

1 **Frequency of disturbance alters diversity, function, and underlying**
2 **assembly mechanisms of complex bacterial communities**

3

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19 **Abstract**

20 Disturbance is known to affect ecosystem structure, but predicting its outcomes remains
21 elusive. Similarly, community diversity is believed to relate to ecosystem functions, yet the
22 underlying mechanisms are poorly understood. Here, we tested the effect of disturbance on
23 the structure, diversity, and ecosystem function of complex microbial communities within an
24 engineered system. We carried out a microcosm experiment where activated sludge
25 bioreactors were subjected to a range of disturbances in the form of a toxic pollutant, tracking
26 changes in ecosystem function. Microbial communities were assessed by combining distance-
27 based methods, general linear multivariate models, α -diversity indices, and null model
28 analyses on metagenomics and 16S rRNA gene amplicon data. A stronger temporal decrease
29 in α -diversity at the extreme, undisturbed and press-disturbed, ends of the disturbance range
30 led to a hump-backed pattern, with the highest diversity found at intermediate levels of
31 disturbance. Undisturbed and press-disturbed levels displayed the highest community and
32 functional similarity across replicates, suggesting deterministic processes were dominating.
33 The opposite was observed amongst intermediately disturbed levels, indicating stronger
34 stochastic assembly mechanisms. Tradeoffs were observed in community function between
35 organic carbon removal and both nitrification and biomass productivity, as well as between
36 diversity and these functions. Hence, not every ecosystem function was favoured by higher
37 community diversity. Our results show that the assessment of changes in diversity, along with
38 the underlying stochastic-niche assembly processes, is essential to understanding the impact
39 of disturbance in complex microbial communities.

40 **Importance**

41 Microbes drive the Earth's biogeochemical cycles, yet how they respond to perturbations like
42 anthropogenic pollutants is poorly understood. As human impact continues to increase
43 worldwide, foreseeing how disturbances will affect microbial communities and the ecosystem
44 services they provide is key for ecosystem management and conservation efforts. Employing
45 laboratory-scale wastewater treatment bioreactors, this study shows that changes in
46 community diversity accompany variations in the underlying deterministic-stochastic
47 assembly mechanisms. Disturbances could promote stochastic community structuring, which
48 despite harboring higher diversity could lead to variable overall function, possibly explaining
49 why after similar perturbations the process outcome differs. A conceptual framework, termed
50 the 'intermediate stochasticity hypothesis' is proposed to theoretically predict bacterial
51 community shifts in diversity and ecosystem function, given a range of possible disturbance
52 types, in a well-replicated time-series experiment. Our findings are relevant for managing
53 complex microbial systems, which could display similar responses to disturbance, like
54 oceans, soils or the human gut.

55 **Introduction**

56 Understanding what drives patterns of community succession and structure remains a
57 central goal in ecology (1, 2) and microbial ecology (3), especially since community diversity
58 and assembly are thought to regulate ecosystem function (4, 5). The factors influencing the
59 balance between mechanisms of community assembly are under debate and require studies
60 across a range of ecosystems (6, 7). Assembly processes can be either stochastic, assuming
61 that all species have equal fitness and that changes in structure arise from random events of
62 ecological drift (8), or deterministic, when communities form as a result of niche diversity
63 shaped by abiotic and biotic factors (9). Deterministic and stochastic assembly dynamics
64 have been proposed to simultaneously act in driving assembly patterns observed in nature
65 (10-14). This has stimulated scientific discourse including modelling of experimental data
66 (15-18) and both observational and manipulative experimentation in a variety of ecosystems,
67 like deserts on a global scale (19), groundwater (6), subsurface environments (2, 20, 21), soil
68 plant-fungi associations (22), rock pools (23), water ponds (24), and sludge bioreactors (7,
69 17, 25). These prior studies emphasized the need to understand what governs the relative
70 balance between stochastic and deterministic processes and what conditions would lead to
71 stochastic processes overwhelming deterministic processes, particularly under disturbance
72 (21). To investigate their roles well-replicated time series experiments are needed (6, 25).

73 Disturbance is defined in ecology as an event that physically inhibits, injures or kills
74 some individuals in a community, creating opportunities for other individuals to grow or
75 reproduce (26). Under disturbance, organisms may benefit from being able to grow,
76 reproduce, and interact with other members of the community in the absence of specific
77 familiarity with the environment. It is deemed a main factor influencing variations in species
78 diversity (27) and structuring of ecosystems (28, 29), but a clear understanding of its
79 outcomes is lacking (30). Particularly, the intermediate disturbance hypothesis (IDH) (31)

80 predicts that diversity should peak at intermediate levels of disturbance due to trade-offs
81 between species' ability to compete, colonize ecological niches, and tolerate disturbance. The
82 IDH has been influential in ecological theory, as well as in management and conservation
83 (32), but its predictions do not always hold true (27, 33). For example, in soil and freshwater
84 bacterial communities different patterns of diversity were observed with increasing
85 disturbance frequency with biomass destruction (34) and removal (35) as disturbance type,
86 respectively. Meanwhile, the effect of varying frequencies of non-destructive disturbances on
87 bacterial diversity remains unknown. Furthermore, the IDH predicts a pattern but it is not a
88 coexistence mechanism as it was originally purported to be (36). Hence, its relevance is being
89 debated (37, 38) with multiple interpretations and simplicity as the main points of critique. To
90 date, the mechanisms behind the observed patterns of diversity under disturbance remain to
91 be elucidated (39, 40).

92 The objective of this work was to test the effect of disturbance on the bacterial
93 community structure, diversity, and ecosystem function of a complex bacterial system, with
94 emphasis on the underlying assembly mechanisms. We employed sequencing batch
95 bioreactors inoculated with activated sludge from an urban wastewater treatment plant, in a
96 laboratory microcosm setup with varying frequencies of augmentation with toxic 3-
97 chloroaniline (3-CA) as disturbance. Chloroanilines are toxic and carcinogenic compounds
98 and few bacteria encode the pathways to degrade 3-CA (41), which is also known to inhibit
99 both organic carbon removal and nitrification in sludge reactors (42). Microcosm studies are
100 useful models of natural systems (43), can be coupled with theory development to stimulate
101 further research (44), and by permitting easier manipulation and replication can allow
102 inference of causal relationships (45) and statistically significant results (46).

103 We analyzed changes in ecosystem function over time by measuring removal of organic
104 carbon, ammonia, and 3-CA, as well as biomass. Changes in community structure were

105 examined at different levels of resolution using a combination of metagenomics sequencing
106 and 16S rRNA gene fingerprinting techniques. We also explored how diversity was related to
107 function, focusing on tradeoffs. Furthermore, the role of stochasticity on community
108 assembly was investigated by employing null model techniques from ecology.

109

110 **Results**

111 *Overall community dynamics and differentiation of clusters*

112 Bacterial community structure displayed temporal changes and varied between
113 disturbance levels (Fig. 1). Constrained ordination showed a defined cluster separation with
114 0% misclassification error of the outermost levels L0 and L7 from the remaining intermediate
115 levels L1-6 (Fig. 1A). Overall community structure differed over time with a dispersion
116 effect after 14 days (Fig. 1B). Levels across disturbance and time factors showed significant
117 differences (PERMANOVA $P = 0.003$, Table S1), with a non-significant interaction effect (P
118 $= 0.15$). Disturbance was the factor responsible for the observed clustering (Fig. 1A),
119 independent of dispersion heterogeneity (PERDIMSP $P = 0.35$).

120 *Ecosystem function dynamics and trade-offs*

121 The undisturbed community (L0) was the only one with complete organic carbon
122 removal and nitrification, while the press-disturbed community (L7) was the only one that
123 could never nitrify and also had the lowest biomass (Fig. 3). Initially, reactors at the disturbed
124 levels showed an inability to remove all of the 3-CA (with the exception of L1). Such lack of
125 3-CA degradation was accompanied by a reduction in organic carbon removal in the first
126 three weeks (Fig. 2A, Fig. S3A-C), and a complete inhibition of nitrification with subsequent
127 accumulation of ammonium (Fig. 2B, Fig. S3F-H). Removal of 3-CA recovered and was
128 above 95% for all disturbed levels after 28 days (Fig. S3D), but COD removal was still not
129 100% despite complete 3-CA removal towards the end of the experiment (Figure 2C).

130 Nitrification was detected on day 21 for L1 and later for other disturbance levels, except
131 for the press-disturbed L7. The dominant NO_x component was nitrite, but some nitrate was
132 also produced (Fig. 2D, Fig. S2H-J). At the end of the study, there was a strong negative
133 correlation between organic carbon removal and nitrification (Fig. S3A-B) in terms of nitrite
134 production ($r = -0.982$, $P = 0.003$) and nitrate production ($\rho = -0.697$, $P = 0.003$). Biomass
135 values on day 35 differed significantly among levels with the highest value at L1 and the
136 lowest at L7 (Fig. 2C). There was a strong negative correlation between biomass and organic
137 carbon removal ($r = -0.418$, $P = 0.042$), and positive between biomass and partial nitrification
138 to nitrite ($r = 0.488$, $P = 0.022$) (Fig. S3C-E).

139 *Intermediate levels of disturbance displayed increased dissimilarity with time*

140 To distinguish the effect of disturbance from temporal community dynamics (Fig. 1),
141 community assembly was assessed at each time point by ordination analysis using PCO (Fig.
142 3A-B), NMDS, and CAP with cluster similarity analysis (Fig. S4). The combination of
143 constrained and unconstrained ordination methods allowed differentiating location from
144 dispersion effects in community assembly (47). L0 was consistently different in all ordination
145 plots and L7 differed after 21 days, both with 0% misclassification error at all time points for
146 CAP plots. Dispersion effects within intermediate levels were evident in the unconstrained
147 ordination plots with higher differentiation of biological replicates after 35 days (Fig. 3B),
148 coinciding with the production of nitrite and low levels of nitrate (Fig. 2D). Community
149 differentiation was statistically significant from day 21 onwards as supported by multivariate
150 tests (Table S1).

151 *Metagenomics community analysis validates observations from fingerprint dataset*

152 β -diversity patterns observed from 16S rRNA gene amplicon T-RFLP data on day 35
153 were significantly similar to those from shotgun metagenomics data. A Mantel test on Bray-

154 Curtis distance matrixes for both datasets ($n = 24$) yielded significant similarity ($r = 0.73$, $P =$
155 0.002). Procrustes tests of comparisons within ordination methods of PCO (Fig. 3C) and
156 NMDS also yielded significant similarities for both datasets ($P = 0.002$, Table S1).
157 Multivariate PERMANOVA tests on the metagenomics dataset produced statistically
158 significant results, but with significant heteroscedasticity as shown by PERMDISP (Table
159 S1). We resolved these mean-variance relationship concerns by running a General Linear
160 Multivariate Models (GLMMs) test to fit the data to a negative binomial distribution. Both
161 residuals vs fitted and mean-variance plots supported the choice of a negative binomial
162 distribution for the regression model (Fig. S5). The analysis of deviance of the regression
163 rejected the null hypothesis of no difference between communities at different disturbance
164 levels, independent of heteroscedasticity ($p = 0.0149$).

165 *Higher α -diversity for intermediately disturbed treatments and its trade-offs with function*

166 The observed patterns in α -diversity were time dependent, as diversity decreased over
167 time with respect to the initial sludge inoculum (Fig. 4A). Such temporal decrease in diversity
168 was higher at the extreme ends of the disturbance range, resulting in a parabolic pattern on
169 day 35 (Fig. 4B-C). The final α -diversity pattern based on Hill number 2D was similar for
170 both T-RFLP and Metagenomics methods (Fig. 4B), although the latter showed higher
171 variability. For the metagenomics dataset we also calculated the lower-order Hill numbers
172 (0D , 1D) which give higher weight to less abundant OTUs. They displayed the same parabolic
173 pattern (Fig. 4C). Welch's ANOVA tests were statistically significant for all Hill numbers (P
174 < 0.01 , $P = 0.022$ for ${}^2D_{\text{Metagenomics}}$). Additionally, there were strong correlations between α -
175 diversity and community function (Fig. S6), focusing on the more robust estimators of
176 microbial diversity 1D and 2D (48). Both 1D and 2D correlated positively with nitrification
177 and biomass productivity ($r > 0.44$, $P < 0.03$), but negatively with organic carbon removal (r
178 < -0.45 , $P < 0.03$).

179 *Null model analyses validate variations in β -diversity and temporal increase in stochastic*
180 *community assembly*

181 To test if the observed changes in β -diversity (Fig. 1A, Fig. 3) were due to underlying
182 stochastic and deterministic mechanisms or due to changes in α - and γ -diversity ratios ($\alpha:\gamma$)
183 alone (49), we employed a null model analysis on the bacterial genus-level metagenomics
184 and T-RFLP datasets on day 35 (Fig. 5A). Linear regression analyses between the calculated
185 β -deviation and the observed γ -diversity yielded non-significant results ($P_{\text{Metagenomics}} = 0.42$,
186 $P_{\text{T-RFLP}} = 0.31$), confirming that observed changes in β -diversity were not only due to changes
187 in $\alpha:\gamma$.

188 Furthermore, another null model analysis was employed to test whether the increase in
189 community dissimilarity observed over time (Fig. 1B, Fig. 3) was related to an increase in
190 stochastic effects driving community assembly. A temporal analysis on the T-RFLP
191 community dataset ($n = 96$) employing the Raup-Crick dissimilarity metric (50, 51) showed
192 that communities became closer to the null expectation over time, with the greater effect of
193 stochasticity observed after 35 days (Fig. 5B). The observed differential heteroscedasticity
194 among disturbance levels was statistically significant (PERMDISP $P = 0.003$).

195

196 **Discussion**

197 *Deterministic and stochastic patterns of assembly amongst different disturbance levels*

198 Niche-structuring at both ends of the disturbance range was suggested by community
199 assembly patterns and ecosystem function. The undisturbed (L0) and press-disturbed (L7)
200 levels were distinct from each other as well as from the remaining intermediate levels, as
201 supported by multivariate tests (both distance-based and GLMMs). The ordination plots and
202 cluster analyses showed a clear separate clustering for the independent replicates of these two
203 disturbance levels along the experiment, and particularly the constrained ordination plots

204 displayed this with 0% misclassification error. Furthermore, community function was clearly
205 differentiated between L0 and L7, as well as being consistent across replicates at each level.
206 We contend that the observed clustering is an indication that both the undisturbed and press-
207 disturbed levels favoured deterministic assembly mechanisms, where the selective pressure
208 due to unaltered succession (L0) or sustained toxic-stress (L7) promoted species sorting
209 resulting in similar community structuring among biological replicates over the course of the
210 experiment.

211 Conversely, the communities from intermediately disturbed levels (L1-6) did not form
212 distinct clusters for any particular level through the experiment. Within-treatment
213 dissimilarity among same-level replicates increased over time, with some replicates being
214 more similar to those of other intermediate levels. Concurrently, ecosystem function
215 parameters also displayed within-treatment variability for L1-6. For example, the conversion
216 of ammonia to NO_x products, which was initially hampered when communities were still
217 adapting to degrade 3-CA, was not the same across all equally handled independent
218 replicates. The observed divergence across independent replicates is considered here as
219 strong indicator of stochasticity in community assembly. Other aspects might promote
220 stochastic assembly, like strong feedback processes (52) that are linked to density
221 dependence and species interactions (53), priority effects (3), and ecological drift (54).
222 Nonetheless, the observed increasing role of stochasticity over time was also supported by
223 null model analyses, which also validated that the observed changes in β -diversity were not
224 due to changes in α : γ diversity ratios alone (49).

225 We argue that there were different underlying neutral-niche mechanisms operating in
226 the resulting community assembly along the disturbance range of our study. Similarly, a
227 study on groundwater microbial communities (6) found through null model analysis that both
228 deterministic and stochastic processes played important roles in controlling community

229 assembly and succession, but their relative importance was time dependent. The greater role
230 of stochasticity we found on day 35 concurred with higher observed variability in community
231 function and structure among replicates for intermediately-disturbed levels. Likewise,
232 previous work on freshwater ponds tracking changes in producers and animals (51) found β -
233 diversity (in terms of dissimilarity) increasing with stochastic processes. These observed
234 patterns are also in accordance with ecological studies proposing deterministic and stochastic
235 processes balancing each other to allow coexistence (12), with communities exhibiting
236 variations in the strength of stabilization mechanisms and the degree of fitness equivalence
237 among species (11). Thus, it is not sufficient to ask whether communities mirror either
238 stochastic or deterministic processes (10), but also necessary to investigate the combination
239 of such mechanisms that in turn explain the observed community structures along a
240 continuum (11).

241 *Diversity-disturbance patterns and trade-offs with function*

242 We observed the highest α -diversity at intermediate levels as predicted by the IDH (31),
243 both in terms of composition (0D) and abundances (1D , 2D). This finding is non-trivial in two
244 aspects. First, Svensson *et al.* (32) have shown that most studies find support for the IDH
245 when using species richness (0D) rather than evenness or other abundance-related indices
246 (like 1D and 2D). They called for the use of logical arguments to support the idea that peaks in
247 compound diversity-indices should be also expected at intermediate levels of disturbance.
248 Second, the use of richness for microbial communities is not reliable (48) since it is heavily
249 constrained by the method of measurement (55), which makes it hard to compare results from
250 different studies using this metric. Hence, for microbial systems, it is more reasonable to
251 assess diversity in terms of more robust compound indices rather than richness, reason why
252 we focused on 1D and 2D for diversity-function analyses.

253 Importantly, the observed pattern in α -diversity was time dependent and resulted in an
254 IDH pattern after 35 days. Temporal dynamics were expected since the sludge community
255 experienced an initial perturbation in all reactors after transfer from a wastewater treatment
256 plant to our microcosm arrangement. For the sludge inoculum this implied changes in reactor
257 volume, frequency of feeding (continuous to batch), type of feeding (sewage to complex
258 synthetic media), immigration rates (open to closed system), and mean cell residence time
259 (low to high). This was a succession scenario in which communities had to adapt to such
260 changes along with the designed disturbance array. For L0 and L7, 2D decreased over time in
261 agreement with niche-dominated processes, probably because such levels represented the
262 most predictable environments within our disturbance range. In contrast, intermediate levels
263 either increased or maintained the same 2D over time (after an initial decrease within the first
264 two weeks), seemingly a case where niche overlap promoted stochastic assembly (10). The
265 emergence of a IDH pattern after time is coherent with findings in previous microcosm
266 studies using synthetic protist (56) and freshwater enrichment (35) microbial communities.
267 Yet, none of these studies evaluated the relative importance of the underlying assembly
268 mechanisms of assembly on the observed diversity dynamics.

269 Additionally, both 1D and 2D were positively correlated with nitrification and
270 productivity, suggesting that higher community evenness favours functionality under
271 selective pressure (57), but were negatively correlated with organic carbon removal. Thus, we
272 cannot affirm that more diverse communities have better functionality without considering
273 tradeoffs. This supports the notion that higher α -diversity does not necessarily imply a
274 “better” or “healthier” system (55). In addition to the observed OTU diversity, there was an
275 evident ecosystem function diversity along the disturbance range studied, a similar finding to
276 that of previous studies with simpler planktonic communities (58).

277 Functional tradeoffs are expected under disturbance since organisms need to allocate
278 resources normally used for other functions to recover after a disturbance (59). In our study,
279 communities with higher biomass had lower organic carbon removal efficiencies, which
280 together with the tradeoffs described for nitrification, suggest the adoption of different
281 community life-history strategies depending on the frequency of disturbance. The results
282 presented here were all taxonomy-independent since our focus was on diversity, function, and
283 mechanisms of community assembly (phylum-level community changes are provided as
284 supplemental material Fig. S7). Taxonomy-independent approaches continue to be useful to
285 describe diversity patterns and mechanisms of community assembly (2, 60). However, it has
286 been proposed that species' traits can predict the effects of disturbance and productivity on
287 diversity (61). Hence, further analysis of the different taxa and their genetic potential paired
288 with the observed tradeoffs in community function will aid in the understanding of potential
289 life-history strategies (59) and their relationship with community aggregated traits (62) in the
290 near future.

291 *Merging mechanisms of community assembly and alpha-diversity patterns: an intermediate*
292 *stochasticity hypothesis*

293 Knowing that the validity of the IDH is still under debate (37, 38) and that many
294 different diversity-disturbance patterns have been reported (27, 30, 33), we asked whether
295 there is a relationship between the peaked pattern in diversity observed and the underlying
296 neutral-niche processes of community assembly. Under purely neutral processes, diversity
297 should vary randomly as all species have equal fitness (63), unless some other mechanism
298 acts to prevent this. We hypothesized that higher α -diversity at intermediate disturbance
299 frequencies is the result of weaker stabilizing mechanisms (niches), which are stronger at
300 extreme ends of the disturbance range. Neutral (stochastic) mechanisms will produce even
301 assemblages (higher α -diversity) at intermediately disturbed levels, whilst infrequent or too-

302 frequent disturbances will favour some species over others (lower α -diversity). We propose
303 this idea as the Intermediate Stochasticity Hypothesis (ISH, Fig. 6) and contend that it should
304 hold particularly for compound α -diversity indices (48), since the underlying assembly
305 mechanisms would affect taxa abundance distributions.

306 It is recognized that, beyond empirical pattern description, an understanding of the
307 underlying mechanisms is necessary to comprehend the outcomes of intermediate disturbance
308 regimes (30, 40, 64). Furthermore, microbes constitute a special case since they evolve
309 quickly, compared with plants and animals microbial turnover is rapid, and selection can
310 cause very rapid shifts in community structure (65, 66). The complementary role of
311 stochastic-niche processes could then be one of the key substantive issues promoting species
312 coexistence (37, 39). This can foster research on disturbance-diversity-relationships by
313 complementing the already proposed modifications to the competition-colonization tradeoffs
314 under succession on which the IDH was originally based (31, 38), like considering effects of
315 competitive exclusion (56, 67), productivity (68, 69), spatial heterogeneity (70), feedbacks
316 between diversity and disturbance (71), and type of α -diversity metric employed (32).

317 *Implications and concluding remarks*

318 The implications of this study relate to both process engineering and environmental
319 management. Sludge communities within wastewater treatment are not only model systems in
320 microbial ecology (72), but also a key driver for water sanitation and the environmental
321 impact of anthropogenic water discharges (73). Disturbances could promote stochastic
322 assemblages of the sludge communities, which despite harbouring higher diversity could lead
323 to variable overall community function. This could be the reason why after similar
324 perturbations the process outcomes differ, causing operational problems for water utilities
325 (74). Furthermore, cases where disturbance temporally favours stochastic assembly could

326 lead to a different final community after the perturbation (29), which could compromise the
327 expected ecosystem function. More research is needed to identify such scenarios in practice.

328 We described how different frequencies of disturbance affected ecosystem function and
329 bacterial community diversity and assembly. Communities were assessed through different
330 molecular methods that nonetheless yielded very similar patterns. Furthermore, besides the
331 wastewater treatment microbial community, other complex microbial systems (e.g. gut
332 microbiome) might display similar responses to disturbance. We argue that changes not only
333 in diversity but also in the underlying deterministic-stochastic assembly mechanisms should
334 be evaluated in studies of the effects of disturbance on such systems. For such an assessment,
335 both replication and wide-enough disturbance ranges are key. This calls for more studies in
336 microcosm (45, 75) and mesocosm settings, as well as meta-analysis from full-scale
337 application studies.

338 Finally, the ISH could conceptually predict bacterial community shifts in diversity and
339 ecosystem function, given a range of possible disturbance types, in a well-replicated time-
340 series experiment, like it did in this study. Since microbes drive the Earth's biogeochemical
341 cycles (76), its potential application in biodiversity conservation efforts and the built
342 environment, should encourage further testing under different disturbance dimensions and
343 microbial systems.

344

345 **Materials and Methods**

346 *Experimental design*

347 We employed sequencing batch microcosm bioreactors (20-mL working volume)
348 inoculated with activated sludge from a full-scale plant (Text S1) and operated for 35 days.
349 The daily complex synthetic feed included toxic 3-CA at varying frequencies. Eight levels of
350 disturbance were set in triplicate independent reactors ($n = 24$), which received 3-CA every

351 day (press-disturbed), every two, three, four, five, six, or seven days (intermediately-
352 disturbed), or never (undisturbed). Level numbers were assigned from 0 to 7 (0 for no
353 disturbance, 1 to 7 for low to high disturbance frequency, Fig. S1). Ecosystem function, in the
354 form of process performance parameters, was measured weekly in accordance with Standard
355 Methods (77) where appropriate, and targeted chemical oxygen demand (COD), nitrogen
356 species (ammonium, nitrite, and nitrate ions), volatile suspended solids (VSS), and 3-CA. On
357 the initial day and from the second week onwards, sludge samples (1 mL) were collected
358 weekly for DNA extraction.

359 *Microbial community analysis and statistical tests*

360 All reported p-values for statistical tests in this study were corrected for multiple
361 comparisons using a False Discovery Rate (FDR) of 10% (78). Community assembly was
362 assessed by a combination of ordination methods (PCO, NMDS, CAP) and multivariate tests
363 (PERMANOVA, PERMDISP) (79) on Bray-Curtis dissimilarity matrixes constructed from
364 square-root transformed normalized abundance data using PRIMER (v.7). Additionally,
365 general linear multivariate models (GLMMs), which deal with mean-variance relationships
366 (80), were employed using the *mvabund* R package (81) fitting the metagenomics dataset to a
367 negative binomial distribution, to ensure that the observed differences among groups were
368 due to disturbance levels and not heteroscedasticity. The 500 most abundant genera (97% of
369 total reads abundances) were employed to ensure random distribution of residuals fitted in the
370 model. Significance was tested using the *anova* function in R with PIT-trap bootstrap
371 resampling ($n = 999$) (82). Hill diversity indices (83) were employed to measure α -diversity
372 as described elsewhere (48, 84), and calculated for normalized non-transformed relative
373 abundance data (Text S1).

374 Abundance data from Terminal Restriction Fragment Length Polymorphism (T-RFLP)
375 analysis of the bacterial 16S rRNA gene (using 530F-1050R primer set targeting V4-V5

376 regions) were employed in combination with shotgun metagenomics sequencing at the genus
377 taxonomic level. Such an approach is valid for the questions asked in this study, since
378 comparisons between NGS and fingerprinting techniques support the use of T-RFLP to detect
379 meaningful community assembly patterns and correlations with environmental variables (60),
380 and such patterns can be validated by NGS on a subset of the fingerprinting dataset (2).

381 *Comparison between Metagenomics and T-RFLP community datasets*

382 Mantel and Procrustes tests (85) were applied to compare metagenomics and T-RFLP
383 datasets from all bioreactors on day 35 (n = 24, subsample of the full T-RFLP dataset).
384 Briefly, Bray-Curtis dissimilarity matrixes were computed using square root transformed T-
385 RFLP data and bacterial genus-level taxa tables generated using a metagenomics approach.
386 Mantel tests were then used to determine the strength and significance of the Pearson
387 product-moment correlation between complete dissimilarity matrices. Procrustes tests
388 (PROTEST) were also employed as an alternative approach to Mantel tests in order to
389 compare and visualize both matrices on PCO and NMDS ordinations. The resultant m2-value
390 is a statistic that describes the degree of concordance between the two matrices evaluated
391 (86). All these statistical tests were performed using the *vegan* R package (functions:
392 *procuste, mantel, metaMDS, vegdist*).

393 *Metagenomics sequencing and reads processing*

394 Around 173 million paired-end reads were generated in total and 7.2 ± 0.7 million
395 paired-end reads on average per sample. Illumina adaptors, short reads, low quality reads or
396 reads containing any ambiguous base were removed using cutadapt (-m 50 -q 20 --max-n 0,
397 v.1.11) (87). Taxonomic assignment of metagenomics reads was done following the method
398 described by Ilott *et al.* (88). High quality reads ($99.2 \pm 0.09\%$ of the raw reads) were
399 randomly subsampled to an even depth of 12,395,400 for each sample prior to further

400 analysis. They were aligned against the NCBI non-redundant (NR) protein database (March
401 2016) using DIAMOND (v.0.7.10.59) with default parameters (89). The lowest common
402 ancestor approach implemented in MEGAN Community Edition v.6.5.5 (90) was used to
403 assign taxonomy to the NCB-NR aligned reads with the following parameters:
404 maxMatches=25, minScore=50, minSupport=20, paired=true. On average, 48.2% of the high-
405 quality reads were assigned to cellular organisms, from which in turn 98% were assigned to
406 the bacterial domain. Adequacy of sequencing depth was corroborated with rarefaction
407 curves at the genus taxonomy level using MEGAN (90). We did not include genotypic
408 information because it was outside the scope of our study, but will do so in future
409 investigations arising from this work.

410 *Null Model Analyses on Diversity*

411 β -diversity is normally seen as the variation in composition of communities among sites
412 (91), although there are several definitions for it (92). It is also known that most changes in β -
413 diversity are determined by variation in the ratio of local (α) and regional (γ) diversity. Thus,
414 to disentangle the roles of stochastic and deterministic processes as drivers of change in β -
415 diversity it is necessary to incorporate statistical null models in the analysis (51, 91), which
416 assume that species interactions are not important for community assembly (93). We
417 employed two null model approaches previously tested in community ecology (49, 51), and
418 more recently for microbial communities (6). To adapt them to handle microbial community
419 data, we considered species as OTUs (i.e. genus for metagenomics and T-RFs for T-RFLP
420 datasets) and each individual count as one read within the metagenomics dataset, or one
421 peak-area unit within the T-RFLP dataset. The first null model approach used here, originally
422 applied to woody plants (49), randomizes the location of each individual within the three
423 independent reactors for each of the eight disturbance treatment levels, while maintaining γ -
424 diversity, the total quantity of individuals per reactor, and the relative abundance of each

425 OTU. The second null model uses a modified Raup-Crick dissimilarity metric (β_{RC}), which
426 indicates the degree to which the observed number of shared OTUs between any two
427 communities deviates from the expected number of shared OTUs in the null model. The
428 closer to the null expectation, the stronger the stochastic effects on the community assembly
429 are, and vice versa (51). This method is robust for variation in α -diversity among
430 communities.

431 **Sequence data and metadata**

432 The microbial DNA metagenomics sequencing datasets supporting the results in this
433 article is available at NCBI BioProjects with accession number: 389377.

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667 **Figure Legends**

668 **FIG 1** Ordination plots differentiating undisturbed (L0) and press-disturbed (L7) levels from
669 intermediately disturbed levels (L1-6) and displaying temporal community dispersion effects
670 for all time points. (A) Canonical Analysis of Principal coordinates (CAP, constrained
671 ordination) plot, with disturbance levels as differentiation criteria, shows niche differentiation
672 for L0 (CAP1 axis) and L7 (CAP2 axis). Disturbance levels: L0[▲], L1[▼], L2[◻], L3[◻], L4[◻]
673 [◻], L5[+], L6[■], and L7[*]. (B) Non-metric Multidimensional Scaling (NMDS,
674 unconstrained ordination) shows dispersion effect after a given number of days. Days: 14[△],
675 21[▽], 28[■], and 35[◆].

676 **FIG 2** Process performance indicators across disturbance levels. Effects include temporal
677 changes and tradeoffs in community function. (A,C) Percentage of organic carbon as
678 chemical oxygen demand (COD, ●) and 3-CA (◇) removal for all levels (negative values
679 represent accumulation). (C) Biomass as volatile suspended solids. (VSS, ◻). (B,D)
680 Concentration of ammonium (◆), nitrite (△), and nitrate (○) as nitrogen for all levels. Data
681 are from days 7 (A-B) and 35 (C-D) of the study (for all time points sampled, see Fig. S2).
682 Mean ± s.d. (n = 3) are shown. Undisturbed L0 replicates had consistent organic carbon
683 removal and complete nitrification, whereas press-disturbed L7 never showed nitrification
684 and had the lowest final biomass. Intermediate levels L1-6 displayed changing functionality
685 with higher s.d. values that increased over time.

686 **FIG 3** Community assembly as assessed by Principal Coordinates Analysis (PCO) plots for
687 all disturbance levels on T-RFLP datasets on days (A) 14 and (B) 35 of the study. Ovals with
688 dashed lines represent 80% similarity calculated by group average clustering. Disturbance
689 levels: L0 [▲], L1 [▼], L2 [◻], L3 [◻], L4 [■], L5 [+], L6 [■], and L7 [*]. (C) Procrustes
690 analysis on PCO at day 35 comparing metagenomics (circles) and T-RFLP (triangles)

691 datasets. Lines unite data points from the same reactor ($n = 24$). Same colour palette as for
692 disturbance levels. Tests comparing both methods were statistically significant (Table S1).
693 Intermediate treatments' (L1-6) within-treatment dissimilarity increased with time. L0 and L7
694 clusters consistently displayed higher similarity after 14 days.

695 **FIG 4** α -diversity patterns based on T-RFLP and metagenomics data. (A) Temporal
696 dynamics of Hill number 2D for abundant OTUs, calculated from T-RFLP data across
697 disturbance levels. (B) Hill number 2D calculated from T-RFLP (black dashed bars) and
698 metagenomics (grey solid bars) data at days 0 (seed) and 35. (C) Hill numbers 0D (black solid
699 bars) and 1D (blue solid bars) from metagenomics data on days 0 (seed) and 35. Values
700 represent mean \pm s.d. ($n = 3$).

701 **FIG 5** Null model analyses of community data. (A) β -deviation versus observed γ -diversity
702 for all disturbance treatment levels on day 35. Each point involved all replicates ($n = 3$) of
703 each disturbance level ($a = 8$) using the model from (49) on the metagenomics (closed circles,
704 bacterial genus taxonomical level), and T-RFLP (open squares) datasets. Non-significant
705 linear regression tests confirmed that changes in β -diversity did not arise from changes in α : γ
706 diversity alone, suggesting other mechanisms underlying community assembly. (B) NMDS
707 ordination plot of community dynamics of common OTUs (T-RFLP dataset, $n = 96$) using
708 the Raup-Crick dissimilarity metric as in (51). Days: 14[▲], 21[▼], 28[■], and 35[◆]. Lines
709 denote the minimum convex shapes around the data for each disturbance level. Points that are
710 further separated represent communities less deviant from the null expectation, whereas those
711 that are closer together are more divergent from the null expectation. Communities on day 35
712 are nearer to the null expectation compared with the ones on previous days, suggesting an
713 increasing role over time of stochastic processes in driving community assembly. Overall,

714 null model analyses validate variations in β -diversity and temporal increase of stochastic
715 community assembly

716 **FIG 6** Intermediate Stochasticity Hypothesis (ISH) for community assembly under varying
717 disturbances. Conceptual representation of the classic relationship between α -diversity and
718 disturbance (31), including the effect of underlying stochastic and deterministic processes
719 driving bacterial community assembly. When intermediate disturbance regimes result in less
720 predictable environments, specialized traits would be advantageous to taxa, and the stochastic
721 equalization of competitive advantages would lead to higher α -diversity. On the contrary,
722 extreme ends of the range where conditions are recurrent would select for adapted organisms
723 whose dominance would result in a lower α -diversity.











