + (\beta_1 + c_{\text{spp}[i]} + c_{\text{spp}[i]}^p) \text{env1}_{\text{site}[i]} + \beta_3 \text{trait1}_{\text{spp}[i]} + \\ 184 \quad \beta_5 \text{env1}_{\text{site}[i]} \times \text{trait1}_{\text{spp}[i]} + e_i$$

$$185 \quad a \sim \text{Gaussian}(\mathbf{0}, \sigma_a^2 \mathbf{I}_n)$$

$$186 \quad a^p \sim \text{Gaussian}(\mathbf{0}, \sigma_{\text{ap}}^2 \mathbf{C})$$

$$187 \quad b \sim \text{Gaussian}(\mathbf{0}, \sigma_b^2 \mathbf{I}_m)$$

$$188 \quad c \sim \text{Gaussian}(\mathbf{0}, \sigma_c^2 \mathbf{I}_n)$$

$$189 \quad c^p \sim \text{Gaussian}(\mathbf{0}, \sigma_{\text{cp}}^2 \mathbf{C}) \tag{3}$$

190

191 Random effect $a_{\text{spp}[i]}^p$ implies closely related species to have similar overall abundance; this will
192 capture the main effects of traits in the simulations (equation 1) if trait2 has phylogenetic signal.
193 Similarly, random effect $c_{\text{spp}[i]}^p$ allows closely related species to have similar responses to env1,
194 thereby capturing the interactive effect of trait2 and env1 in the simulations if trait2 has
195 phylogenetic signal.

196 Results

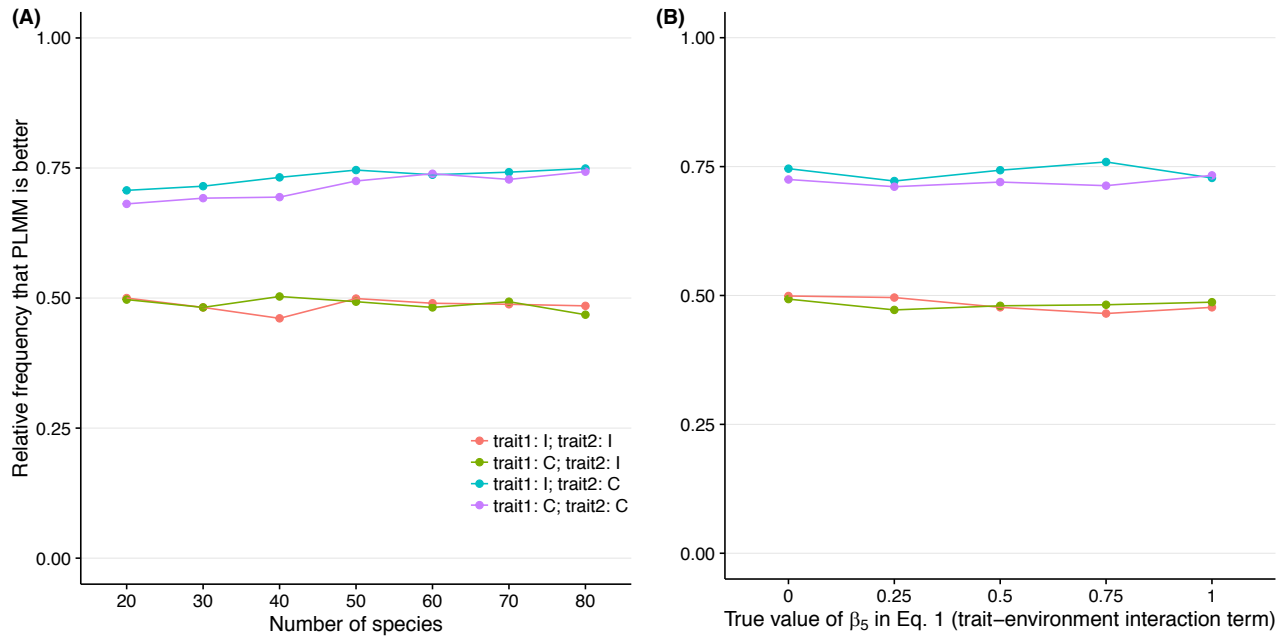
197 To compare LMMs and PLMMs, we focused on the regression coefficient β_5 for the interaction
198 between env1 and trait1. For each simulated dataset, we compared the accuracy of LMM and
199 PLMM by determining the frequency with which one gave a more accurate estimate of β_5 than
200 the other, and also by calculating the means and standard deviations of the estimates of β_5 . We
201 also counted the number of estimates that were scored as significant at the $\alpha = 0.05$ level for both
202 models to determine their type I errors (when the true value of $\beta_5 = 0$) and statistical power
203 (when the true value of $\beta_5 > 0$).

204

205 *No phylogenetic signal in trait2*

206 When the unmeasured trait2 did not have phylogenetic signal (trait1: **I**; trait2: **I**, and trait1: **C**;
207 trait2: **I**), implying no phylogenetic signal in the unexplained variation in species abundances
208 among sites, LMM and PLMM had similar estimation accuracy (Fig. 1-2), type I error rates, and
209 power (Fig. 3). Averaged across all simulation scenarios, in roughly 50% of simulations LMM
210 produced better estimates (closer to the true value) of β_5 (Fig. 1). The estimators of β_5 from
211 LMM and PLMM had similar means and standard deviations (Fig. 2A, Fig. 2B, Fig. A1).
212 Furthermore, LMM and PLMM had almost identical type I error rates and power across all
213 simulation scenarios (Fig. 3). They also gave very similar estimates when $\beta_5 > 0$ (Fig. A2). These
214 results are explained, in part, by the fact that in about 65% of simulations across all scenarios we
215 investigated with no phylogenetic residual variation (trait2: **I**), the estimates of both σ_{ap}^2 and σ_{cp}^2
216 in the PLMM were zero, so the PLMM collapsed to the LMM and estimates of β_5 were the same
217 (\pm numerical accuracy in the REML optimizations).

218



219

220 **Figure 1.** The fraction of simulations in which PLMM yielded a better estimate of β_5 (i.e., closer to its

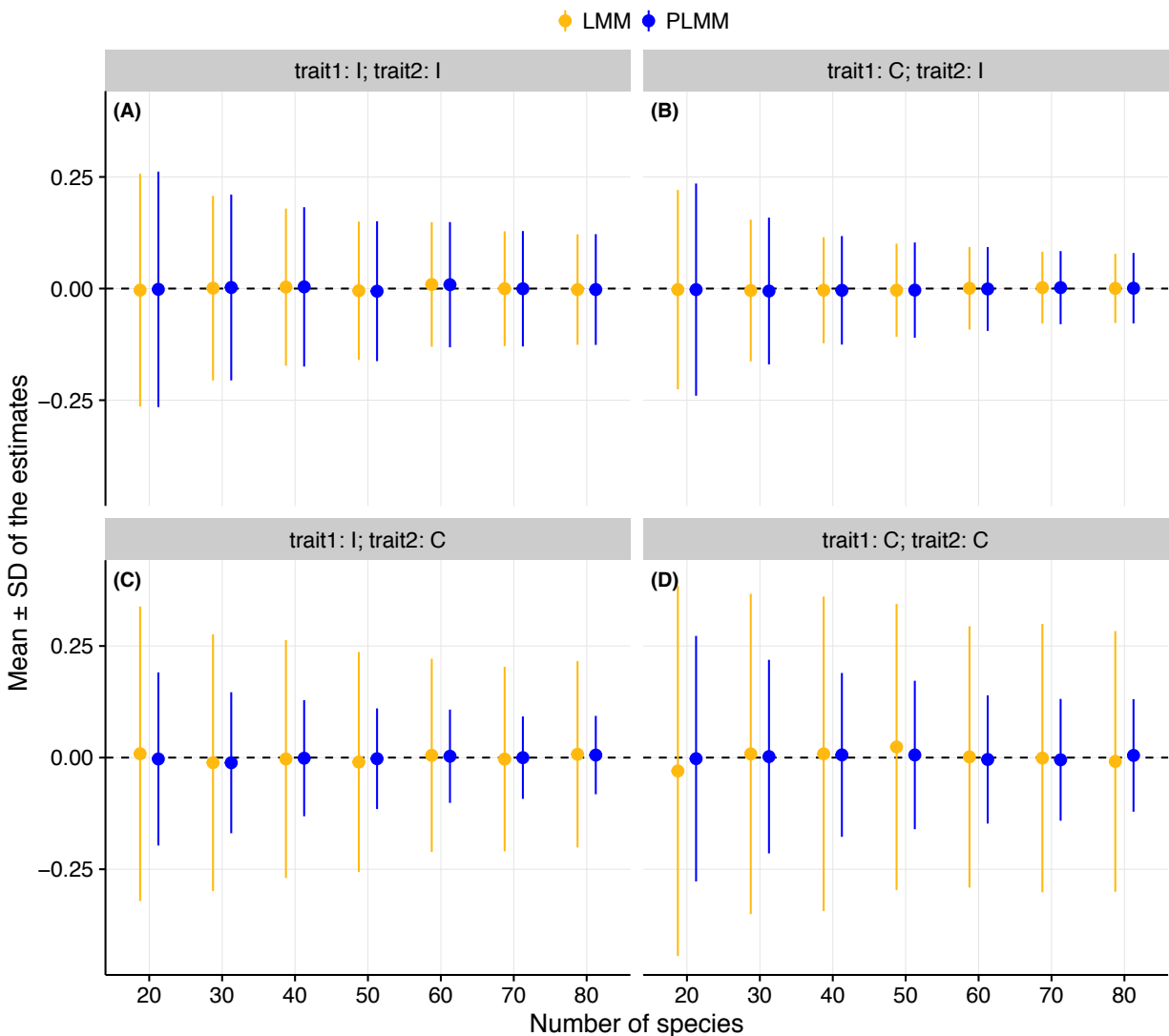
221 true value) than LMM versus (A) the number of simulated species and (B) the true value of β_5 (Eq. 1).

222 The performance of PLMM was consistently better than LMM whenever there was phylogenetic signal in

223 the residual variation (caused by unmeasured trait2). Abbreviations: trait1: **I** – measured trait1 does not

224 have phylogenetic signal; trait1: **C** – measured trait1 has phylogenetic signal; trait2: **I** – unmeasured trait2

225 does not have phylogenetic signal; trait2: **C** – unmeasured trait2 has phylogenetic signal.



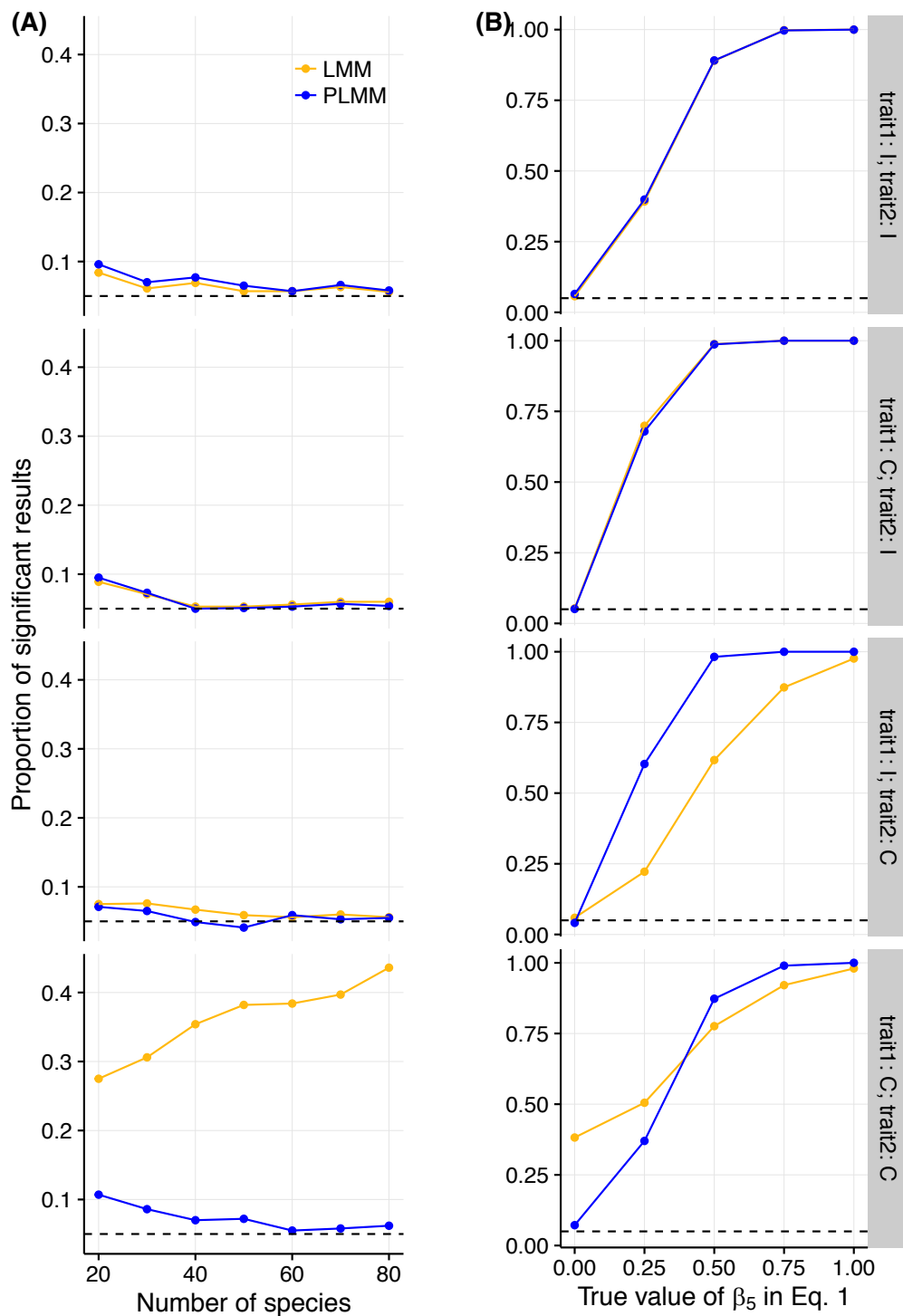
226

227 **Figure 2:** Mean (\pm standard deviation) of simulated estimates of β_5 (Eq. 1) using LMM and PLMM

228 versus the number of species in the simulations for cases (A) trait1: **I**; trait2: **I**, (B) trait1: **C**; trait2: **I**,

229 (C) trait1: **I**; trait2: **C**, and (D) trait1: **C**; trait2: **C**. Horizontal dash lines represent the true value of the

230 parameter. Abbreviations are as in Fig. 1.



231

232 **Figure 3.** (A) Type I error rates and (B) statistical power of LMMs and PLMMs under four scenarios of
 233 simulated functional traits (abbreviations as in Fig. 1). For all tests, a significance level of $\alpha = 0.05$ is used
 234 (horizontal dashed lines).

235 *Phylogenetic signal in trait2*

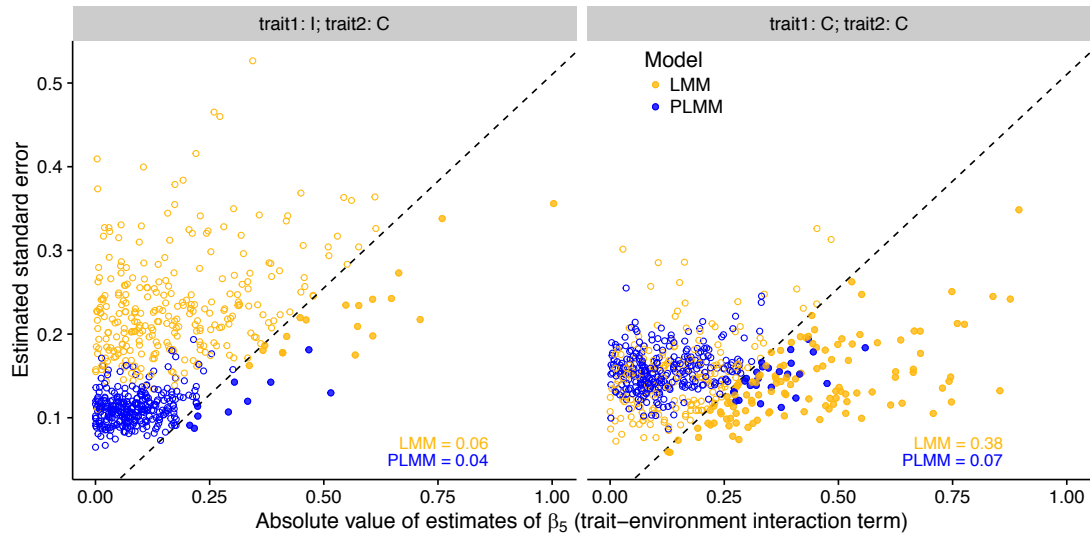
236 When the unmeasured trait2 had phylogenetic signal (trait1: **I**; trait2: **C**, and trait1: **C**; trait2: **C**),
237 PLMM had substantially higher estimation accuracy (Fig. 1-2), better type I error control (Fig.
238 3A), and higher power (Fig. 3B) than LMM. Type I error control and power were particularly
239 poor for LMM when trait1 also had phylogenetic signal (i.e., trait1: **C**; trait2: **C**).

240
241 Averaged across all simulation conditions, in about 75% simulations PLMM produced more
242 accurate estimates of β_5 (Fig. 1), and the variance of the estimator of β_5 (Fig. 2 and A1) was
243 consistently lower than LMM. This was true regardless of the number of species, the true value of
244 β_5 , and the status of the measured trait1 (with or without phylogenetic signal) used in
245 simulations. In addition, for type I error control and power, LMM had particularly poor
246 performance when the measured trait1 had phylogenetic signal (trait1: **C**; trait2: **C**). For
247 simulations with $\beta_5 = 0$ (Fig. 3), LMM rejected $H_0: \beta_5 = 0$ at the $\alpha = 0.05$ level in ~25% of the
248 datasets with 20 species, and type I error control became worse as the number of species
249 increased. When there was no phylogenetic signal in trait1 (trait1: **I**; trait2: **C**) and type I error
250 control was only slightly elevated, LMM had much lower power than PLMM (Fig. 3B)

251
252 We investigated further the particularly poor type I error control of LMM when there is
253 phylogenetic signal in both the measured trait and the unexplained residual variation (trait1: **C**;
254 trait2: **C**). Poor type I error control occurs when the estimate of the standard error of β_5 is smaller
255 than the true standard error. For cases both with phylogenetic signal in trait 1 (trait1: **C**; trait2: **C**)
256 and without (trait1: **I**; trait2: **C**), we plotted the estimate of the standard error of β_5 for each
257 simulated dataset against the estimate of β_5 using both LMM and PLMM (Fig. 4). For the case

258 (trait1: **I**; trait2: **C**), the decrease in accuracy of LMM relative to PLMM is seen in the greater
259 variance in the estimates of β_5 (variance in the horizontal direction). Despite this increase in the
260 variance in the estimates of β_5 , false positives (given by values to the right of the dashed line of
261 Fig. 4) from the LMM are only slightly inflated, because the LMM estimates of the standard
262 error of β_5 are larger than those from PLMM. However, for the case (trait1: **C**; trait2: **C**), the
263 decrease in accuracy of LMM relative to PLMM is not accompanied by an appropriate increase
264 in the LMM estimates of the standard error, thereby leading to high type I error rates. In contrast
265 to LMM, even though the variance in the estimates of β_5 from PLMM increases when there is
266 phylogenetic signal in trait1 (Fig. 4A vs. 4B), the estimates of the standard error also increase,
267 leading to much better type I error control than LMM. In summary, the poor type I error control
268 for LMM when there is phylogenetic signal in trait1 occurs because, as phylogenetic signal in
269 trait1 increases the variance in the LMM estimates of β_5 , phylogenetic signal in trait1 decreases
270 the LMM estimates of this variance. The decrease in power of LMM relative to PLMM for the
271 case without phylogenetic signal in trait1 (trait1: **I**; trait2: **C**) is caused by the increase in
272 variance in the estimator of β_5 , that is, decreased accuracy. Given the very poor type I error
273 control for LMM for the case with phylogenetic signal in trait1 (trait1: **C**; trait2: **C**), it is
274 inappropriate to assess power for this case.

275



276

277 **Figure 4.** The relationships between the absolute value of the estimates of β_5 and the estimates of the
278 standard errors from LMM (yellow) and PLMM (blue). Solid points to the right of the dashed line would
279 (approximately) reject the null hypothesis $H_0: \beta_5 = 0$ at the 0.05 α significance level. When unmeasured
280 trait2 has phylogenetic signal but not measured trait1 (trait1: **I**; trait2: **C**), LMM estimates are more
281 variable (horizontal axis) and have greater estimated standard errors (vertical axis) than PLMM, leading
282 to only slightly inflated type I error control. When both trait1 and trait2 have phylogenetic signal (trait1:
283 **C**; trait2: **C**), LMM estimates are more variable, but this is not correctly captured by increasing estimates
284 of standard errors, leading to very high type I error rates. The dashed line has intercept of zero and slope
285 of 1/1.96. Simulations were performed with 50 species; the fractions of simulations rejecting the null
286 hypothesis (text in the panels) were calculated from 1000 simulations, of which only 300 are presented
287 for clarity.

288

289 Discussion

290 Our simulations have demonstrated the importance of incorporating phylogeny into the study of
291 how species functional traits interact with the environment to affect their abundance. In

292 simulations in which there was phylogenetic signal in the residual variation in abundances
293 caused by an unmeasured (latent) trait, we showed that LMMs have lower accuracy, poor type I
294 error control, and lower power than PLMMs in identifying the trait \times environment interaction.
295 The performance of LMMs was particularly poor in terms of type I error control and power when
296 there was also phylogenetic signal in the measured trait. In contrast, PLMMs had better
297 accuracy, generally good type I error control (except when the number of species was small), and
298 good power.

299
300 Our results mirror the results of Revell (2010) who studied the performance of LMs and PLMs
301 applied to regression for phylogenetic comparative data. The model he considered that most
302 closely corresponds to our PLMM is a phylogenetic least-squares model in which Pagel's λ
303 branch-length transform is used. Pagel's λ transformation can be constructed by adding a
304 phylogenetic and a non-phylogenetic covariance matrix with λ scaling between them (i.e., $(1 -$
305 $\lambda)\mathbf{I} + \lambda\mathbf{C}$). In our PLMM (Eq. 3), covariance terms are similarly combined; for example, the
306 covariance for species-specific slopes across environmental variable 1 is $\sigma_c^2\mathbf{I}_n + \sigma_{cp}^2\mathbf{C}$. Revell
307 (2010) found that PLMs outperformed LMs when there was phylogenetic signal in the residual
308 variation, with the performance of LMs particularly poor when there was also phylogenetic
309 signal in the independent variable. Thus, we found similar results in the more-complex problem
310 of identifying trait \times environment interactions in community data.

311
312 The better performance of PLMMs over LMMs is not surprising on theoretical grounds. For the
313 special, hypothetical case in which the variance parameters σ_a^2 , σ_{ap}^2 , σ_b^2 , σ_c^2 , and σ_{cp}^2 are known, the
314 PLMM in equation 3 will be the minimum variance estimator of the regression coefficients

315 (fixed effects), including the trait \times environment interaction β_5 ; this is a consequence of the
316 Cramer-Rao Theorem applied to Generalized Least Squares (GLS) models (Judge et al. 1985).
317 This explains why PLMMs provide more accurate estimates of β_5 than LMMs, and the increase
318 in accuracy explains the increase in power of PLMMs relative to LMMs.

319
320 A particular warning derived from our simulations is the poor type I error control for LMMs
321 when there is phylogenetic signal in both the residual variation and in the independent variable.
322 When there is also phylogenetic signal in the measured trait1, the variance in the estimates of β_5
323 greatly increases. Nonetheless, the LMM estimates of the standard error of β_5 do not increase as
324 they should, leading to false rejections of the null hypothesis that $\beta_5 = 0$. Because PGLMMs are
325 close to the minimum variance estimators of β_5 , the variance in its estimates of β_5 does not
326 increase as much as LMMs when there is phylogenetic signal in the independent variable, and
327 what increase occurs is correctly given by the estimates standard errors of β_5 ; thus, there is
328 generally good type I error control.

329
330 When the number of species is small (<60), however, PLMM had inflated type I error rates; for
331 simulations with 20 species and phylogenetic signal in both independent variable (measured
332 trait1) and residual variation (unmeasured trait2), the null hypothesis $H_0: \beta_5 = 0$ was rejected in
333 10% of the datasets at the α significance level of 0.05. In analyses with small numbers of species
334 and P-values computed from the data that are close to the significance level selected by the
335 researcher, we suggest using parametric bootstrapping. This can be performed by estimating
336 parameters from the data under $H_0: \beta_5 = 0$ (i.e., without the trait \times environment interaction),
337 simulating a large number (e.g., 2000) datasets with these parameter values, fitting each dataset

338 with the full model (i.e., with the trait \times environment interaction), and for each dataset recording
339 the Z-score of the estimate of β_5 . The bootstrap approximate P-value of β_5 under an α
340 significance level of 0.05 is then given by the proportion of bootstrap Z-scores whose absolute
341 values exceed the absolute value of the Z-score from the observed data. Code for performing this
342 bootstrap is provided in Appendix S1.

343
344 Our analyses have been confined to abundance as a continuous dependent variable.
345 Presence/absence (incidence) community data can also be analysed with phylogenetic
346 information using PGLMM (Ives and Helmus 2011), and results will likely be similar. We did
347 not pursue this here, however, because the computational burden of PGLMMs with existing
348 software makes simulation studies difficult. Nonetheless, if tests of the existence of relationships
349 (i.e., testing $H_0: \beta_5 = 0$) are all that is needed, applying PLMMs to binary data generally provides
350 good type I error control, although at the expense of some power (Ives 2015; Warton et al. 2016).

351
352 Even when there was no phylogenetic signal in the residual variation, PLMMs performed as well
353 as LMMs. In part, this is because, when PLMMs detected no phylogenetic signal in the residual
354 variation, they give the same results as the corresponding LMMs (although their AIC values are
355 still penalized by the variance term that equals zero). The fact that PLMMs often collapse exactly
356 to LMMs as a special case suggests that PLMMs should be always used in analyses of trait \times
357 environment interactions, since there is no cost in the absence of phylogenetic signal and
358 considerable benefits when there is (which is likely).

359

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363 **Data accessibility:**

364 No data were used in this article.

365 **References:**

- 366 Blomberg, S. P., Garland, T., & Ives, A.R. (2003). Testing for phylogenetic signal in comparative data:
367 behavioral traits are more labile. *Evolution*, **57**:717-745.
- 368 Bolker, B.M., Brooks, M.E., Clark, C.J., Geange, S.W., Poulsen, J.R., Stevens, M.H.H. & White, J.-S.S.
369 (2009). Generalized linear mixed models: A practical guide for ecology and evolution. *Trends in Ecology*
370 *& Evolution*, **24**, 127–135.
- 371 Brown, A.M., Warton, D.I., Andrew, N.R., Binns, M., Cassis, G. & Gibb, H. (2014). The fourth-corner
372 solution - using predictive models to understand how species traits interact with the environment.
373 *Methods in Ecology and Evolution*, **5**, 344–352.
- 374 Cavender-Bares, J., Kozak, K.H., Fine, P.V.A. & Kembel, S.W. (2009). The merging of community
375 ecology and phylogenetic biology. *Ecology Letters*, **12**, 693–715.
- 376 Dray, S. & Legendre, P. (2008). Testing the species traits-environment relationships: The fourth-corner
377 problem revisited. *Ecology*, **89**, 3400–3412.
- 378 Felsenstein, J. (1985). Phylogenies and the Comparative Method. *The American Naturalist*, **125**, 1–15.
- 379 Garamszegi, L.Z. (2014). Modern phylogenetic comparative methods and their application in
380 evolutionary biology. *Concepts and Practice*. London, UK: Springer.
- 381 Garland, T., Bennett, A.F. & Rezende, E.L. (2005). Phylogenetic approaches in comparative physiology.
382 *Journal of experimental Biology*, **208**, 3015–3035.

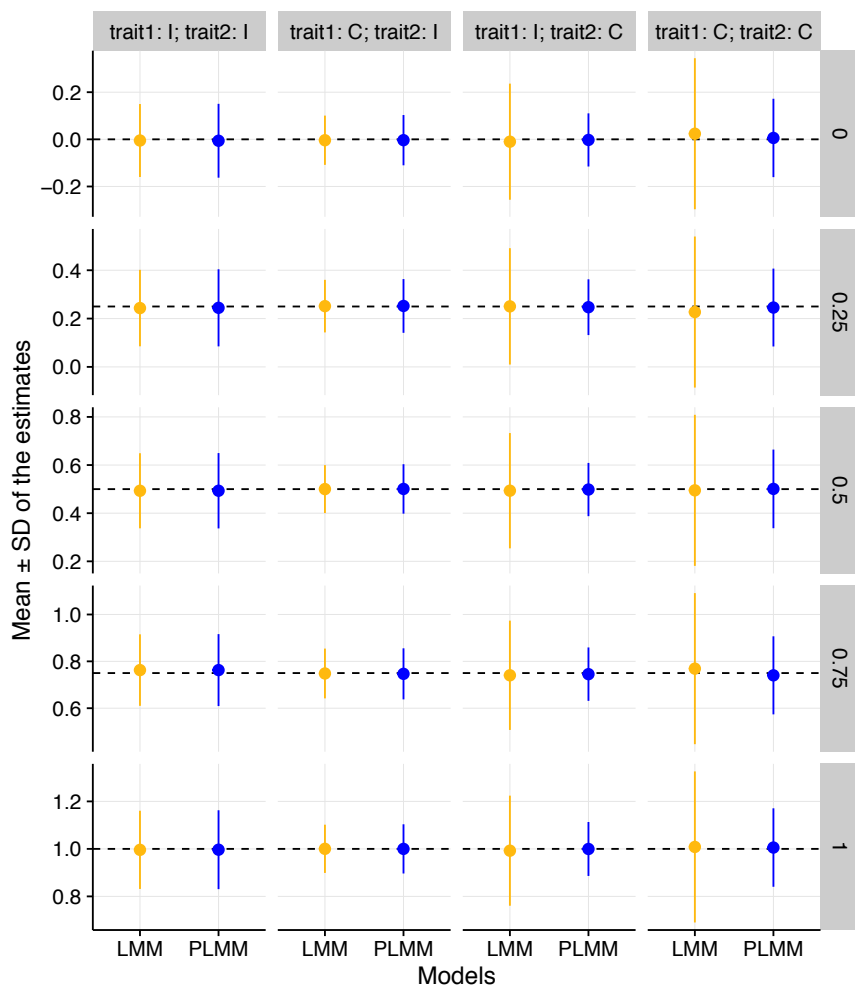
- 383 Gelman, A. & Hill, J. (2007). *Data analysis using regression and multilevel/hierarchical models*.
384 Cambridge University Press.
- 385 Grafen, A. (1989). The phylogenetic regression. *Philosophical Transactions of the Royal Society of*
386 *London. Series B, Biological Sciences*, **326**, 119–157.
- 387 Harmon, L., Weir, J., Brock, C., Glor, R. & Challenger, W. (2008). GEIGER: Investigating evolutionary
388 radiations. *Bioinformatics*, **24**, 129–131.
- 389 Harvey, P.H., Pagel, M.D. & others. (1991). *The comparative method in evolutionary biology*. Oxford
390 university press Oxford.
- 391 Ives, A.R. (2015). For testing the significance of regression coefficients, go ahead and log-transform
392 count data. *Methods in Ecology and Evolution*, **6**, 828–835.
- 393 Ives, A.R. & Helmus, M.R. (2011). Generalized linear mixed models for phylogenetic analyses of
394 community structure. *Ecological Monographs*, **81**, 511–525.
- 395 Jackson, M.M., Turner, M.G., Pearson, S.M. & Ives, A.R. (2012). Seeing the forest and the trees:
396 Multilevel models reveal both species and community patterns. *Ecosphere*, **3**, art79.
- 397 Jamil, T., Ozinga, W.A., Kleyer, M. & ter Braak, C.J. (2013). Selecting traits that explain
398 speciesenvironment relationships: A generalized linear mixed model approach. *Journal of Vegetation*
399 *Science*, **24**, 988–1000.
- 400 Judge, G.G., Griffiths, W., Hill, R.C., Lutkepohl, H. & Lee, T.C. (1985). Introduction to the theory and
401 practice of econometrics.
- 402 Legendre P., M.L.H.-V., René Galzin. (1997). Relating behavior to habitat: Solutions to the fourth-corner
403 problem. *Ecology*, **78**, 547–562.
- 404 Martins, E.P. & Hansen, T.F. (1997). Phylogenies and the Comparative Method: A General Approach to
405 Incorporating Phylogenetic Information into the Analysis of Interspecific Data. *The American Naturalist*,
406 **149**, 646–667.
- 407 McGill, B.J., Enquist, B.J., Weiher, E. & Westoby, M. (2006). Rebuilding community ecology from
408 functional traits. *Trends in Ecology & Evolution*, **21**, 178–185.
- 409 Ovaskainen, O., De Knecht, H.J. & Mar Delgado, M. del. (2016). *Quantitative ecology and evolutionary*
410 *biology: Integrating models with data*. Oxford University Press.

- 411 Paradis, E. (2011). *Analysis of phylogenetics and evolution with r*. Springer Science & Business Media.
- 412 Pearse, W.D., Cadotte, M.W., Cavender-Bares, J., Ives, A.R., Tucker, C.M., Walker, S.C., & Helmus,
413 M.R. (2015). pez: phylogenetics for the environmental sciences. *Bioinformatics*, **31**, 2888-2890.
- 414 Pollock, L.J., Morris, W.K. & Vesk, P.A. (2012). The role of functional traits in species distributions
415 revealed through a hierarchical model. *Ecography*, **35**, 716–725.
- 416 R Core Team. (2015). R: A Language and Environment for Statistical Computing.
- 417 Revell, L.J. (2010). Phylogenetic signal and linear regression on species data. *Methods in Ecology and*
418 *Evolution*, **1**, 319–329.
- 419 Revell, L.J. (2012). Phytools: An r package for phylogenetic comparative biology (and other things).
420 *Methods in Ecology and Evolution*, **3**, 217–223.
- 421 Soudzilovskaia, N.A., Elumeeva, T.G., Onipchenko, V.G., Shidakov, I.I., Salpagarova, F.S., Khubiev,
422 A.B., Tekeev, D.K. & Cornelissen, J.H.C. (2013). Functional traits predict relationship between plant
423 abundance dynamic and long-term climate warming. *Proceedings of the National Academy of Sciences*,
424 **110**, 18180–18184.
- 425 Warton, D.I., Foster, S.D., De'ath, G., Stoklosa, J. & Dunstan, P.K. (2014). Model-based thinking for
426 community ecology. *Plant Ecology*, 1–14.
- 427 Warton, D.I., Lyons, M., Stoklosa, J. & Ives, A.R. (2016). Three points to consider when choosing a LM
428 or GLM test for count data. *Methods in Ecology and Evolution*, **7**, 882–890.
- 429 Webb, C.O., Ackerly, D.D., McPeck, M.A. & Donoghue, M.J. (2002). Phylogenies and community
430 ecology. *Annual Review of Ecology and Systematics*, **33**, 475–505.
- 431 Westoby, M. & Wright, I.J. (2006). Land-plant ecology on the basis of functional traits. *Trends in*
432 *Ecology & Evolution*, **21**, 261–268.

433 **Supporting information:**

- 434 Appendix S1: R scripts used for the simulations and analyses (in the Rcode folder). It is time
435 consuming to re-run all simulations and model fittings. Therefore, we also include all model

436 results for fixed effects (as `rds.zip` and `rds2.zip` files). (We cannot upload these two files
437 through manuscript central as they include >50 files; one can download them from here:
438 <https://uwmadison.box.com/s/11nsmqlymg53xyxdarfkjsjrtegvuzci>)

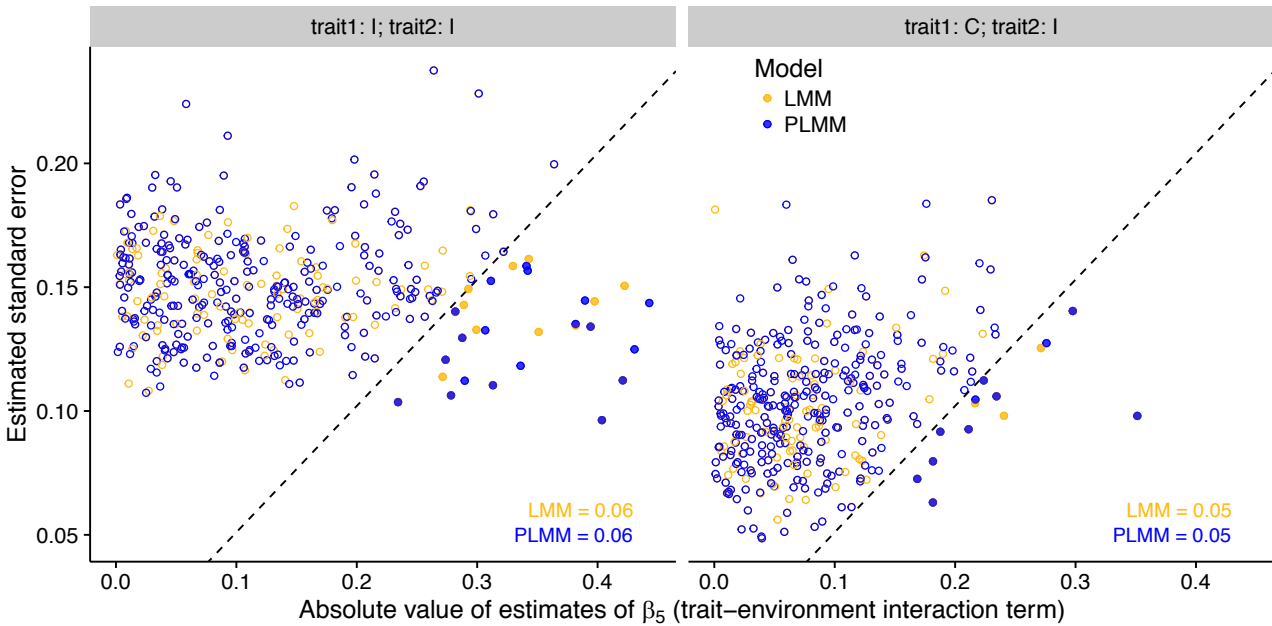


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440

441 **Figure A1.** Mean (\pm standard deviation) of estimates of β_5 (Eq. 1) using LMM and PLMM varying the
442 true value of β_5 (0, 0.25, 0.5, 0.75, and 1). Horizontal dash lines represent the true value of the parameter.

443 Abbreviations are as in Fig. 1.



444

445 **Figure A2.** The relationships between the absolute value of the estimates of β_5 and the estimates of the
446 standard errors for LMM (yellow) and PLMM (blue). Solid points to the right of the dashed line would
447 reject the null hypothesis $H_0: \beta_5 = 0$. When neither trait1 nor trait2 has phylogenetic signal (trait1: **I**;
448 trait2: **I**), LMM and PLMM estimates have similar variability (horizontal axis) and similar estimated
449 standard errors (vertical axis). When trait1 has phylogenetic signal (trait1: **C**; trait2: **I**), both LMM and
450 PLMM estimates become less variable, which is correctly captured by decreasing estimates of standard
451 errors, leading to appropriate type I error rates. The dashed line has intercept of zero and slope of 1/1.96.
452 Simulations were performed with 50 species; the fractions of simulations rejecting the null hypothesis
453 (text in the panels) were calculated from 1000 simulations, of which only 300 are presented for clarity.

454