Trait-based approach to bacterial growth efficiency

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ABSTRACT

Bacterial growth efficiency (BGE) is the proportion of assimilated carbon that is converted into biomass and reflects the balance between growth and energetic demands. Often measured as an aggregate property of the community, BGE is highly variable within and across ecosystems. To understand this variation, we focused on how species and resource identity affect BGE using a trait-based approach with 20 bacterial isolates enriched from north temperate lakes. We used phenotypic and genomic approaches to characterize the metabolism of each isolate and test for predicted trade-offs between growth rate and efficiency. Regardless of resource identity, 20 % of the variation in BGE could be attributed to the taxonomic class of the bacterial species, and 58 % of the variation could be explained by species identity. Meanwhile, resource identity explained a relatively small amount of variation (7 %) in BGE across species but accounted for > 60 % of the variation within a species. Metabolic trade-offs and genomic features suggest that BGE is a predictable trait that is independent of resource type. As such, genomic and phylogenetic information from microbiomes may help predict aggregate community functions like BGE and the fate carbon in ecosystems.

34 INTRODUCTION

36 In most ecosystems, heterotrophic bacteria play a pivotal role in determining whether
organic carbon is respired as CO₂ or converted into biomass and retained in food webs (Pomeroy
et al. 1998; Bardgett et al. 2008; Ducklow 2008). Many factors control how bacteria process
38 carbon, but perhaps the most important is reflected by bacterial growth efficiency (BGE). BGE is
the proportion of assimilated organic carbon that is converted into bacterial biomass (del Giorgio
40 and Cole 1998). When BGE is high, more carbon is turned into bacterial biomass where it can be
retained for longer periods of time while also serving as a resource for other members of the food
42 web. In contrast, when BGE is low, microbially assimilated carbon has a shorter residence time
and is released into the environment as CO₂. When measured at the community scale, BGE is
44 notoriously variable among habitats and has proven difficult to predict (del Giorgio and Cole
1998).

46 While BGE can be influenced by a range of chemical and physical properties (Apple and
del Giorgio 2007; Hall and Cotner 2007; del Giorgio and Newell 2012; Sinsabaugh et al. 2013;
48 Geyer et al. 2016), it may also reflect taxon-specific differences in microbial metabolism. For
example, BGE results from the physiological balance between cellular growth and energetic
50 demands. As such, bacterial growth strategy is predicted to constrain BGE via physiological
trade-offs (Litchman et al. 2015), and it has been hypothesized that oligotrophs have higher
52 maximum growth efficiency than copiotrophs (Roller and Schmidt 2015). In addition, species
that specialize on only a few resources are predicted to be more efficient at using those resources
54 than generalist species (Dykhuizen and Davies 1980; Glasser 1984). As such, traits such as
maximum growth rate and the number of resources used (i.e., niche breadth) could underlie
56 species-specific differences in BGE.

Variation in BGE may also be influenced by the properties of resources that are used to meet energetic and growth demands. Different resources can affect cellular ATP yield depending on the metabolic pathways used (Fuhrer et al. 2005; Flamholz et al. 2013). For example, glucose is metabolized via glycolysis, but growth on aromatic compounds, such as protocatechuate, requires the β -ketoadipate pathway which yields less ATP (Gottschalk 1986). Furthermore, energy-producing catabolic processes and biomass-producing anabolic processes are not independent (Russell and Cook 1995; Kempes et al. 2012). For example, cells have the potential to produce > 30 ATP from a single glucose molecule if it is completely oxidized, but there would be no remaining carbon to yield new biomass. Instead, cells must use the intermediate products of glycolysis to form proteins and other cellular material, which diminishes the maximum ATP yield (Gottschalk 1986; Flamholz et al. 2013). In addition, biomass production requires materials and energy. For example, the synthesis of proteins, which constitute ~70 % of cellular dry mass, requires amino acid building blocks and 4 ATP per peptide bond (Tempest and Neijssel 1984; Gottschalk 1986; Lynch and Marinov 2015). Therefore, because resources differ in their potential energy yield and bacteria differ in their ability to extract energy and form biomass from a given resource, BGE should vary based on the resources available to bacteria.

In this study we measured BGE in a set of bacterial isolates growing on three different carbon resources that varied in chemical structure and metabolic pathway (Fig. 1). In addition to partitioning variation in BGE based on species and resource identity, we tested for hypothesized trade-offs with growth rate and niche breadth. Furthermore, using the genomes of each isolate, we evaluated whether metabolic pathways could explain differences in BGE among diverse representatives of aquatic bacteria from north temperate lakes. Last, to test if resources have differential effects on the metabolic traits that underlie BGE (i.e., production and respiration), we

tested for resource-specific relationships between respiration and production rates for each resource. Together, our trait-based approach provides a framework for understanding linkages between community structure and function due to the physiological constraints on BGE and suggest that large changes in community composition or available resources may alter BGE and therefore carbon cycling in predictive ways.

RESULTS

Bacterial Growth Efficiency — Using measures of bacterial productivity (BP) and respiration (BR), we calculated bacterial growth efficiency (BGE) for 20 isolates growing on three different resources: glucose, succinate, and protocatechuate (Fig. 1). All isolates belonged to the Proteobacteria phylum with representatives from the Alpha-, Beta-, and Gammaproteobacteria subphyla (Fig. 2, Fig. S1). Across isolates, we found that BGE ranged from <0.01 to 0.32 (Fig. 2). Based on linear mixed-effects models we found that isolate and resource identity explained substantial amount of variation in BGE. Across resources, isolate identity explained 58 % of the variation in BGE, and 67 % of the variation within resource, and the taxonomic order of each isolate explained 20 % of the variation in BGE across resources, and 28 % of the variation within resource. Across isolates, resource identity only explained 7 % of the variation in BGE across all isolates, but 63 % of variation within isolate.

Next, we determined if there was any evidence of phylogenetic signal for BGE among our isolates. Based on Blomberg's K (a measure of phylogenetic signal strength), there was weak phylogenetic signal for BGE. We did not detect phylogenetic signal for BGE when isolates used succinate ($K = 0.002$, $p = 0.24$) or protocatechuate ($K = 0.001$, $p = 0.146$), but there was a

significant phylogenetic signal when isolates used glucose ($K = 0.002$, $p = 0.04$). However, the low K value suggests that BGE is over-dispersed (i.e., less phylogenetic signal than expected under Brownian motion). Furthermore, when we use Pagel's λ (a measure of phylogenetic signal compared to Brownian motion), we found no evidence of a signal (Glucose: $\lambda = 0.10$, $p = 0.76$; Succinate: $\lambda = 0.13$, $p = 0.66$; Protocatechuate: $\lambda = < 0.01$, $p = 0.99$).

Last, we determined if the values of BGE observed across isolates and resources were unimodal. Based on Hartigan's dip test, we found that there was a bimodal distribution of BGE among our isolates ($D = 0.07$, $p = 0.58$, Fig. S2). Using this distribution, we split isolates into two groups, which we define as the "high BGE", and "low BGE" groups.

Phenotypic Comparisons — Using linear models, we found phenotypic differences between isolates that were related to BGE (Fig. 3). While there was no relationship between BGE and maximum growth rate ($\mu_{\max}$; $F_{1,7} = 0.035$, $r^2 < 0.01$, $p = 0.86$) in the low-BGE group of bacteria, we identified a significant negative relationship between BGE and $\mu_{\max}$ for the high-BGE group ($F_{1,7} = 7.79$, $r^2 = 0.53$, $p = 0.027$). Based on this model we would predict a 2.6 % decrease in BGE for each per minute increase in $\mu_{\max}$ in the high-BGE group. In contrast to our predictions, there was no relationship between niche breadth (Levin's Index) and BGE for the low-BGE group ($F_{1,7} = 1.42$, $r^2 = 0.17$, $p = 0.27$) or the high-BGE group ($F_{1,7} = 0.92$, $r^2 = 0.11$, $p = 0.37$).

Genomic Comparisons — We found genomic differences between isolates that were related to BGE. First, isolates in the high-BGE group had 26 % more annotated metabolic pathways (based on an 80 % module completion ratio cutoff) than isolates in the low BGE-group ($m_{\text{high}} = 66$, $m_{\text{low}} = 652$, t-test: $t_{18} = -2.64$, $p = 0.02$). Second, we found that the number of metabolic

pathways was related to the observed BGE when grown on glucose, but the relationship depended on the group. For the high-BGE group there was a negative relationship between BGE and the number of pathways ($\beta = -0.006$, $r^2 = 0.48$, $p = 0.04$), but for the low-BGE group there was a positive relationship ($\beta = 0.0003$, $r^2 = 0.37$, $p = 0.05$). Next, we found that differences in the metabolic pathway composition of a genome could help explain which group an isolate belongs, and pathway composition was related to the observed value of BGE. Specifically, we found three pathways that were indicators of an isolate being in the high-BGE group (Table 1). Likewise, for the high-BGE group we found that we could explain 24 % of the variation in BGE based on the composition of metabolic pathways (dbRDA: $F_{1,7} = 2.17$, $R^2 = 0.24$, $p = 0.05$). We found eight pathways with significant positive or negative correlations ($|\rho| > 0.7$) with BGE (Table 2). However, we not find a relationship between pathway composition and BGE for the low-BGE group ($p = 0.45$).

Resource Effects — Indicator variable linear regression revealed a positive relationship between respiration and production rates (Fig. 4, $F_{9,42} = 8.07$, $R^2 = 0.63$, $p < 0.001$) with there being a higher y-intercept for the high-BGE group of isolates ($\beta_{\text{Group}} = 2.7$, $p < 0.001$). Resource identity, however, had no effect on the BR-BP relationship or the effect of group (all $p > 0.25$). Last, we did not find evidence that the slope of the BR-BP relationship was different from one ($t_{42} = 0.76$, $p = 0.45$); therefore, the two measures of bacterial metabolism appear to scale proportionately with one another.

DISCUSSION

We measured bacterial growth efficiency (BGE) in 20 environmental bacterial isolates grown on three resources that varied in their bioavailability, structure, and pathways required for metabolism (Fig. 1). While BGE varied across isolates, there was no evidence for a strong phylogenetic signal in efficiency. However, a substantial amount (20 %) of the variation in BGE could be explained by an isolate's taxonomic order while a much smaller amount of the variation (7 %) could be attributed to resource identity (Fig. 3). We found evidence for the predicted trade-off between maximum growth rate and efficiency, but only on the most labile resource (glucose) (Fig. 3). Even though we found that resource identity explained 63 % of the variation in BGE within an isolate, resource identity did not alter the relationship between respiration and production rate observed across isolates (Fig. 4) suggesting that resource identity has a stronger effect on BGE within a species than across species. Together, we propose that growth efficiency is a physiological trait independent of resource identity, but resource characteristics may modify species-specific physiological performances. We propose that taxonomic groups of bacteria may have fundamentally different growth efficiencies such that changes in community composition may alter the fate of carbon resources (i.e., biomass versus CO₂) within the ecosystem.

Bacterial Growth Efficiency as a Trait

Our results indicate that there are species-specific properties regulating BGE, which may be conserved at higher taxonomic levels. This conclusion is consistent with the view that BGE represents a complex bacterial trait (i.e., aggregate property of numerous cellular functions) with ecological significance, and that different groups of bacteria have fundamentally different strategies for carbon allocation. Our phylogenetic analyses suggest that BGE may be an over-dispersed trait (at least with glucose) such that the efficiency of closely related bacteria may be

less similar than expected. One potential explanation for this pattern of over-dispersion is that our culture collection lacked phylogenetic resolution within some of our taxonomic groups (e.g., Betaproteobacteria) or that the variation in BGE within a taxonomic group (e.g., order) may not be the same across taxonomic groups. Alternatively, BGE may not be a phylogenetically conserved trait. Though some traits such as phosphorus acquisition, photosynthesis, and methanogenesis are phylogenetically conserved deep in the microbial tree of life (Martiny et al. 2006, 2013), others such as complex carbon metabolism are not (Zimmerman et al. 2013). Therefore, it is possible that BGE may be similar to traits such as complex carbon metabolism that are not deeply conserved, which appears to be common among complex traits (Martiny et al. 2015). Regardless, our data reveal that BGE is a complex bacterial trait that is influenced by taxonomic affiliation. As such, it may be possible to make predictions about BGE and other ecosystem functions given information about composition of resident microbiomes (Goberna and Verdú 2016).

Bacterial Growth Efficiency on Different Resources

Differences in resource complexity and the metabolic pathways required for degradation may explain species-specific differences in BGE due to resource identity. Within an isolate, resource identity accounted for 63 % of the variation in BGE. Given that different resources are processed via different metabolic pathways, resource-based variation in BGE within a species is expected. For example, BGE was higher when isolates were supplied with glucose compared to when they were supplied with protocatechuate. Glucose is a simple sugar that is able to be metabolized by numerous pathways and converted to acetyl-CoA (Neidhardt 2007). Protocatechuate, on the other hand, is a complex aromatic compound that requires a specific metabolic pathway to be

converted to acetyl-CoA. Furthermore, because protocatechuate is chemically more complex, it requires more energy (i.e., ATP) to be degraded than more labile resources such as glucose (Harwood and Parales 1996). Therefore, resource complexity and the metabolic pathways required may explain the within-isolate variation in BGE. However, across isolates, we did not find resource-specific differences in the relationship between respiration and production rate. Such findings suggest that energetic demands required to use different resources may also be species-specific trait. That is, the energetic demands for individual species may be highly constrained and therefore not change much when growing on different resources. Together, these findings suggest that the effect of resources on the efficiency of entire microbiomes may depend on the composition of bacteria consuming those resources.

Bacterial Growth Efficiency Groups

Across all isolates, we found a bimodal distribution of BGE suggesting that there were two distinct groups with contrasting efficiencies. One group had low BGE ($< 5\%$) across all treatments, and the other group ranged in BGE from 7-30 % (Fig. 2 & 3). Although, the range of BGE measured across isolates is similar to the range observed in many ecosystems (del Giorgio and Cole 1998), our results suggest that some species of bacteria grow relatively inefficiently, irrespective of resource quality. One explanation is that the minimum cellular energetic demand (i.e., cellular maintenance costs) is higher in some bacteria than others (Russell and Cook 1995). Furthermore, energetic demand may be higher when bacteria are grown in minimal media where they must produce all cellular components from a single carbon resource (Tao et al. 1999). Alternatively, nutrient concentrations (e.g., phosphorus) and other physical properties (e.g., temperature) may regulate efficiency (Smith and Prairie 2004; Frey et al. 2013) and the effects of

these properties may be species-specific. As such, it is possible that maintenance costs, resource
 218 imbalances, and the physical growth conditions affected BGE of our isolates. Differences in low-
 BGE and high-BGE isolates was also reflected in genomic content, including the number and
 220 presence-absence of metabolic pathways. Together, these findings suggest that there are
 fundamental differences between bacterial species that determine BGE. However, these
 222 genomic features seem to best explain large-scale rather than fine-scale differences in BGE.

224 **Physiological Trade-Offs**

We found evidence to support a trade-off between maximum growth rate and BGE (Fig. 4),
 226 which is predicted in microbial and non-microbial systems (Glasser 1984; Roller and Schmidt
 2015). For example, theoretical models of microbial communities predict a rate-efficiency trade-
 228 off (Allison 2014), and this trade-off has been observed across microbial taxa (Lipson 2015).
 Physiologically, the trade-off is based on allocation constraints imposed by the balance between
 230 energy requirements and biomass yield: organisms with higher maximum growth rates may have
 more energetic requirements and thus lower BGE (Russell and Baldwin 1979; Russell and Cook
 232 1995). Furthermore, processes that limit respiration, such as oxygen availability, have been
 shown to suppress bacterial growth rate (Meyenburg and Andersen 1980). Therefore, respiration
 234 rate is likely a major control on biomass production and BGE. Consistent with this, we observed
 a scaling relationship between respiration and production rates (Fig. 4). Furthermore, this
 236 relationship between respiration and production and the non-zero intercept suggest that there is a
 minimum respiration rate required before any biomass can be produced, which is commonly
 238 interpreted as the cellular maintenance requirement. Therefore, it is possible that the maintenance

energy demand of bacterial species explains the physiological trade-off between maximum
 240 growth rate and growth efficiency.

Theory also predicts a trade-off between resource niche-breadth and growth efficiency
 242 (Glasser 1984). This trade-off is based on the assumption that there is an energetic cost to
 maintaining numerous metabolic pathways (Johnson et al. 2012). As such, species with more
 244 metabolic pathways should have more energetic requirements and thus lower BGE; although, the
 effects of genome reduction has been debated (Giovannoni et al. 2005; Livermore et al. 2014). In
 246 this study, we did not find evidence of a trade-off between resource niche breadth and BGE (Fig.
 4). One possible explanation is that the resources used in our phenotypic assay (i.e. Ecolog
 248 plates) did not reflect the full metabolic potential of our isolates. Alternatively, there may not be
 a strong trade-off between niche breadth and efficiency, but further experiments with additional
 250 isolates and resources would be required to test this prediction more rigorously.

252 **Genomic Signatures**

In addition to the physiological differences documented among our isolates, we found genomic
 254 evidence of metabolic pathways that are associated with BGE. Specifically, we found genomic
 differences between isolates that belong to low-BGE and high-BGE groups. We discovered that
 256 isolates in the high-BGE group had 26 % more metabolic pathways than the low-BGE group.
 Furthermore, we identified three pathways that were unique to the high-BGE group (Table 1)
 258 and a number of pathways that were correlated with the observed BGE (Table 2). Together, our
 findings suggest that there are genomic features that may contribute to or regulate BGE.
 260 In general, the genomic composition of BGE groups appear to reflect differences in cellular
 biosynthesis. It is possible that species with particular biosynthesis pathways may generate

essential cellular components with less energetic demand. For example, the low-BGE isolates lacked some metabolic pathways, including pyridoxal biosynthesis and histidine degradation, which were present in the high-BGE group. The pyridoxal biosynthesis pathway produces vitamin B₆ from erythrose-4-phosphate (Mukherjee et al. 2011). Because vitamin B₆ is essential for growth, the isolates lacking the pyridoxal pathway must use alternatives such as uptake from the environment if they are auxotrophic (i.e., unable to synthesize) or other synthesis pathways such as the Deoxyxylulose-5-phosphate synthase (DXS) pathway (Mukherjee et al. 2011). However, the DXS pathway requires pyruvate (a precursor for Krebs cycle) and thus may limit central metabolism and possibly lead to lower BGE. Likewise, the histidine degradation pathway is used to breakdown histidine into ammonium and glutamate (Bender 2012). Alternatively, glutamate can be synthesized from α -ketoglutarate; however, because α -ketoglutarate is an intermediate component of Krebs cycle this may limit central metabolism and possibly lead to reduced BGE.

Conclusion

At the cellular level, BGE reflects the balancing energetic and cellular growth demands. We find evidence of this based on physiological trade-offs (i.e., maximum growth rate) as well as metabolic pathways. As such, changes in community composition and resource availability have the potential to alter food web and ecosystem function due to changes in BGE. For example, communities dominated by species with low BGE should yield a net release of CO₂ from the ecosystem. Alternatively, communities comprised of individuals with high BGE should yield a net increase in ecosystem productivity. However, variation in BGE can arise within a species due to the ways in which it processes different resources. Therefore, changes in the resource supply will alter the performance of individual taxa, but we predict that these changes will not be as

strong as changes in BGE that arise owing to differences in community composition. A trait-based approach can be used to provide a mechanistic link between the structure and function of bacterial communities.

METHODS

Bacterial Isolates — Using a novel cultivation approach, we isolated 20 bacterial strains from lakes in the Huron Mountain Research Preserve (Powell, MI, USA) by incubating inert carbon beads (Bio-Sep Beads) in the water column for one week. Prior to the incubations, the beads were saturated with a sterile complex carbon substrate, i.e., Super Hume (CropMaster, United Agricultural Services of America, Lake Panasoffkee, Florida, USA). Super Hume is a lignin-rich resource comprising 17 % humic and 13 % fulvic acids, and has been shown to be an analog of terrestrial DOC in aquatic ecosystems that can be used by diverse bacteria (Lennon et al. 2013). We used this enrichment technique to select for bacteria with a range of metabolic potentials (Ghosh et al. 2009). After the incubation, beads were rolled on R2 agar plates (BD Difco, Sparks Maryland, USA) and incubated at 25 °C. We picked random colonies from plates and serially transferred until axenic. All isolates were preserved in 25 % glycerol at -80 °C.

We identified each bacterial strain by direct sequencing the 16S rRNA gene. We obtained genomic DNA from log phase cultures using the FastPrep DNA extraction kit according to the manufacturer's specifications (MP Biomedical). We used 10 ng of genomic DNA to amplify the 16S rRNA gene using the 27F and 1492R bacterial primers (See Supplemental for primer sequences and PCR conditions). We sequenced the PCR products at the Indiana Molecular Biology Institute (IMBI) at Indiana University (Bloomington, Indiana, USA). Raw sequence reads were quality-trimmed based on a Phred quality score of 25. Forward and reverse reads

were manually merged after aligning sequences to the Silva 16S SSU rRNA reference alignment (release 132) using SINA v. 1.2.11 and the Bacteria variability profile (Pruesse 2011). After merging into full length 16S rRNA sequences, we used mothur (Schloss et al. 2009) to check the quality of sequences and alignments were checked using the ARB software package (Ludwig et al. 2004). Finally, sequences were compared to the Silva All-Species Living Tree Project database (Yilmaz et al. 2014) for taxonomic identification (Fig. S1).

Bacterial Growth Efficiency — To test for differences in bacterial growth efficiency (BGE), we measured BGE for each isolate when grown on each of three different carbon substrates: glucose, succinate, or protocatechuate (Fig. 2). These carbon sources (i.e., resources) were chosen based on differences in their bioavailability and structure but also the required pathways for metabolism (see Fig. 2).

To determine BGE we measured bacterial respiration and production rates and then calculated BGE as $BP/(BP + BR)$, where BP is bacterial productivity and BR is bacterial respiration (del Giorgio and Cole 1998). BP and BR were measured using triplicate cultures of each isolate. Cultures of each isolate were grown in R2 broth (BD Difco, Sparks Maryland, USA) until mid-log phase. We then transferred 100 μ L of culture into 10 mL of M9 broth (Green and Sambrook 2012) with the appropriate carbon source (25 mM C) and allowed 24 h for the cultures to acclimate. We then transferred 100 μ L of culture into 10 mL of fresh carbon-amended M9 broth and incubated 1-3 h to replenish nutrients. Using these transfers, we were able to establish populations of each isolate at target cell densities between 10^4 and 10^5 cells mL^{-1} . We used the populations to measure BP and BR, which were normalized to cell density using plate counts of colony forming units. We measured BP using the ^3H -Leucine assay (Smith and Azam

1992) with 1.5 mL of culture. We added ^3H -Leucine to a final concentration of 50 mM and
 332 incubated for 1 h. Following incubation, we terminated production with trichloroacetic acid
 (final concentration 3 mM) and measured leucine incorporation using a liquid scintillation
 334 counter. We measured BR using an automated O_2 measurement system (PreSens Sensor Dish
 System, PreSens, Regensburg, Germany) on 5 mL of culture based on the rate O_2 consumption
 336 during three-hour incubations. We estimated BR as the slope of O_2 concentration during the
 incubation using linear regression. We used theoretical respiratory quotients for each resource to
 338 convert O_2 depletion into C respiration assuming aerobic growth.

Taxonomic and Phylogenetic Relationships — We compared differences in BGE across
 isolates and resources using linear models. First, we used a taxonomic framework to compare
 342 BGE between isolates (Lennon et al. 2012). Isolates were classified into taxonomic groups based
 on the species tree constructed in ARB. We then used mixed linear models to compare BGE
 344 across taxonomic groups and resources. To test the hypothesis that taxonomy (i.e., at the class
 level) affects BGE, we nested resource identity within isolate in the linear model. To test the
 346 hypothesis that resource identity affects BGE, we nested isolate identity within resource. We
 identified the best statistical models based on the variation explained (R^2) and AIC values.
 348 Second, we tested if phylogenetic relationships between isolates explained differences in BGE
 across isolates. We created a phylogenetic tree based on the full-length 16S rRNA gene
 350 sequences. We aligned sequences using the SINA aligner (Pruesse et al. 2012) and checked
 alignments using ARB. We generated a phylogenetic tree using the CIPRES science gateway
 352 (Miller et al. 2010). The phylogenetic tree was created using RAxML v.8.2.12 (Stamatakis
 2006). We used the GTRGAMMA DNA substitution model and the rapid hill-climbing

algorithm to build our maximum likelihood trees, and we used the extended majority rule to find the consensus tree. We used Blomberg's K and Pagel's Lambda to compare trait variation across the tree and test if phylogenetic relationships between isolates could explain differences in traits (Pagel 1999; Blomberg et al. 2003). Blomberg's K is a test for phylogenetic signal that determines if trait variation is better explained by phylogenetic relationships or Brownian motion. Pagel's Lambda is a test of phylogenetic signal that determines if trait variation differs from Brownian motion. Last, to determine if the distribution of BGE across isolates was unimodal, we used Hartigan's dip test for unimodality (Hartigan and Hartigan 1985). Hartigan's dip test is used to determine if a distribution is unimodal by testing the null hypothesis that there is a dip in the distribution. A significant Hartigan's dip test would suggest that the distribution is unimodal. Alternatively, the distribution has an internal "dip" (reported as D). All statistical tests were conducted in the R statistical environment (R Core Development Team 2013). We used the nlme package (Pinheiro and Bates 2011) for the mixed linear models, the picante package (Kembel et al. 2015) for the phylogenetic methods, and the diptest package (Maechler 2015) for Hartigan's dip test.

Phenotypic Comparisons and Trade-offs — To test the hypothesis that phenotypic differences and physiological trade-offs underlie BGE variation, we compared the maximum growth rate (μ_{max}) and niche breadth of each isolate. First, to test whether BGE was affected by growth strategy (i.e., copiotrophs vs. oligotrophs), we measured the maximum growth rate of each isolate. Bacterial growth rates were measured based on changes in optical density during 18-h incubations. Bacterial isolates were grown in R2 broth in 48-well plates. We incubated plates with continuous shaking and measured optical density every 15 min using a plate reader (BioTek

MX). Growth curves were analyzed by fitting a modified Gompertz growth model (Zwietering et al. 1990; Lennon 2007) to the observed growth curves using maximum likelihood fitting. We used the model fit as our estimate of $\mu_{\max}$.

Second, to test whether BGE was affected by niche breadth, we generated carbon usage profiles using BioLog EcoPlatesTM (Garland and Mills 1991). The EcoPlate is a phenotypic profiling tool consisting of 31 unique carbon sources. In addition to the carbon source, each well contains a tetrazolium dye, which in the presence of NADH will be reduced and change color. We used this colorimetric assay to generate carbon usage profiles for each strain. We standardized profiles for each strain by subtracting water blanks (average water blank + 1 SD), and relativizing across substrates. Using these data, we calculated resource niche breadth using Levin's Index (Colwell and Futuyma 1971).

Genomic Comparisons — To test the hypothesis that variation in metabolic pathways could explain differences in BGE, we compared the genomes of each isolate. First, we determined the metabolic pathways found in the genome of each isolate. We characterized each isolate using whole genome sequencing. Genomic DNA libraries for each isolate were prepared using the Illumina TruSeq DNA sample prep kit using an insert size of 250 base pairs (bp). Libraries were sequenced on an Illumina HiSeq 2500 (Illumina, San Diego, GA) using 100-bp paired-end reads at the Michigan State University Research Technology Support Facility. We processed raw sequence reads (FASTQ) by removing the Illumina TruSeq adaptors using Cutadapt (Martin 2011), interleaving reads using Khmer (McDonald and Brown 2013), and quality-filtering based on an average Phred score of 30 using the FASTX-toolkit (Hannon Lab 2010). Finally, we normalized coverage to 25 based on a k-mer size of 25 using Khmer. We assembled the genomes

using Velvet (Zerbino and Birney 2008) after optimizing assembly parameters for each isolate with Velvet Optimizer (Gladman and Seemann 2012). We annotated contigs larger than 200 bp using Prokka (Seemann 2014), and predicted metabolic and physiological functions using MAPLE with bidirectional best-hit matches (Takami et al. 2012). We identified functional pathway based on the presence of intermediate genes within a pathway. We scored pathways as functional if more than 80 % of the intermediate genes were recovered in the genomes based on module completion ratios.

To test the hypothesis that metabolic pathways affect BGE, we used multivariate methods to compare the pathways of each isolate. First, we used PERMANOVA to determine if there were differences in pathways associated with different levels of BGE. When significant differences were found, we used indicator species analysis (Dufrene and Legendre 1997) to determine which metabolic pathways contributed to group differences in BGE. Next, to determine if metabolic pathways could explain differences in BGE within a group, we used distance-based redundancy analysis (dbRDA) which is a multivariate technique that tests if a quantitative predictor can explain differences in multivariate datasets (Legendre and Legendre 2012). Because we scored pathways as present or absent, metabolic distances between isolates were calculated using the Jaccard Index. We tested for significance using a permutation test. If the dbRDA model was significant, we used Spearman's rank-order correlation to test for correlations between BGE and individual metabolic pathways. We used the vegan R package (Oksanen et al. 2012) for multivariate analyses.

Resource Effects — To test the hypothesis that resources have different effects on components of metabolism that affect BGE, we used a linear model to test for a relationship between BR and

BP. Because BP required energy through respiration, we used production rate as the dependent
 424 variable and respiration rate as the independent variable. We used an indicator variable linear
 regression to test for changes in BP rate due to BR. We included resource identity and group
 426 (high- versus low-BGE) as the categorical predictors and BR as the continuous predictor
 (Lennon and Pfaff 2005). In addition, we included all interactions terms. Respiration and
 428 production rates were $\log_{10}$ -transformed to meet model assumptions. Last, to determine if the
 relationship between BR and BP rates was isometric (proportional scaling, slope = one) or
 430 allometric (disproportional scaling, slope $\neq$ one), we used a one-sample t-test to determine if the
 slope was different from one. All statistical tests were conducted in the R statistical environment.

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 GitHub repository (<https://www.github.com/LennonLab/MicrobialCarbonTraits>). Isolate
 440 genomes are available on NCBI (BioProject PRJNA420393).

REFERENCES

- 444 Allison, S. D. 2014. Modeling adaptation of carbon use efficiency in microbial communities.
Frontiers in Microbiology 5:1–9.
- 446 Apple, J. K., and P. A. del Giorgio. 2007. Organic substrate quality as the link between
bacterioplankton carbon demand and growth efficiency in a temperate salt-marsh estuary. The
448 ISME Journal 1:729–742.
- Bardgett, R. D., C. Freeman, and N. J. Ostle. 2008. Microbial contributions to climate change
450 through carbon cycle feedbacks. The ISME Journal 2:805–14.
- Bender, R. A. 2012. Regulation of the histidine utilization (Hut) system in bacteria.
452 Microbiology and Molecular Biology Reviews 76:565–584.
- Blomberg, S. P., T. Garland, and A. R. Ives. 2003. Testing for phylogenetic signal in
454 comparative data: behavioral traits are more labile. Evolution 57:717–745.
- Buchan, A., L. S. Collier, and E. L. Neidle. 2000. Key aromatic-ring-cleaving enzyme,
456 protocatechuate 3,4-dioxygenase, in the ecologically important marine *Roseobacter* lineage.
Applied and Environmental Microbiology 66:4662–4672.
- 458 Colwell, R. K., and D. J. Futuyma. 1971. On the Measurement of Niche Breadth and Overlap.
Ecology 52:567–576.
- 460 del Giorgio, P. A., and J. J. Cole. 1998. Bacterial growth efficiency in natural aquatic systems.
Annual Review of Ecology and Systematics 29:503–541.
- 462 del Giorgio, P. A., and R. E. I. Newell. 2012. Phosphorus and DOC availability influence the
partitioning between bacterioplankton production and respiration in tidal marsh ecosystems.
464 Environmental Microbiology 14:1296–307.
- Ducklow, H. W. 2008. Bacterial Production and Biomass in the Oceans. Pages 1–47 in Microbial

466 Ecology of the Oceans.

Dufrene, M., and P. Legendre. 1997. Species assemblages and indicator species: the need for a

468 flexible asymmetrical approach. *Ecological Monographs* 67:345–366.

Dykhuizen, D., and M. Davies. 1980. An Experimental Model: Bacterial Specialists and

470 Generalists Competing in Chemostats. *Ecology* 61:1213–1227.

Flamholz, A., E. Noor, A. Bar-Even, W. Liebermeister, and R. Milo. 2013. Glycolytic strategy as

472 a tradeoff between energy yield and protein cost. *Proceedings of the National Academy of Sciences* 110:10039–44.

474 Frey, S. D., J. Lee, J. M. Melillo, and J. Six. 2013. The temperature response of soil microbial efficiency and its feedback to climate. *Nature Climate Change* 3:395–398.

476 Fuhrer, T., E. Fischer, and U. Sauer. 2005. Experimental identification and quantification of glucose metabolism in seven bacterial species. *Journal of Bacteriology* 187:1581–1590.

478 Garland, J. L., and A. L. Mills. 1991. Classification and characterization of heterotrophic microbial communities on the basis of patterns of community-level sole-carbon-source

480 utilization. *Applied and Environmental Microbiology* 57:2351–9.

Geyer, K. M., E. Kyker-Snowman, A. S. Grandy, and S. D. Frey. 2016. Microbial carbon use

482 efficiency: accounting for population, community, and ecosystem-scale controls over the fate of metabolized organic matter. *Biogeochemistry* 127:173–188.

484 Ghosh, D., K. Roy, V. Srinivasan, T. Mueller, O. H. Tuovinen, K. Sublette, A. Peacock, et al. 2009. In-situ enrichment and analysis of atrazine-degrading microbial communities using

486 atrazine-containing porous beads. *Soil Biology and Biochemistry* 41:1331–1334.

Giovannoni, S. J., H. J. Tripp, S. Givan, M. Podar, K. L. Vergin, D. Baptista, L. Bibbs, et al.

488 2005. Genome streamlining in a cosmopolitan oceanic bacterium. *Science* 309:1242–5.

Gladman, S., and T. Seemann. 2012. Velvet Optimizer.

490 Glasser, J. W. J. 1984. Evolution of efficiencies and strategies of resource exploitation. *Ecology*
65:1570–1578.

492 Goberna, M., and M. Verdú. 2016. Predicting microbial traits with phylogenies. *The ISME*
Journal 10:959–967.

494 Gottschalk, G. 1986. *Bacterial Metabolism* (2nd ed.). Springer-Verlag, New York, NY.

Green, M. R., and J. Sambrook. 2012. *Molecular Cloning: A Laboratory Manual* (4th ed.). Cold
496 Spring Harbor Laboratory Press, Cold Spring Harbor, New York.

Hall, E. K., and J. B. Cotner. 2007. Interactive effect of temperature and resources on carbon
498 cycling by freshwater bacterioplankton communities. *Aquatic Microbial Ecology* 49:35–45.

Hannon Lab. 2010. FASTX Toolkit.

500 Hartigan, J. A., and P. M. Hartigan. 1985. The dip test of unimodality. *The Annals of Statistics*
13:70–84.

502 Harwood, C. S., and R. E. Parales. 1996. The beta-ketoadipate pathway and the biology of self-
identity. *Annual Review of Microbiology* 50:553–90.

504 Johnson, D. R., F. Goldschmidt, E. E. Lilja, and M. Ackermann. 2012. Metabolic specialization
and the assembly of microbial communities. *The ISME Journal* 6:1985–1991.

506 Kembel, S. W., D. D. Ackerly, S. P. Blomberg, W. K. Cornwell, M. R. Helmus, M. Helene, and
C. O. Webb. 2015. *Picante: R tools for integrating phylogenies and ecology*.

508 Kempes, C. P., S. Dutkiewicz, and M. J. Follows. 2012. Growth, metabolic partitioning, and the
size of microorganisms. *Proceedings of the National Academy of Sciences* 109:495–500.

510 Legendre, P., and L. F. J. Legendre. 2012. *Numerical Ecology* (Vol. 24). Elsevier.

Lennon, J. T. 2007. Diversity and metabolism of marine bacteria cultivated on dissolved DNA.

512 Applied and Environmental Microbiology 73:2799–2805.

Lennon, J. T., Z. T. Aanderud, B. K. Lehmkuhl, and D. R. J. Schoolmaster. 2012. Mapping the

514 niche space of soil microorganisms using taxonomy and traits. Ecology 93:1867–1879.

Lennon, J. T., S. K. Hamilton, M. E. Muscarella, A. S. Grandy, K. Wickings, and S. E. Jones.

516 2013. A source of terrestrial organic carbon to investigate the browning of aquatic ecosystems. PloS One 8:e75771.

Lennon, J. T., and L. E. Pfaff. 2005. Source and supply of terrestrial organic matter affects

aquatic microbial metabolism. Aquatic Microbial Ecology 39:107–119.

520 Lipson, D. A. 2015. The complex relationship between microbial growth rate and yield and its implications for ecosystem processes. Frontiers in Microbiology 6:1–5.

522 Litchman, E., K. F. Edwards, and C. A. Klausmeier. 2015. Microbial resource utilization traits and trade-offs: implications for community structure, functioning, and biogeochemical impacts at

524 present and in the future. Frontiers in Microbiology 06:1–10.

Livermore, J. A., S. J. Emrich, J. Tan, and S. E. Jones. 2014. Freshwater bacterial lifestyles

526 inferred from comparative genomics. Environmental Microbiology 16:746–758.

Ludwig, W., O. Strunk, R. Westram, L. Richter, H. Meier, Yadhukumar, A. Buchner, et al. 2004.

528 ARB: a software environment for sequence data. Nucleic Acids Research 32:1363–71.

Lynch, M., and G. K. Marinov. 2015. The bioenergetic costs of a gene. Proceedings of the

530 National Academy of Sciences 112:201514974.

Maechler, M. 2015. diptest: Hartigan’s dip test statistic for unimodality.

532 Martin, M. 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads. EMBnet.journal 17:10.

534 Martiny, A. C., M. L. Coleman, and S. W. Chisholm. 2006. Phosphate acquisition genes in

Prochlorococcus ecotypes: evidence for genome-wide adaptation. Proceedings of the National

536 Academy of Sciences 103:12552–7.

Martiny, A. C., K. Treseder, and G. Pusch. 2013. Phylogenetic conservatism of functional traits

538 in microorganisms. The ISME Journal 7:830–838.

Martiny, J. B. H., S. E. Jones, J. T. Lennon, and A. C. Martiny. 2015. Microbiomes in light of

540 traits: A phylogenetic perspective. Science 350:aac9323–aac9323.

McDonald, E., and C. T. Brown. 2013. Khmer: working with big data in bioinformatics. arXiv

542 1303.2223:1–18.

Meyenburg, K. V. O. N., and K. B. Andersen. 1980. Are growth rates of *Escherichia coli* in

544 batch cultures limited by respiration? Journal of Bacteriology 144:114–123.

Miller, M. A., W. Pfeiffer, and T. Schwartz. 2010. Creating the CIPRES Science Gateway for

546 inference of large phylogenetic trees. 2010 Gateway Computing Environments Workshop.

Mukherjee, T., J. Hanes, I. Tews, S. E. Ealick, and T. P. Begley. 2011. Pyridoxal phosphate:

548 biosynthesis and catabolism. Biochimica et Biophysica Acta - Proteins and Proteomics
1814:1585–1596.

550 Neidhardt, F. C. 2007. The Physiology and Biochemistry of Prokaryotes, 3rd ed. Microbe
Magazine (4th ed., Vol. 2). Oxford University Press, New York, NY.

552 Oksanen, A. J., F. G. Blanchet, R. Kindt, P. Legendre, P. R. Minchin, R. B. O. Hara, G. L.
Simpson, et al. 2012. Community Ecology Package. ... Ecology Package

554 Pagel, M. 1999. Inferring the historical patterns of biological evolution. Nature 401:877–884.

Pinheiro, J., and D. Bates. 2011. Linear and nonlinear mixed effects models (nlme). Package of

556 the R software for statistical computing on the CRAN repository.

Pomeroy, L. R., P. J. Williams, F. Azam, and J. E. Hobbie. 1998. The microbial loop.

558 Oceanography 20:28–33.

Pruesse, E. 2011. SINA: high throughput multiple sequence alignment.

560 Pruesse, E., J. Peplies, and F. O. Glöckner. 2012. SINA: accurate high-throughput multiple
sequence alignment of ribosomal RNA genes. *Bioinformatics* 28:1823–1829.

562 R Core Development Team. 2013. A language and environment for statistical computing.

Roller, B. R., and T. M. Schmidt. 2015. The physiology and ecological implications of efficient
564 growth. *The ISME Journal* 9:1481–1487.

Russell, J. B., and R. L. Baldwin. 1979. Comparison of maintenance energy expenditures and
566 growth yields among several rumen bacteria grown on continuous culture. *Applied and
Environmental Microbiology* 37:537–543.

568 Russell, J. B., and G. M. Cook. 1995. Energetics of bacterial growth: balance of anabolic and
catabolic reactions. *Microbiological Reviews* 59:48–62.

570 Schloss, P. D., S. L. Westcott, T. Ryabin, J. R. Hall, M. Hartmann, E. B. Hollister, R. A.
Lesniewski, et al. 2009. Introducing mothur: open-source, platform-independent, community-
572 supported software for describing and comparing microbial communities. *Applied and
Environmental Microbiology* 75:7537–7541.

574 Seemann, T. 2014. Prokka: rapid prokaryotic genome annotation. *Bioinformatics* 30:2068–9.

Sinsabaugh, R. L., S. Manzoni, D. L. Moorhead, and A. Richter. 2013. Carbon use efficiency of
576 microbial communities: stoichiometry, methodology and modelling. (J. Elser, ed.) *Ecology
Letters* 16:930–939.

578 Smith, D. C., and F. Azam. 1992. A simple, economical method for measuring bacterial protein
synthesis rates in seawater using ³H-leucine. *Marine Microbial Food Webs* 6:107–114.

580 Smith, E. M., and Y. T. Prairie. 2004. Bacterial metabolism and growth efficiency in lakes: the

importance of phosphorus availability. *Limnology and Oceanography* 49:137–147.

Stamatakis, A. 2006. RAxML-VI-HPC: maximum likelihood-based phylogenetic analyses with thousands of taxa and mixed models. *Bioinformatics* 22:2688–2690.

Takami, H., T. Taniguchi, Y. Moriya, T. Kuwahara, M. Kanehisa, and S. Goto. 2012. Evaluation method for the potential functionome harbored in the genome and metagenome. *BMC Genomics* 13:699.

Tao, H., C. Bausch, C. Richmond, F. R. Blattner, and T. Conway. 1999. Functional genomics: expression analysis of *Escherichia coli* growing on minimal and rich media. *Journal of Bacteriology* 181:6425–6440.

Tempest, D. W., and O. M. Neijssel. 1984. The status of YATP and maintenance energy as biologically interpretable phenomena. *Annual Reviews in Microbiology* 459–486.

Yilmaz, P., L. W. Parfrey, P. Yarza, J. Gerken, E. Pruesse, C. Quast, T. Schweer, et al. 2014. The SILVA and “All-species Living Tree Project (LTP)” taxonomic frameworks. *Nucleic Acids Research* 42:643–648.

Zerbino, D. R., and E. Birney. 2008. Velvet: algorithms for de novo short read assembly using de Bruijn graphs. *Genome Research* 18:821–9.

Zimmerman, A. E., A. C. Martiny, and S. D. Allison. 2013. Microdiversity of extracellular enzyme genes among sequenced prokaryotic genomes. *The ISME Journal* 7:1187–1199.

Zwietering, M. H., I. Jongerburger, F. M. Rombouts, and K. van’t Riet. 1990. Modeling of the bacterial growth curve. *Applied and Environmental Microbiology* 56:1875–1881.

TABLES

Table 1: Genetic pathways unique to the high BGE isolates. Pathways are functional metabolic pathways identified from genome sequencing and predicted using Maple. Prob. = probability statistic from indicator species analysis: the probability that the “species” (i.e., pathway) is not unique to the group.

Group	Prob.	Pathway	Reference Function
High BGE	0.02	M00045	Histidine degradation (histidine → N-formiminoglutamate → glutamate)
	0.02	M00060	Lipopolysaccharide biosynthesis (Kdo2-lipid A biosynthesis)
	0.03	M00565	Trehalose biosynthesis (D-glucose-1P → trehalose)

Table 2: Genetic pathways correlated with BGE in the high-BGE group. Correlations are spearman's rank correlations between BGE and the pathway presence. Pathways with correlations greater than 0.70 or less than -0.70 were considered significant.

Relationship	ρ	Pathway	Reference Function
Positive	0.82	M00025	Tyrosine biosynthesis (chorismate $\rightarrow$ tyrosine)
	0.72	M00034	Methionine salvage pathway
Negative	-0.73	M00117	Ubiquinone biosynthesis (chorismate $\rightarrow$ ubiquinone)
	-0.82	M00044	Tyrosine degradation (tyrosine $\rightarrow$ homogentisate)
	-0.82	M00053	Pyrimidine deoxyribonucleotide biosynthesis (CDP/CTP $\rightarrow$ dCDP/dCTP, dTDP/dTTP)
	-0.82	M00549	Nucleotide sugar biosynthesis (glucose $\rightarrow$ UDP-glucose)
	-0.82	M00568	Catechol ortho-cleavage (catechol $\rightarrow$ 3-oxoadipate)
	-0.82	M00637	Anthranilate degradation (anthranilate $\rightarrow$ catechol)

FIGURE LEGENDS

Fig. 1: Carbon resources used to study BGE variation in environmental isolates. **A:** Glucose – the baseline resource used to compare BGE across isolates. Glucose can be degraded by the Embden-Meyerhof-Parnas, pentose phosphate, or Entner-Doudoroff pathway. Ultimately, these pathways produce pyruvate (and then acetyl-CoA), which enters Krebs cycle and is used to produce energy and intermediate for biomass synthesis, when cells are grown aerobically. Alternatively, glucose can be fermented into organic acids (e.g., lactate), but these reactions yield less energy (Gottschalk 1986). **B:** Succinate – is a simple organic acid. Succinate is an intermediate of Krebs cycle and thus it does not require previous degradation. Additionally, succinate can be used directly to produce energy via succinate dehydrogenase (Neidhardt 2007). **C:** Protocatechuate – is a complex resource with an aromatic core. Typically, it is degraded to acetyl-CoA and Succinyl-CoA via the β -ketoadipate pathway (Harwood and Parales 1996). Protocatechuate is commonly used to study aromatic resource degradation in ecosystems, and the β -ketoadipate pathway is commonly found in bacteria across the phylum Proteobacteria (Buchan et al. 2000).

Fig. 2: Bacterial growth efficiency (BGE) of each isolate for each resource. BGE was calculated based on measured production (BP) and respiration (BR) rates using the following equation:

$BGE = BP / (BP + BR)$. Cladogram is based on the consensus phylogeny. Taxonomic class and order are included based on the ribosomal database taxonomy: α = Alphaproteobacteria, β =

Betaproteobacteria, γ = Gammaproteobacteria, *Xan.* = Xanthomonadales, *Aero.* = Aeromonadales, *Pseu.* = Pseudomonadales.

Fig. 3: Phenotypic traits associated with BGE. A: Maximum growth rate, a measure of growth

strategy, demonstrates a trade-off (negative relationship) with BGE in the high BGE group ($F_{1,7} = 9.52$, $r^2 = 0.54$, $p = 0.015$), but not the low BGE group ($F_{1,7} = 0.51$, $r^2 = 0.06$, $p = 0.50$). B:

Levin's Index, a measure of niche breadth, does not demonstrate a trade-off with BGE in either the high or low BGE groups (high: $F_{1,7} = 0.92$, $r^2 = 0.11$, $p = 0.37$; low: $F_{1,7} = 1.47$, $r^2 = 0.17$, $p = 0.27$). High and low BGE groups were determined based on bimodal distribution of BGE.

Fig. 4: Relationship between respiration and production rates. Respiration and production rates were compared using an indicator variable linear regression ($F_{9,42} = 4.92$, $R^2 = 0.51$, $p < 0.001$).

According to the regression model, production rate increases with respiration rate proportionally (i.e., slope not significantly different from one, $t_{42} = -0.26$, $p = 0.79$). In addition, group (high vs.

low BGE) was a significant factor and isolates in the high BGE group had a greater y-intercept ($p = 0.02$). Symbols indicate isolate group (high and low BGE), and color indicates the resource

being used. Symbol size is scaled by growth efficiency.

FIGURES

Fig. 1:

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Fig. 2:

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672 **Fig. 3:**

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Fig. 4: