

1 An evaluation model for aboveground biomass based 2 on Hyperspectral Data from field and TM8 in Khorchin 3 grassland, China

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11 Abstract

12 Biomass is an important indicator for monitoring vegetation degradation and productivity. This
13 study tests the applicability of Hyperspectral Remote-Sensing in situ measurements for
14 high-precision estimation aboveground biomass (AGB) on regional scales of Khorchin grassland
15 landscape in Inner Mongolia, China. Field experiments were carried out which collected
16 hyperspectral data with a portable visible/NIR hyperspectral spectrometer (SOC 710), and
17 collected aboveground net primary productivity (ANPP). Ground spectral models were then
18 developed to estimate ANPP from the normalized difference vegetation index (NDVI), which was
19 measured in the field following the same method as that of the Thematic Mapper(TM) from the
20 Landsat 8 land imager (TM_NDVI). Regression analysis was used to assess the relationship
21 between ANPP and NDVI based on coefficients of determination (R^2) and error analysis. The
22 estimation of ANPP had unique optimal regression models. By comparing the different spectral
23 inversion models, we selected an exponential model associating ANPP with NDVI (ANPP =
24 $12.523 \times e^{3.370 \times (0.462 \times TM_NDVI + 0.413)}$, standard error = 24.74 g m⁻², $R^2 = 0.636$, $P < 0.001$).
25 This study suggests that the model can be used to monitor the condition and estimate the
26 productivity of grassland at regional scales. The results still show a high potential to map
27 grassland degradation proxies on the ground hyperspectral model. Thus, this study presents
28 biomass hyperspectral inversion technology to remotely detect and monitor grassland degradation
29 and productivity at high precision.

30 **Key words:** Biomass, Field hyperspectral, Remote sensing, Khorchin

31 Introduction

32 The tools for remotely sensing of vegetation have evolved significantly in recent decades[1], and
33 spectral imaging has become increasingly popular in remote-sensing research for correlating
34 spectral data with the biophysical properties of vegetation. Hyperspectral remote-sensing data
35 have subsequently been widely used to estimate vegetation biomass[2-5], vegetation cover
36 (VC)[6-7], nitrogen content[8], and the leaf area index of vegetation[1,9].

37 The accurate estimation of aboveground net primary productivity (ANPP) is an active area of
38 research and can provide valuable information about the productivity and ecosystem service value

39 of grassland [10]. ANPP is an important impact factor for desertification and are often used as
40 indicators for monitoring and evaluating grassland productivity and degradation [11]. The present
41 study aimed to develop models for estimating biomass and VC based on satellite data, which
42 allow an assessment over large areas at a low cost [12]. Desertification in the Khorchin grassland
43 is becoming worse with the rapid expansion of the population, overgrazing and the warmer and
44 drier trend associated with climate change [13]. The accurate estimation of the biomass of the
45 grassland over large areas using remotely sensed data is thus very important for monitoring
46 desertification and for improving the scientific management of grassland ecological resources.

47 Remote-sensing data have been transformed and combined into various spectral vegetation indices
48 that are used as predictors of parameters, such as the normalised difference vegetation index
49 (NDVI), ratio vegetation index, perpendicular vegetation index, soil adjusted vegetation index,
50 and transformed soil adjusted vegetation index [14-17]. Most studies apply NDVI because it
51 minimises the effects of topography[18] and is more reliable for the estimation of biomass of
52 ecosystems/habitats dominated by grasses when the grasses are actively growing [9,19]. The
53 NDVI is thus widely used to characterise grass ecosystems and to estimate biomass and
54 VC[3,4,19-23].

55 NDVI have been determined from data sets collected by various satellite instruments, such as the
56 Landsat Thematic Mapper (TM), the National Oceanic and Atmospheric
57 Administration/Advanced Very High Resolution Radiometer (NOAA/AVHRR), MODIS,
58 Gaofen-2 and the Moderate Resolution Imaging Spectro radiometer[9,24]. The Landsat 8 with
59 narrowband indices are highly suitable to be chosen to map AGB accurately, because the
60 narrowband indices have led to significant improvements in the predictive capability of models,
61 and hyperspectral data from aerial imagery or field spectrometry have the potential to estimate the
62 biophysical properties of rangeland or steppe vegetation with a greater accuracy than broadband
63 indices [3,4,23,25-26].

64 In order to improve the accuracy of biomass estimation, the measurements of ground reflectance
65 have been used to estimate biomass in grasslands and steppes since the 1970s [27], but ground
66 spectral reflectance can be influenced by variable factors of the landscape such as the distribution
67 of plant communities [28], soil colour [29], hydrology[30], and topography[31], and sensor
68 radiance may be strongly affected by atmospheric scattering[32]. For these reasons, various
69 regression models have been established for associating vegetation indices with biomass in
70 different areas. Relationships have been established between remote-sensing data and the
71 biophysical properties of vegetation, mostly linear or non-linear, that have greatly improved the
72 accuracy of biomass estimates and that have determined patterns of grassland productivity in
73 various regions [6,33-37].

74 Taking into account the advantages and disadvantages of the current remote sensing sources to
75 estimate AGB, In this study, we present a remote sensing approach for estimating and monitoring
76 AGB in meadows and pastures during the growing season. We used remote sensing of Landsat 8
77 and ground hyperspectral to calculate the normalized difference vegetation index (NDVI), and
78 on-field aboveground net primary production (ANPP) measurements to establish an empirical
79 exponential model to estimate spatial ANPP across the entire Khorchin grassland.

80 The main objectives of this study are: (i) to analyse the relationship of the ground narrowband
81 NDVI with ANPP and then to develop the most suitable ground spectral models for evaluating
82 ANPP over a large area of the Khorchin steppe, and (ii) to use the ANPP model to understand the

83 biocapacity of the Khorchin grassland for providing technical support for determining a
84 reasonable grazing intensity and guiding the development of the livestock industry.

85 **Materials and methods**

86 **Study area**

87 The Khorchin grassland is located in the eastern section of the ecotone between crop production
88 and animal husbandry, we choose the region of Bairin Youqi in Inner Mongolia of China as the
89 study area (Fig. 1), which is an important component of the Khorchin grassland and is typically
90 sensitive and fragile. Bairin Youqi has a semiarid, temperate, continental monsoon climate with
91 mean annual temperature of 4.9 °C and mean annual precipitation of 358 mm (precipitation is less
92 than evaporation). From low to high elevations, the distribution of vegetation is meadow, sandy
93 vegetation, and low mountain grassland respectively. The dominant grassland species include
94 *Achnatherum splendens* (Trin.) Nevskia, *Stipa capillata* Linn., *Leymus chinensis* (Trin.) Tzvel.,
95 and *Agropyron cristatum* (Linn.) Gaertn.

97 **Fig. 1 Map of Inner Mongolia (left) and the location of the sampling sites in Bairin Youqi (right)**

98 This study was carried out in the field of Khorchin grassland which was State-owned Land and did not
99 involve endangered or protected species. Meanwhile, because this study supported by National
100 Environmental Conservation Research Program, so the government of Bairin Youqi permitted and
101 approved this study.

102 **Collection of field data**

103 **Experimental setup**

104 Field work was conducted during 15-30 July 2016, coinciding with the most productive period of
105 vegetational growth. Based on the topography and land use, 39 plots were established that
106 included large, homogeneous patches of vegetation and representative vegetational communities
107 with different types of vegetation. The plot size was set at 30 × 30 m, equivalent to the size of a
108 TM8 pixel. The plots contained a total of 173 quadrats of 1 × 1 m. The data collected were divided
109 into 2 groups. Group one, which contained 153 quadrats were used to build the ground spectral
110 model; group two, which contained 20 quadrats, were used for the accuracy test of the spectral
111 inversion model. Meanwhile, within group one, the data of approximately two thirds of the total
112 quadrats (n=115) were chosen randomly to build the model while the rest were used for testing the
113 terrain model in terms of selecting the best fitting function and precision.

114 **Field spectral data**

115 The field data were collected using the SOC710 Hyperspectral Imaging System which
116 Manufactured by Surface Optics Corporation in America. The SOC710 is a precision instrument
117 with an integrated scanning system and analysis software that can quickly obtain high-quality
118 hyperspectral images at visible to near-infrared (NIR) wavelengths in the range 0.4-1.0 μm. The
119 system can be used under normal lighting conditions at variable exposures and gains. The SOC
120 spectra were collected with a 10° field of view and at 1.2 m above the grass canopy. All spectral

121 measurements were collected between 9:00 and 15:00 Beijing time under clear skies. Three
122 measurements were taken for each sample of grass canopy. These spectra were standardised to
123 spectra measured at approximately 10-minute intervals with a white board. The average of three
124 replicates for each sample was used for the analysis.

125 **Biomass measurements**

126 After the spectral data had been recorded, the standing biomass was collected in the quadrats at
127 each sample location. The fresh weight of green herbaceous material was recorded soon after
128 clipping, the samples were then dried at 80 °C for 10-12 hours, and the dry masses of the samples
129 were determined.

130 **Image data acquisition, satellite data, and preprocessing**

131 Biomass was assessed using TM8 data from the Landsat 8 land imager of the United States
132 Geological Survey. The satellite data were acquired within the same time frame in which the field
133 data had been collected, and the images were free of clouds and haze. Four suitable TM8 satellite
134 scenes at PATH/ROWS 123/29, 123/30, 122/29, and 122/30 were analysed. The satellite data were
135 geometrically rectified by a digital elevation model and ground-control points from Land Survey.
136 The four TM8 scenes were processed for atmospheric correction with the Fast Line-of-sight
137 Atmospheric Analysis of Spectral Hypercubes software package.

138 **Data analysis**

139 **NDVI calculation**

140 NDVI are commonly calculated from RED and NIR reflectance data [38]. We calculated the
141 SOC_NDVI of the samples from SOC710 spectral reflectance using the ENVI 5.0 image analysis
142 software. The method for calculating NDVI was the same as that used for calculating the
143 TM_NDVI:

$$144 \quad NDVI = \frac{NIR - RED}{NIR + RED} \quad (1)$$

145 Where the RED and NIR bands correspond to wavelengths of 630-680 and 845-885 nm,
146 respectively. Spectral reflectance data should be resampled within the scope of the RED and NIR
147 bands.

148 **Regression analyses**

149 The regression analyses were carried out for the scatter diagrams of ANPP vs. SOC_NDVI, and
150 SOC_NDVI vs. TM_NDVI. In study area, The 173 quadrats data were employed to obtain the
151 regression model for ANPP vs. SOC_NDVI. Mean value of NDVI within a specific plot was
152 calculated, and then the data of the total 39 plots (Fig. 1) were used in the regression analysis for
153 SOC_NDVI vs. TM_NDVI.

154 The coefficient of determination (R^2) and the adjusted R^2 were used to test the strength and
155 significance of the relationships between the field data and the corresponding data extracted from
156 the satellite scenes. The standard error (SE, Eq. 2) of the prediction based on the independent test

157 data and the coefficient of mean error (MEC, Eq. 3) were calculated to assess the accuracy of the
 158 developed models.

$$159 \quad SE = \sqrt{\frac{\sum_{i=1}^n (y - y')^2}{n}} \quad (2)$$

$$160 \quad MEC = \frac{\sum_{i=1}^n \left| \frac{y - y'}{y} \right|}{n} \quad (3)$$

161 where y is a measured biomass, y' is an estimated biomass for the test data, and n is the number of
 162 samples.

163 Results

164 Optimal ground spectral models and tests of model accuracy

165 The optimal ground spectral models for biomass

166 From the analysis and evaluation of the relationships between ANPP and SOC_NDVI computed
 167 from reflectance data obtained by the SOC710 in the field, we chose linear, logarithmic, power,
 168 and exponential functions to fit and optimise the regression equations for selecting the best
 169 regression model (Fig.2).

170

171 **Fig.2 The Simulation Curves of the Regression Equation of the Training Samples**

172 The relationships between ANPP and SOC_NDVI was significant ($P < 0.001$) for all functions
 173 and met the assumptions of the statistical analyses. The exponential model was superior for ANPP,
 174 with an R^2 of 0.636, indicated by bold type in Tables 1.

175 **Table 1** Comparison of the regression equations between ANPP and SOC_NDVI.

| | Linear | Logarithmic | Power | Exponential |
|------------------|--------------------------|-------------------------------|--------------------------|---------------------------------|
| n | 115 | 115 | 115 | 115 |
| Equation | $y = 443.297x - 166.610$ | $y = 284.562\ln(x) + 248.525$ | $y = 299.611x^{2.216}$ | $y = \mathbf{12.523e^{3.370x}}$ |
| R^2 | 0.617 | 0.579 | 0.626 | 0.636 |
| Adjusted R^2 | 0.614 | 0.575 | 0.623 | 0.633 |
| $F(\alpha=0.01)$ | 182.255 ($P < 0.001$) | 155.181 ($P < 0.001$) | 802.5746 ($P < 0.001$) | 1089.7635 ($P < 0.001$) |

176 Note: n , number of samples; R^2 , coefficients of determination; the best regression model is highlighted in bold.

177 Tests of model accuracy

178 The accuracy of the models was tested to obtain the best regression models. We used test sets of
 179 all field samples to analyse and evaluate the errors in the regression models (Table 2). A
 180 comparison of the predictive performances of the regression equations indicated by SE and MEC
 181 are presented in Table 2.

182 **Table 2** Comparison of the errors of the regression equations.

| | Linear | Logarithmic | Power | Exponential |
|--|--------|-------------|-------|-------------|
|--|--------|-------------|-------|-------------|

| | | | | | |
|-------------|------------------------|-------|-------|-------|--------------|
| <i>ANPP</i> | <i>n</i> | 38 | 38 | 38 | 38 |
| | SE (gm ⁻²) | 52.64 | 55.46 | 51.32 | 49.22 |
| | MEC (%) | 34.77 | 37.09 | 31.64 | 30.01 |

183 Note: *n*, number of samples; SE, standard error of prediction; MEC, coefficient of mean error; the SE and MEC of
 184 the best regression model for each vegetational parameter are highlighted in bold.

185

186 We determined the best models for ANPP based on R^2 and the independent validations. The
 187 exponential equation was optimal for ANPP ($R^2 = 0.636$, SE = 49.22gm⁻², MEC = 30.01%; Tables
 188 1 and 2). Models with the following equations (Eqs. 4) were selected and used as the optimal
 189 ground spectral models for ANPP of the entire Khorchin grassland:

190
$$ANPP = 12.523 * e^{3.370 * SOC_NDVI} \quad (4)$$

191 **The relationship between TM_NDVI and SOC_NDVI**

192 The linear regression equation was selected based on the analysis of the TM_NDVI/SOC_NDVI
 193 scatter plot. The relationship between TM_NDVI and SOC_NDVI was significant, with an R^2 of
 194 0.656 ($P < 0.001$) and met the assumptions of the statistical analyses (Table 4, Fig.3). The model
 195 with the following equation (Eq. 5) was selected for the relationship between TM_NDVI and
 196 SOC_NDVI for the entire Khorchin grassland:

197
$$SOC_NDVI = 0.462 * TM_NDVI + 0.413 \quad (5)$$

198 **Table 3** Regression equations between TM_NDVI and SOC_NDVI.

| Equation | <i>n</i> | <i>R</i> | R^2 | <i>F</i> ($\alpha=0.01$) |
|------------------|----------|----------|-------|----------------------------|
| $y=0.462x+0.413$ | 39 | 0.810 | 0.656 | 70.63, $P < 0.001$ |

199 Note: *n*, number of samples; *R*, R^2 , coefficients of determination.

200

201 **Fig. 3 Fitted curve of the best model for the relationship between SOC_NDVI and**
 202 **TM_NDVI.**

203

204 **Spectral inversion models**

205 The spectral inversion models of TM8 for ANPP was calculated by Eqs. 4-5:

206
$$ANPP = 12.523 * e^{3.370 * (0.462 * TM_NDVI + 0.413)} \quad (6)$$

207 To test the agreement between measured and predicted values, we applied Eqs. 6 to the TM8
 208 NDVI greyscale image and obtained the patterns of ANPP distribution in the study area by grid
 209 computing. The test data sets were then converted into vector diagrams defined by geographic
 210 coordinates by geographic information system. The values at the test points were recorded in the
 211 distribution patterns as the corresponding pixels predicting values of ANPP. The relationship
 212 between actual and predicted values was used to evaluate the accuracy of model.

213 The correlation between the predicted and actual values was significant, as were the independent
 214 validations for predicting biomass (SE = 24.74, MEC = 18.61%; Fig.4). This study suggested that
 215 the spectral inversion models could be used to monitor grassland biomass at regional scales.

216

217 **Fig. 4 Independent validation for predicting biomass ($n = 20$, $P < 0.001$). SE, standard error of**

218 **predicted biomass; MEC, coefficient of mean error.**

219 **Discussion**

220 The main goal of this study was to establish more accurate models for estimating ANPP of the
221 grassland in Khorchin. We chose the NDVI vegetation index, which can be calculated from
222 spectral reflectance data acquired in the field and from data from Landsat TM7 Band 4 (TM4;
223 760-900 nm) and Band 3 (TM3; 630-690 nm) or from NOAA/AVHRR Channel 1 (580-680 nm)
224 and Channel 2 (720-1100 nm)[39-41]. In a previous study, we also calculated the NDVI from data
225 collected by a FieldSpec3 spectroradiometer (Analytical Spectral Devices, Boulder, USA), at
226 spectral reflectances of 620-670 (RED) and 841-876 (NIR) nm [9]. To further improve the
227 accuracy in the present study, we chose satellite data from Landsat 8(TM8), which have a higher
228 geometric precision and signal-to-noise ratio than the other Landsat data, and used the SOC710
229 Hyperspectral Imaging System, which is more accurate than the FieldSpec3 spectroradiometer.
230 The TM8 remotely sensed imaging data were only released in 2013, so they have not yet been
231 widely applied to monitor vegetational biomass. This study applied the field data for monitoring
232 the vegetation, thereby providing an informational baseline for this study area. The spectral
233 inversion model was ideal, indicating that TM8 remote imaging can be used for research on
234 vegetation biomass on a regional scale.

235 ANPP have their own optimal regression models based on the processing and statistical analysis
236 of experimental data in the study area. The optimal equations for the estimation of ANPP (Fig.5)
237 indicate that the relationship between SOC_NDVI and ANPP weakens at biomass $>350 \text{ g m}^{-2}$ for
238 grassland. Estimates of biomass above these levels are inaccurate or unreliable and may be
239 affected by the NDVI lower saturation phenomenon in areas of dense vegetation cover. When
240 biomass exceed these levels, factors such as grass height and leaf area index must be considered,
241 or a modified SOC_NDVI should be derived[42].

242

243 **Fig. 5 Fitted curve of the best model for the relationship between ANPP and SOC_NDVI for**
244 **the calibration sets**

245 The ground spectral models for ANPP can be applied to TM8 images, because measured spectral
246 characteristics of plants on the ground are intrinsically linked to those obtained by TM8 remote
247 sensing. Grassland yield over large areas can be estimated based on the ground spectral model.
248 The models, however, could be more accurate if field and satellite data are collected over several
249 years rather than only for one year. Also, the field and satellite data should be acquired at the same
250 time for maximal correspondence. In future field experiments, we will assess the collective
251 influence of these vegetational characteristics and the NDVI on biomass prediction and will seek
252 to obtain a modified NDVI for estimating the biomass of dense vegetation under natural
253 conditions.

254 **Conclusions**

255 This study developed a relatively accurate model for estimating AGB and tests the applicability of
256 hyperspectral data from field and TM8 to map AGB on regional scales by a regression analysis
257 method. The methodology we adopted in the study was a first attempt to Retrieval of vegetation
258 biomass from ground hyperspectral remote sensing in Khorchin grassland.

259 The accuracy of ground spectral inversion is affected by many factors, and the quality of the selected

260 remote sensing image data has the greatest impact on the fitting accuracy of the model. Landsat 8
261 satellite data is selected for remote sensing data, which has higher geometric accuracy and
262 signal-to-noise ratio than previous Landsat data, which effectively expands the application range of
263 image data. In the aspect of imaging mode, the sweep pendulum design of OLI imager has good
264 stability and improves the image quality, and in the aspect of geometric accuracy, L1T data product is a
265 data product after precise correction, and the product accuracy has been greatly improved. In this paper,
266 TM8 data is used to retrieve vegetation biomass, and the results show that calculated R^2 and SE and
267 MEC values for various regression models vary among ground spectral models. By comparison,
268 the exponential regression models we developed show a stronger relationship between spectral
269 reflectance and ANPP. An exponential equation was optimal for estimating ANPP in the Khorchin
270 grassland. Accuracy verification indicated that the relationship between the actual and predicted
271 biomass was significant. Estimating ANPP with high accuracy based on NDVI derived from TM8
272 satellite data is thus possible, which accumulates experience for the application of TM8 data in
273 vegetation monitoring field.

274 The accuracy of this technique depends on living, green biomass and not on senesced or dead
275 biomass, so the timing of the acquisition of NDVI data is critical, and the model can possibly be
276 improved if models are developed per vegetation types and using a larger range of ground data. In
277 brief, this research shows the usefulness of hyperspectral data from field and TM8 to evaluate
278 aboveground biomass at very high precision to provide theoretical and data support for RS
279 monitoring, grassland governance and ecological restoration.

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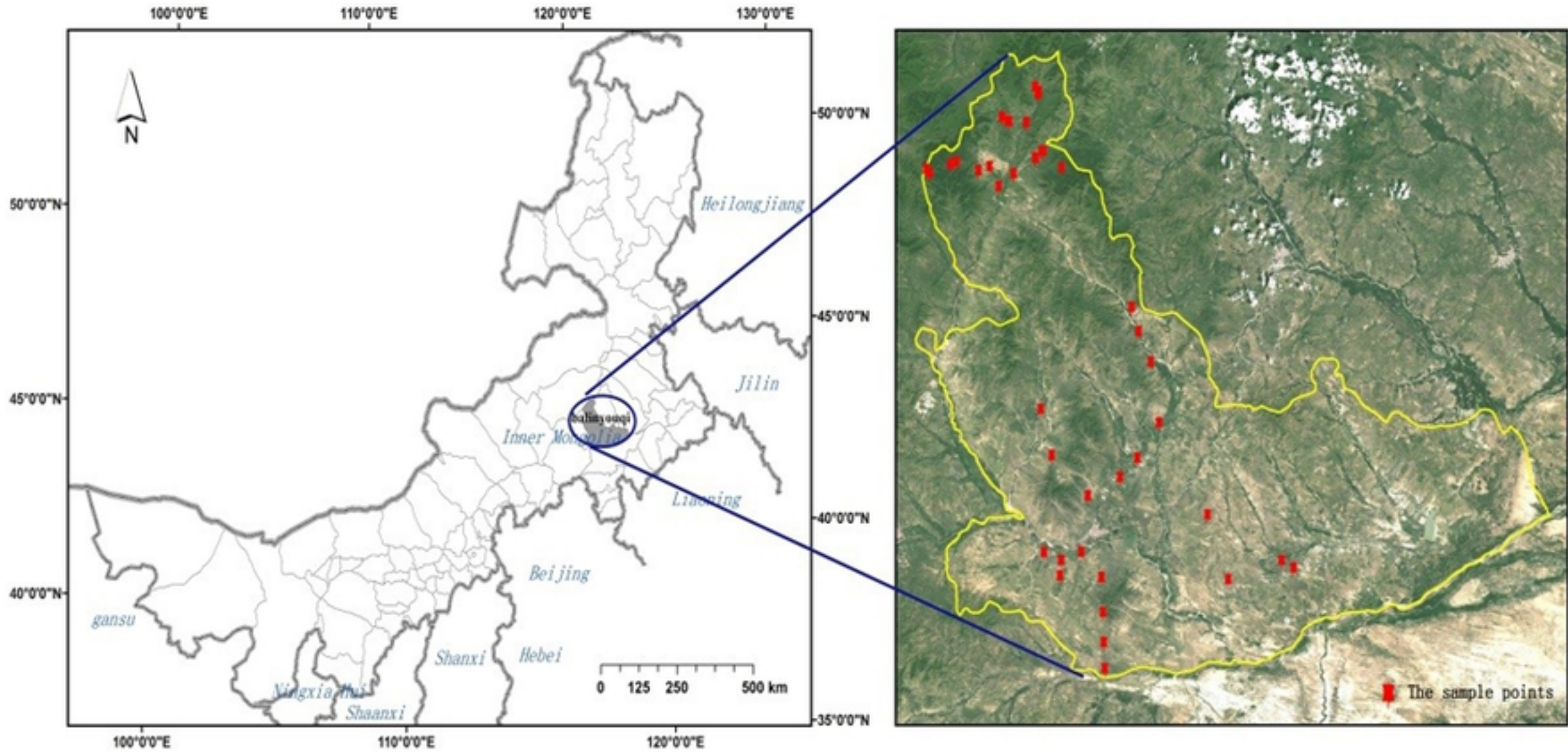
284 **References**

- 285 1. Darvishzadeh, R, Skidmore, A, Atzberger, C, Wieren, S. V. Estimation of vegetation LAI from hyperspectral
286 reflectance data: effects of soil type and plant architecture. *International Journal of Applied Earth
287 Observation and Geoinformation*. 2008;10(3):358-373. <https://doi.org/10.1016/j.jag.2008.02.005>.
- 288 2. Mutanga O, Skidmore A K. Hyperspectral band depth analysis for a better estimation of grass biomass
289 (*Cenchrus ciliaris*) measured under controlled laboratory conditions. *International Journal of Applied Earth
290 Observation and Geoinformation*. 2004;5(2):87-96. <https://doi.org/10.1016/j.jag.2004.01.001>.
- 291 3. M.A.Cho, A.Skidmore, F.Corsi, S.E.Van wieren, I.Sobhan. Estimation of green grass/herb biomass from
292 airborne hyperspectral imagery using spectral indices and partial least squares regression. *International
293 Journal of Applied Earth Observation and Geoinformation*. 2007;9(4):414–424
294 <https://doi.org/10.1016/j.jag.2007.02.001>.
- 295 4. I.Numata, D.A.Roberts, O.A.Chadwick, J.P.Schimel, L.S.Galvao, J.V.Soares. Evaluation of hyperspectral
296 data for pasture estimate in the Brazilian Amazon using field and imaging spectrometers. *Remote Sensing
297 of Environment*. 2008; 112(4):1569-1583. <https://doi.org/10.1016/j.rse.2007.08.014>.
- 298 5. Gaia Vaglio Laurin, Johannes Balling, Piermaria Corona, Walter Mattioli, Dario Papale, Nicola Puletti,
299 Maria Rizzo, John Truckenbrodt, Marcel Urban. Above-ground biomass prediction by
300 Sentinel-1multitemporal data in central Italy with integration of ALOS2 and Sentinel-2 data. *Appl. Remote
301 Sens*. 2018;12(1):016008. <https://doi.org/10.1117/1.JRS.12.016008>

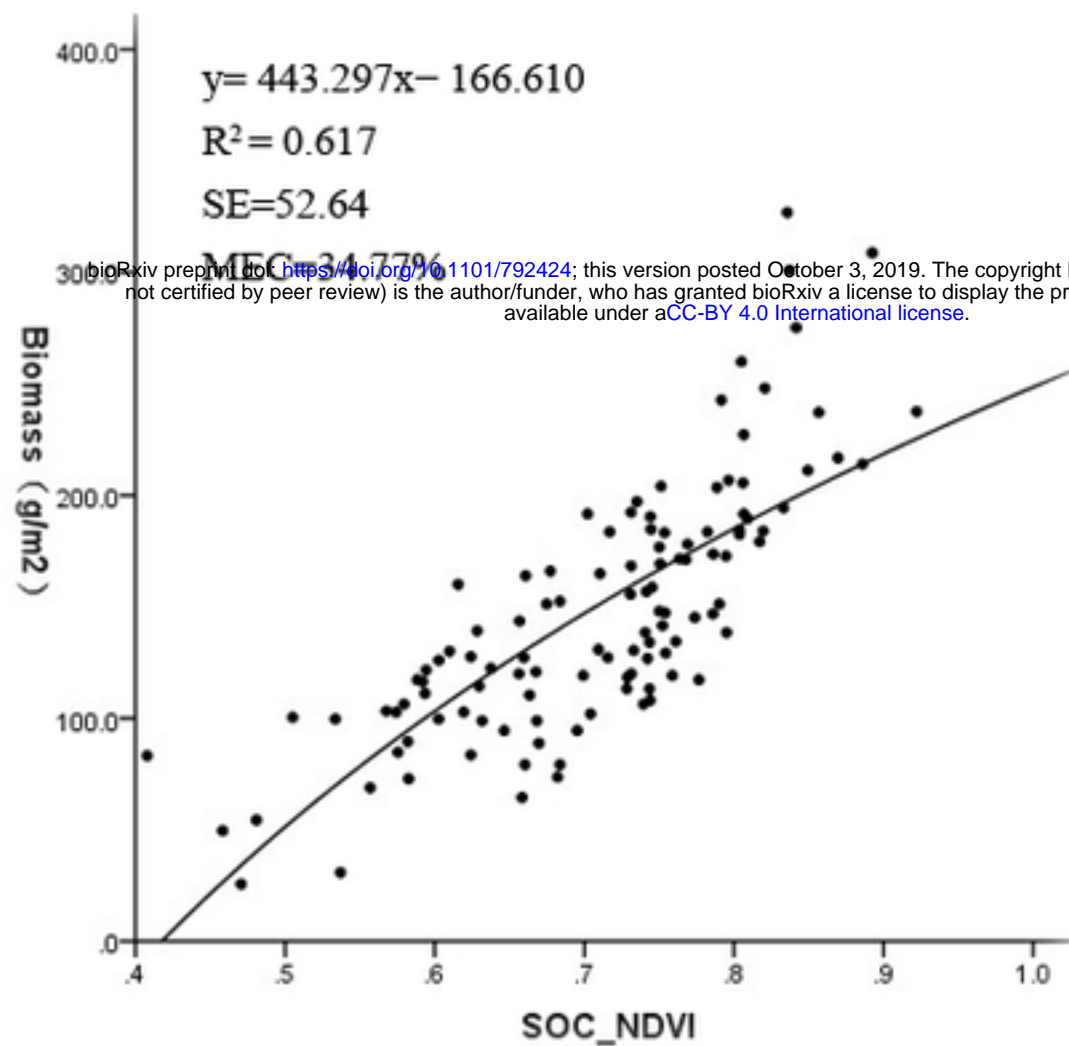
- 302 6. M. Kenneth, M. Timothy, F. Lynn. Hyperspectral Mixture Modeling for Quantifying Sparse Vegetation Cover
303 in Arid Environments. *Remote Sensing of Environment*. 2000; 72(3):360–374.
304 [https://doi.org/10.1016/S0034-4257\(99\)00112-1](https://doi.org/10.1016/S0034-4257(99)00112-1).
- 305 7. X.F. Zhang, C.H. Liao, J. Li, Q. Sun. Fractional vegetation cover estimation in arid and semi-arid
306 environments using HJ-1 satellite hyperspectral data. *International Journal of Applied Earth Observation*
307 *and Geoinformation*. 2013; 21:506–512. <https://doi.org/10.1016/j.jag.2012.07.003>.
- 308 8. A. Ramoelo, A.K. Skidmore, M.A. Cho, R. Mathieu, I.M.A. Heitkönig, N. Dudeni-Tlhone, M. Schlerf,
309 H.H.T. Prins. Non-linear partial least square regression increases the estimation accuracy of grass nitrogen
310 and phosphorus using in situ hyperspectral and environmental data. *ISPRS Journal of Photogrammetry and*
311 *Remote Sensing*. 2013;82:27–40. <https://doi.org/10.1016/j.isprsjprs.2013.04.012>.
- 312 9. S. Gao, Z. Niu, N. Huang X.H. Hou. Estimating the Leaf Area Index, height and biomass of maize using
313 HJ-1 and RADARSAT-2. *International Journal of Applied Earth Observation and Geoinformation*. 2013;
314 24, 1–8. <https://doi.org/10.1016/j.jag.2013.02.002>.
- 315 10. J.X. Gao, Y.M. Chen, Sh. H. Lü, Ch. Y. Feng, X.L. Chang, Sh. X. Ye, J.D. Liu. A ground spectral model for
316 estimating biomass at the peak of the growing season in Hulunbeier grassland, Inner Mongolia, China.
317 *International Journal of Remote Sensing*. 2012 ;33(13): 4029–4043.
318 <https://doi.org/10.1080/01431161.2011.639401>.
- 319 11. T.J. Cai, C.Y. Ju, Y.F. Yao. Quantitative estimation of vegetation coverage in Mu Us sandy land based on
320 RS and GIS. *Chinese Journal of Applied Ecology*. 2005;16(12):2301-2305 (in Chinese with English
321 abstract). http://en.cnki.com.cn/Article_en/CJFDTotal-STBC200302005.htm.
- 322 12. S. Eckert, M. Engesser. Assessing vegetation cover and biomass in restored erosion areas in Iceland using
323 SPOT satellite data. *Applied Geography*. 2013;40: 179-190.
324 <https://doi.org/10.1016/j.apgeog.2013.02.015>.
- 325 13. C.Y. Cao, D.M. Jiang, L.H. Zhu, Y.H. Nan. Degradation and diversity changes of meadow grassland in
326 Keerqin Sandy Land. *Acta Prata Agriculturae Sinica*. 2006;15(3):18-26 (in Chinese with English abstract).
327 http://en.cnki.com.cn/Article_en/CJFDTotal-CYXB200603002.htm.
- 328 14. T.N. Carlson, D.A. Ripley. On the relation between NDVI, fractional vegetation cover, and leaf area index.
329 *Remote Sensing of Environment*. 1997;62:241–252. [https://doi.org/10.1016/S0034-4257\(97\)00104-1](https://doi.org/10.1016/S0034-4257(97)00104-1).
- 330 15. P.S. Thenkabail, R.B. Smith, E. De Pauw. Hyperspectral vegetation indices and their relationships with
331 agricultural crop characteristics. *Remote Sensing of Environment*. 2000; 71:158–182.
332 [https://doi.org/10.1016/S0034-4257\(99\)00067-X](https://doi.org/10.1016/S0034-4257(99)00067-X).
- 333 16. C.K. Wright, K.M. Beurs, G.M. Henebry. Combined analysis of land cover change and NDVI trends in the
334 Northern Eurasian grain belt. *Frontiers of Earth Science*. 2012; 6(2):177-187.
335 <https://doi.org/10.1007/s11707-012-0327-x>.
- 336 17. Y.Z. An, W. Gao, Z.Q. Gao, C.S. Liu, R.H. Shi. Trend analysis for evaluating the consistency of Terra
337 MODIS and SPOT VGT NDVI time series products in China. *Frontiers of Earth Science*.
338 2015;9(1):125-136. <https://doi.org/10.1007/s11707-014-0428-9>.
- 339 18. B. N. Holben, C.O. Justice. An examination of spectral band ratioing to reduce the topographic effect on
340 remotely sense data. *International journal of remote sensing*. 1981; 2:115-133.
341 <https://doi.org/10.1080/01431168108948349>.
- 342 19. Y.M. Chen, J.X. Gao, Z.Y. Diao, S.H. Lü, X.L. Chang, Z.Y. Feng. Spectral models for estimating grassland
343 vegetation coverage on Hulunbeier grassland, Inner Mongolia, China. *China Environmental Science*.
344 2010;30: 1287–1292 (in Chinese with English abstract).
345 http://en.cnki.com.cn/Article_en/CJFDTotal-ZGHJ201009033.htm.

- 346 20. P. Sellers. Canopy reflectance, photosynthesis and transpiration. *Remote Sensing of Environment*. 1985;
347 21:143–183. <https://doi.org/10.1080/01431168508948283>.
- 348 21. P. Sellers. Canopy reflectance, photosynthesis and transpiration. II. The role of biophysics in the linearity of
349 the interdependence. *International journal of remote sensing*. 1987;6(8):1335–1372.
- 350 22. K. McGwire, Ti. Minor, L. Fenstermaker. Hyperspectral Mixture Modeling for Quantifying Sparse
351 Vegetation Cover in Arid Environments. *Remote sensing of environment*. 2000; 72:360–374.
352 [https://doi.org/10.1016/S0034-4257\(99\)00112-1](https://doi.org/10.1016/S0034-4257(99)00112-1).
- 353 23. O. Mutanga. Hyperspectral remote sensing of tropical grass quality and quantity. PhD thesis, Wageningen
354 University, Wageningen, The Netherlands. Available online(2004) at:
355 <http://library.wur.nl/wda/abstracts/ab3565.html>.
- 356 24. Chengming Zhang, Jiping Liu, Fan Yu, Shujing Wan, Yingjuan Han, Jing Wang, Gang Wang. Segmentation
357 model based on convolutional neural networks for extracting vegetation from Gaofen-2 images. *Appl*
358 *Remote Sens*. 2018;12(4):042804. <https://doi.org/10.1117/1.JRS.12.042804>.
- 359 25. O. Beerli, R. Phillips, J. Hendrickson, A.B. Frank, S. Kronberg. Estimating forage quantity and quality using
360 aerial hyperspectral imagery for northern mixed-grass prairie. *Remote Sensing of Environment*. 2007;110:
361 216–225. <https://doi.org/10.1016/j.rse.2007.02.027>.
- 362 26. F. Fava, R. Colombo, S. Bocchi, M. Meroni, M. Sitzia, N. Fois, C. Zucca. Identification of hyperspectral
363 vegetation indices for Mediterranean pasture Characterization. *International Journal of Applied Earth*
364 *Observation and Geoinformation*. 2009; 11:233–243. <https://doi.org/10.1016/j.jag.2009.02.003>.
- 365 27. C.J.Tucker. Red and photographic infrared linear combinations for monitoring vegetation. *Remote Sensing*
366 *of Environment*. 1979; 8: 127–150. [https://doi.org/10.1016/0034-4257\(79\)90013-0](https://doi.org/10.1016/0034-4257(79)90013-0).
- 367 28. N.T. Boelman, M. Stieglitz, K.L. Griffin, G.R. Shaver. Interannual variability of NDVI in response to
368 long-term warming and fertilization in wet sedge and tussock tundra. *Oecologia*. 2005; 143:588–597.
369 <https://doi.org/10.1007/s00442-005-0012-9>.
- 370 29. X. Gao, A. Huete, W. Ni, T. Miura. Optical-biophysical relationships of vegetation spectra without
371 background contamination. *Remote Sensing of Environment*. 2000; 74: 609–620.
372 [https://doi.org/10.1016/S0034-4257\(00\)00150-4](https://doi.org/10.1016/S0034-4257(00)00150-4).
- 373 30. S.W. Todd, R.M. Hoffer. Responses of spectral indices to variations in vegetation cover and soil background.
374 *Photogrammetric Engineering & Remote Sensing*. 1998; 64:915–921.
- 375 31. K. Kawamura. Quantifying grazing intensities using geographic information system and satellite remote
376 sensing in the Xilingol steppe region, Inner Mongolia, China. *Agriculture, Ecosystems and Environment*.
377 2005; 107:83–93. <https://doi.org/10.1016/j.agee.2004.09.008>.
- 378 32. A.R. Huete, C. J.Tucker. Investigation of soil influences in AVHRR red and near-infrared vegetation index
379 imagery. *International journal of remote sensing*. 1991;12: 1223–1242.
380 <https://doi.org/10.1080/01431169108929723>.
- 381 33. B.R. Zhao, C. Liu, A.J. Liu, Z.X. Wang. Estimate the yield of grassland using MODIS–NDVI – a case study
382 of the grassland in Xilinguole in Inner Mongolia. *Pratacultural Science*. 2004; 21:12–14 (in Chinese with
383 English abstract). http://en.cnki.com.cn/Article_en/CJFDTOTAL-CYKX200408003.htm.
- 384 34. Z.X. Wang, C. Liu, B.R. Zhao, A.J. Liu. ANPP estimate from Modis-EVI for the grassland region of
385 Xilingol, China. *Journal of Lanzhou University (Natural Sciences)*. 2005; 41:10–16 (in Chinese with
386 English abstract).
387 http://en.cnki.com.cn/article_en/cjfdtotal-ldzk200502003.htm.
- 388 35. Z.Q. Du, J. Wang, Y.D. Shen. Estimating model of grassland above-ground biomass based on Remote
389 SensedData in Shandan County. *Remote Sensing Technology and Application*. 2006;21: 338–343(in

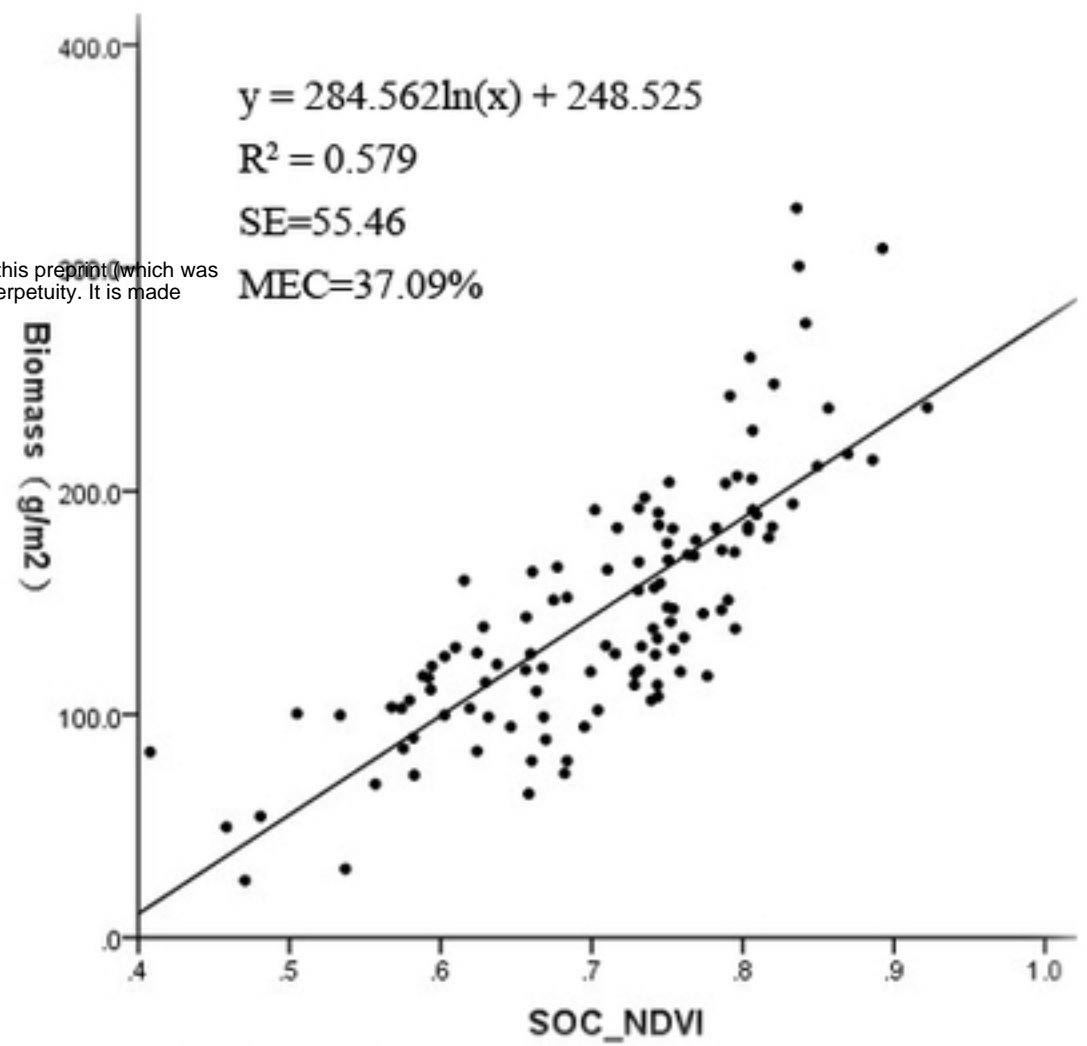
- 390 Chinese with English abstract). http://en.cnki.com.cn/Article_en/CJFDTOTAL-YGJS200604012.htm.
- 391 36. B.Xu, X.C. Yang. Calculation of grass production and balance of livestock carrying capacity in rangeland
392 region of Northeast China. Geographical Research. 2009; 28:402–408 (in Chinese with English abstract).
393 http://en.cnki.com.cn/Article_en/CJFDTotal-DLYJ200902017.htm.
- 394 37. A. Swatantran, R. Dubayah, D. Roberts, M. Hofton, J.B. Blair. Mapping biomass and stress in the Sierra
395 Nevada using lidar and hyperspectral data fusion. Remote Sensing of Environment.2011;115:2917–2930.
396 <https://doi.org/10.1016/j.rse.2010.08.027>.
- 397 38. J.W. Rouse, R.H. Haas, D.W. Deering, J.A. Schell, J.C. Harlan. Monitoring the vernal advancement and
398 retrogradation (green wave effect) of natural vegetation. NASA/GSFC Type III Final Report, Greenbelt,
399 (1974). <https://ntrs.nasa.gov/search.jsp?R=19730017588>.
- 400 39. L.F. Jin, X.R.Xu, M.Zhang. The spectral reflectance model for estimation of green biomass of the typical
401 grassland zone in inner Mongolia. Journal of Inner Mongolia University (Acta Scientiarum Naturalium
402 Universitatis Neimongol).1986;17:735–740 (in Chinese with English abstract).
403 http://en.cnki.com.cn/Article_en/CJFDTotal-NMGX198604019.htm.
- 404 40. P.J. Shi, B.Li, Z.H.Li, T.Hu. A research on the production estimation by remote sensing for large area
405 grassland. Acta Agrestia Sinca.1994; 2: 9–12.(in Chinese with English abstract).
406 http://en.cnki.com.cn/Article_en/CJFDTOTAL-CDXU199401001.htm.
- 407 41. Y.R.Wang, W.Y Yong. A correlative analysis between near-ground spectral reflectance and yield for Stipa
408 gobica Desert-steppe. Journal of Inner Mongolia University (Acta Scientiarum Naturalium Universitatis
409 Neimongol). 1996;27: 664–669 (in Chinese with English
410 abstract).http://en.cnki.com.cn/Article_en/CJFDTOTAL-NMGX605.017.htm.
- 411 42. Mutanga, A.K.SKIDMORE. Narrow band vegetation indices overcome the saturation problem in biomass
412 estimation. International Journal of Remote Sensing. 2004;25: 3999–4014.
413 <https://doi.org/10.1080/01431160310001654923>.



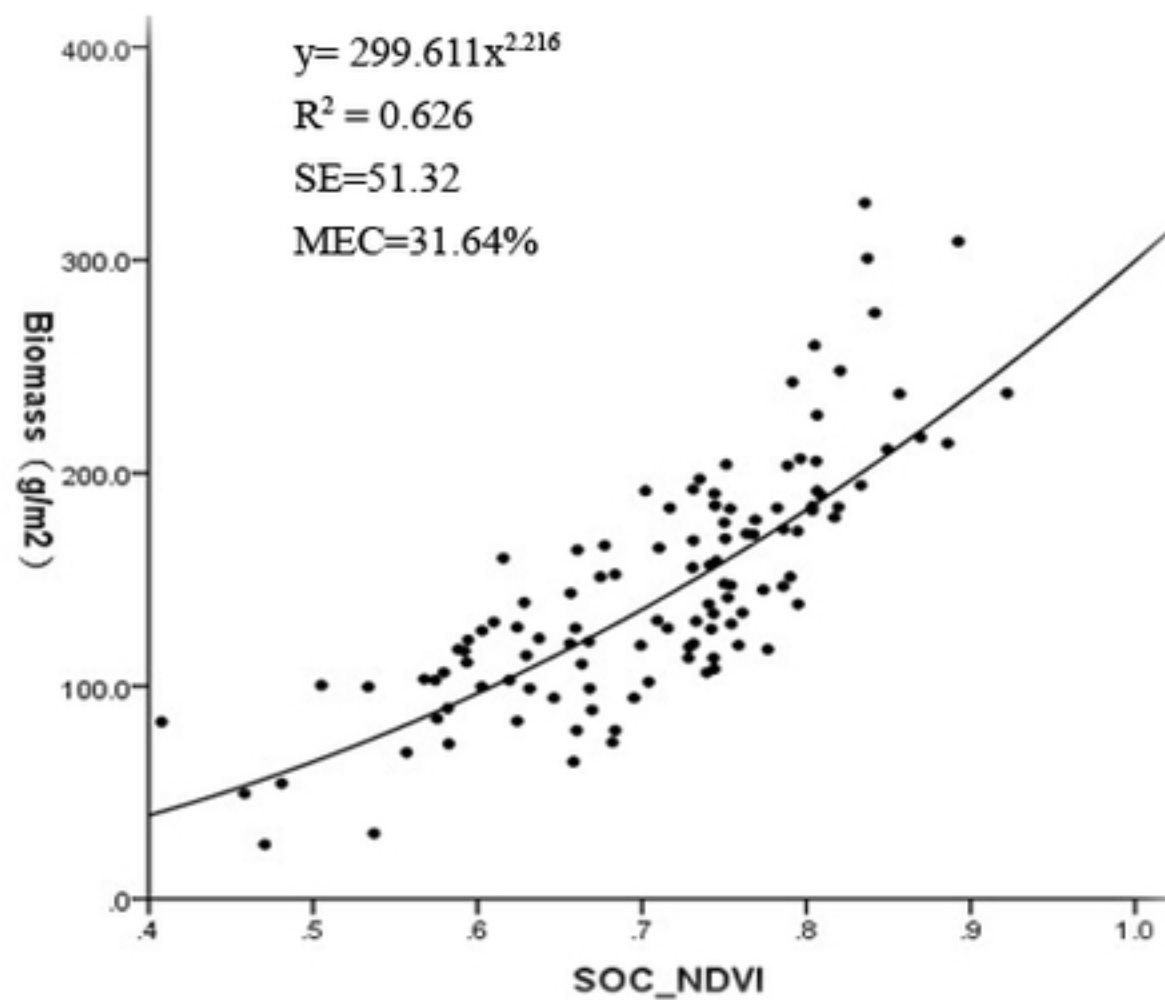
Figure



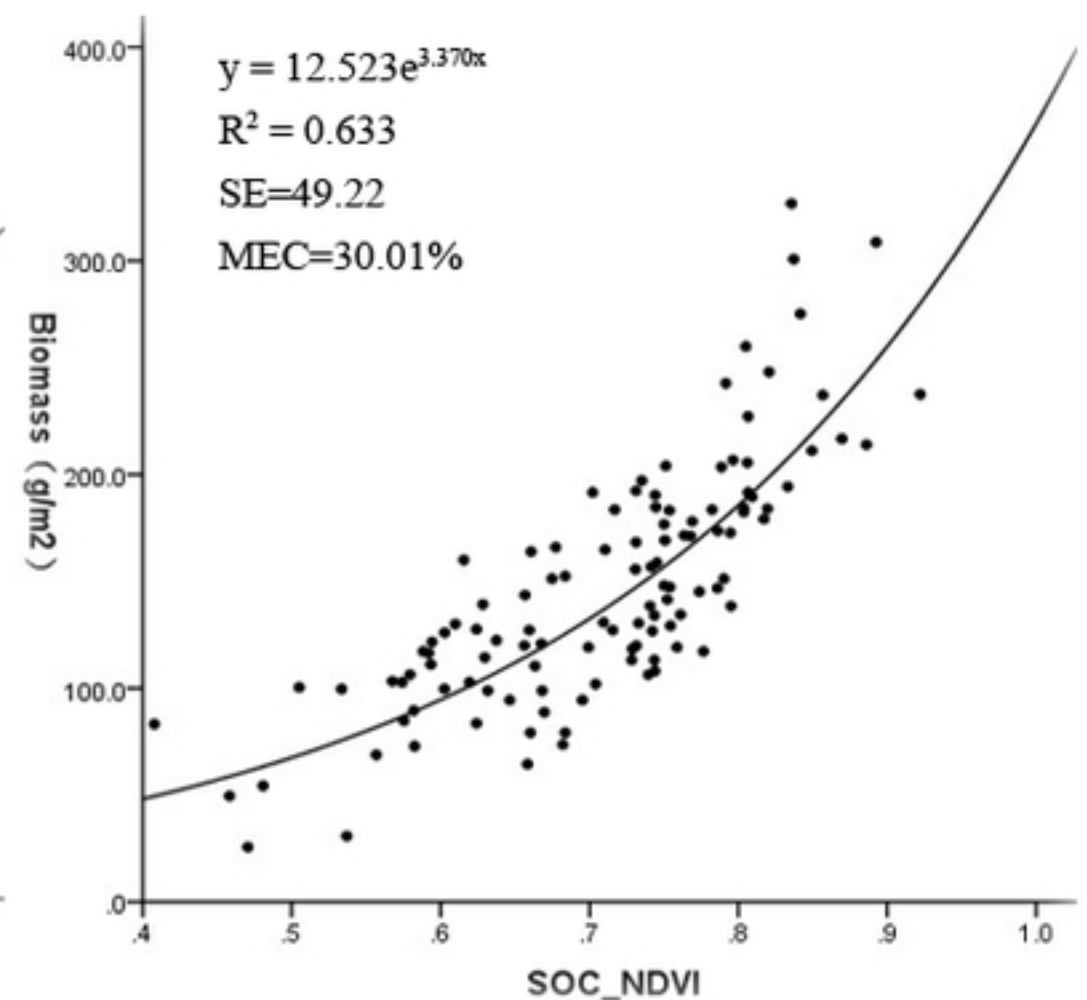
A. Linear



B. Logarithmic



C. Power



D. Exponential

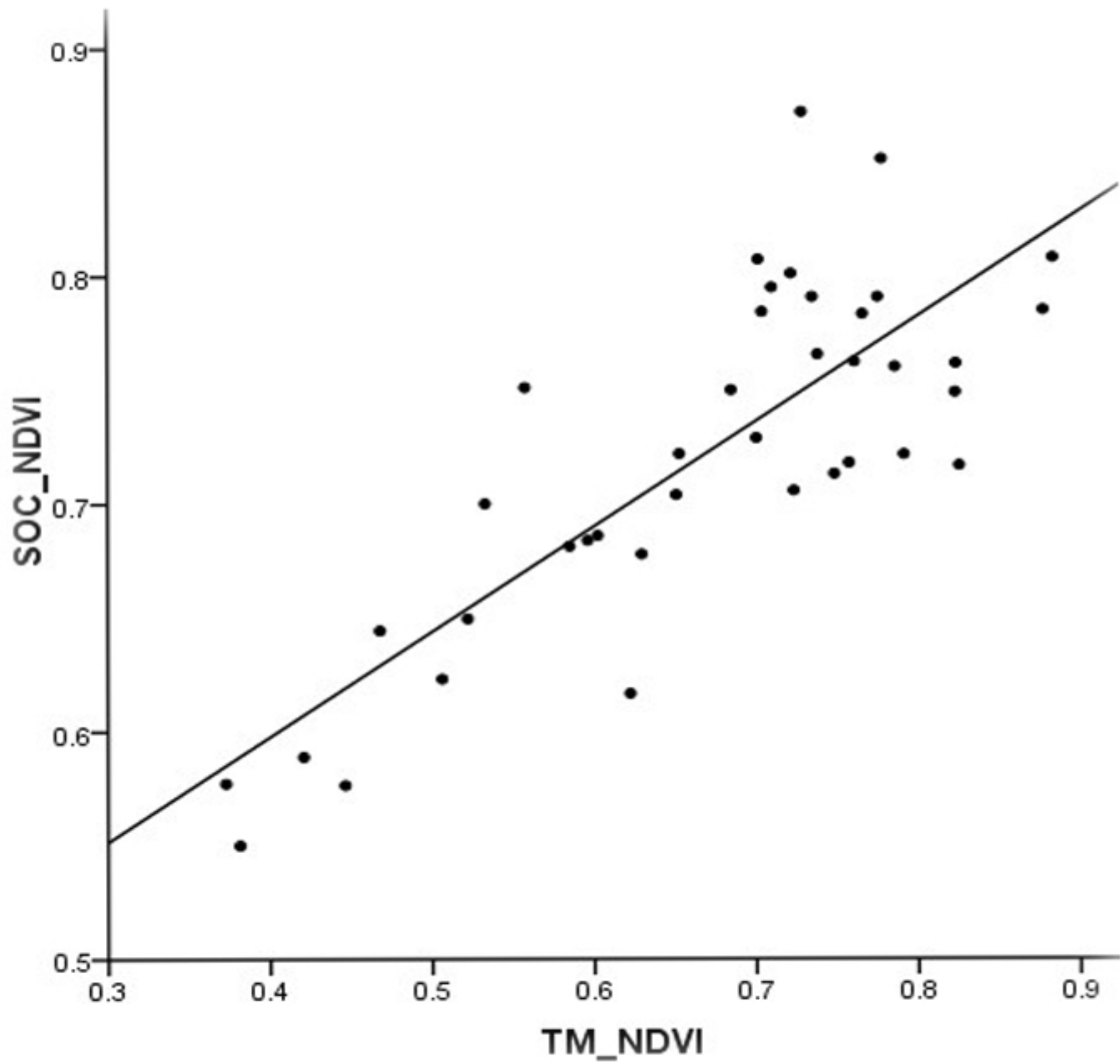


Figure3

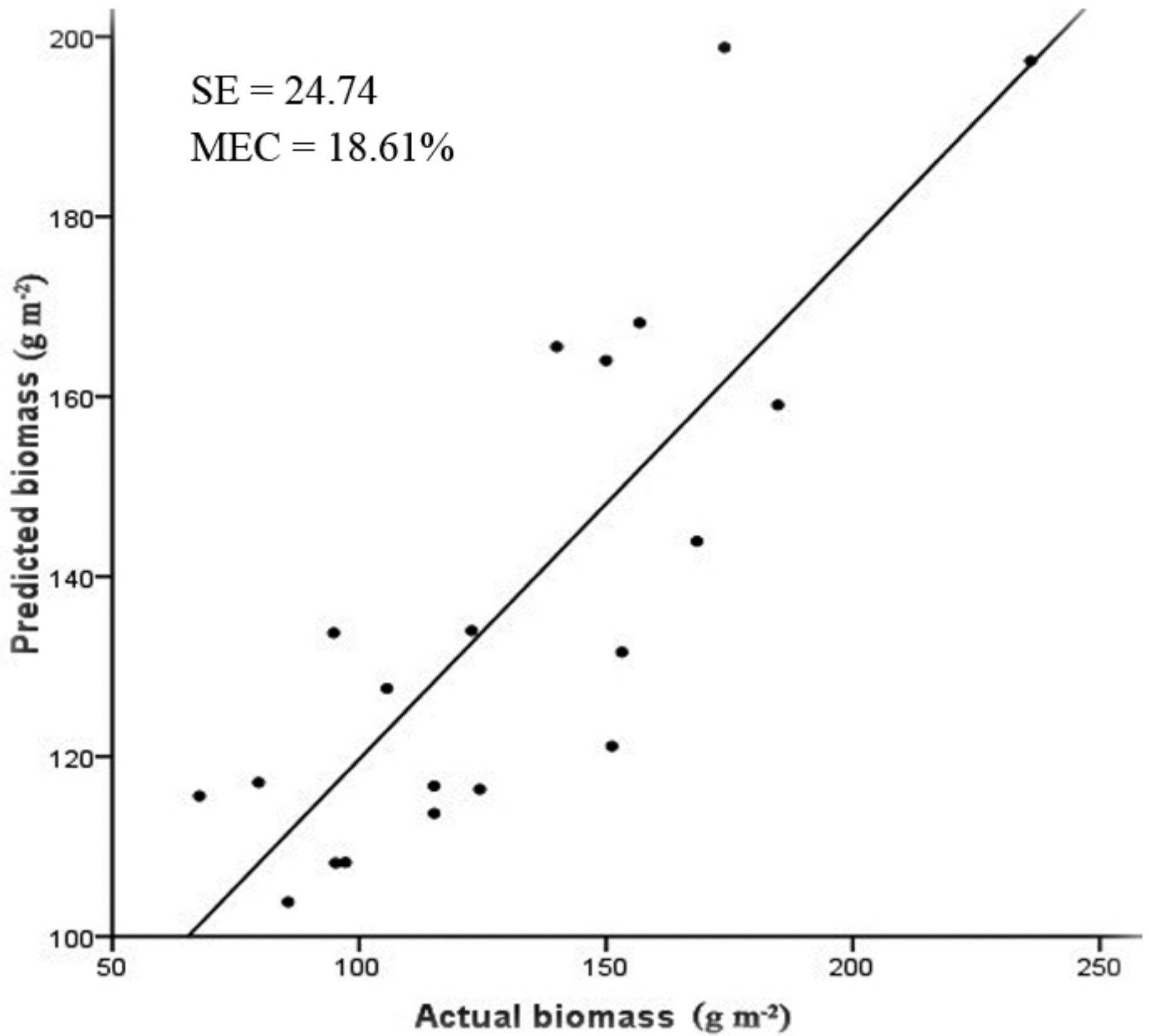


Figure4

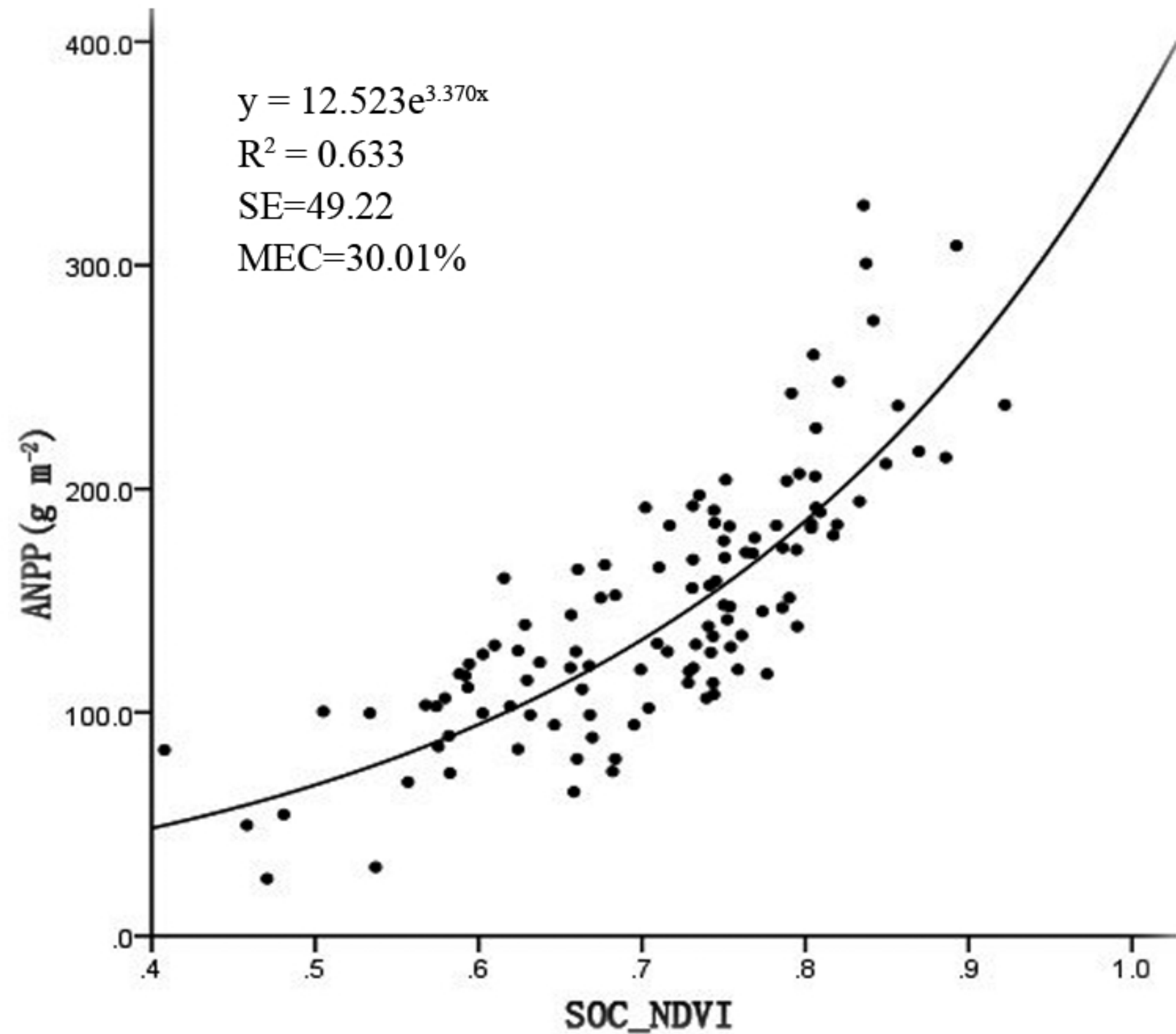


Figure5