

1 Title: What's left in the tank? Identification of non-ascribed aquarium's coral collections with DNA barcodes
2 as part of an integrated diagnostic approach

3

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22

23 Abstract

24 The unprecedented threats to coral reef ecosystems from global climate change (GCC) require an
25 urgent response from the aquarium community, which is becoming an increasingly vital coral
26 conservation resource. Unfortunately, many hermatypic corals in aquaria are not identified to
27 species level, which hinders assessment of their conservation significance. Traditional methods of
28 species identification using morphology can be challenging, especially to non-taxonomists. DNA
29 barcoding is an option for species identification of Scleractinian corals, especially when used in
30 concert with morphology based assessment. This study uses DNA barcodes to try to identify
31 aquarium specimens of the diverse reef-forming family *Acroporidae* from 127 samples. We
32 identified to species name 44% of the analysed samples and provided provisional identification for
33 80% of them (101/127, in the form of a list of species names with associate confidence values). We
34 highlighted a sampling bias in public nucleotide sequences repertories (e.g.: GenBank) towards
35 more charismatic and more studied species, even inside a well-studied genus like *Acropora*. In
36 addition, we showed a potential “single observer” effect with over a quarter of the reference
37 sequences used for these identifications coming from the same study. We propose the use of
38 barcoding and query matching as an additional tool for taxonomic experts and general aquarists to
39 increase their chances of making high confidence species-level identifications. We produce a
40 standardised and repeatable methodology to increase the capacity of aquariums and other facilities
41 to assess non-ascribed species.

42

43 Introduction

44 It is widely acknowledged that coral reef ecosystems are facing unprecedented threats from global
45 climate change (GCC) e.g.(Pratchett et al. 2013; Hoegh-Guldberg et al. 2017; Cowburn et al. 2018;
46 Hughes et al. 2018; Masson-Delmotte et al. 2018; Brondizio et al. 2019; Sheppard et al. 2020).
47 Warm water reef-building corals create the most biodiverse ecosystems on the planet and have been

48 estimated to support 830'000 species of multi-cellular plants and animals worldwide (Bellwood and
49 Hughes 2001; Mora et al. 2008, 2011; Fisher et al. 2015), providing a variety of habitats for fish,
50 invertebrates and other taxa in shallow tropical seas (Bellwood and Hughes 2001). Corals in the
51 family *Acroporidae* play a key functional ecosystem role as they are the major reef builders in the
52 majority of warm water reefs ecosystems (Fukami et al. 2000).

53

54 [Corals in aquaria](#)

55 Given the severity of the species and ecosystem level climate change threats to corals (Masson-
56 Delmotte et al. 2018), and their reliant biodiversity, *ex situ* conservation assistance is likely to play
57 an important part in ensuring the survival of many, perhaps the majority, of coral species. Aquaria
58 institutions therefore have a vitally important conservation remit to support coral restoration
59 strategies allowing for both sexual and asexual coral recruitment, biobanking and a wide range of
60 research, education and thus contribution towards *in situ* conservation activities. Early attempts in
61 the 1980s at keeping corals in aquaria faced difficulties (Borneman 2008; Brunner 2012) it was
62 until 1980 that *Acropora* was the the first genus to be successfully cultured by Stüber (Borneman
63 2008). From the 1990s onwards coral husbandry boomed, reaching a level where virtually all
64 families of zooxanthellate corals were not only being maintained for many years, but were being
65 propagated and traded between private and public aquaria (Borneman 2008). Specimens conserved
66 in Aquaria thus, became a highly valuable resource for restoration strategies (Rinkevich 1995).

67

68 [Identification of corals](#)

69 Coral taxonomy is a traditionally difficult discipline, subject to same challenges of modern
70 taxonomy (Godfray 2002). The genus *Acropora* is among the most diverse and geographically
71 widespread reef building corals (Wallace 1999). Research on *Acropora* suggests that species limits

72 may be sometimes narrower, sometimes broader, than generally perceived (Wallace and Willis
73 1994). Currently the number of officially recognised *Acropora* species are between 135 (WORMS
74 (Horton et al. 2020)) and 163 (Corals of the World (Veron et al. 2020)), reaching 186 if we include
75 the “*taxon inquirendum*” on WORMS (Horton et al. 2020). There have been different attempts at
76 grouping these species within the genus *Acropora* (Wallace 1999). These attempts lead to the
77 creation of the concept of species group (aka syngameon). The concept of species groups, first
78 defined for convenience of identification and without implying taxonomic relation, has been revised
79 to reflect the evolution and phylogeny of the genus (Wallace 1999; Wallace et al. 2012). The use of
80 genetic tools for species identification has provided another line of evidence for species
81 delimitation, further adding to the blurred boundaries between species groups, species and sub-
82 species.

83

84 Traditionally, identification of coral has relied on morphological skeletal features, with the great
85 majority of taxa originally described following typological species concepts defined around a
86 century ago (Best et al. 1984; Todd 2008). The relationship between morphological variation
87 within-species and the environment has been a source of contention since the late 1800s (Todd
88 2008). Whether this variability is due to different underlying genotypes or plastic phenotypes is
89 unclear. Evidence supports both scenarios (difference in phenotype e.g.: (Willis 1985; Ayre and
90 Willis 1988; Todd et al. 2004)– Plastic phenotypes e.g.: (Foster 1979; Miller 1994; Bruno and
91 Edmunds 1997; Muko et al. 2000; Todd et al. 2004)) and these scenarios are not necessary mutually
92 exclusive but probable to operate simultaneously (Foster 1979; Amaral 1994; Todd 2008). This has
93 particular relevance in the face of continuous GCC, that is creating an increasingly unpredictable
94 wild reef environment.

95

96 Notwithstanding the taxonomic difficulties, the identification of coral species can be challenging
97 even in optimal ‘wild’ conditions. However, identification requirements are far more than just for

98 well documented field collections, but also cover aquaria collections, that in some cases are held in
99 conditions dramatically different from the wild. Furthermore customs officials rely in proper
100 taxonomic identification even at the genera level for threatened taxa by the international wildlife
101 trade (AC25 Doc. 23 CITES 2011). A status of these identification difficulties in aquaria can be
102 demonstrated by an analysis of Species360 Zoological Information Management Software (ZIMS),
103 adopted by more than 95 aquariums in 24 countries. Assuming correct identification ZIMS
104 indicates c. 42.9% of corals of the orders Scleractinia, Alcyonacea, Helioporacea, Antipatharia,
105 Corallimorpharia, Pennatulacea, and Zoanthidea in aquaria are identified to genus or higher
106 taxonomic level (ZIMS list of species holdings – 4th June 2020). Most of the records correspond to
107 the order Scleractinia, which is also the order with more species represented in aquariums.
108 Corallimorpharia is, however, the order with the highest number of reported individuals in
109 aquariums members of the Species360 network (Fig. 1). In addition to this the non-species
110 associated genera (e.g.: *Acropora* sp.) is usually recorded once in the ZIMS database and most other
111 aquarium inventory systems, which possibly significantly underestimates the actual number of non-
112 identified specimens. Moreover, there are currently no confidence values ascribed to any taxonomic
113 level a collection allocates to a specimen which means that there is likely to mean that a significant
114 number of species ascribed specimens are actually at the lower end of the confidence spectrum and
115 would benefit from being reassessed.

116 In many aquariums' corals come from three sources: fragments from other aquaria, via ornamental
117 trade, or from confiscated and seized shipments (which aquariums are usually asked to
118 accommodate). All three sources present identification challenges as often exact knowledge from
119 their wild origin is absent. This specially a challenge for confiscated specimens since they have
120 none or relatively poor source records, sample integrity, and any other additional background
121 information that might be of assistance for the identification process. This, together with the
122 difficulties caused by environmental phenotypic plasticity makes coral identification in aquariums a
123 particularly challenging task. Data from the Convention on the International Trade of Endangered

124 Species of Flora and Fauna (CITES) shows a total of 4557 instances of coral confiscation from
125 1982 to 2018 worldwide, for a total of 27027 individual items and 5461 kg per year (CITES data -
126 <https://trade.cites.org> 02/07/2020). Of these, less than 20% are reported at species level, meaning
127 that for most cases we know only the genus, family or even order of the specimens confiscated.
128 Given these challenges CITES requires ideally identification to the order level.

129

130 DNA Barcoding

131 An alternative or complementary diagnostic option to morphological identification is genetic
132 analysis. DNA Barcodes have been proposed by Hebert et al. (Hebert et al. 2003) as a way to
133 circum-vent the “limitations inherent in morphology-based identification systems”. The core
134 concept of DNA barcoding is the use of a standard sequence corresponding to a single homologous
135 region that can be amplified with universal primers that is sufficiently variable to distinguish
136 between species – usually represented by the mitochondrially encoded cytochrome c oxidase I (MT-
137 CO1 or CO1 or COX1) (Hebert et al. 2003).

138

139 DNA barcoding is especially attractive as an option for species identification of Scleractinian, as
140 this method bypasses the problems caused by phenotypic plasticity or life stage (larva, juvenile or
141 adult). Unfortunately, the proposed ‘universal’ Barcoding region CO1 is at least 10-20 times slower
142 in Cnidaria than the standard vertebrate mtDNA rate (Van Oppen et al. 1999; Shearer et al. 2002)
143 with zero interspecific divergence for many species (Shearer and Coffroth 2008). This has led to the
144 use of a number of different barcoding regions for corals (Shearer et al. 2002).

145 The diverse, reef forming *Acropora* group are widely grown in aquaria and have been a target for
146 barcoding work. Potential barcodes were reviewed by Shearer et al. (Shearer et al. 2002) examining
147 both nuclear and mitochondrial regions. Substitution rates of mitochondrial genes in *Acropora* tend
148 to be lower relative to the nuclear counterparts mentioned by Shearer (i.e. Pax-C intron; internal
149 transcribed spacer 1 and 2 (ITS) (Shearer et al. 2002). Although, Van Oppen et al. (van Oppen et al.

150 1999) and Vollmer et al. (Vollmer 2002) have highlighted the mitochondrial putative control region
151 (mtCR) is more variable than the other mitochondrial regions. Subsequently Shearer et al. (Shearer
152 et al. 2002) compared the genetic divergence of mitochondrial and nuclear regions in Scleractinian,
153 highlighting how the mtCR region has a similar range of divergence to nuclear coding regions like
154 the Pax-C intron, but is much slower than the hyper-variable ITS1-2.

155 This study seeks to build on these studies by investigating the utility of DNA barcoding as a method
156 for species identification of *Acropora* collections in aquaria as part of an integrated diagnostic
157 approach.

158

159 Material and Methods

160 Study Area and Collection

161 As part of the “idCoral” project we collected 224 samples across seven institutions (Aquarium de
162 Paris (France), Chester’s Zoo (UK), Haus des Meeres-Vienna (Austria), Horniman Museum and
163 Gardens (UK), Royal Burgers’ Zoo (Netherlands), Tierpark Hagenbeck Aquarium (Germany), ZSL
164 London Zoo (UK)). Branch tip fragments of 2 cm were collected and stored in single ethanol
165 (+95%) vials labelled with a unique identifier. During the sample collection the colonies were
166 tagged and photographed with collection dates and tank locations were recorded. The photographs
167 and tags were also recorded in the “idCoral” database (database - idcoral.org (2020)). After
168 collection samples were stored at -20°C until extraction, with an initial change of ethanol after 1-2
169 days aimed at reducing degradation. We analysed 127 *Acropora* samples from the above mentioned
170 collection.

171

172 DNA Extraction, Amplification and Sequencing

173 The DNA extractions were undertaken using the DNeasy PowerSoil Pro Kit from Qiagen[®] with a
174 modified version of the manufacturer's protocol. An incubation step with proteinase K was added
175 and the physical breaking of the cell structure with a chemical dissolution was removed from the
176 protocol as it resulted in excessive shearing of the DNA (Online Resource 1). Extracted DNA was
177 then preserved at -20°C, and the DNA concentration and quality tested with both Nanodrop assay
178 and gel electrophoresis. The DNA was diluted to standardise the concentration to between 5 to 20
179 µg/ml.

180 Two genes, PaxC and mtCR were selected as barcodes based on an assessment of their variability
181 and availability as reference sequences (Van Oppen et al. 2001; Shearer et al. 2002). The ITS1-2
182 region was excluded due to the low availability of reference sequences. PCR amplification was
183 conducted with Q5[®] High-Fidelity DNA Polymerase mixed following manufacturer's indication.
184 3µl of DNA template and 0.5 µl of Bovine Serum Albumin (BSA)1µg were added, reaching a final
185 volume of 25µl with RNA free water. For the PaxC gene the primer used were PaxC FP1 5'-TCC
186 AGA GCA TTA GAG ATG CTG G-3' and PaxC RP1 5'-GGC GAT TTG AGA ACC AAA CCT
187 GTA-3' (Van Oppen et al. 2001) with a protocol of 98°C for 30s, followed by 31 cycles at 98°C for
188 10 s, 62°C for 30 s and 72°C for 60 s, ending with a final phase of 72°C for 2min. For the mtCR the
189 primer used were RNS2: 5'-CAG AGT AAG TCG TAA CAT AG-3' and GR: 5'-AAT TCC GGT
190 GTG TGT TCT CT-3' (Suzuki et al. 2008) with a protocol of 98°C for 30s, followed by 40 cycles
191 at 98°C for 10 s, 62°C for 15 s and 72°C for 60 s, ending with a final phase of 72°C for 5min.

192 Successful PCR amplifications (assessed by gel electrophoresis) were sent to Eurofins Genomics
193 for PCR cleaning and custom Sanger sequencing on both the forward and the reverse primers using
194 cycle sequencing technology (dideoxy chain termination / cycle sequencing) on ABI 3730XL

195 sequencing machines. The returned sequences were manually trimmed in order to remove poor
196 quality sections and assembled using Geneious Prime[®] (2019.1.2.).

197

198 Identification

199 Taxonomy

200 Both the barcode gap and specimen identifications were initially performed at the species level, but
201 were also performed at the level of species group based on the revision of *Acropora* by Wallace et
202 al. (Wallace et al. 2012).

203

204 Barcoding Gap

205 References sequences for the two targeted regions were collated. FASTA files were created with the
206 records matching the following query on GenBank: mtCR="Acroporidae"[Organism] AND
207 mitochondrial "complete genome" OR "control region" OR "putative control region" OR
208 mitochondrial control region"; PaxC="Acroporidae"[Organism] AND PaxC[gene] OR pax-
209 c[gene]". Two multiple alignments were performed, one for each gene, using the Geneious[®]
210 algorithm (progressive pairwise aligner) and trimmed. For both the produced references alignments
211 distance matrices were calculated using the function `dist.dna` (R package `ape` (Jombart et al. 2017))
212 and with different models (Raw ; JC69; K80; K81; F84; BH87; paralin; indel Y; indelblock).
213 Successively the frequency distribution of intra/inter-specific distances was plotted. The quantile
214 function (R package "stats") was used to define the 99; 50; 1% quantile confidence interval used in
215 the blast match analysis. Note: the two alignments were constructed and used only for the specific
216 purposes of examining the barcoding gap for the targeted region and determining potential
217 confidence interval for the blast match analysis.

218

219 **Blast match**

220 Query based specimen identification with blast match (Benson 2000) was determined using a
221 custom Blastn database. All records from the query “Acroporidae”[Organism]” on GenBank were
222 downloaded and included in the local database built with the packages rBLAST (Hahsler and Nagar
223 2019) and rentrez (Winter 2017). A local Blast match (Camacho et al. 2009) was ultimately
224 performed. To filter for species match with identical scores a decision tree (Fig. 2) was designed
225 relying on Bit score, Percentage identity, and number of mismatches. Matches shorter than 2/3 of
226 the original sample sequence query were removed. The decision tree prioritised 100% pairwise
227 identity matches were possible, then to the highest bit score, then lowest number of base
228 mismatches. 100% pairwise identity matches to a species group rather than a single species name
229 were prioritised over single species name matches with lower pairwise identity with higher bit
230 score. We confronted the match from the two targeted regions to obtain a single most probable ID.
231 To account for conflicting matches between the two region we relied on the 99; 50; 1% quantile
232 distribution of the intraspecific distances obtained from the barcoding gap analysis. Using pairwise
233 identity threshold as confidence intervals to discern which region offers the better match for the ID
234 (R-scripts as Online Resource 3 and 4).

235

236 **Results**

237 From the 127 collected samples all were successfully extracted and, 107 mtCR and 100 PaxC had a
238 successful DNA amplification, respectively 84% and 79% with a 76 samples overlap between the
239 two region (Table 1). 348 reads were successfully sequenced, 84 mtCR sample and 90 PaxC. The
240 median sequence lengths were 734 post assembly for the mtCR region and 649 for the PaxC region
241 (Table 1).

242

243 **Barcoding**

244 Our GenBank query produced a total of 1348 mtCR sequences and 427 PaxC sequences. The mtCR
245 references query, once cleaned for clones, produced a 579 sequences alignment of 914bp. The PaxC
246 references query produced a 352 sequence alignment of 770 bp.

247

248 The mtCR alignment of 579 sequences resulted in a matrix of 335 241 pairwise genetic distances,
249 grouped by 57 species names. 6.4% of these (21 471) are intraspecific distances, while 93.6% (313
250 770) are interspecific distances (Fig. 3A). The PaxC alignment of 352 sequences resulted in a
251 matrix of 123 904 genetic distances, grouped by 55 species names. 3% of these (3 726) are
252 intraspecific distances, while 97% (120 178) are interspecific distances (Fig. 3B). Between the two
253 groups there is an overlap of 49 species names. Table 3 and 4 with the sources of the sequences
254 used in these alignments.

255

256 Fig. 3 highlights in blue the intraspecific variation (sequences with the same name on GenBank)
257 and in the light green the interspecific distances (sequences with different names on GenBank). In
258 both Fig. 3A) mtCR and 3B) PaxC we see no discernable barcoding gap.

259

260 The quantile within species distribution produce the following distance threshold for our confidence
261 interval: mtCR 0.027 (99%); 0.0068 (50%); 0 (1%) and PaxC 0.17 (99%); 0.014 (50%); 0 (1%).

262

263 In Fig. 4 and 5, for the mtCR and the PaxC respectively, we can see the above mentioned uneven
264 distribution of the GenBank records towards specific species and species groups. With example like
265 species group *humilis* showing 4 times the frequency of less represented species group like *loripes*.

266

267 Blast match

268 The local database for blast matching contained 510826 sequences. In Table 2 we see that 23 of our
269 76 samples (with both regions sequenced) were uniquely identified to a single species, while 13
270 samples matched to conflicting names. 20 of our samples had a unique species name match for the
271 PaxC region but mtCR region matched to 2 or more names with all but 6 of these including the
272 matching PaxC name. Very few samples matched to multiple species for both regions (4).

273

274 We have 47 % of samples matching to a single species name with both regions, 5% with the same
275 identical match. While we have 42% of the matches to a single species name that are in conflict,
276 25% of these conflicts are resolvable (i.e.: one of the two region with a match better fitting than the
277 other one, based on the defined confidence thresholds of each region). Similar results for the other
278 sample matching for two or more in at least one of the two region. Out of the 76 samples, in 15
279 cases we observe the second region adding information to the first ones (Table 5 as Online
280 Resource 2). Mainly from a mtCR multi name ID to a single PaxC name (i.e.: sample P269 has 7
281 mtCR identical matches to 7 different species names, and the PaxC matches at a lower rate to only
282 one of them). In addition, the second region functions as backup option whenever one of the two
283 falls too short (5/76 sample have too low-quality match to be trusted independently).

284

285 The complete list of our current best species name match, with species group information and
286 detailed genetic info is provided as Online Resource (Table 5 – Online Resource).

287

288 Discussion

289 Barcoding gap

290 Using DNA barcodes to identify unknown samples is only achievable if a well- studied, well-
291 sampled reference sequence database is available (Meyer and Paulay 2005), and even though

292 databases such as GenBank (which comprises the DNA DataBank of Japan (DDBJ), the European
293 Nucleotide Archive (ENA), and GenBank at NCBI - <http://www.ncbi.nlm.nih.gov/genbank>) and
294 BOLD (Barcode Of Life Data System - <http://www.boldsystems.org>) contain an ever increasing
295 number of species-level identified sequences (i.e.: 216 531 829 total sequences on GenBank as
296 April 2020), these databases are far from complete. The open nature of GenBank is destined to
297 produce a bias towards species with a higher research interest (i.e.: Human sequences constitute
298 56% of the total sequences on GenBank (Benson 2000)). The genus focus of our study, *Acropora*, is
299 one of the more commonly studied and widespread hard corals in the ocean (Wallace 1999; Fukami
300 et al. 2000) making it one of the most covered coral genera in GenBank. The reality is that even for
301 a widely studied group such as *Acropora* we can observe a disparity in the number of sequences
302 available for each species name (Fig. 4 and 5). The best example of this is the delta between two
303 species group: group *humilis* being the most sampled (251 mtCR and 61 PaxC references sequences
304 from our query) and group *robusta* one of the least sampled (11 mtCR and 13 PaxC references
305 sequences from our query). Note that both species group are formed by of the same number of
306 species (8 species each (Wallace et al. 2012)). This disparity is intrinsic to the nature of these public
307 repertoires, as species of higher interest will always be more studied than others, creating the
308 necessity to acknowledge the potential bias when using these resources for identification purposes.

309

310 From our analysis we could not identify a clear barcoding gap in both our targeted regions (Fig. 3),
311 and furthermore we can observe a total overlap of the intra-specific distances by the inter-specific
312 distances. For both our targeted regions, even if more evident in the mtCR (Fig. 3A), we see that the
313 inter-specific distances follow a distribution reminiscent of a more common and more expected
314 barcoding distribution. We performed the same barcoding gap analysis at the higher taxonomic
315 level using the species group information without obtaining a noticeable difference. With the lack
316 of a clear Barcoding gap, and accounting for the extensive criticism present in the literature: For the
317 use of distances in taxonomy and systematics (Ferguson 2002; Lee 2004; Moritz and Cicero 2004;

318 DeSalle et al. 2005; Prendini 2005; Knapp et al. 2005; Meyer and Paulay 2005; Cognato 2006;
319 Hickerson et al. 2006; Little and Stevenson 2007; Meier et al. 2008); for the drawbacks of using
320 GenBank data (Meier et al. 2006); for the evidence of different rates of evolution in Scleractinian
321 (Van Oppen et al. 1999; Shearer et al. 2002; Kitahara et al. 2010). We cannot establish a simple
322 distance base threshold to use for our identification purposes, leading us to depend more on the
323 blast match.

324

325 Another potential issue with the use of these methods of identification is the source of the reference
326 sequences. With databases like GenBank, with little or no review on the uploaded information, the
327 question is the confidence that the initial identification of the now available reference sequences is
328 indeed correct (James Harris 2003). In our specific case after the processing we performed for the
329 mtCR reference alignment, out of the total 579 included 42% (242 sequences) come from a single
330 doctoral thesis project, while 5.4% (33 sequences) are from unpublished papers (Table 3). For the
331 PaxC 16% (58 sequences) are from unpublished papers, and another 39% (138 sequences) from a
332 single paper (Richards et al. 2008) (Table 4). Highlighting clearly the need for more study on the
333 subject to increase the number of potential sources, and alleviate the possible bias of a “single
334 observer”.

335

336 **Blast match**

337 The use of query match with blast is far from being a complete solution, with 26 out of 76 samples
338 without a single species or species group resolved ID (Table 5), we have proposed a theoretical
339 decision tree that standardise the choice when there is a need of sorting through the output of a
340 multi-species name blast match (Fig. 2). In addition we used two different regions, a mitochondrial
341 one and a nuclear one to increase the confidence in our results (Suzuki et al. 2008). Our proposed
342 species ID are provisional and subject to revision when more species and sequences are available,
343 hence the value of a repeatable process for matching and re-matching.

344

345 All methods of identification that relay on DNA barcodes (e.g.: Blast matching, Barcoding Gap
346 analysis, etc.) use genetic divergence thresholds to assign individuals to correct species. This based
347 on two fundamental assumptions 1) monophyly of species with respect of the molecular marker
348 used, and 2) intraspecific genetic divergence is much smaller than interspecific genetic differences
349 (Toffoli et al. 2008). The first assumption is more dependent on the experimental choices made (i.e.
350 sequence region selected). The second assumption is where most of the criticism of barcodes
351 methods lies, on not comparing sister species or geographical distribution thus, underestimating
352 intraspecific genetic variability (Moritz and Cicero 2004; Toffoli et al. 2008). Fundamentally, the
353 issue of using thresholds lies in the fact that species naturally embodies an evolutionary process,
354 being subject to demographic and selective processes that will act on the genetic diversity (e.g.
355 (Avice et al. 2000; Hey 2001; Coyne and Orr 2004; Toffoli et al. 2008)). Since species are real
356 evolutionary groups and not categories which are created as a direct function of perceived
357 distinction (Hey 2001), the use of thresholds in “discovering” new species is overly simplistic, and
358 in some cases even misleading (Toffoli et al. 2008). We argue against simply substituting
359 morphology based taxonomy with DNA barcoding based taxonomy, as each system is more adapt
360 to answer different questions (Toffoli et al. 2008). Correct assignment is thus only possible by
361 complementing DNA barcoding with other data types, such as morphological, and ecological
362 characters as part of an integrated diagnostic approach (Toffoli et al. 2008).

363

364 While the simple blast match is inheritably less complex than other traditional DNA barcodes
365 identification methods (i.e.: the above mentioned barcoding gap analysis and, the potentially more
366 direct Automatic Barcoding Gap Discovery (ABGD) (Puillandre et al. 2012)). The argument we
367 make is that given the shared limitation and the common need for additional data in an integrated
368 approach, a faster and simpler blast match provides a relatively similar additional amount of
369 information for the species identification while being more easily repeatable, especially considering

370 the value standardized data among aquariums (da Silva et al. 2019) and in light of the ever
371 increasing quantity of reference materials being added to GenBank and BOLD.

372

373 The reality is that barcoding as an independent tool of identification has its clear limitation (Van
374 Oppen et al. 1999; Shearer et al. 2002; Ferguson 2002; Moritz and Cicero 2004; Lee 2004; DeSalle
375 et al. 2005; Meyer and Paulay 2005; Prendini 2005; Knapp et al. 2005; Cognato 2006; Hickerson et
376 al. 2006; Meier et al. 2006, 2008; Little and Stevenson 2007; Kitahara et al. 2010), and coral
377 barcoding is no exception. With some of the concern on the use of genetic repertoires in common of
378 both the methods explored in this study, we strongly recommend an integrated diagnostic approach,
379 combining morphological and genetic means of identification, in agreement with C. Moritz in
380 “DNA Barcoding: Promise and Pitfalls” (Moritz and Cicero 2004). Proposing the use of barcoding
381 and query matching as an additional tool for identification, increasing confidence of experts and
382 both confidence and capacity of non-experts taxonomist. Producing this way reliably identified
383 sequences that can potentially become new reference, adding to the repertoires already freely
384 available. Considering the specific needs of the aquarium community and their critical role to
385 support corals conservation (i.e., by hosting at least 24% of the corals assessed as highly vulnerable
386 to climate change (da Silva et al. 2019)), we emphasise the value of integrating standardised
387 barcoding analysis, and specimen morphology photographs to optimise usage of authoritative
388 identification guides and expert opinion.

389

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394 **Conflict of interest**

395 On behalf of all authors, the corresponding author states that there is no conflict of interest.

396

397 *Fig. 1: Number of records from member institutions of the Species360 network at genus or higher*
398 *level (i.e. order, family and genus) and at species or lower level (i.e. species and subspecies). This*
399 *information is based on the “rank” field from the holdings information reported by each institution*
400 *through the ZIMS software. Information downloaded from ZIMS (Species360, 2020) on 4th June*
401 *2020.*

402

403 *Fig. 2: Decision tree for conflict solving based on Pairwise percentage identity, Bit score and*
404 *number of mismatches.*

405

406 *Fig. 3: Barcoding gap (raw distance matrix frequency distribution). A) mtCR max distance =*
407 *0.0877 minimum distance = 0 (Bin size=0.002) B) PaxC max distance = 0.5895 minimum distance*
408 *= 0 (Bin size =0.05)*

409

410 *Fig. 4: Number of mtCR references sequences with the same species name by species group*

411

412 *Fig. 5: Number of PaxC references sequences with the same species name by species group*

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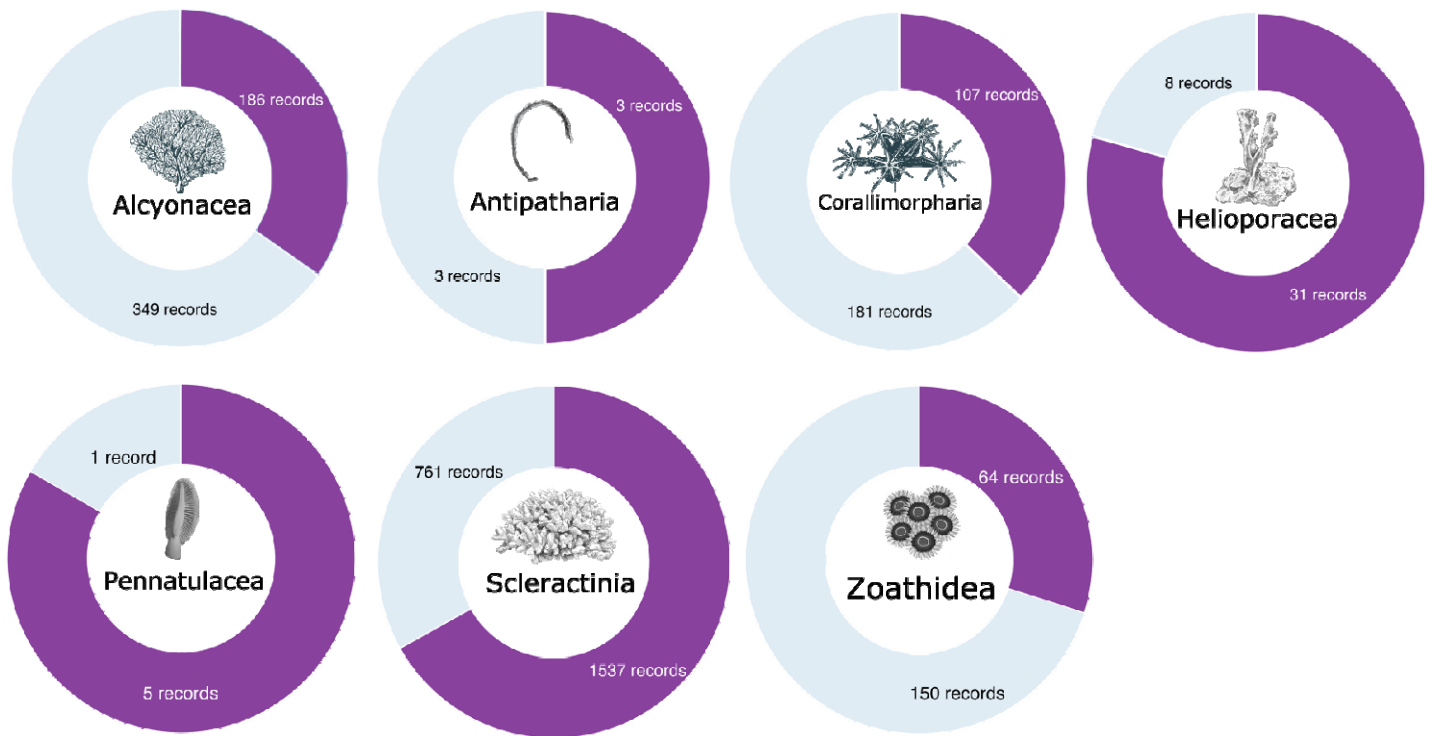


Figure 1: Number of records from member institutions of the Species360 network at genus or higher level in light blue (i.e. order, family and genus) and at species or lower level in purple (i.e. species and subspecies). This information is based on the rank field from the holdings information reported by each institution through the ZIMS software. Information downloaded from ZIMS (Species360, 2020) on 4th June 2020.

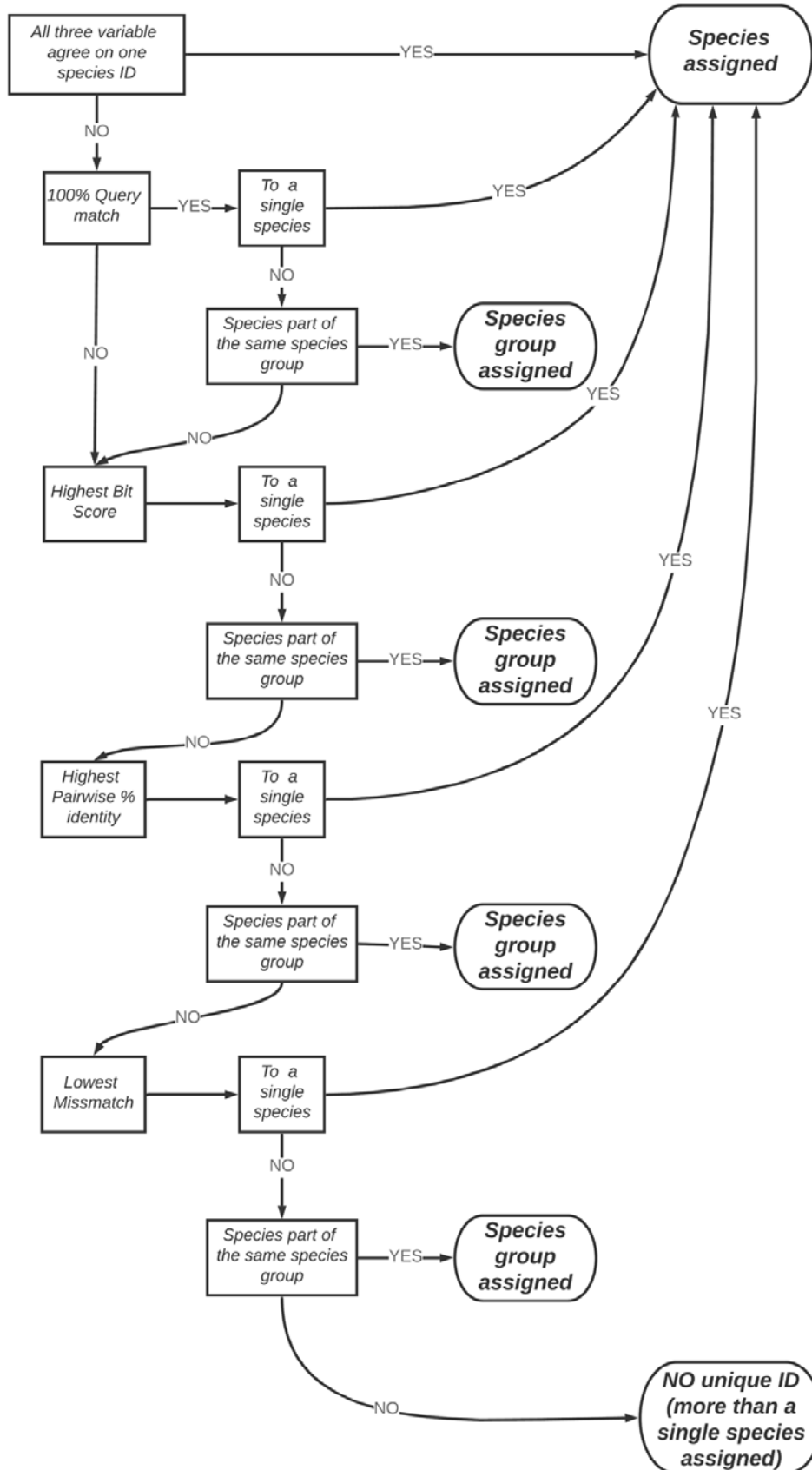


Figure 2: Decision tree for conflict solving based on Pairwise percentage identity, Bit score and number of mismatches.

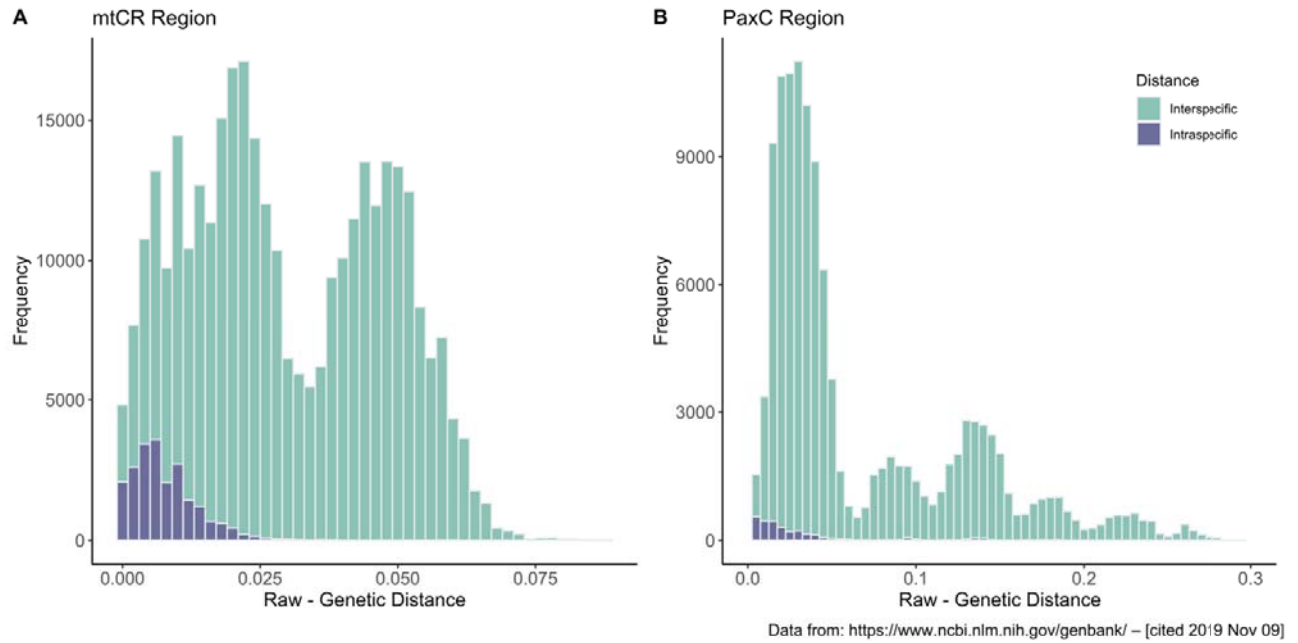


Figure 3: Barcoding gap (raw distance matrix frequency distribution). A) mtCR max distance = 0.0877 minimum distance = 0 (Bin size=0.002) B) PaxC max distance = 0.5895 minimum distance = 0 (Bin size =0.05)

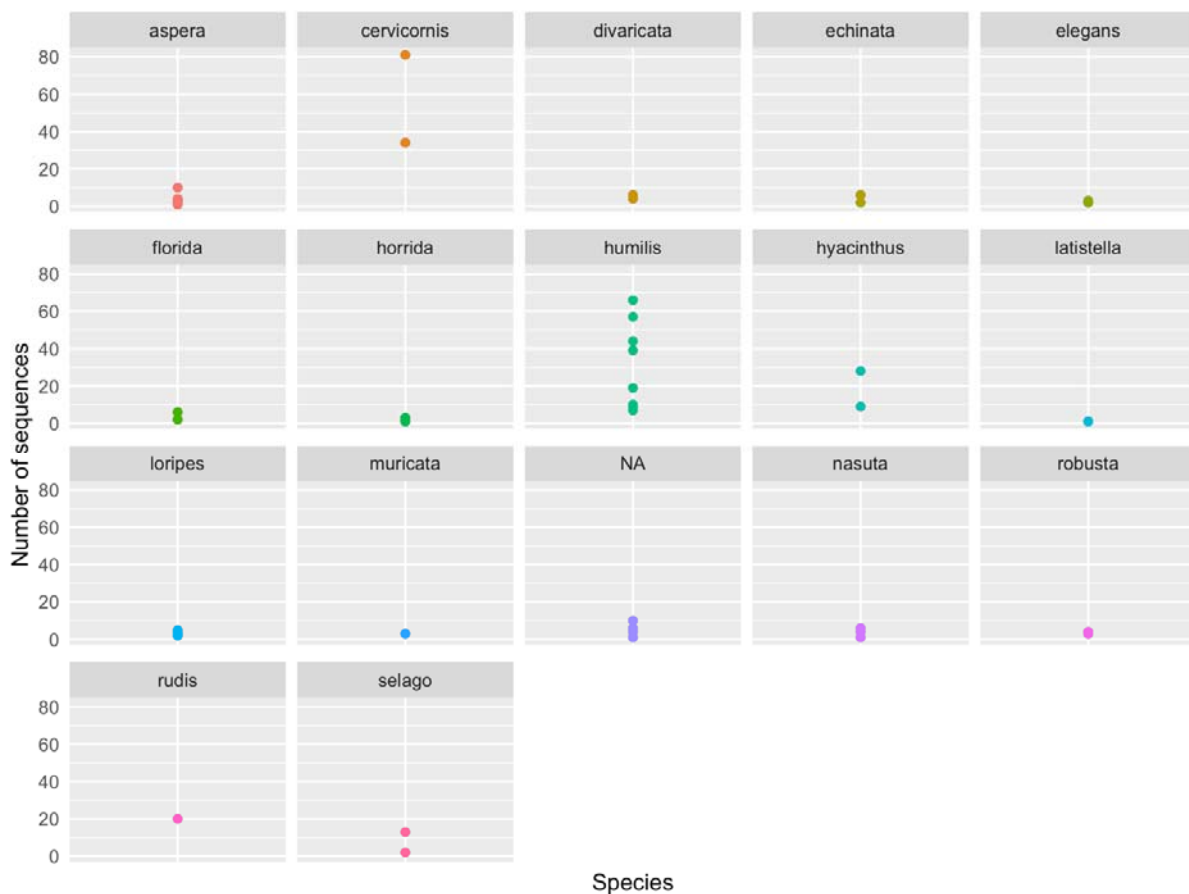


Figure 4: Number of mtCR references sequences with the same species name by species group

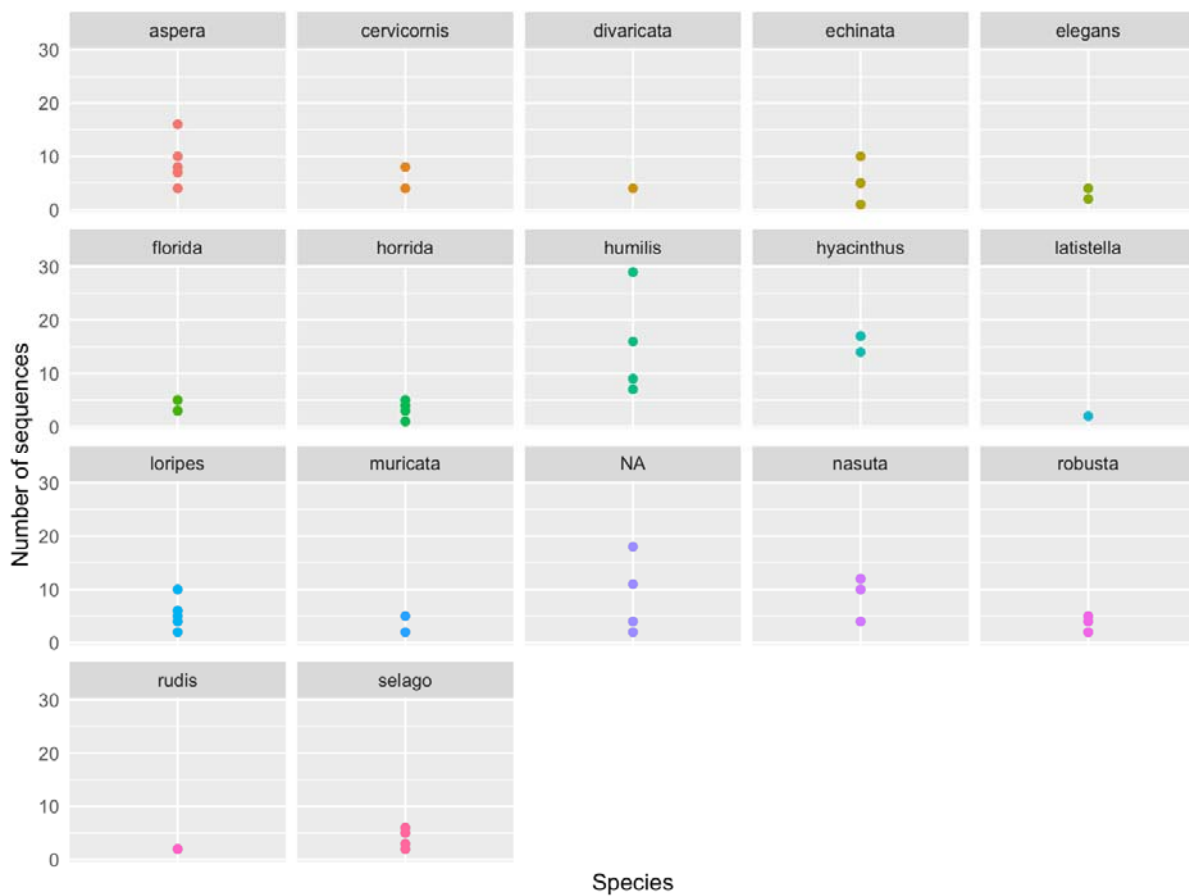


Figure 5: Number of PaxC references sequences with the same species name by species group

Table 1: Success rate across genes

N samples	N samples overlapping	Gene	N Sequences	Minimum sequence length	Maximum sequence length	Median sequence length
127	76	<i>mtCR</i>	84	563	904	734
		<i>PaxC</i>	90	429	869	649

Table 2: Number of sample matching to a single or more (2+) species name on GenBank. Above the diagonal line match that agree between the two region. Below the diagonal line matches that disagree between the two region

		mtCR	
		1	2+
PaxC	1	23 / 13	20 / 6
	2+	6 / 4	1 / 3

Table 3: Sources of *mtCR* references sequences by Paper title

Journal	Authors	Title	N	%
Doctoral Thesis (2003) James Cook University, Townsville, Queensland, Australia	Wolstenholme JK	Species boundaries in scleractinia n corals: a case study of the <i>Acropora humilis</i> species group	242	41.8
PLOS ONE 3 (9), E3240 (2008)	Z. T. Richards, M. J. H. van Oppen, C. C. Wallace, B. L. Willis, and D. J. Miller	Some rare Indo-Pacific coral species are probable hybrids	79	13.6
Conserv. Genet. 18 (4) 825–835 (2017)	M. D. Waterhouse, C. Blair, K. W. Larsen, and M. A. Russello	Genetic Variation and Fine-Scale Population Structure in the Threatened <i>Acropora palmata</i> and <i>Acropora cervicornis</i> around Puerto Rico	48	8.3
Science 296 (5575), 2023-2025 (2002)	S. V. Vollmer	Hybridization and the evolution of reef coral diversity	42	7.3
Mol. Biol. Evol. 18 (7), 1315-1329 (2001)	M. J. H. van Oppen, B. J. McDonald, B. Willis, and D. J. Miller	The evolutionary history of the coral genus <i>Acropora</i> (Scleractinia, Cnidaria) based on a mitochondrial and a nuclear marker: reticulation, incomplete lineage sorting, or morphological convergence?	41	7.1
PLoS One 5 (1) e8652 (2010)	E. M. Hemond and S. V. Vollmer	Genetic Diversity and Connectivity in the Threatened Staghorn Coral (<i>Acropora cervicornis</i>) in Florida	32	5.5
Mol. Ecol. 11 (8), 1339-1349 (2002)	L. M. Márquez, M. J. H. Van Oppen, B. L. Willis, A. Reyes, and D. J. Miller	The highly cross-fertile coral species, <i>Acropora hyacinthus</i> and <i>Acropora cytherea</i> , constitute statistically distinguishable lineages	21	3.6
Zool. Sci. 29 (2), 134-140 (2012)	G. Suzuki and H. Fukami	Evidence of genetic and reproductive isolation between two morphs of subtropical-dominant coral <i>Acropora solitaria</i> ensis in the non-reef region of Japan	13	2.2
Mol. Ecol. 24 (19), 5006-5019 (2015)	N. L. Rosser	Asynchronous spawning in sympatric populations of a hard coral reveals cryptic species and an ancient genetic lineages	8	1.4
PeerJ 7, e6429 (2019)	H. Fukami, K. Iwao, N. H. Kumagai, M. Morita, and N. Isomura	Maternal inheritance of F1 hybrid morphology and colony shape in the coral genus <i>Acropora</i>	6	1.0
Syst. Biodivers. 8 (2), 281–288 (2010)	Z. T. Richards, C. C. Wallace, and D. J. Miller	Archetypal 'elkhorn' coral discovered in the Pacific Ocean	6	1.0
Mol. Phylogenet. Evol. 63 (2), 527-531 (2012)	Y. Nakajima, A. Nishikawa, A. Iguchi, and K. Sakai	The population genetic approach delineates the species boundary of reproductively isolated corymbose acroporid corals	4	0.7
Mar. Ecol. Prog. Ser. 355, 149–159 (2008)	G. Suzuki, T. Hayashibara, Y. Shirayama, and H. Fukami	Evidence of species-specific habitat selectivity of <i>Acropora</i> corals based on identification of new recruits by two molecular markers	3	0.5
J. Mol. Evol. 55 (1), 1-13 (2002)	M. J. H. Van Oppen, J. Catmull, B. J. McDonald, N. R. Hislop, P. J. Hagerman, and D. J. Miller	The mitochondrial genome of <i>Acropora tenuis</i> (Cnidaria; Scleractinia) contains a large group I intron and a candidate control region	1	0.2
Mitochondrial DNA, 1-2 (2015) In press	Y. Zhang, X. Yu, Z. Zhou, and B. Huang	The complete mitochondrial genome of <i>Acropora aculeus</i> (Cnidaria, Scleractinia, Acroporidae)	1	0.2
Proc. R. Soc. Lond., B, Biol. Sci. 266 (1415), 179-183 (1999)	M. J. H. Van Oppen, B. L. Willis, and D. J. Miller	Atypically low rate of cytochrome b evolution in the scleractinian coral genus <i>Acropora</i>	1	0.2
Unpublished	Santacruz-Castro, A. and Dai, C.F.	Latitudinal cline in the reproductive traits and local adaptation of the <i>Acropora hyacinthus</i> species complex	20	3.5
Unpublished	Chan, C.L. and Chen, C.A.	Multiplex next generation sequencing of scleractinian mitochondrial genomes	11	1.9

Table 4: Sources of PaxC references sequences by Paper title

Journal	Authors	Title	N	%
PLoS ONE 3 (9), E3240 (2008)	Z. T. Richards, M. J. H. van Oppen, C. C. Wallace, B. L. Willis, and D. J. Miller	Some rare Indo-Pacific coral species are probable hybrids	138	39.2
Mol. Biol. Evol. 18 (7), 1315-1329 (2001)	M. J. H. van Oppen, B. J. McDonald, B. Willis, and D. J. Miller	The evolutionary history of the coral genus <i>Acropora</i> (Scleractinia, Cnidaria) based on a mitochondrial and a nuclear marker: reticulation, incomplete lineage sorting, or morphological convergence?	87	24.7
Mol. Ecol. 11 (8), 1339-1349 (2002)	L. M. Márquez, M. J. H. Van Oppen, B. L. Willis, A. Reyes, and D. J. Miller	The highly cross-fertile coral species, <i>Acropora hyacinthus</i> and <i>Acropora cytherea</i> , constitute statistically distinguishable lineages	34	9.7
Mol. Ecol. 24 (19), 5006-5019 (2015)	N. L. Røsser	Asynchronous spawning in sympatric populations of a hard coral reveals cryptic species and ancient genetic lineages	16	4.5
Syst. Biodivers. 8 (2), 281-288 (2010)	Z. T. Richards, C. C. Wallace, and D. J. Miller	Archetypal 'elkhorn' coral discovered in the Pacific Ocean	10	2.8
Evol. Ecol. (2011) In press	S. R. Palumbi, S. Vollmer, S. Romano, T. Oliver, and J. Ladner	The role of genes in understanding the evolutionary ecology of reef building corals	3	0.9
Galaxea, J. Coral Reef Stud. 10 (2), 91 (2008)	M. Hatta and K. Matsushima	Presumed natural hybrids between <i>Acropora donei</i> and <i>Acropora tenuis</i>	3	0.9
Mol. Ecol. 9 (9), 1363-1373 (2000)	M. J. H. Van Oppen, B. L. Willis, H. W. J. A. Van Vugt, and D. J. Miller	Examination of species boundaries in the <i>Acropora cervicornis</i> group (Scleractinia, cnidaria) using nuclear DNA sequence analyses	3	0.9
Unpublished	Wei, N.V., Tang, P.-C. and Chen, A.C.	Hybridization does not occur among common <i>Acropora</i> species in a marginal coral assemblage	29	8.2
Unpublished	Ohki, S., Kowalski, R.K., Kitanobou, S. and Morita, M.	Different timing of the spawning is related to reproductive isolation and gamete species recognition in the broadcast spawning coral <i>Acropora</i>	27	7.7
Unpublished	van Oppen, M.J.H. and Miller, D.J.	A new single-locus nuclear DNA marker for evolutionary studies in the scleractinian coral genus <i>Acropora</i>	2	0.6