

1 **The soil microbial food web revisited with metatranscriptomics - predatory**

2 ***Myxobacteria* as keystone taxon?**

3

4 Sebastian Petters¹, Andrea Söllinger^{1,2}, Mia Maria Bengtsson¹, Tim Urich^{1,*}

6 *¹Institute of Microbiology, Center for Functional Genomics of Microbes, University of*

7 *Greifswald, Greifswald, Germany*

8 *²Department of Ecogenomics and Systems Biology, University of Vienna, Vienna,*

9 *Austria*

10 ** corresponding author*

11 ***email: tim.urich@uni-greifswald.de***

12 **Abstract**

13 Trophic interactions in the microbial food web of soils are crucial for nutrient and
14 carbon cycling. Traditionally, protozoa are considered the major micropredators of
15 bacteria in soil. However, some prokaryotes, such as *Myxobacteria* and *Bdellovibrio*
16 are also famous for bacterivorous life style. Until recently, it was impossible to assess
17 the abundance of pro- and eukaryotic micropredators in soils simultaneously. Using a
18 metatranscriptomic three-domain profiling of small subunit ribosomal RNA we
19 investigated the abundance of bacterivores in 28 datasets from eleven European
20 mineral and organic soils of different climatic zones. In all soils, *Myxobacteria*
21 comprised a significant proportion from 4 – 19% of prokaryotic 16S rRNA transcripts
22 and more than 60% of all bacterivores in most soils. *Haliangiaceae* and
23 *Polyangiaceae* were most abundant, while the name-giving *Myxococcaceae* were
24 barely present. Other bacterial predators like *Bdellovibrio* were low abundant. Also
25 Protozoan micropredator 18S rRNA transcripts, e.g. from *Cercozoa*, *Amoebozoa* and
26 *Ciliophora*, were on average less abundant, especially in mineral soils. *Nematodes*
27 were even less abundant. In addition, we applied a longitudinal approach to identify
28 bacterivores during beech litter colonisation. Here, *Myxobacteria* showed prey-
29 dependent, protozoa-like community dynamics during colonisation. Thus, their broad
30 prey range and high abundance suggests a major influence of *Myxobacteria* on
31 structuring the prokaryotic community composition in soil, and might warrant their
32 classification as keystone taxon. Our results suggest the presence of an ecologically
33 important “bacterial loop” in soil food webs, independent of protozoa and nematodes.
34

35 **Introduction**

36 Predation of predators on prey is a key process in structuring community composition
37 in ecosystems and in maintaining high biodiversity. Predator - prey interactions and
38 dynamics among animals and consequences for ecosystem functioning have been
39 studied extensively since the early days of ecology. While less visible and thus less
40 acknowledged, predation is not foreign to the microbial world. Eukaryotic as well as
41 prokaryotic microorganisms are known to prey on other microorganisms in marine,
42 aquatic and terrestrial habitats as part of the microbial food web (Clarholm, 1985;
43 Azam et al., 1983).

44 Protozoa are traditionally considered the main microbial predators of bacteria, a
45 notion that stems from the fact that, unlike in bacteria, where it is somewhat “exotic”,
46 predation is a common lifestyle among protozoa. Predatory protozoa are known from
47 both aquatic and soil environments and have been considered a key-component of the
48 “microbial loop” responsible for the remineralisation of nutrients (Bonkowski, 2004;
49 Clarholm, 1985). Whereas protozoa in the aquatic system have been well
50 characterised, both in terms of their identity and population size, research in soils has
51 been much more hampered, since no adequate molecular tools have been available for
52 a long time, cultivation is often difficult, and direct microscopic observations are
53 impossible (Geisen et al., 2015).

54 Much fewer prokaryotic species are considered predatory, although a predatory
55 lifestyle in prokaryotes probably evolved prior to its development in eukaryotes.
56 Several bacterial predators have been identified, with more and more taxa exhibiting a
57 predatory lifestyle being recognized recently. These include *Myxobacteria*,
58 *Lysobacter*, *Bdellovibrio* and like organisms (BLO), *Vampirococcus*, and

59 *Dapterobacter*, among others (Reichenbach, 1999). Especially the *Myxobacteria*, with
60 their ‘wolf pack hunting’ strategy, are known micropredators since more than 70 years
61 ago and have been isolated from soils world-wide (Keane & Berleman, 2016;
62 Reichenbach, 1999).

63 It has been until recently impossible to assess bacterial and protist community
64 composition with the same methodology. Although PCR amplicon approaches
65 enabled the study of both groups separately, a direct comparison of their relative
66 abundances was not possible due to the absence of universal primers that would tackle
67 all groups without bias. However, these obstacles are avoided when applying random
68 hexamer-primed reverse transcription as in metatranscriptomics approaches that target
69 SSU rRNA of organisms from all three domains of life (Urich et al., 2008).
70 Furthermore, these rRNA transcripts are indicative of ribosomes and thus are likely
71 derived from metabolically active cells and can be considered markers for living
72 biomass. The generated cDNA fragments originate from different regions of the SSU
73 rRNA molecule unlike PCR primed specific sites, and are therefore insensitive to the
74 presence of introns or primer mismatches, when PCR primers are applied.

75 We have recently used this PCR-free metatranscriptomics approach to reveal the
76 diversity of the active soil protist communities within five different natural soil
77 systems in Europe, including forest, grassland and peat soils as well as beech litter
78 (Geisen et al., 2015).

79 Here we have focused on other groups of microbial predators - predatory bacteria. We
80 have assessed the relative abundance of SSU rRNAs from bacterial groups known to
81 exhibit a predatory lifestyle in these soils. Metatranscriptomics enabled the direct
82 comparison of SSU rRNA transcripts from bacterial and protozoan micropredators

83 and revealed that potentially predatory bacteria, especially *Myxobacteria*, were
84 abundantly detected in all soils, while protozoa abundances were much more variable.
85 The underlying causes and consequences for our perception of microbial predation in
86 soils are discussed and an alternative model of the soil microbial loop is put forward.

87

88 **Material and Methods**

89 *Data acquisition*

90 The investigated metatranscriptomes had been obtained from different previous
91 studies on a range of European soils (Table 1). These included 4 samples from organic
92 peatland, 3 samples from organic floodplain, 3 samples from gleic fluvisol, 3 samples
93 from mineral grassland, 2 samples from organic forest litter, 4 samples from mineral
94 forest soil, and 3 samples each from 3 different mineral shrubland soils. RNA, cDNA
95 and sequences were obtained as previously described (Beulig et al., 2016; Epelde et al.,
96 2015; Geisen et al., 2015; Tveit et al., 2013; Urich et al., 2008).

97 Furthermore, metatranscriptomic data were obtained from four different beech litter
98 types (K, A, O and S), which had been incubated with the same microbial community
99 in mesocosms (see Wanek et al., 2010 for details of the experimental setup). Litter
100 samples were taken at three time points: after two weeks and after three and six
101 months after inoculation, flash-frozen in liquid nitrogen and stored at -80 °C. RNA
102 was extracted and double-stranded cDNA was prepared as previously described
103 (Urich et al., 2008). 454 pyrosequencing was performed at the Norwegian Sequencing
104 center, CEFS, University of Oslo (Norway). Raw sequence data were submitted to the
105 NCBI Sequence Read Archive (SRA) under the accession number SRP134247.

106 *Bioinformatic analysis*

107 Raw sequence datasets were filtered to a minimum length of 200 nucleotides and a
108 minimum mean quality score of 25 using prinseq-lite (Schmieder & Edwards, 2011).
109 SSU rRNA sequences were identified via SortMeRNA (Kopylova et al., 2012).
110 USEARCH (Edgar, 2010) was used to randomly subsample datasets to a maximum of
111 50 000 - 100 000 sequences. The datasets were mapped against the CREST database
112 silva123.1 by blastn (Altschul et al., 1990; Lanzén et al., 2012). The obtained blastn
113 files were taxonomically analysed using MEGAN (Huson et al., 2011, min score 155;
114 top percent 2.0; min support 1). The number of SSU rRNA reads of the investigated
115 organisms was normalized in MEGAN to the total number of read counts.
116 Investigated taxa with predatory lifestyle were *Myxococcales*, *Bdellovibrionales*,
117 *Lysobacter*, *Dapterobacter*, *Vampirococcus*, *Amoebozoa*, *Cercozoa*, *Ciliophora*,
118 *Foraminifera*, *Euglenozoa*, *Heterolobosea*, and *Nematoda*. Different Nematoda taxa
119 were not investigated separately. The read counts of the analysed bacterial
120 micropredators were subtracted from the total bacterial SSU rRNA resulting in prey
121 bacterial rRNA. The read counts of each analysed bacterivorous group were then
122 normalized to the prey bacterial SSU rRNA reads.
123 Results for organic, excluding mofette (MO) samples, and mineral soils were tested
124 for differentially expressed sequences with the R package edgeR (McCarthy et al.,
125 2012; functions glmFit and glmLRT), using the non-normalized total read counts
126 MEGAN file.

127 **Results**

128 *Abundance of bacterivores in soil microbiomes*

129 We screened the SSU (16S and 18S) rRNA fraction of 28 soil metatranscriptome
130 datasets obtained from eleven different soils across Europe (Table 1) for bacterivorous
131 pro- and eukaryotes. *Myxococcales* SSU rRNA reads comprised a high proportion of
132 prey bacterial SSU rRNAs, ranging from 3.5 to 18.9% (9% on average), and higher
133 than all other investigated bacterivores (Figure 1a). Their highest proportion in
134 relation to bacteria was detected in peat soils. Additionally, an organic fluvisol and a
135 beach litter layer showed *Myxococcales* abundances above 10%. The latter came up as
136 the only exception in the pattern, i.e. here the Protozoa were the most abundant
137 bacterivorous group. Overall, SSU rRNAs of protozoa were the second most abundant
138 (Figure 1a). Like the *Myxococcales* they were generally more abundant in organic
139 soils than in mineral soils. The only two cases where their proportion was above 10%
140 of prey bacterial SSU rRNA reads were a peatland and a forest litter sample. While
141 *Myxococcales* abundance never dropped below 3.4%, protozoa abundance was much
142 lower in mineral soils (down to 0.7%). The third most abundant group were the
143 *Nematoda* (Figure 1a). They showed greater variation in abundance compared to the
144 aforementioned taxa, especially in organic soils, where they showed both their highest
145 abundance (8.5%), namely in the forest litter horizon, and also their lowest abundance
146 (< 0.1%), which occurred in the suboxic mofette soil. This was the only sampling site,
147 where their abundance dropped below 0.1% of the prey bacterial SSU rRNA reads.
148 The only other soils which showed *Nematoda* abundances above 1% were the organic
149 peatland samples and the mineral Rothamsted soil. All mineral soils showed fractions
150 of *Nematoda* SSU rRNAs within 0.1 – 1%. The *Bdellovibrionales* comprised even

151 lower SSU rRNA abundances. Similar to the aforementioned, highest relative
152 abundance of *Bdellovibrionales* was observed in organic soils (0.9% in peatland soil).
153 We did not detect *Vampirococcus* and *Dapterobacter* in any of the investigated
154 samples. *Lysobacter* comprised the lowest SSU rRNA abundances of all detected
155 micropredators, namely 0.07% or lower.

156 Looking at the composition of potential prey bacteria revealed that SSU rRNA from
157 gram-negative bacteria comprised approximately 80%, while SSU rRNA from gram-
158 positive were approximately 20%. The latter were slightly more abundant in mineral
159 soils (supplementary figure S1).

160

161 *Myxococcales* dominate bacterivorous taxa

162 Comparing all investigated bacterivorous groups, the *Myxococcales* were highest
163 abundant in every sampling site, except for the forest litter (Figure 1b). In fact, in nine
164 of the eleven sites, including all mineral sampling sites, the proportion of
165 *Myxococcales* SSU rRNAs was more than 60% of all micropredators. In the forest
166 litter their proportion of the bacterivorous groups was below 30%. Correspondingly,
167 the protozoa were the most abundant group in that site, comprising up to more than
168 half of all bacterial predators. However, in all the other sampled sites, the proportion
169 of the protozoa was below 40%, in three cases even below 20%. Those were namely
170 the organic mofette samples as well as the mineral Rothamsted site, and mine M from
171 Spain, where the lowest percentage of all micropredators was observed. All of the
172 sampling sites had *Nematoda* SSU rRNA below 20% (Figure 1b). Their highest
173 proportions occurred in the organic forest litter samples. Moreover, the only other two
174 sites where their proportions were above 10%, were Rothamsted, where they were

175 even more than the protozoa, and mineral mine H. All other sites showed proportions
176 below 10%, with the lowest proportions in samples from mofette. The
177 *Bdellovibrionales* were below 10% of micropredators in all sampling sites
178 (Figure 1b).

179

180 *Community composition of Myxobacteria*

181 We analysed the community composition of *Myxococcales* in more detail (Figure 2a).
182 The most dominant family was *Haliangiaceae*, followed by *Polyangiaceae* and
183 Blrii41, a family level group in the SILVA taxonomy that is currently devoid of
184 cultured representatives. These three together comprised more than 2/3 of
185 *Myxobacteria* SSU rRNAs in all but one site. *Haliangiaceae* and *Polyangiaceae* were
186 more abundant in mineral soils, while Blrii41 was more characteristic for organic
187 soils. The name-giving family *Myxococcaceae*, which is comprised, among others, of
188 the most frequently isolated genera *Myxococcus* and *Coralloccoccus*, was barely
189 present.

190

191 *Protozoa community composition*

192 As previously found (Geisen et al., 2015), *Amoebozoa*, *Cercozoa* and *Ciliophora* were
193 the three most abundant protist groups (Figure 2b). While *Amoebozoa* and *Cercozoa*
194 dominated in mineral soils, the *Ciliophora* were most abundant in organic soils. The
195 remaining predatory groups *Foraminifera*, *Euglenozoa*, and *Heterolobosea* accounted
196 for low abundances on average. The mofette reference (MR) samples were an
197 exception, with *Foraminifera* comprising more than 20% of protists.

198

199 *Dominance of Myxobacteria among bacterivores in mineral soils*

200 We compared the average micropredator abundance (normalized to the prey bacteria)
201 between mineral and organic soils (excluding mofette samples) based on SSU rRNA
202 reads (Figure 3). *Lysobacter* data are not shown due to low abundances. Remarkably,
203 micropredator SSU rRNAs comprised 32.3% of prey SSU rRNAs in organic soils, as
204 compared to only 7.9% in mineral soils. Although concomitantly lower in abundance
205 in mineral soils, *Myxococcales* comprised the highest micropredator proportions in
206 both soil types (16.1% in organic vs. 5.7% in mineral soil). While protozoa were
207 almost equally abundant in organic soil, they comprised approx. 1/4 of *Myxobacteria*
208 in mineral soils. In fact, the percentage of *Myxococcales* within the bacterivores was
209 remarkably higher in mineral soils, i.e. 72% compared to 50% in organic soils. Thus,
210 the decrease in abundance of *Myxococcales* in mineral soils was not as strong as seen
211 in the other micropredators.

212 To statistically verify the observed differences in abundance between organic and
213 mineral soils, we tested the data for differentially expressed SSU rRNAs of
214 micropredators. While the protozoa ($p < 0.01$), *Lysobacter* ($p < 0.01$), and
215 *Bdellovibrionales* ($p = 0.02$) were significantly differently abundant between organic
216 and mineral soils, no significant differences were detected for *Myxococcales*
217 ($p = 0.31$) and *Nematoda* ($p = 0.78$). This supports the observed phenomenon, where
218 the *Myxococcales* remained dominant in mineral soils, while the SSU rRNA
219 abundances of other investigated micropredators significantly decreased in mineral
220 soils.

221

222 *Temporal dynamics of bacterivores during community succession*

223 It has been shown that *Myxobacteria* can have a saprotrophic life style next to
224 bacterivory (reviewed in Reichenbach, 1999). We therefore analysed micropredator
225 dynamics in a litter colonisation experiment. In a longitudinal experiment four types
226 of sterilized beech litter differing in their C:N:P ratio were colonized by the same
227 microbial community taken from beech forest soil (Supplementary Table 1 in Wanek
228 et al., 2010). Metatranscriptome data were obtained from three time points over the
229 course of six months. SSU rRNA abundances of protozoa, *Nematoda*, and
230 *Myxococcales*, as well as total bacteria and fungi were assessed to investigate their
231 temporal dynamics (Figure 4). Microbiomes on litters K and S, which had a higher
232 nitrogen content, were strongly dominated by fungal SSU rRNAs as compared to
233 bacterial rRNAs, while litters A and O, which had lower nitrogen content, had higher
234 proportions of bacterial reads. The fungal:bacterial ratio stayed rather constant for
235 each litter type over time. SSU rRNAs of bacterivores generally increased in relative
236 abundance over time, especially from two weeks to three months. Remarkably, the
237 bacterivores (including *Myxococcales*) appeared earlier and in higher relative
238 abundance in litters with more prey bacteria. With few exceptions, protozoa
239 comprised the most abundant predator of bacteria, with *Myxococcales* and *Nematoda*
240 being second and third most abundant respectively. It appeared that after three months
241 a rather stable predator:prey ratio had established.

242

243 **Discussion**

244 *Metatranscriptomics-enabled holistic assessment of soil micropredators*

245 It has been until recently impossible to assess soil bacterial and protist community
246 composition with the same methodology. Although PCR amplicon approaches

247 enabled the study of both groups separately, a direct comparison of their relative
248 abundances was not possible due to the absence of universal primers that would tackle
249 all groups without bias. The rRNA fraction of metatranscriptomics data enables broad
250 three-domain community profiling of abundant bacteria, archaea, and eukaryotes via
251 rRNA (Urich et al., 2008). This innovative approach has constantly been developed
252 further and recently come to maturation due to lower-cost NGS sequencing
253 technologies and bioinformatic tools (e.g. (Bengtsson et al., 2018; Schwab et al.,
254 2014; Tveit et al., 2015)). Using this approach, we have recently created a first
255 molecular census of active protists in soils (Geisen et al., 2015). The study showed
256 that protozoa usually not detected with general PCR primers such as *Amoebozoa* and
257 *Foraminifera* are abundantly present and provided the most comprehensive picture of
258 active protist communities in soils to date. The strength of rRNA transcripts for
259 comparatively unbiased views into community composition and assessment of unseen
260 microbial diversity has recently gained popularity (e.g., Karst et al., 2018).

261

262 *Predatory Myxobacteria as key-stone taxon in mineral soils?*

263 In all investigated soils the *Myxobacteria* comprised a significant proportion of the
264 overall bacterial SSU rRNA transcripts. This confirms a recent PCR / 16S rRNA gene
265 based survey where *Myxobacteria* also comprised a substantial fraction (4.1%; Zhou
266 et al., 2014). In our study, lower micropredator abundances were detected in mineral
267 soils as compared to organic soils, which may be due to less available carbon resulting
268 in lower prey cell density in mineral soils. However, given the predatory traits of most
269 myxobacterial taxa, their high relative abundance hints towards an important role in
270 the microbial food web of soils. Moreover, with only one exception, the

271 *Myxobacteria* exhibited the highest abundances when compared to all other
272 bacterivores.

273 Traditionally, protists are considered to be the dominant group preying on bacteria
274 (e.g. Geisen et al., 2016; Trap et al., 2016). In contrast to this, our data suggest an
275 importance, possibly even dominance of *Myxococcales*. In fact, *Myxococcales*
276 comprised approx. 3/4 of all micropredators in mineral soils. Possibly, the smaller
277 pore sizes in mineral soils provide restricted access of protists to their bacterial prey,
278 as compared to the smaller *Myxococcales*. Thus, microorganisms inhabiting non-
279 continuous capillary pores could be protected from predation by Protozoa and
280 *Nematoda*, but not from the similarly-sized *Myxobacteria*. The prokaryotes inhabiting
281 the organic soil horizons, with unprotected macro-pore space, would in turn be
282 subjected to higher grazing pressure. The *Myxobacteria* and protozoa exhibit
283 fundamentally different predation strategies, with the much smaller *Myxobacteria*
284 being famous for their social 'wolf-pack' hunting combined with the secretion of lytic
285 enzymes, as compared to the larger phagotrophic protozoa (Reichenbach, 1999). The
286 more similar cell size of *Myxobacteria* and prey bacteria could thus favour
287 myxobacterial predation in mineral soils with small pores. Given the broad prey range
288 of *Myxobacteria*, their high abundance in soils suggests a major influence on
289 structuring the prokaryotic community composition, and might warrant their
290 classification as key-stone taxon.

291 Interestingly, the majority of *Myxobacteria* was not related to the well-studied and
292 easy-to-isolate *Myxococcaceae*, but belonged to families with a less well
293 characterised prey spectrum, such as *Haliangiaceae*, *Kofleriaceae*, and
294 *Polyangiaceae* (see Figure 2a), similar to findings of Zhou et al. (2014). The data in

295 this study hint toward biases in culturability within the *Myxococcales*. In fact, one
296 large family-level group abundant in organic soils, represented in the SILVA database
297 by clone Blrii41, is currently without any cultured representatives.

298

299 *A bacterial loop within the microbial loop*

300 The soil survey could not give direct proof of whether the *Myxobacteria* (or any
301 presumed micropredator) actually showed bacterivorous behavior *in situ*. In a
302 colonisation experiment with sterilised beech litter we compared their succession with
303 other bacterivores, such as the protozoa. In general, the abundance of potential
304 bacterivores was positively associated with abundance of prey bacteria, and increased
305 over time indicating a developing food web during litter colonisation. *Myxococcales*
306 and protozoa abundances developed similarly over time. This observation hints at
307 *Myxococcales* indeed having a predominantly predatory and not saprotrophic lifestyle,
308 thus rendering the *Myxococcales* a prominent predatory taxon feeding on other
309 bacteria.

310 There is direct *in situ* evidence for myxobacterial bacterivory from RNA-stable
311 isotope probing studies. The hallmark study of Lueders and colleagues
312 (2006) introduced the use of isotopically labelled prey bacteria to target the general
313 diversity of micropredators in soil *in situ* and follow the carbon flow through the
314 bacterial channel of the soil food web. The authors reported the detection of labelled
315 sequences related to *Myxococcus*, *Lysobacter*, and *Bacteroidetes*. However, due to the
316 limited technology at that time, they were not able to assess the contribution of
317 predatory protozoa. Another shortcoming was the use of *E. coli* cells as prey because
318 the survival of *E. coli* cells added to soil is rather low. In a more recent follow-up

319 study with labelled native soil bacteria, Zhang and Lueders (2017) provided evidence
320 for niche partitioning between bacterial and eukaryotic micropredators in soil, driven
321 by the soil compartment. Interestingly, the *Myxobacteria* preyed on both gram-
322 positive and gram-negative bacteria. Like in their previous study, the relative
323 contribution of pro- and eukaryotic micropredators could not be assessed due to
324 methodological limitations. Another recent study also included fungi and bacteria
325 (Kramer et al., 2016), and indicated that the commonly accepted split of energy
326 channels does not exist.

327 The dominance of *Myxobacteria*, especially in mineral soils, suggests their important
328 role in the soil microbial food web (Figure 5). The so-called microbial loop in soil
329 (Bonkowski, 2004) is important for the remineralisation of nutrients, where especially
330 protozoa and *Nematoda* feed on bacteria and by this set free nutrients, which are in
331 turn provided for the bacteria as well as plants (Coleman, 1994). Our observations
332 hint at the presence of an ecologically important 'bacterial loop', especially in mineral
333 soils, within the prokaryotes and independent of protozoa, that has been overlooked
334 until today (Figure 5). As a consequence, bacterial micropredators might not only be
335 important for shaping microbial communities but might also prove to be important for
336 the recycling of nutrients in soils, as it has been shown for protozoa (Bonkowski,
337 2004; Koller et al., 2013), and thus potentially for nutrient and carbon cycling.

338 Although the *Myxococcales* order comprised a significantly high proportion of
339 bacterial SSU rRNA in all sites, differences at family-level were observed among the
340 soils. Nevertheless, the two most abundant families were the *Polyangiaceae* and
341 *Haliangiaceae*. The high dominance of the latter may well be due to the
342 *Haliangiaceae* taxon being clustered together with the *Kofleriaceae* taxon in the

343 current SILVA database. Interestingly the name-giving group, the *Myxococcaceae*,
344 comprised a very low proportion of all *Myxococcales*. The *Myxococcaceae* are known
345 to be easily cultivable from a variety of environmental samples. Our data show that
346 other, less well characterised families are in fact much more abundant in soil. Thus,
347 efforts should be undertaken to investigate their biology and in particular their prey
348 spectrum.

349

350 *Methodological considerations*

351 The micropredator abundance data in this study are derived from the abundance of
352 SSU rRNA in metatranscriptomes. This does not reflect organismic abundance but is
353 rather a proxy of living biomass (Urich & Schleper, 2011). In fact, several factors need
354 to be taken into account when comparing the SSU rRNA from different pro- and
355 eukaryotic organisms. Results of various studies suggest differences in RNA contents
356 per biomass (1) between organisms and (2) between growth phases, respectively. The
357 ribosome density in prokaryotic cells is generally considered higher than in
358 eukaryotes. However, few data are available to our knowledge. The RNA content of
359 *E. coli* was determined to be 15.7% of dry mass (dm) (Stouthamer & van
360 Leeuwenhoek, 1973), of *Bacillus subtilis* between 8.5% and 14% dm⁻¹
361 (Tempest et al., 1968), *Saccharomyces cerevisiae* 23% dm⁻¹ (Parada & Acevedo, 1983),
362 *Aspergillus* 5.9% (Carlsen et al., 2000) and *Penicillium chrysogenum* between 5% and
363 8% (Henriksen et al., 1996). Furthermore, prokaryotic cells in an exponential growth
364 phase are known to contain more RNA than cells in stationary phase (e.g. Tempest et
365 al., 1968). Preliminary data (Petters & Urich, unpublished) hint to correction factors
366 to be applied when comparing rRNA based abundances from metatranscriptomes

367 between pro- and eukaroytes. Nevertheless, a recent study showed that rRNA
368 correlates better with cell counts than ribosomal RNA genes (Giner et al., 2016).
369 Thus, our metatranscriptomics data identify the predatory *Myxobacteria* as important
370 players in the midst of the soil food web and suggests a prominent role in the soil
371 microbial loop in particular.

372

373 **Acknowledgements**

374 Andreas Richter, Wolfgang Wanek and colleagues (University of Vienna) are thanked
375 for providing beech litter samples. Daniela Teichmann and Sylvia Klaubauf
376 (University of Vienna) are acknowledged for excellent technical assistance in RNA
377 extraction. We thank Ave Tooming-Klunderud, Lex Nederbragt and others at the
378 Norwegian High-Throughput Sequencing Centre (University of Oslo) for 454
379 pyrosequencing. Andrea Söllinger acknowledges funding from the University of Vienna
380 (uni:docs) and the OeAD (Austrian agency for international mobility and cooperation
381 in education, science and research; Marietta-Blau-Fellowship).

382

383 **Authors' contributions.** The study was designed by TU. Data analysis was
384 performed by SP and TU, supported by AS and MB. The manuscript was written by
385 SP and TU, assisted by all co-authors.

386

387 **References**

388 Altschul, S. F., Gish, W., Miller, W., Myers, E. W., & Lipman, D. J. (1990). Basic
389 local alignment search tool. *Journal of Molecular Biology*, 215(3), 403–10.
390 [http://doi.org/10.1016/S0022-2836\(05\)80360-2](http://doi.org/10.1016/S0022-2836(05)80360-2)

- 391 Azam, F., Fenchel, T., Field, J., Gray, J., Meyer-Reil, L., & Thingstad, F. (1983). The
392 Ecological Role of Water-Column Microbes in the Sea. *Marine Ecology*
393 *Progress Series*, 10, 257–263. <http://doi.org/10.3354/meps010257>
- 394 Bengtsson, M. M., Wagner, K., Schwab, C., Urich, T., & Battin, T. J. (2018). Light
395 availability impacts structure and function of phototrophic stream biofilms across
396 domains and trophic levels. *Molecular Ecology*, 27(14), 2913–2925.
397 <http://doi.org/10.1111/MEC.14696>
- 398 Beulig, F., Urich, T., Nowak, M., Trumbore, S. E., Gleixner, G., Gilfillan, G. D., ...
399 Kusel, K. (2016). Altered carbon turnover processes and microbiomes in soils
400 under long-term extremely high CO₂ exposure. *Nature Microbiology*, 1(2),
401 15025. <http://doi.org/10.1038/nmicrobiol.2015.25>
- 402 Bonkowski, M. (2004). Protozoa and plant growth: The microbial loop in soil
403 revisited. *New Phytologist*. <http://doi.org/10.1111/j.1469-8137.2004.01066.x>
- 404 Carlsen, M., Spohr, A. B., Nielsen, J., & Villadsen, J. (1996). Morphology and
405 physiology of an α -amylase producing strain of *Aspergillus oryzae* during batch
406 cultivations. *Biotechnology and Bioengineering*, 49(3), 266–276.
407 [http://doi.org/10.1002/\(SICI\)1097-0290\(19960205\)49:3<266::AID-](http://doi.org/10.1002/(SICI)1097-0290(19960205)49:3<266::AID-BIT4>3.0.CO;2-I)
408 [BIT4>3.0.CO;2-I](http://doi.org/10.1002/(SICI)1097-0290(19960205)49:3<266::AID-BIT4>3.0.CO;2-I)
- 409 Choma, M., Barta, J., Šantrrkov, H., & Urich, T. (2016). Low abundance of
410 Archaeorhizomycetes among fungi in soil metatranscriptomes. *Scientific*
411 *Reports*, 6(1), 38455. <http://doi.org/10.1038/srep38455>
- 412 Clarholm, M. (1985). Interactions of bacteria, protozoa and plants leading to
413 mineralization of soil nitrogen. *Soil Biology and Biochemistry*, 17(2), 181–187.
414 [http://doi.org/10.1016/0038-0717\(85\)90113-0](http://doi.org/10.1016/0038-0717(85)90113-0)
- 415 Coleman, D. C. (1994). The microbial loop concept as used in terrestrial soil ecology
416 studies. *Microbial Ecology*, 28(2), 245–250. <http://doi.org/10.1007/BF00166814>
- 417 Edgar, R. C. (2010). Search and clustering orders of magnitude faster than BLAST.
418 *Bioinformatics*, 26(19), 2460–2461. <http://doi.org/10.1093/bioinformatics/btq461>
- 419 Epelde, L., Lanzen, A., Blanco, F., Urich, T., & Garbisu, C. (2015). Adaptation of soil
420 microbial community structure and function to chronic metal contamination at an
421 abandoned Pb-Zn mine. *FEMS Microbiology Ecology*, 91(1), 1–11.
422 <http://doi.org/10.1093/femsec/fiu007>
- 423 Geisen, S., Koller, R., Hunninghaus, M., Dumack, K., Urich, T., & Bonkowski, M.
424 (2016). The soil food web revisited: Diverse and widespread mycophagous soil
425 protists. *Soil Biology and Biochemistry*, 94, 10–18.
426 <http://doi.org/10.1016/j.soilbio.2015.11.010>

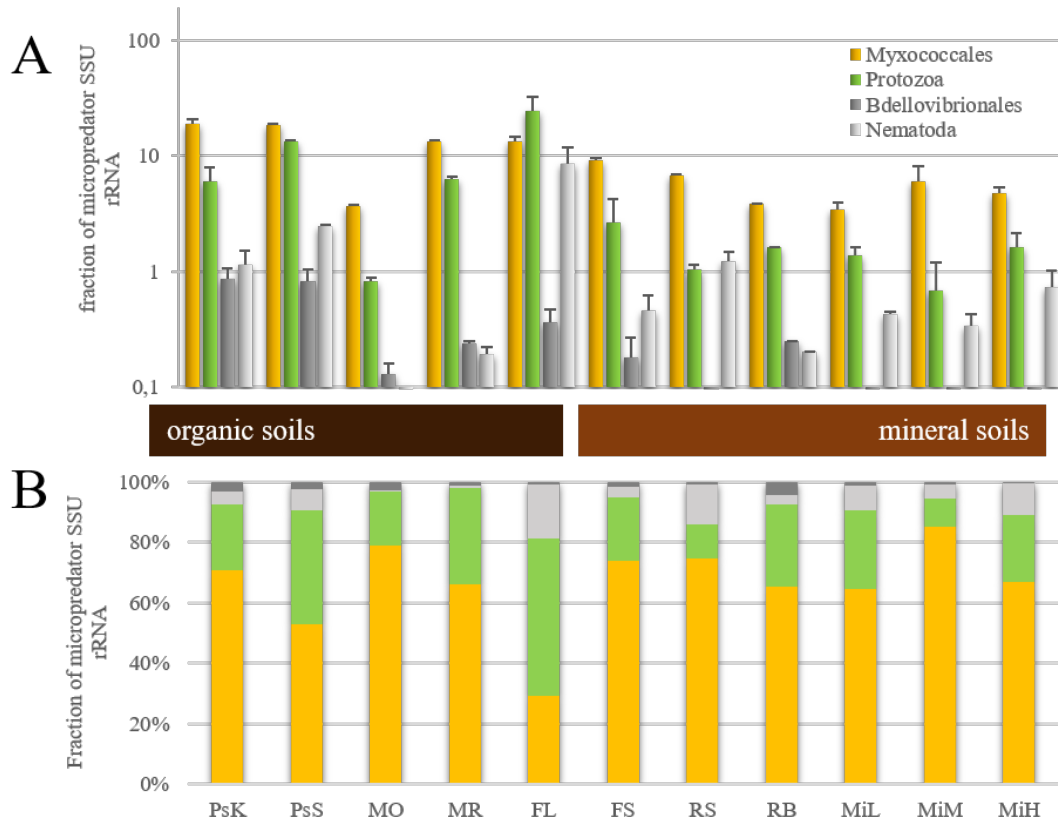
- 427 Geisen, S., Rosengarten, J., Koller, R., Mulder, C., Urich, T., & Bonkowski, M.
428 (2015). Pack hunting by a common soil amoeba on nematodes. *Environmental*
429 *Microbiology*, 17(11), 4538–4546. <http://doi.org/10.1111/1462-2920.12949>
- 430 Geisen, S., Tveit, A. T., Clark, I. M., Richter, A., Svenning, M. M., Bonkowski, M., &
431 Urich, T. (2015). Metatranscriptomic census of active protists in soils. *ISME*
432 *Journal*, 9(10), 2178–2190. <http://doi.org/10.1038/ismej.2015.30>
- 433 Giner, C. R., Forn, I., Romac, S., Logares, R., de Vargas, C., & Massana, R. (2016).
434 Environmental sequencing provides reasonable estimates of the relative
435 abundance of specific picoeukaryotes. *Applied and Environmental Microbiology*,
436 (May), AEM.00560-16. <http://doi.org/10.1128/AEM.00560-16>
- 437 Henriksen, C. M., Christensen, L. H., Nielsen, J., & Villadsen, J. (1996). Growth
438 energetics and metabolic fluxes in continuous cultures of *Penicillium*
439 *chrysogenum*. *Journal of Biotechnology*, 45(2), 149–64. Retrieved from
440 <http://www.ncbi.nlm.nih.gov/pubmed/9147448>
- 441 Huson, D. H., Mitra, S., Ruscheweyh, H.-J., Weber, N., & Schuster, S. C. (2011).
442 Integrative analysis of environmental sequences using MEGAN4. *Genome*
443 *Research*, 21(9), 1552–60. <http://doi.org/10.1101/gr.120618.111>
- 444 Karst, S. M., Dueholm, M. S., McIlroy, S. J., Kirkegaard, R. H., Nielsen, P. H., &
445 Albertsen, M. (2018). Retrieval of a million high-quality, full-length microbial
446 16S and 18S rRNA gene sequences without primer bias. *Nature Biotechnology*,
447 36(2), 190–195. <http://doi.org/10.1038/nbt.4045>
- 448 Keane, R., & Berleman, J. (2016). The predatory life cycle of *Myxococcus xanthus*.
449 *Microbiology (United Kingdom)*. <http://doi.org/10.1099/mic.0.000208>
- 450 Koller, R., Rodriguez, A., Robin, C., Scheu, S., & Bonkowski, M. (2013). Protozoa
451 enhance foraging efficiency of arbuscular mycorrhizal fungi for mineral nitrogen
452 from organic matter in soil to the benefit of host plants. *New Phytologist*, 199(1),
453 203–211. <http://doi.org/10.1111/nph.12249>
- 454 Kopylova, E., Noé, L., & Touzet, H. (2012). SortMeRNA: fast and accurate filtering
455 of ribosomal RNAs in metatranscriptomic data. *Bioinformatics (Oxford,*
456 *England)*, 28(24), 3211–7. <http://doi.org/10.1093/bioinformatics/bts611>
- 457 Kramer, S., Dibbern, D., Moll, J., Huenninghaus, M., Koller, R., Krueger, D., ...
458 Kandler, E. (2016). Resource partitioning between bacteria, fungi, and protists
459 in the detritusphere of an agricultural soil. *Frontiers in Microbiology*, 7(SEP), 1–
460 12. <http://doi.org/10.3389/fmicb.2016.01524>

- 461 Lanzén, A., Jørgensen, S. L., Huson, D. H., Gorfer, M., Grindhaug, S. H., Jonassen, I.,
462 ... Urich, T. (2012). CREST--classification resources for environmental sequence
463 tags. *PloS One*, 7(11), e49334. <http://doi.org/10.1371/journal.pone.0049334>
- 464 Lueders, T., Kindler, R., Miltner, A., Friedrich, M. W., & Kaestner, M. (2006).
465 Identification of bacterial micropredators distinctively active in a soil microbial
466 food web. *Applied and Environmental Microbiology*, 72(8), 5342–5348.
467 <http://doi.org/10.1128/AEM.00400-06>
- 468 McCarthy, D. J., Chen, Y., & Smyth, G. K. (2012). Differential expression analysis of
469 multifactor RNA-Seq experiments with respect to biological variation. *Nucleic
470 Acids Research*, 40(10), 4288–4297. <http://doi.org/10.1093/nar/gks042>
- 471 Parada, G., & Acevedo, F. (1983). On the relation of temperature and RNA content to
472 the specific growth rate in *Saccharomyces cerevisiae*. *Biotechnology and
473 Bioengineering*, 25(11), 2785–2788. <http://doi.org/10.1002/bit.260251120>
- 474 Reichenbach, H. (1999, February). The ecology of the myxobacteria. *Environmental
475 Microbiology*. Wiley/Blackwell (10.1111). [http://doi.org/10.1046/j.1462-
476 2920.1999.00016.x](http://doi.org/10.1046/j.1462-2920.1999.00016.x)
- 477 Schmieder, R., & Edwards, R. (2011). Quality control and preprocessing of
478 metagenomic datasets. *Bioinformatics*, 27(6), 863–864.
479 <http://doi.org/10.1093/bioinformatics/btr026>
- 480 Schwab, C., Berry, D., Rauch, I., Rennisch, I., Ramesmayer, J., Hainzl, E., ... Urich,
481 T. (2014). Longitudinal study of murine microbiota activity and interactions with
482 the host during acute inflammation and recovery. *ISME Journal*, 8(5), 1101–
483 1114. <http://doi.org/10.1038/ismej.2013.223>
- 484 Tempest, D. W., Dicks, J. W., & Ellwood, D. C. (1968). Influence of growth condition
485 on the concentration of potassium in *Bacillus subtilis* var. niger and its possible
486 relationship to cellular ribonucleic acid, teichoic acid and teichuronic acid. *The
487 Biochemical Journal*, 106(1), 237–43. Retrieved from
488 <http://www.ncbi.nlm.nih.gov/pubmed/4976492>
- 489 Trap, J., Bonkowski, M., Plassard, C., Villenave, C., & Blanchart, E. (2016).
490 Ecological importance of soil bacterivores for ecosystem functions. *Plant and
491 Soil*, 398(1–2), 1–24. <http://doi.org/10.1007/s11104-015-2671-6>
- 492 Tveit, A., Schwacke, R., Svenning, M. M., & Urich, T. (2013). Organic carbon
493 transformations in high-Arctic peat soils: key functions and microorganisms. *The
494 ISME Journal*, 7(2), 299–311. <http://doi.org/10.1038/ismej.2012.99>
- 495 Tveit, A. T., Urich, T., Frenzel, P., & Svenning, M. M. (2015). Metabolic and trophic
496 interactions modulate methane production by Arctic peat microbiota in response

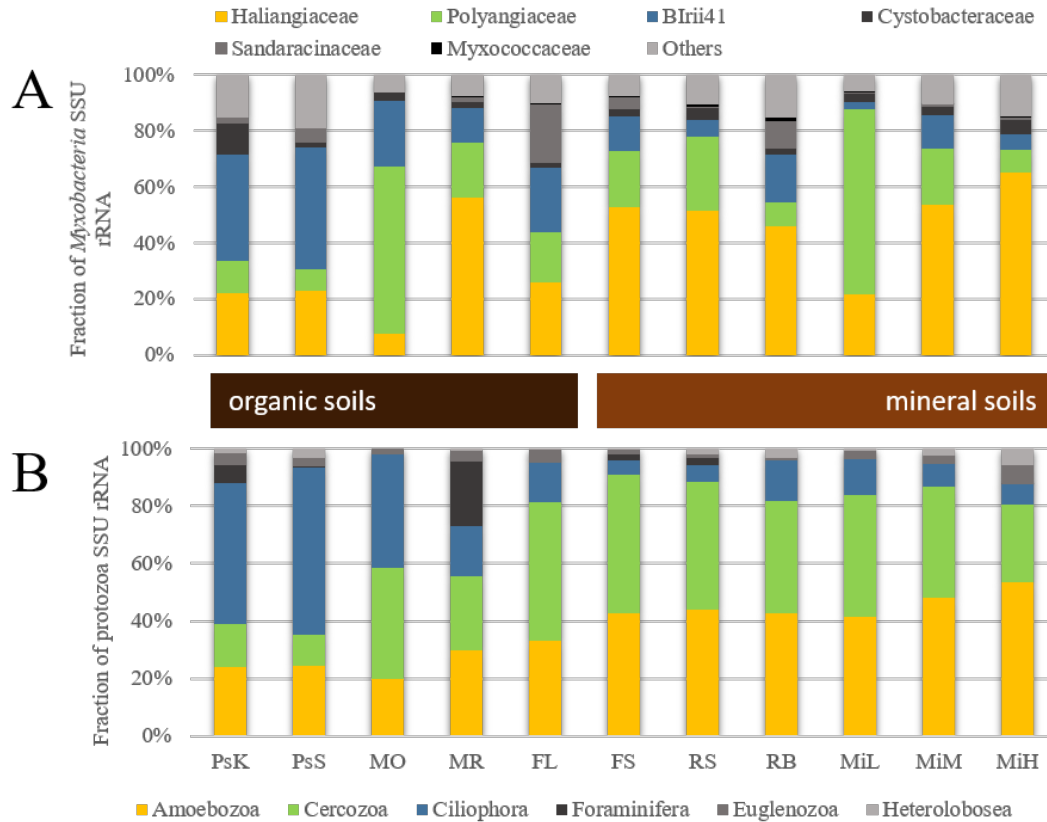
- 497 to warming. *Proceedings of the National Academy of Sciences*, 112(19), E2507–
498 E2516. <http://doi.org/10.1073/pnas.1420797112>
- 499 Urich, T., Lanzén, A., Qi, J., Huson, D. H., Schleper, C., & Schuster, S. C. (2008).
500 Simultaneous assessment of soil microbial community structure and function
501 through analysis of the meta-transcriptome. *PloS One*, 3(6), e2527.
502 <http://doi.org/10.1371/journal.pone.0002527>
- 503 Urich, T., & Schleper, C. (2011). The “Double-RNA” Approach to Simultaneously
504 Assess the Structure and Function of a Soil Microbial Community BT - ... and
505 Complementary Approaches. In ... and Complementary Approaches (pp. 587–
506 596). John Wiley & Sons, Inc. Retrieved from
507 [http://doi.wiley.com/10.1002/9781118010518.ch64%5Cnpapers2://publication/d](http://doi.wiley.com/10.1002/9781118010518.ch64%5Cnpapers2://publication/doi/10.1002/9781118010518.ch64)
508 [oi/10.1002/9781118010518.ch64](http://doi.wiley.com/10.1002/9781118010518.ch64)
- 509 Wanek, W., Mooshammer, M., Blöchl, A., Hanreich, A., & Richter, A. (2010).
510 Determination of gross rates of amino acid production and immobilization in
511 decomposing leaf litter by a novel ^{15}N isotope pool dilution technique. *Soil*
512 *Biology and Biochemistry*, 42(8), 1293–1302.
513 <http://doi.org/10.1016/j.soilbio.2010.04.001>
- 514 Zhang, L., & Lueders, T. (2017). Micropredator niche differentiation between bulk
515 soil and rhizosphere of an agricultural soil depends on bacterial prey. *FEMS*
516 *Microbiology Ecology*, 93(9). <http://doi.org/10.1093/femsec/fix103>
- 517 Zhou, X. W., Li, S. G., Li, W., Jiang, D. M., Han, K., Wu, Z. H., & Li, Y. Z. (2014).
518 Myxobacterial community is a predominant and highly diverse bacterial group in soil
519 niches. *Environmental Microbiology Reports*, 6(1), 45–56.
520 <http://doi.org/10.1111/1758-2229.12107>

521 **Table 1.** Context data for relevant sampling sites (changed from Choma *et al.*, 2016)

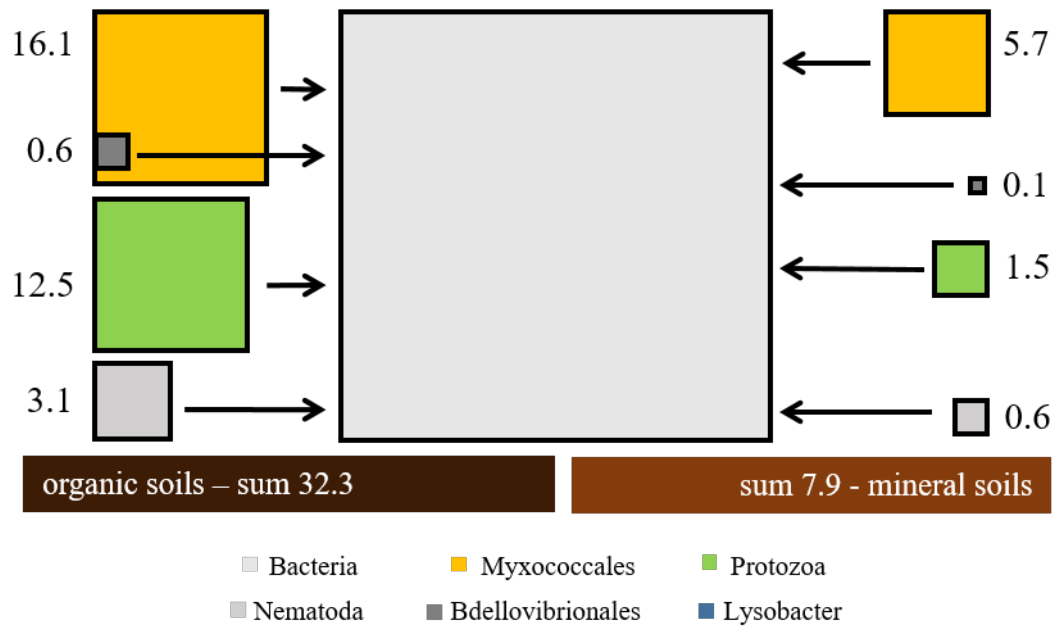
Site	Peatland soil "Knudsenheia"	Peatland soil "Solvatn"	Mofette	Mofette reference	Rothamsted grassland	Rotböhl	Forest Litter	Forest Soil	Mine L	Mine M	Mine H
Abbreviation	PsK	PsS	MO	MR	RS	RB	FL	FS	MiL	MiM	MiH
Location	Ny-Ålesund, Norway (Svalbard)	Ny-Ålesund, Norway (Svalbard)	Hartoušov, Czech Republic	Hartoušov, Czech Republic	Rothamsted, United Kingdom	Darmstadt, Germany	Vienna woods, Austria	Vienna woods, Austria	Coto Txomin, Spain	Coto Txomin, Spain	Coto Txomin, Spain
Climatic zone	Arctic	Arctic	Temperate	Temperate	Temperate	Temperate	Temperate	Temperate	Temperate	Temperate	Temperate
Biome	Fen wet land	Fen wet land	Floodplain	Floodplain	Grassland	Grassland	Temperate deciduous forest	Temperate deciduous forest	Shrubland	Shrubland	Shrubland
Dominant vegetation	Mosses	Mosses	<i>Filipendula ulmaria</i>	<i>Deschampsia cespitosa, Eriophorum vaginatum</i>	N.A.	N.A.	<i>Fagus sylvatica</i>	<i>Fagus sylvatica</i>	<i>Ulex europaeus</i>	<i>Festuca rubra</i>	<i>Festuca rubra</i>
Substrate type / Horizon	Organic peat (Top layer)	Organic peat (Top layer)	Organic soil	Gleic fluvisol	Mineral soil	Mineral soil	Litter horizon	Mineral soil (A horizon)	Mineral soil	Mineral soil	Mineral soil
pH	7.3	7.6	4.7	5.3	4.9	7.1	N.A.	4.5-5.1	3.9	5.6	5.9
Moisture (% soil dry weight)	1010	900	N.A.	N.A.	33	32	18	43-64	52	49	30
# of replicates	2	2	3	3	2	1	2	4	3	3	3
Sampling time	August 2009	August 2009	July 2013	July 2013	July 2009	January 2006	May 2008	May 2008	March 2011	March 2011	March 2011
Sequencing method	454 GS FLX Titanium	454 GS FLX Titanium	Illumina HiSeq 2500	Illumina HiSeq 2500	454 GS FLX Titanium	454 GS 20	454 GS FLX	454 GS FLX	Illumina HiSeq 2000	Illumina HiSeq 2000	Illumina HiSeq 2000



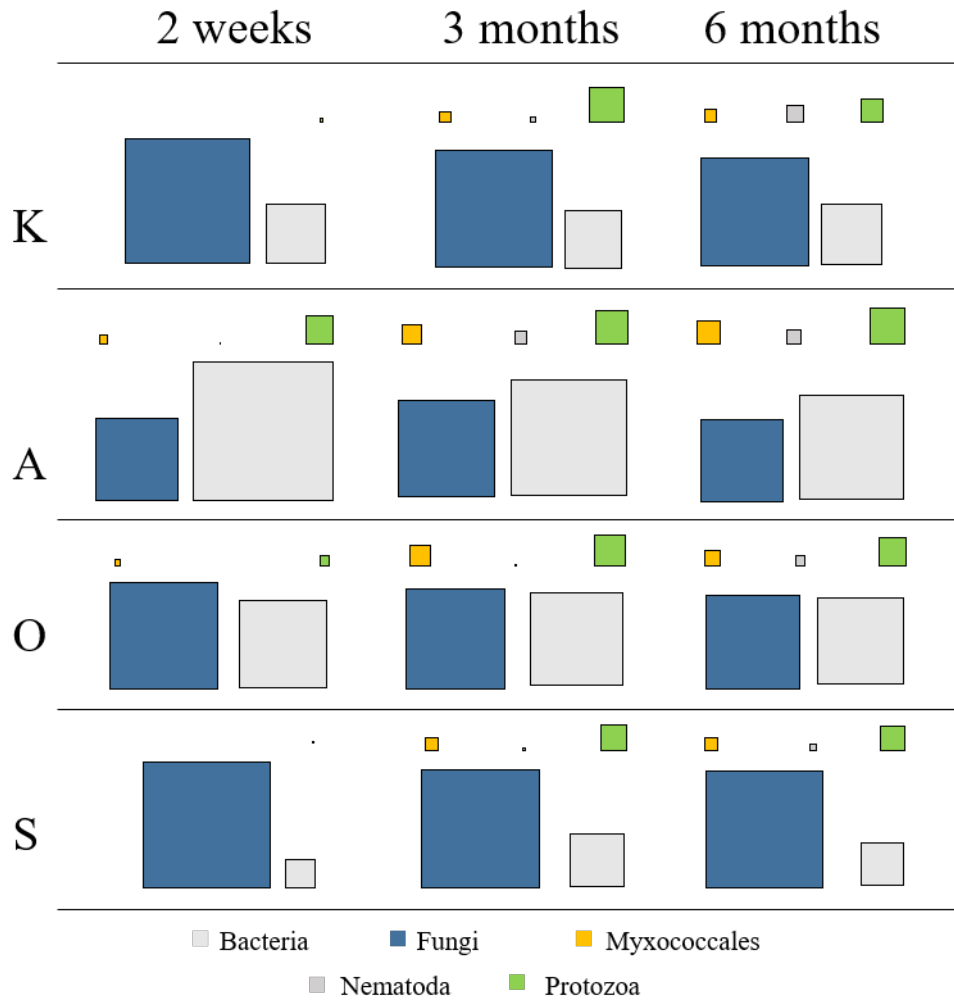
522 **Figure 1. Screening of pro- and eukaryotic micropredators.** (A) Fraction of major
523 identified micropredator SSU rRNA normalized to SSU rRNA of prey bacteria. (B)
524 Fraction of major identified micropredator SSU rRNA normalized to total
525 micropredator SSU rRNA. Error bars show standard deviation of replicates. For sites
526 see Table 1.



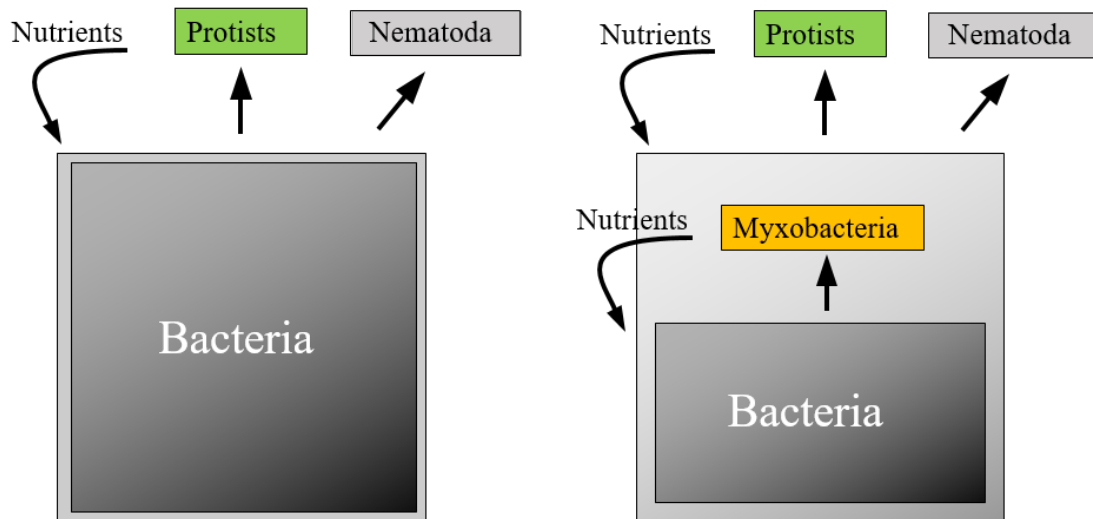
527 **Figure 2. Screening of Myxococcales and protozoa taxa.** (A) Fraction of identified
528 *Myxococcales* SSU rRNA normalized to overall *Myxococcales* SSU rRNA. (B)
529 Fraction of protozoa SSU rRNA normalized to total protozoa SSU rRNA. For sites
530 see Table 1.
531



532 **Figure 3. Comparison of organic and mineral soils.** Fraction of major identified
533 micropredator SSU rRNA normalized to SSU rRNA of prey bacteria. Average in
534 organic soils (excluding MO samples) on the left; average in mineral soils in the right.
535 Area of boxes resembles abundance of SSU rRNA. Numbers show proportions [%] of
536 prey bacterial SSU rRNA. *Lysobacter* data are not shown due to low abundances.



537 **Figure 4. Colonisation of beech litter.** Fraction of major identified micropredator,
 538 fungi, and prey bacteria SSU rRNA. Area of boxes resembles abundance of SSU
 539 rRNA. *Lysobacter* data are not shown due to low abundances. For litter types see
 540 Wanek *et al.*, 2010.



541 **Figure 5.** Simplified soil microbial loop. Left: Traditional microbial loop with
542 separate roles of prokaryotic and eukaryotic organisms. Right: Microbial loop
543 containing a bacterial loop independent of eukaryotic organisms. Straight arrows:
544 links between trophic levels. Bent arrows: provision of nutrients.