

1 **Spectral niches reveal taxonomic identity and complementarity in plant**
2 **communities**

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24 **Summary**

- 25 • Plants' spectra provide integrative measures of their chemical, morphological, anatomical, and
26 architectural traits. We posit that the degree to which plants differentiate in n-dimensional
27 spectral space is a measure of niche differentiation and reveals functional complementarity.
- 28 • In both experimentally and naturally assembled communities, we quantified plant niches using
29 hypervolumes delineated by either plant spectra or 10 functional traits. We compared the niche
30 fraction unique to each species in spectral and trait spaces with increasing dimensionality, and
31 investigated the association between the spectral space occupied, plant growth and community
32 productivity.
- 33 • We show that spectral niches differentiated species better than their functional trait niches. The
34 amount of spectral space occupied by individuals and plant communities increased with plant
35 growth and community productivity, respectively. Further, community productivity was better
36 explained by inter-individual spectral complementarity than by productive individuals
37 occupying large spectral niches.
- 38 • The degree of differentiation in spectral space provides the conceptual basis for identifying
39 plant taxa spectrally. Moreover, our results indicate that the size and position of plant spectral
40 niches reflect ecological strategies that shape community composition and ecosystem function,
41 with the potential to reveal insight in niche partitioning over large areas with spectroscopy.

42

43 **Keywords**

44 complementarity, functional traits, ecosystem function, Hutchinsonian niche, plant species
45 identification, spectroscopy, spectral space, remote sensing

46 **Introduction**

47 The Hutchinsonian niche concept (Hutchinson, 1957) describes the niche as an n-dimensional
48 hypervolume delineated by axes of environmental conditions (micro-habitat, abiotic factors,
49 resources, predators etc.) within which populations of organisms can survive in the long term
50 (Leibold, 1995; Bazzaz & Bazzaz, 1996; Silvertown, 2004). This understanding of niches
51 combines two main lines of theory (Leibold, 1995; Chase & Leibold, 2003): the “habitat” niche,
52 the conceptual space delineated by all environmental requirements of an organism (Grinnell,
53 1917), and the “functional” niche, the ecological role organisms play in their environment (Elton,
54 1927). These habitat- and function-centric concepts are connected, because organisms inevitably
55 consume and supply resources when they occur in an environment supporting their needs. The
56 realized niche thus entails both the requirements of a population of organisms in a particular
57 environment (habitat factors) and their impact on that environment (biological interactions;
58 Leibold, 1995; Chase & Leibold, 2003). However, quantifying plant niches in an n-dimensional
59 environmental space that captures all biotic and abiotic factors important for determining the
60 fitness of populations of organisms and their impacts on the environment and other trophic levels
61 is challenging, especially outside of laboratory conditions (Bazzaz & Bazzaz, 1996; Violle &
62 Jiang, 2009). The trait niche delineated by functional traits related to the performance of
63 organisms and species along key environmental gradients (McGill *et al.*, 2006; Violle & Jiang,
64 2009) has been proposed as a means to overcome the problem of niche hyper-dimensionality.
65 Niches of plants in trait space provide a mechanistic link to ecosystem function, since only
66 certain combinations of traits appear to be successful (Reich *et al.*, 1997; Wright *et al.*, 2004) or
67 are biologically possible. A key first step in defining the trait niche is prioritizing environmental
68 factors that significantly influence plant fitness and determining which traits or trait syndromes
69 best reflect the adaptations necessary for success in a particular environment (McGill *et al.*,
70 2006). But determining all environmental factors important for structuring plant communities
71 and their relative importance is difficult, and there are practical and methodological limits to the
72 (number of) traits or trait syndromes we can measure (Petchey & Gaston, 2006).

73 Plants partition resources in time and space as a result of contrasting ecological strategies,
74 giving rise to biochemical, structural and phenological differences that affect optical properties
75 of leaves and canopies. Spectral profiles of plants (the curves resulting from measuring plant
76 spectra at high spectral resolution) are influenced by leaf traits (Gates *et al.*, 1965; Knipling,

77 1970; Jacquemoud & Baret, 1990; Ustin *et al.*, 2004; Cavender-Bares *et al.*, 2017), including
78 pigment composition, micro- and macronutrient content, water content, specific leaf area (SLA),
79 leaf surface and internal structure, and they are, when measured from a distance, also influenced
80 by plant traits related to canopy architecture (Jacquemoud & Baret, 1990; Curran *et al.*, 2001;
81 Slaton *et al.*, 2001; Sims & Gamon, 2002; Ollinger, 2011). Spectral profiles thus capture key
82 differences in foliar chemistry, leaf anatomy, plant morphology, life history strategies, and
83 responses to environmental variation, which have evolved over time and reflect ecological
84 strategies (Ustin *et al.*, 2004; Cavender-Bares *et al.*, 2017) with consequences for ecosystem
85 structure and function above- and belowground (Madritch *et al.*, 2014; Cavender-Bares *et al.*,
86 2017). The concept of spectral or optical types states that functionally distinct species (or other
87 clades) occupy unique spectral spaces because of their specific chemical and structural
88 characteristics (Palmer *et al.*, 2002; Rocchini *et al.*, 2010; Asner *et al.*, 2014; Féret & Asner,
89 2014), and there is evidence that spectral dissimilarity among species increases with functional
90 dissimilarity (Asner & Martin, 2008; McManus *et al.*, 2016; Schweiger *et al.*, 2018) and
91 evolutionary divergence time (Cavender-Bares *et al.*, 2016; McManus *et al.*, 2016; Schweiger *et*
92 *al.*, 2018).

93 Here we combine the optical type concept with classical niche theory and present plant
94 “spectral niches” - the spectral spaces occupied by plants - as an integrative approach to
95 calculating niche size and its consequence for ecosystem function. We define the spectral niche
96 as the n-dimensional hypervolume occupied by plants and delineated by spectral axes (which can
97 be spectral bands, spectral indices, or other expressions of spectral variation) along which plants
98 can vary. We quantify plant niches using hypervolumes estimated from convex hulls, which are
99 the multivariate equivalents of range and independent of the shape of the distribution and
100 correlation structure among measured variables (Cornwell *et al.*, 2006; Blonder *et al.*, 2014), and
101 test three hypotheses associated with the degree of plant differentiation in spectral space: (H1)
102 Plant species occupy distinct spectral niches as a consequence of contrasting functional attributes
103 that are linked to resource acquisition; (H2) Plant individuals that occupy greater spectral niche
104 space have higher performance in terms of growth due to increased intra-individual foliar
105 variation in more complex canopies; (H3) More productive plant communities occupy greater
106 spectral niche space, because of optical and functional complementarity.

107 We tested these hypotheses using data collected in a grassland and a forest biodiversity
108 experiment – the Cedar Creek biodiversity (BioDIV) experiment (Tilman *et al.*, 1997) and the
109 Forest and Biodiversity (FAB) experiment (Grossman *et al.*, 2017). We used leaf spectra and
110 functional traits of 902 individuals from 14 grassland–savanna perennials sampled in 35 plots in
111 BioDIV and aboveground biomass determined in the same plots in July 2015 (H1, H3); and leaf
112 spectra, plant height and diameter measurements of 537 individuals from 12 tree species sampled
113 in 68 plots in FAB in July 2016 (H2, H3). In addition, we were interested in the stability of
114 species’ spectral niches in space, time and across measurement scales. We thus tested the degree
115 to which species identification models developed in BioDIV correctly differentiate species
116 measured in naturally assembled grassland communities, and at a different time and
117 measurement scale using three additional datasets: Leaf spectra and functional traits of 281
118 individuals from seven grassland–savanna perennials collected in July 2016 in 19 plots of an old
119 fields chronosequence taken out of agricultural use between 1928 and 2015 (Inouye *et al.*, 1987;
120 Clark *et al.*, 2019); leaf spectra and functional traits of 243 individuals from nine species
121 collected in 14 plots in BioDIV in August 2015; and remotely sensed spectra collected with an
122 imaging spectrometer mounted on a robotic cart (Wang *et al.*, 2016) in 18 plots of the BioDIV
123 experiment in July 2015.

124

125 **Material and Methods**

126 Spectral data

127 We measured leaf spectra using a leaf-clip assembly and two portable field spectrometers (SVC
128 HR-1024i, Spectra Vista Corp., Poughkeepsie, NY; and PSR +, Spectral Evolution Inc.,
129 Lawrence, MA) covering the wavelength range from 350 nm to 2500 nm in 1024 spectral bands.
130 We used the SVC instrument for measuring grassland-savanna perennials and the PSR+ for
131 measuring tree species. To characterize one individual spectrally, we measured the reflectance of
132 either three or five mature, healthy leaves per individual depending on plant height. Spectra were
133 automatically calibrated for dark current and stray light, and referenced to the white calibration
134 disc of the leaf clip approximately every 10 minutes. Spectral data processing included
135 correcting discontinuities at the sensor overlap regions between the Si and first InGaAs sensor
136 (around 1000 nm) and between the first and second InGaAs sensor (around 1900 nm), removing

137 noisy regions at the beginning and end of the spectrum, and interpolating spectra to 1 nm
138 resolution. In addition, we collected leaf level spectral with an imaging spectrometer (E Series,
139 Headwall Photonics, Fitchburg, MA) mounted on an automated tram (also referred to as spectral
140 images; see Gamon *et al.*, 2006 for details). Each processed image consisted of 1,000 x 1,000
141 pixels with 1 mm spatial resolution covering the visible and near-infrared regions (400–990 nm)
142 in 924 spectral bands. Again, spectra were resampled to 1 nm spectral resolution and noisy
143 regions at the beginning and end of the spectrum were excluded. For spectral processing we used
144 the spectrolab (Meireles *et al.*, 2017) package in R (R Core Team, 2019). The BioDIV July
145 dataset consisted of 902 individuals (ind.) from 14 species: *Achillea millefolium* L. (49 ind.),
146 *Amorpha canescens* Pursh (28 ind.), *Andropogon gerardii* Vitman (162 ind.), *Asclepias tuberosa*
147 L. (70 ind.), *Lespedeza capitata* Michx. (99 ind.), *Liatris aspera* Michx. (49 ind.), *Lupinus*
148 *perennis* L. (121 ind.), *Panicum virgatum* L. (49 ind.), *Petalostemum candidum* (Willd.) Michx.
149 (28 ind.), *Petalostemum purpureum* (Vent.) Rydb. (52 ind.), *Petalostemum villosum* Nutt. (42
150 ind.), *Schizachyrium scoparium* (Michx.) Nash (76 ind.), *Solidago rigida* L. (50 ind.),
151 *Sorghastrum nutans* (L.) Nash (27 ind.). The old fields dataset consisted of 281 individuals from
152 seven species: *Agropyron repens* (L.) P. Beauv. (48 ind.), *Berteroa incana* (L.) DC. (29 ind.),
153 *Lespedeza capitata* Michx. (29 ind.), *Panicum virgatum* L. (27 ind.), *Poa pratensis* L. (67 ind.),
154 *Rumex acetosella* L. (27 ind.), *Schizachyrium scoparium* (Michx.) Nash (54 ind.). The BioDiv
155 August dataset consisted of 243 individuals from nine species: *Achillea millefolium* L. (20 ind.),
156 *Amorpha canescens* Pursh (18 ind.), *Andropogon gerardii* Vitman (27 ind.), *Lespedeza capitata*
157 Michx. (30 ind.), *Liatris aspera* Michx. (18 ind.), *Petalostemum purpureum* (Vent.) Rydb. (33
158 ind.), *Schizachyrium scoparium* (Michx.) Nash (41 ind.), *Solidago rigida* L. (32 ind.),
159 *Sorghastrum nutans* (L.) Nash (24 ind.). The spectral image data consisted of 2237 pixels from
160 seven species: *Achillea millefolium* L. (264 pixels), *Andropogon gerardii* Vitman (283 ind.),
161 *Asclepias tuberosa* L. (422 pixels), *Lespedeza capitata* Michx. (149 pixels), *Monarda fistulosa*
162 L. (242 pixels), *Panicum virgatum* L. (388 pixels), *Schizachyrium scoparium* (Michx.) Nash
163 (241 pixels), *Solidago rigida* L. (248 pixels). And the FAB dataset consisted of 537 individuals
164 from 12 species: *Acer negundo* L. (30 ind.), *Acer rubrum* L. (47 ind.), *Betula papyrifera*
165 Marshall (44 ind.), *Juniperus virginiana* L. (39 ind.), *Pinus banksiana* Lamb. (47 ind.), *Pinus*
166 *resinosa* Aiton (52 ind.), *Pinus strobus* L. (47 ind.), *Quercus alba* L. (42 ind.), *Quercus*

167 *ellipsoidalis* E.J.Hill (49 ind.), *Quercus macrocarpa* Michx. (50 ind.), *Quercus rubra* L. (39
168 ind.), *Tilia americana* L. (51 ind.).

169

170 Functional traits

171 We determined the following functional traits for all individuals of grassland–savanna perennials
172 measured spectrally with the leaf clip: foliar nitrogen, carbon, non-structural carbohydrate,
173 hemicellulose, cellulose, and lignin concentration (%), and the content of chlorophyll a and b,
174 beta-carotene, lutein, violaxanthin, antheraxanthin, and zeaxanthin pigments ($\mu\text{mol m}^{-2}$). Foliar
175 traits were predicted using partial least squares regression (PLSR) models (Wold *et al.*, 1983)
176 developed from chemical assays of leaf tissue samples and corresponding level spectra. Leaf
177 tissue samples were collected in the summers of 2015 and 2016 at the CCESR and encompassed
178 62 species; all chemical trait analyses were performed at the University of Minnesota following
179 methods described in (Schweiger *et al.*, 2018). We summarized chlorophyll content as
180 chlorophyll a plus chlorophyll b; and we expressed beta-carotene and lutein content, and the size
181 of the xanthophyll pigment pool (VAZ = violaxanthin plus antheraxanthin plus zeaxanthin) as
182 ratios relative to chlorophyll content to indicate contrasting photosynthetic behavior and
183 photoprotective capacity among plants (Gamon *et al.*, 1997; Gamon & Berry, 2012). We tested
184 for differences among species functional traits using Tukey’s honest significant difference
185 (HSD) post hoc tests and the R package agricolae (de Mendiburu, 2017). Further, since we
186 expected species separability to increase with phylogenetic distance (Schweiger *et al.*, 2018), we
187 tested for phylogenetic signal of each trait using Blomberg’s K statistic (Blomberg *et al.*, 2003)
188 as implemented in the R package picante (Kembel *et al.*, 2010) and the phylogeny reconstructed
189 by (Kothari *et al.*, 2018) with one missing species (*Petalostemum candidum* (Willd.) Michx.)
190 added manually with phytools (Revell, 2012; for details, see also Table S5). For all trees in our
191 study we used individual tree height (cm) and basal diameter (cm) as measures of growth
192 (Grossman *et al.*, 2017).

193

194 Species spectral and functional niches

195 We calculated the niche fraction unique to each of the 14 species sampled in BioDIV in spectral
196 space and functional trait space of increasing dimensionality using the R package hypervolume

197 (Blonder *et al.*, 2014). We randomly selected between 2 and 21 spectral bands and between 2
198 and 10 functional traits as axes delineating species' spectral and functional niches, respectively,
199 and repeated each selection 50 times; functional traits were z-standardized. We projected all
200 individuals into the resulting spectral and functional trait spaces, and calculated the fraction of
201 the hypervolume unique to each species (i.e. the hypervolume that is occupied by the focal
202 species and not overlapped by any other species).

203 Since it is difficult to show more than three or four niche dimensions in one graph, we
204 used linear discriminant analysis (LDA) to illustrate species spectral and functional dissimilarity
205 in niche spaces delineated by the main axes of spectral and functional variation. In our case,
206 linear discriminants (LDs) are linear combinations of all band-wise reflectance and functional
207 trait values, respectively, which re-project observations into a new coordinate system while
208 maximizing the differences between groups; our grouping variable was species identity. For
209 LDA we used the R package MASS (Venables & Ripley, 2002); for interactive 3D graphics
210 illustrating species niche shifts with changes in community diversity we used plotly (Sievert *et*
211 *al.*, 2017).

212

213 Species identification models

214 We tested the degree to which plant species can be correctly identified based on spectra and
215 functional traits with partial least squares discriminant analysis (PLSDA) as implemented in the
216 R packages pls (Mevik *et al.*, 2018) and caret (Kuhn, 2018). We chose the number of samples for
217 model training depending on sample size: For the BioDIV model based on data sampled in July
218 and for the old fields chronosequence model we used random draws of 20 individuals per species
219 for model training; for the BioDIV model based on data sampled in August we used random
220 draws of 10 individuals per species; and for the imaging spectroscopy model we used random
221 draws of 50 pixels per species. The remaining data were used for validation and for evaluating
222 model fit; all statistics and graphs are based on the validation results. For each dataset, we
223 performed 100 PLSDA model iterations always using new random draws of training samples and
224 selected the optimal number of components based on the minimum of the root mean squared
225 error of prediction (RMSEP) for the test samples. We tested for significant differences in
226 RMSEP values using Tukey's HSD test as implemented in the R package agricolae (de
227 Mendiburu, 2017), and used the smaller number of components when models performed

228 similarly ($p > 0.05$). We investigated which wavelengths and functional traits contributed most to
229 species separability using PLSDA loadings. For the species identification models based on
230 spectral images, we extracted sunlit pixels from seven clearly identifiable species using ENVI
231 5.4 (Exelis Visual Information Solutions, Boulder, CO). To investigate model transferability, we
232 subset each dataset to match the species sampled in both datasets. For each of the subsets, we
233 calibrated new PLSDA models as described above, always using 20 randomly selected samples
234 per species for training, and the remaining data for testing and model evaluation. We assessed
235 model transferability i) in space by applying the BioDIV July model to the old fields
236 chronosequence data and vice versa, ii) in time by applying the BioDIV July model to the
237 BioDIV August data and vice versa, and iii) across sampling scales by applying the BioDIV July
238 model to spectral image data.

239

240 Spectral space occupied by individuals and plant communities

241 We tested the degree to which the spectral space occupied by individual plants predicts plant
242 growth by fitting regression models between the spectral space occupied by trees sampled in
243 FAB and two measures of growth, tree height (cm) and basal diameter (cm). We assessed model
244 performance based on the coefficient of determination (R^2) and the RMSEP. Next, we tested the
245 degree to which the spectral space occupied by plant communities predicts aboveground
246 productivity. In BioDIV we used biomass (g m^{-2} , dry weight) determined in clipstrips as a
247 measure of aboveground net primary productivity. In FAB we used overyielding as a measure of
248 the net biodiversity effect (NBE), which we partitioned into complementary (CE) and selection
249 effects (SE), following (Loreau & Hector, 2001). Biomass for NBE, SE and CE calculations was
250 determined from allometrically derived incremental stem biomass (kg y^{-1} ; see Grossman *et al.*,
251 2017 for details). We did not calculate and partition the NBE in BioDIV because monocultures
252 are not replicated in this experiment. Niche size, as other measures of the spread of variables, is
253 known to be positively correlated with sample size (Cornwell *et al.*, 2006). We thus used the
254 same number of randomly selected individuals per community to calculate the occupied spectral
255 space. For the BioDIV data collected with the leaf clip, we used spectra of 12 randomly selected
256 individuals per plot resulting in a total of 30 communities used for analysis. For the BioDIV data
257 collected with the imaging spectrometer, we randomly extracted 30 pixels per plot; spectra were
258 corrected for soil effects following (Gholizadeh *et al.*, 2018). For FAB, we used spectra of nine

259 randomly selected individuals per plot, resulting in a total of 68 communities used for analysis.
260 We reduced data dimensionality to the first three principal component (PC) axes, which
261 explained more than 98% of the total spectral variation in the leaf clip data (BioDIV and FAB)
262 and more than 96% of the total spectral variation in the proximal remote sensing data,
263 respectively, and calculated the spectral space occupied per community using the hypervolume
264 package (Blonder *et al.*, 2014) in R. We tested the association between the spectral space
265 occupied by plant communities and community productivity using linear regression models, and
266 assessed model performances based on the R^2 and RMSEP. In addition, we tested the degree to
267 which the spectral space occupied by plant communities increases with species richness.

268

269 Data availability

270 The data used in this manuscript are publicly available through EcoSIS <https://ecosis.org>
271 (spectral data), the Cedar Creek Ecosystem Science Reserve
272 <http://www.cedarcreek.umn.edu/research/data> (biomass data), and LPDAAC
273 https://lpdaac.usgs.gov/dataset_discovery/community/community_products_table/hwhyppccmn1
274 [mm_v001](https://lpdaac.usgs.gov/dataset_discovery/community/community_products_table/hwhyppccmn1mm_v001) (spectral images, DOI: 10.5067/Community/Headwall/HWHYPCCMN1MM.001).

275

276 Results

277 Species spectral niches were more distinct than their trait-based niches calculated from the 10
278 chemical and structural foliar traits (carbon, nitrogen and carbon fraction concentration and the
279 contents of chlorophyll and carotenoid pigments; see Methods) measured in our study. The
280 fraction of the niche space unique to each of the 14 species of grassland–savanna perennials
281 increased with the dimensionality of spectral and functional trait space (Fig. 1). However, while
282 each focal species occupied a hypervolume that was at least 90% unique to the species after
283 including 15 randomly selected spectral bands as axes (Fig. 1a, Table S1), not all focal species
284 reached the same level of uniqueness in functional trait space after including all 10 functional
285 traits as axes (Fig. 1b, Table S2). Greater distinctiveness of species spectral niches as compared
286 to their functional niches was confirmed when projecting species' positions into spectral and
287 functional spaces reduced to their main axes of variation. In spectral space, all non-graminoids
288 species clearly separated along the first four linear discriminate axes (LDs, Fig. 2a-b), and LDs

289 11 and 12 separated the graminoids (Fig. 2c). In functional trait space, however, only a few
290 species formed distinct clusters, and we found no combination of LDs that separated the four
291 graminoids from each other (Fig. 2d-f). The likely reason being the substantial degree of
292 intraspecific variation across all measured traits (Fig. S1) and the lack of a particular trait or
293 traits that differed significantly among all species (Fig. S2, Table S3). This is also illustrated by
294 the high degree of overlap among species niches in two-dimensional trait space (Fig. S3b);
295 although legumes and graminoids clustered somewhat separately from other species along
296 carbon content, and nitrogen content and carbon fraction axes, respectively. Species' niche
297 overlap was also pronounced in two-dimensional spectral space delineated by the 10 most
298 variable spectral bands, but spectrally distinct species already started emerging (Fig. S3a).

299 Species identification models based on spectra (65% - 98% accuracy per species, Figs 3a,
300 S4a, Table S4) consistently outperformed species identification models based on functional traits
301 (47% - 97% accuracy per species, Figs 3b, S4b, Table S4), probably due to the smaller overlap
302 (Fig. 1) and greater distance (Fig. 2) between species' niches in spectral space compared to trait
303 space. The spectral bands contributing most to species' separability aligned with absorption
304 features related to leaf chlorophyll, carotenoid, lignin and protein content (Fig S5a). These foliar
305 traits also contributed most to species separability in functional trait space (Fig. S5b) and all of
306 them, except for chlorophyll content, showed evidence of phylogenetic signal (Table S5, Fig.
307 S6). The better performance of species identification models based on spectra compared to
308 models based on functional traits was confirmed in the old fields chronosequence (93%
309 compared to 69% overall accuracy) and in the BioDIV experiment sampled later in the season
310 (96% compared to 74% overall accuracy), and species were also to 93% correctly identified from
311 the spectral images (Figs S7-S9, Table S4). Generally, spectral models were more transferable
312 across sites and functional trait models were more transferable in time (Table S6). However, it
313 was not possible to successfully transfer a species identification model calibrated using leaf-clip
314 data to the spectral images collected at the same time, likely because the sampling method
315 normalized for illumination and removed effects of plant architecture and atmosphere, which are
316 important sources of spectral variation at the canopy level.

317 The spectral space occupied by individual trees - a measure of intra-individual spectral
318 variation - was linked to the total variation in the tree height and diameter (Fig. 4), likely because
319 trees that grow more tend to have larger, more complex canopies and higher foliar plasticity than

320 trees that grow less. Over time, taller trees seem to be able to sustain and perhaps even increase
321 the benefits gained from harnessing a more diverse light environment (Figs 4b, 4d), pointing
322 towards size-asymmetric (size-dependent) light competition (Schwinning & Weiner, 1998)
323 among the young trees in FAB. The spectral space occupied by individuals was more closely
324 associated with tree diameter than tree height. One likely explanation being that once trees are
325 taller than their neighbors, it may be more advantageous to invest in mechanical stability and
326 horizontal canopy extension than in vertical growth to maximize light interception.

327 The spectral space occupied by plant communities in BioDIV explained 44% of the total
328 variation in aboveground productivity when estimated from spectra measured with the leaf clip
329 (Fig. 5a) and 31% when estimated from the spectral images (Fig. 5b). Likewise, the spectral
330 space occupied increased with the number of species per plant community (Fig. 5c). Thus, the
331 size of the spectral niche occupied by plant communities provides an alternative measure of
332 spectral diversity that is similarly predictive of productivity as other measures of biodiversity
333 reported in this experiment (see Schweiger *et al.*, 2018 and references therein). In FAB, we used
334 overyielding (i.e. the access biomass produced by mixed species plots compared to what would
335 be expected based on their monoculture yields) as a measure of the net biodiversity effect
336 (NBE); and we partitioned the NBE into its two components, complementary (the positive
337 effects of diverse resource use strategies and positive interactions among plants on productivity)
338 and selection effects (the influence of plants with particular traits on productivity, following
339 Loreau & Hector, 2001). The spectral space occupied by tree communities in FAB explained
340 42% of overyielding (Fig. 6a) and increased with the number of species per community (Fig. 6b).
341 Partitioning the NBE, into its two components revealed a positive relationship between the
342 spectral space occupied by communities and complementarity (Fig. 6c), while the association
343 with the selection effect was negative (Fig. 6d). In other words, compared to less productive
344 communities more productive communities did on average not harbor more highly productive
345 individuals which occupy larger spectral spaces, but rather more spectrally dissimilar and
346 complementary species that collectively contributed to the large spectral space occupied by these
347 communities.

348

349 Discussion

350 Plants display themselves towards the sky with contrasting optical patterns linked to
351 biochemical, anatomical and morphological plant traits that have evolved over time and
352 influence their spectral response (Ustin & Gamon, 2010; Cavender-Bares *et al.*, 2017). Not all
353 functional traits of plants, including hydraulic and root traits, and traits specific to organs that
354 interact minimally with light, including seeds, can be spectrally detected. However, plant spectra
355 integrate many aspects of plant form and function offering an effective and novel way to
356 quantify plant niches in n-dimensional space (*sensu* Hutchinson, 1957). Here we show that
357 spectral differentiation – the degree to which species and individuals occupy distinct spectral
358 spaces – is an important and useful measure of complementarity and provides a practical means
359 of identifying plant species or other clades or functional groups with spectra.

360 Spectra of plants captured more of the total variation in plant characteristics than the set
361 of functional traits measured in our study. Projecting species into functional and spectral spaces
362 with increasing dimensionality gradually reduced the overlap among species (Fig. 1). However,
363 species functional niches contracted to a lesser degree than their spectral niches, leading to
364 spectral species identification models consistently outperforming species identification models
365 based on functional traits (Table S4). To some extent, this effect could be due to redundancy in
366 our metrics of function. Plant functional traits reflect their phylogenetic legacies as well as
367 environmental adaptations. In our case, light gradients are probably the dominating source of
368 environmental variation, and all leaf traits measured in our study are to some degree influenced
369 by variation in light. For instance, the degree of correlation between chlorophyll and carotenoid
370 pigment levels reflects biochemical adaptation to different light environments; with tight
371 correlations indicating their common role in light harvesting and weak correlations indicating a
372 stress response (Gamon & Berry, 2012). Likewise, the contents of different carbon fractions are
373 tied to morphological adaptations (such as leaf thickness and SLA) to light gradients within
374 canopies (Niinemets, 2007). In this way, what we think of as multiple traits can also be thought
375 of as different proxies for the same or overlapping traits.

376 Misclassifications occurred more often among closely related than among distantly
377 related species (Figs 3, S4a, S7a, S8a), likely due to the similarity in functional traits (Tables S3,
378 S5, Fig. S6) and spectra among close relatives (Cavender-Bares *et al.*, 2016; Schweiger *et al.*,
379 2018). The spectral regions that contributed most to species separability aligned with absorption

380 features for proteins and lignin content (Fig. S5a; see e.g. Curran, 1989), functional traits
381 associated with the trade-off between fast and slow return on investment (Reich *et al.*, 1997;
382 Wright *et al.*, 2004) suggesting the relevance of different resource use strategies for structuring
383 plant communities and community productivity (Reich *et al.*, 1997). In addition, chlorophyll
384 content and xanthophyll cycle pigment pool size contributed substantially to species separability,
385 highlighting that prairie ecosystems harbor species with different strategies for light capture and
386 photoprotection (Kothari *et al.*, 2018). These sets of traits also relate to canopy structure (e.g.
387 leaf display) and so can tie to higher level spectral and plant traits at the canopy scale.

388 The images collected by the mobile tram capture the spectra of leaves in their natural
389 orientation and illumination, as would be seen in remote sensing. The high classification
390 accuracy of spectral species identification models based on spectral images is a promising result
391 for high resolution remote sensing approaches to species identification (e.g. using unmanned
392 aerial vehicles, UAV's). For instance, imaging spectroscopy that can resolve canopies, i.e. with
393 pixel sizes smaller than the typical species crown (Arroyo-Mora *et al.*, 2019), has tremendous
394 potential for detecting biodiversity and quantifying spectral niches, given that a training dataset
395 of plants or plant leaves can be identified directly from the images, or on the ground and geo-
396 referenced.

397 Spectral models were more transferable in space, while functional trait models were more
398 transferable in time (Table S6), pointing towards differences in the relative importance of
399 biochemical vs. morphological/anatomical trait variation for species differentiation. Chemical
400 characteristics of leaves change as leaves age and spectra capture these phenological differences
401 more completely than a set of foliar traits (Chavana-Bryant *et al.*, 2017), making it easier to
402 transfer species identification models based on a limited number of foliar traits in time compared
403 to species identification models based on spectra. Leaf morphology and anatomy, including
404 surface structure and leaf thickness, are often less variable within species than leaf chemistry
405 (Valladares *et al.*, 2000). Morphological and anatomical leaf traits were not included in the set of
406 functional traits we measured, but they are captured by spectral measurements (Cavender-Bares
407 *et al.*, 2017), likely making it easier to transfer species identification models based on spectra in
408 space. However, it is worth noting that our study sites were spatially close and similar in their
409 environmental characteristics, including soil, and we expect decreasing spatial transferability of
410 spectral models with increasing environmental dissimilarity among sites.

411 The positive relationship between the intra-individual spectra variation and tree height
412 and diameter (Fig. 4) supports our hypothesis that plants occupying more spectral space have
413 greater performance in terms growth. Light is likely the most variable resource for the young
414 trees in the FAB experiment, leading to adaptations in leaf traits, including SLA, pigment and
415 nitrogen content (Valladares *et al.*, 2000; Rozendaal *et al.*, 2006), which all influence the spectral
416 response. The foliar plasticity in response to diverse light environments might create a self-
417 reinforcing system leading to increased plant growth, which, in turn, modifies and likely
418 increases intra-individual spectral variation through increased leaf area index (LAI) generating
419 even more pronounced light gradients, as found in other tree diversity experiments (Williams *et al.*,
420 2017). Notably, while tree growth in FAB was positively correlated with intra-individual
421 spectral variation (Fig. 4), the size of the spectral space occupied by productive individuals did
422 not explain well the association between productivity and the spectral space occupied by
423 communities (Fig. 6a). In our case, it appears that the spectral space occupied by productive
424 plant communities is dominated by spectral complementarity, which has been identified as an
425 important factor for light resource partitioning among evergreen and deciduous trees (Gamon *et al.*,
426 2016; Springer *et al.*, 2017). Again, the size of the spectral space occupied by spectrally
427 diverse communities can be explained by a positive feedback, where spectral dissimilarity results
428 in and is a consequence of greater resource use and increased growth. The spectral space
429 occupied by plant communities can thus be interpreted as a measure of functional complexity,
430 because more spectrally dissimilar species tend to be more functionally dissimilar (Asner &
431 Martin, 2008; Schweiger *et al.*, 2018).

432 Investigating changes in species spectral niche size, overlap and position in relation to
433 changes in environmental conditions presents an interesting avenue for the future studies. Niches
434 of species can be distinct from each other due to niche separation, niche differentiation or by
435 chance, due to stochastic factors influencing the occurrence of a particular set of individuals with
436 distinct traits in one particular place and time (Cornwell *et al.*, 2006). Investigating species
437 trajectories and niche sizes in spectral space and identifying the absorption features that
438 contribute most to niche shifts and that separate species best under certain environmental
439 conditions can indicate temporal and spatial resource partitioning (Pickett & Bazzaz, 1978).
440 Empty volumes in the spectral space occupied by plant communities might indicate colonization
441 or invasion potential, or, alternatively, biologically unrealized spectral types. In our case, species

442 niche sizes and positions in trait space changed depending on the diversity level of the plant
443 community (Fig. S10), which was in line with species-specific trait changes with community
444 diversity (Fig. S12, Tables S7, S8, Notes S1). Species' niche sizes in spectral space changed
445 depending on community diversity as well, but their positions remained more stable (Fig. S11),
446 providing an additional explanation for the high classification accuracies of spectral species
447 identification models (Table S4). Also, it is likely that biophysical trade-offs in plant functional
448 traits, such as between nitrogen content or photosynthetic capacity and SLA (see e.g. Reich *et*
449 *al.*, 1997; Wright *et al.*, 2004) impose boundaries on spectral profiles limiting the degree of
450 spectral variation possible. This might result both in species spectra diverging or converging in
451 particular cases, and in species, lineages or functional groups following unique trajectories in
452 spectral space depending on environmental conditions and phenology. Incorporating spectral
453 niche shift of plants in identification models could allow groups of plants that are spectrally
454 similar at one point in time to be differentiated with time series of spectral data, which would be
455 particularly useful in diverse environments and for large-scale studies.

456 The concept of the niche integrates many biological characteristics of organisms,
457 providing a conceptual synthesis for the functioning of ecosystems (Leibold, 1995). The
458 integrative nature of plant spectra and their links to plant form, function and phylogeny (Ustin &
459 Gamon, 2010; Cavender-Bares *et al.*, 2017; Schweiger *et al.*, 2018) make the spectral niche
460 concept particularly relevant to ecology. Plants with contrasting functional attributes occupy
461 distinct spectral spaces allowing plant species, lineages and functional groups to be identified
462 spectrally. Furthermore, evaluation of the occupied hypervolumes and the position of
463 individuals, taxonomic groups and communities in spectral space with leaf-level and imaging
464 spectroscopy provides a novel way of assessing plant-plant and plant-resource interactions. The
465 spectral niche concept unites ecological theory and biology with the physics of light capture and
466 distribution to reveal mechanisms of plant species coexistence, their distribution, abundance and
467 diversity.

468

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482

483 **Author contributions**

484 The ideas presented in this article originated through many discussions during the Dimensions of
485 Biodiversity project "Linking remotely sensed optical diversity to genetic, phylogenetic and
486 functional diversity to predict ecosystem processes" conceptualized by JCB, PAT, SEH, MDM
487 and JAG. AKS, JCB, PAT and JAG planned data collection in BioDIV. JCB, JJG, SK and SEH
488 planned data collection in FAB. JAG and RW designed the spectral tram system. AKS, RW, and
489 JAG collected data in BioDIV. JJG and SK collected data in FAB and calculated measures of
490 tree growth. AKS analyzed the data and wrote the first draft of the manuscript, with input from
491 JCB and JAG. All authors contributed substantially to revisions.

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654

Figures

Fig. 1 The unique niche fraction per species increases with increasing dimensionality of the spectral space **(a)** and functional trait space **(b)**. Curves represent second order polynomials fitted to 50 estimates of species niche sizes calculated with increasing number of dimensions (2-21 spectral bands and 2-10 functional traits, respectively); 95% confidence interval are shown in grey. The species are *Sorghastrum nutans* (L.) Nash (27 ind., SORNU), *Schizachyrium scoparium* (Michx.) Nash (76 ind., SCHSC), *Andropogon gerardii* Vitman (162 ind., ANDGE), *Panicum virgatum* L. (49 ind., PANVI), *Lespedeza capitata* Michx. (99 ind., LESCA), *Petalostemum villosum* Nutt. (42 ind., PETVI), *Petalostemum purpureum* (Vent.) Rydb. (52 ind., PETPU), *Petalostemum candidum* (Willd.) Michx. (28 ind., PETCA), *Amorpha canescens* Pursh (28 ind., AMOCA), *Lupinus perennis* L. (121 ind., LUPPE), *Solidago rigida* L. (50 ind., SOLRI), *Liatris aspera* Michx. (49 ind., LIAAS), *Achillea millefolium* L. (49 ind., ACHMI), *Asclepias tuberosa* L. (70 ind., ASCTU). The phylogenetic relationships among species are displayed on the right; graminoids are coded in blue, legumes in purple-red-orange and forbs in yellow-grey-black colors.

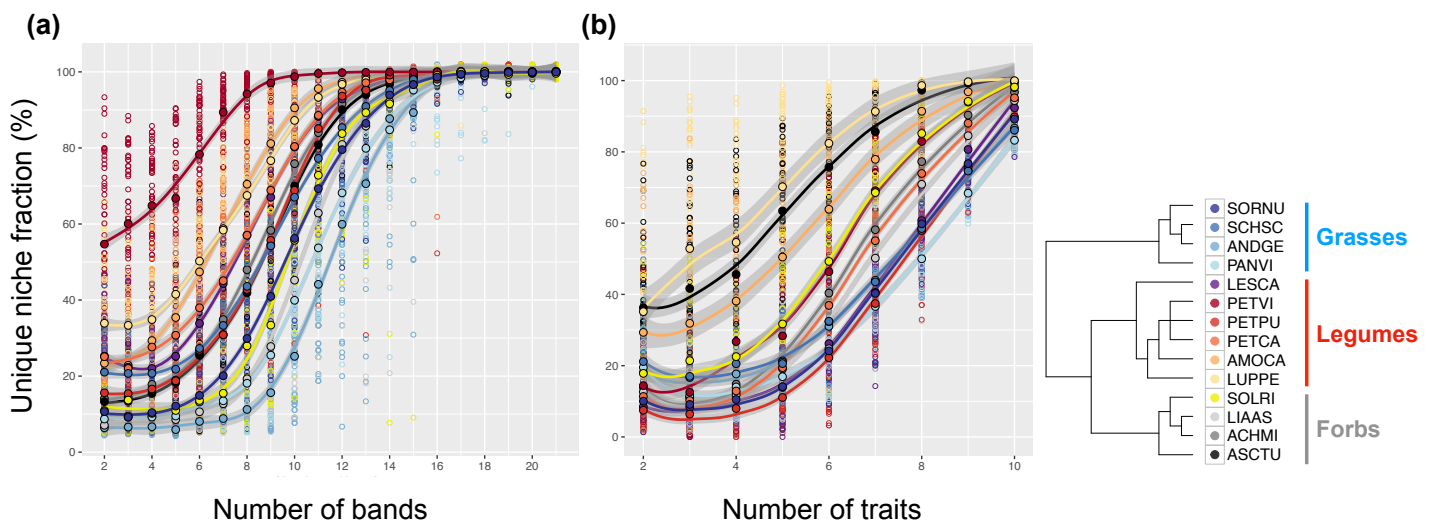


Fig. 2 Species clustering along linear discriminant (LD) axes maximizing the differences among species based on spectra (**a-c**) and functional traits (**d-f**). The amount of the total variation explained by each LD axis is shown in parentheses; for species abbreviations and number of individuals per species see Fig. 1.

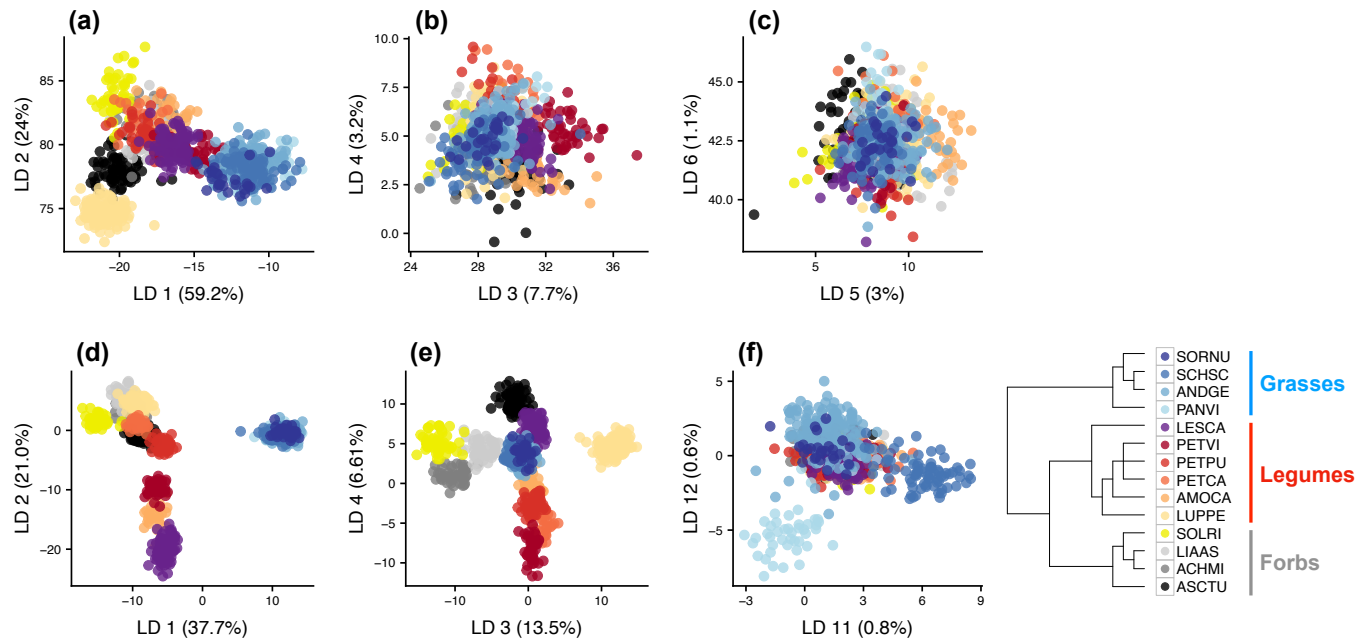


Fig. 3 Probabilities for each species to be classified correctly, or incorrectly classified as another species based on partial least squares discriminant analysis (PLSDA) models using species spectra **(a)** and functional traits **(b)** sampled in the BioDIV experiment in July. PLSDA assigns to every sample the classification probabilities to belong to each class (i.e. species). Colored bars that match the species color, shown at the tip of the phylogeny, indicate the averaged probability that the species was correctly classified. Bars that show a different color indicate the averaged probability that the species was misclassified as a different species. The percentages of correctly classified samples per species in PLSDA are shown in parentheses. For species abbreviations and number of individuals per species see Fig. 1; the graph shows model validation results, i.e. n = the number of individual per species - 20 individuals for model calibration.

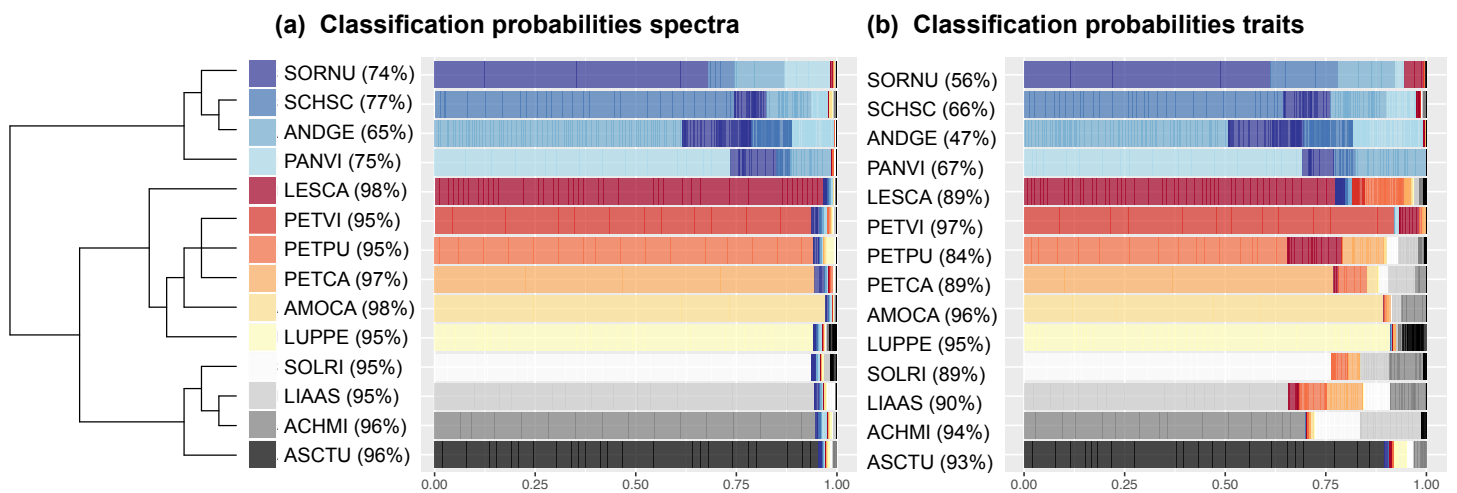


Fig. 4 The spectral space occupied by individual trees in the FAB experiment predicting tree growth. Tree height [(a), $n = 532$, $R^2 = 0.11$, $F_{1,530} = 63.2$, $P < 0.001$; (b), $n = 524$, $R^2 = 0.14$, $F_{1,522} = 81.6$, $P < 0.001$], and stem diameter [(c); $n = 532$, $R^2 = 0.20$, $F_{1,530} = 135.5$, $P < 0.001$; (d), $n = 396$, $R^2 = 0.25$, $F_{1,394} = 130.2$, $P < 0.001$] were measured in 2016 and 2017, respectively; spectral space sizes are log-transformed; the number of species per plot is indicated with different symbols.

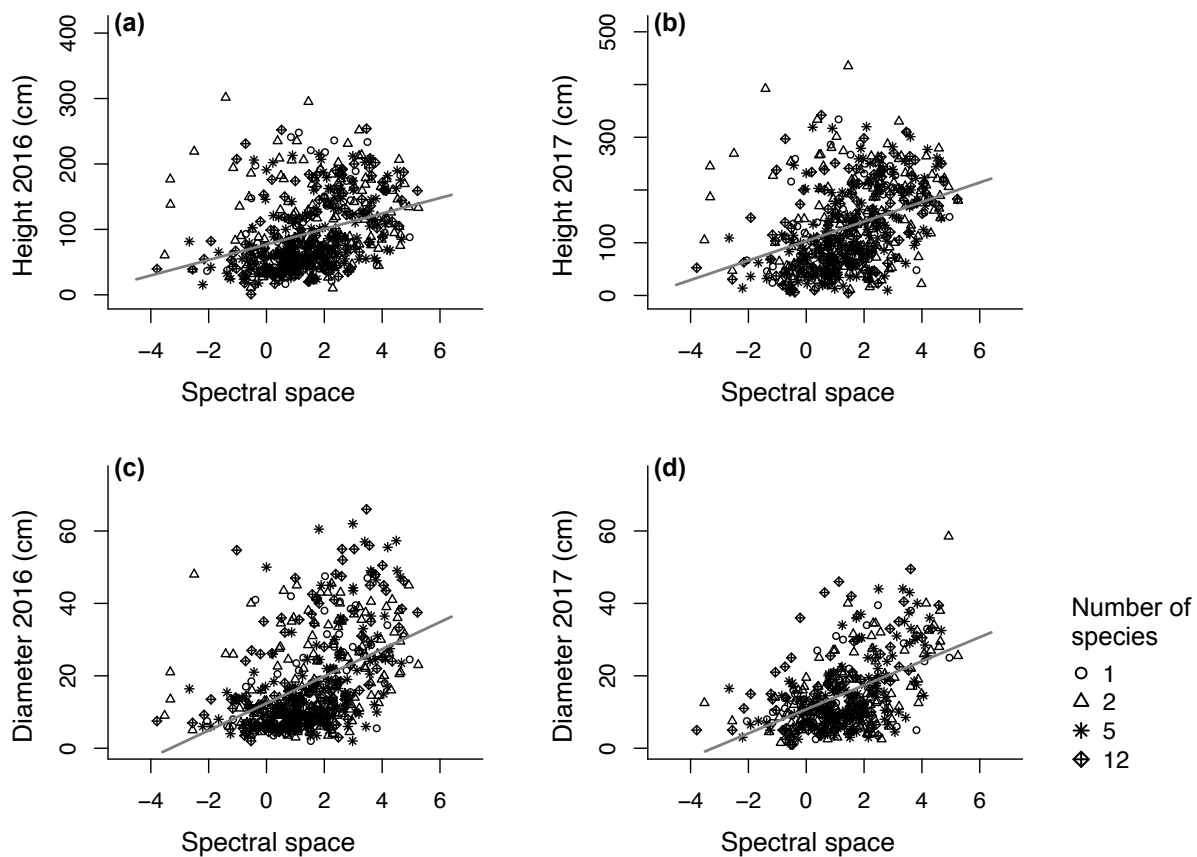


Fig. 5 The spectral space occupied by plant communities predicting aboveground productivity and species richness in the BioDIV experiment. The size of the spectral space occupied per community is calculated from contact measurements of leaf spectra in July [(a), $n = 30$, $R^2 = 0.44$, $F_{1,28} = 22.6$, $P < 0.001$]; (c), $n = 30$, $R^2 = 0.25$, $F_{1,28} = 9.1$, $P < 0.001$] and from imaging spectroscopy data [(b), $n = 18$, $R^2 = 0.31$, $F_{1,16} = 7.1$, $P < 0.017$]; species richness (indicated with different symbols) is the number of species planted per plot; aboveground productivity (g m^{-2}) was determined in clip strips the same month; spectral space sizes are log-transformed.

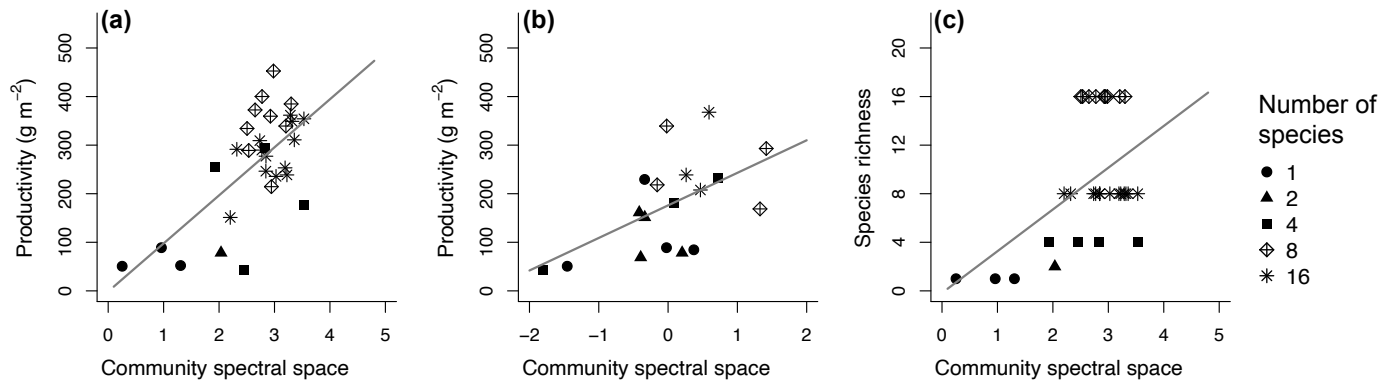


Fig. 6 The spectral space occupied by plant communities predicting the net biodiversity effect [NBE, **(a)**, $n = 43$, $R^2 = 0.37$, $F_{1,41} = 24.3$, $P < 0.001$] and its two components complementarity [CE, **(c)**, $n = 43$, $R^2 = 0.36$, $F_{1,41} = 23.4$, $P < 0.001$], and selection effect [SE, **(d)**, $n = 43$, $R^2 = 0.20$, $F_{1,41} = 10.3$, $P < 0.003$], and species richness [(b), $n = 43$, $R^2 = 0.20$, $F_{1,41} = 10.5$, $P < 0.003$] in the FAB experiment. The size of the spectral space occupied per community is calculated from contact measurements of leaf spectra; species richness (indicated with different symbols) is the number of species planted per plot; NBE, SE and CE are square root-transformed with signs retained (see Grossman *et al.*, 2017); spectral space sizes are log-transformed.

