

1 **Commercial fishery disturbance of the global open-ocean carbon sink**

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27 **Primary production in the global oceans fuels multiple ecosystem services including**
28 **fisheries, and the open-ocean biological carbon sink, which support food security and**
29 **livelihoods¹, and the regulation of atmospheric CO₂ levels² respectively. The spatial**
30 **distributions of these two services are driven by primary production and it is likely that**
31 **ecosystem disturbance from fishing impacts both the carbon sink and atmospheric CO₂.**
32 **Yet the extent of these impacts from past, present and future fishing is unknown. Here**
33 **we show that 23% of global export and 40% of fishing effort are concentrated in zones**
34 **of intensive overlap representing 7% of the global ocean area. This overlap is**
35 **particularly evident in the Northeast Atlantic and Northwest Pacific. Small pelagic fish**
36 **dominate catches in these regions and globally, and their exploitation will reduce faecal**
37 **pellet carbon sinks and may cause tropic cascades affecting plankton communities**
38 **important in sinking carbon. There is an urgent need to address how fisheries affect**
39 **carbon cycling, and for policy objectives to include protecting the carbon sink,**
40 **particularly in areas where fishing intensity and carbon export and storage are high.**

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42 The open-ocean carbon sink and store via the biological pump^{2,3}, hereafter ‘carbon sink’, is
43 an important regulator of atmospheric CO₂ levels, which would otherwise be 50 % higher⁴.
44 Estimates of organic carbon exported out of the top 100 m of the global ocean range from 4 –
45 12 Gt C yr⁻¹ ^{5,6}. Exported carbon sinks down to the deep ocean (> 1000 m) where ~ 1 % is
46 locked away on timescales from decades to millennia, with the rest being recycled and
47 eventually converted back to CO₂ by microbes and zooplankton³. This 1 % of carbon export
48 equates to deep ocean carbon sequestration of up to 0.12 Gt C yr⁻¹ , which is on a par with
49 coastal blue carbon sequestration (0.11 Gt C yr⁻¹ from mangroves, salt marshes and
50 seagrass⁷), or 1.1 % of anthropogenic carbon release (10 Gt C yr⁻¹)⁸. The open-ocean carbon
51 sink is predominantly driven by phyto- and zooplankton at the base of ocean food-webs³. The
52 faecal pellets of current and potential fishery species, including anchovy⁹, krill¹⁰ and
53 mesopelagic fish¹¹, are particularly important in sinking. Any marine ecosystem change
54 resulting in changes in abundance or community composition of species responsible for
55 sinking and storing carbon could result in a positive feedback increasing atmospheric CO₂
56 levels¹².

57 Marine fishing currently removes ~ 0.10 Gt yr⁻¹ of biomass¹³ and has profoundly altered
58 ecosystems throughout the global ocean. These impacts can propagate through foodwebs in
59 trophic cascades which produce sequential changes in the abundance of successive trophic

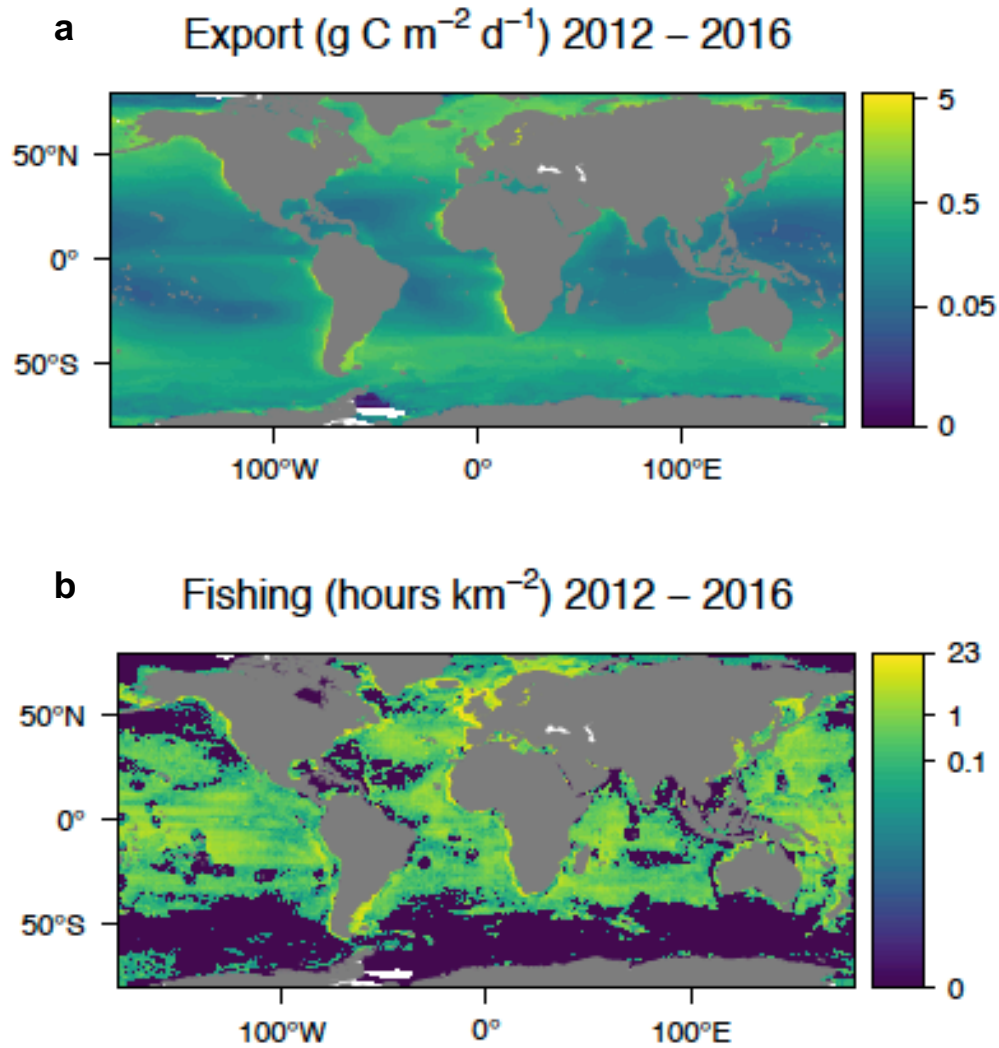
60 levels¹⁴. Fishing also affects the physical habitat, such as through the removal of oyster
61 beds¹⁵. These ecological alterations can affect the lower trophic levels responsible for the
62 majority of carbon fixation and export, and those that contribute to deeper faecal carbon
63 sinks. The reliance of both fish biomass and the carbon sink on phytoplankton^{1,2} creates the
64 potential for significant spatial overlap between the two and for the fishing to disturb the
65 carbon sink. Although there is some acknowledgement of potential interactions between the
66 two¹⁶, the impact of past and current fishing on the carbon sink and atmospheric CO₂ has not
67 been investigated, nor is fishery disturbance factored in to forecasts of future changes to the
68 global carbon cycle¹⁷.

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70 The main reason for the lack of attention to this topic is likely a discipline divide between
71 biogeochemistry and marine ecology. This divide is reflected in models; the biogeochemical
72 modules of the Earth System Models (ESMs) which inform Intergovernmental Panel on
73 Climate Change (IPCC) assessment reports do not include trophic levels above
74 zooplankton¹⁸. While ecological modellers are working to better link ESMs and models of
75 fished species¹⁹, the primary motivation is to investigate the bottom-up impacts of climate
76 change on these species²⁰, rather than top-down controls on the global carbon sink.

77

78 The current study uses global scale satellite data to assess the spatial overlap between
79 commercial fishing effort²¹ and the carbon sink (specifically particulate carbon export at 100
80 m depth)⁶, thereby mapping the risk of impact. We analyse these data at two scales, namely a
81 1° x 1° grid and the nineteen major fishing areas (hereafter ‘fishing area’) used by the UN
82 Food and Agricultural Organisation (FAO) for recording catch statistics. We also identify the
83 routes by which different fishing practices might impact the carbon sink.

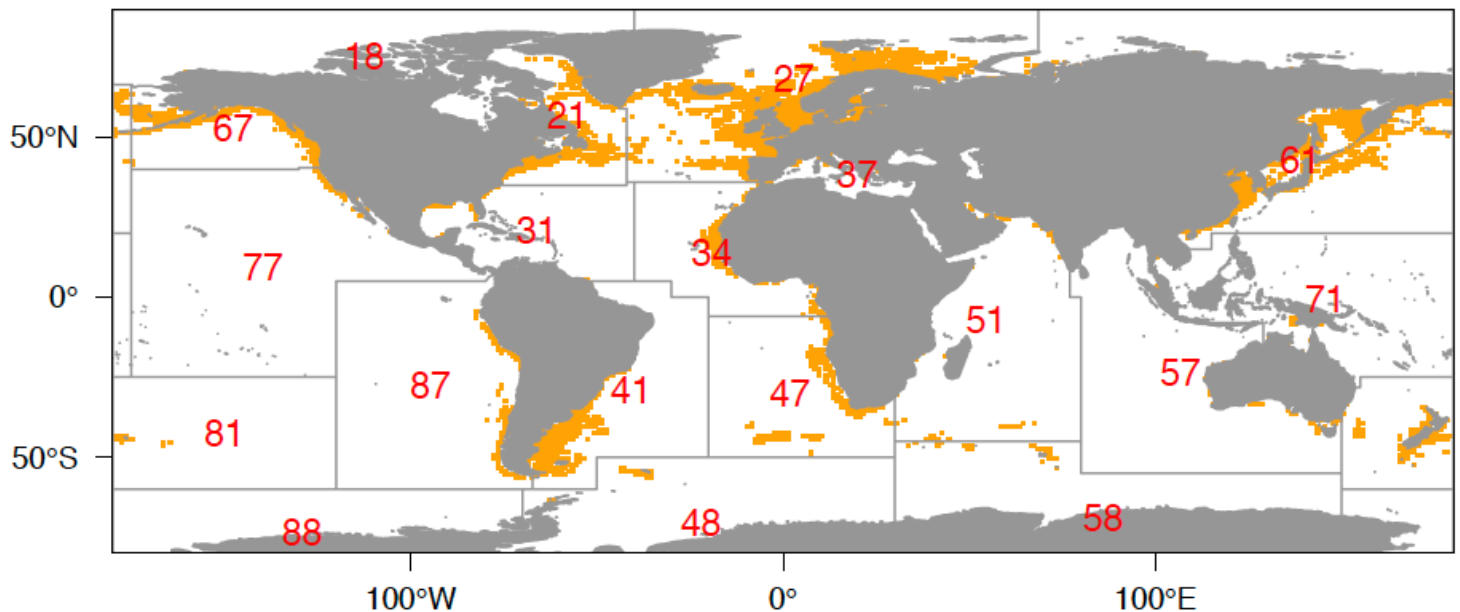


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85 **Fig. 1. Global annual carbon sink (export) and fishing intensity.** a) Average annual
86 particulate organic carbon export (g C m⁻² d⁻¹) from 100 m depth estimated using satellite
87 primary production and sea surface temperature according to the algorithm in Henson et al⁶.
88 b) Average annual commercial (vessels 6 – 146 m in length) fishing intensity (hours fished
89 km⁻²), data downloaded from Global Fishing Watch²¹. Both datasets are averaged over a 5-
90 year period from 2012 – 2016, note the log z-scale.

91
92 **Regions of high carbon sink and fishing**

93
94 Both carbon export and fishing intensity are highest around coastlines (Fig. 1), which is
95 reflected in the map showing areas of combined high carbon export and high fishing intensity
96 (Fig. 2). Both ecosystem services are concentrated in coastal regions where primary
97 production is highest²². The spatial overlap (orange pixels in Fig. 2) represents 7% of the

98 global oceans by area, but 23% of carbon export and 40 % of fishing effort globally. The two
99 highest ranking areas, for both carbon export and fishing intensity, are the Northeast Atlantic
100 (fishing area 27, Fig. 2) and the Northwest Pacific (fishing area 61). These areas are
101 respectively responsible for 14% and 9% (0.46 and 0.32 Gt C yr⁻¹) of global carbon export
102 and 15% and 14% (33.26×10^6 and 29.99×10^6 hours yr⁻¹) of global fishing effort (Fig. 3,
103 Supplementary Table 1).
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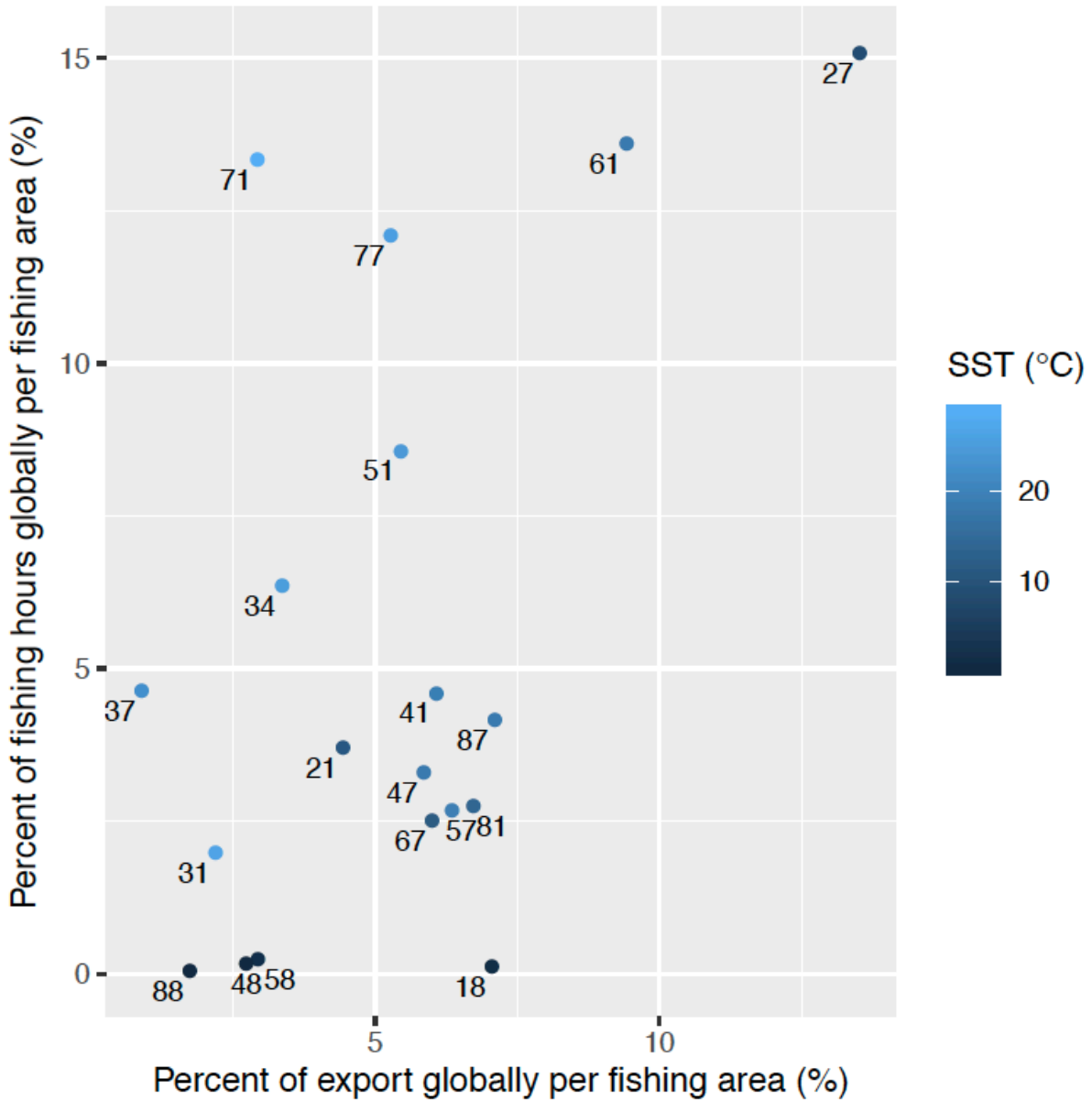


105 **Fig. 2. Regions of high fishing and carbon export intensity.** 1° by 1° grid cells where
106 carbon export (Fig. 1a) and fishing hours (Fig. 1b) values are in the upper quartile of both
107 data sets, emphasising the importance of coastal regions at higher latitudes, particularly the
108 Northeast Atlantic (fishing area 27) and Northwest Pacific (fishing area 61) (see Fig. 3 and
109 Supplementary Table 1). Grey grid lines and red numbers indicate the FAO major fishing
110 areas.

111

112 Fishing intensity increases with total carbon export at the fishing area scale (Fig. 3). The
113 Arctic fishing area (18) does not follow this pattern as it has high carbon export but relatively
114 little fishing effort due to seasonal ice cover, although melting sea ice may change this in the
115 future²³. Subtropical fishing areas (Central West Pacific, 71, Central East Pacific, 77, and
116 West Indian, 51) have high total fishing intensity (13 %, 12 % and 9 % of global total
117 respectively), but fairly low carbon export (≤ 5 %) (Fig. 3). Fishing areas which contain
118 coastal upwelling regions (e.g. Southeast Pacific, 87, and Southeast Atlantic, 47) make
119 relatively low contributions to global fishing and export (Fig. 3) because they are dominated

120 by low productivity oligotrophic gyres (Fig. 1a). The high localised primary production, and
121 thus carbon export and fishing, in coastal upwellings is nonetheless highlighted by our upper
122 quartile analysis of both data sets (Fig. 2).
123



124

125 **Fig. 3. Relationship between carbon export and fishing intensity across fishing areas.**

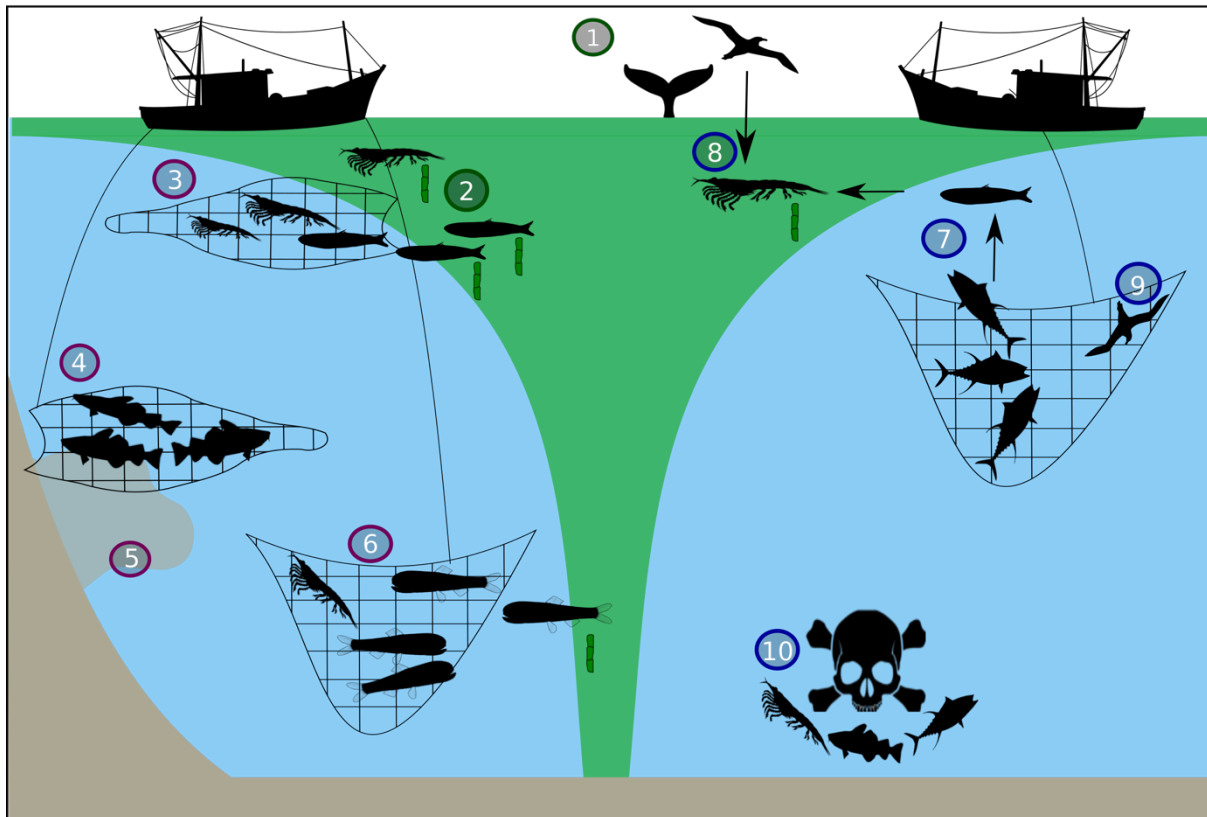
126 *Percent of global particulate organic carbon export and fishing intensity in each fishing area*
127 *averaged over 2012 - 2016. Colour of points present the mean sea surface temperature (SST)*
128 *of each fishing area (Supplementary Table 1) and the labels refer to fishing area number.*

129 *Fishing areas 27 (Northeast Atlantic) and 61 (Northwest Pacific) have highest carbon export*
130 *and fishing intensity. Fishing area 18 is the Arctic where fishing is minimal, but the export is*
131 *relatively high due to high primary production and low temperatures.*

132 **Impacts of fishing on the carbon sink**

133 From our analysis of FAO catch data, we identified small and medium (< 60 cm length,
134 hereafter small) pelagic fish as the dominant fished group globally, with trawls the dominant
135 gear type. In the Northeast Atlantic where fishing intensity and carbon export are highest,
136 Atlantic mackerel and Atlantic herring dominate the catch, and in the Northwest Pacific
137 Japanese anchovy is the main fished small pelagic. Fishing small pelagics can have both
138 direct and indirect impacts on the carbon sink. These fish contribute to the carbon sink
139 through releasing carbon-rich and fast sinking faecal pellets that can sink at $> 700 \text{ m d}^{-1}$ ⁹
140 (Fig. 4). For example Peruvian anchoveta may be responsible for around 7 % of local carbon
141 export²⁴. Reducing the biomass of these species will reduce the carbon faecal pellet sink,
142 which is one of the most important routes to sink organic carbon²⁵. Whether the removal of
143 small pelagics indirectly impacts the lower trophic levels through trophic cascades remains
144 uncertain. Cod fishing in the Baltic Sea increased small pelagic (sprat) biomass, which led to
145 a reduction in its zooplankton prey as part of a more extensive trophic cascade²⁶ (Fig. 4).
146 However, specific evidence of the existence or extent of indirect impacts caused by trophic
147 cascades is lacking for major fished species, including for Atlantic herring, mackerel and
148 Japanese anchovy
149
150 Groundfish such as Atlantic cod and Alaska pollock (caught in fishing areas 27 and 61
151 respectively) are the next most important catch category after small pelagics (Supplementary
152 Table 1), but their contribution to the carbon sink is also currently unknown. Groundfish
153 fisheries could have the greatest impacts on the carbon sink through trophic cascades as
154 described above in the Baltic Sea²⁶ and physical disturbance of the seabed^{27,28} (Fig. 4). The
155 demersal trawls used in these fisheries create plumes of resuspended material that can remove
156 seabed carbon at a rate that counteracts sinking carbon²⁸. As groundfish reside near the
157 seabed, the pellets they egest would be subjected to less water column degradation prior to
158 sedimentation of the carbon. Similarly, mesopelagic fish that live permanently or migrate
159 daily into this depth realm can increase the sink of carbon to the deep sea and seabed¹¹; any
160 carbon they release below the permanent thermocline (winter mixed layer depth) will not be
161 subject to water column mixing and remain sequestered for decades or centuries¹⁰. Thus
162 targeted or incidental harvesting of mesopelagic species is likely to increase the rate at which
163 CO₂ returns to the atmosphere (Fig. 4). Other mechanisms by which fishing for any species
164 could impact the carbon sink include the harvesting or by-catch of fertilising species such as

165 krill²⁹, whales³⁰ or seabirds³¹, and the release of discards causing localized dead zones (see
166 Supplementary Information) or re-routing carbon through different trophic cycles e.g.
167 through scavenging seabirds³² (Fig. 4).



168

169 **Fig. 4. Direct and indirect impacts of fishing to the carbon sink.** Phytoplankton (green
170 shading in the surface) stimulate fish biomass production and the export of carbon out of the
171 upper ocean, of which ~ 1 % sinks to the deep ocean. The carbon sink is enhanced by (1)
172 fertilising species and (2) those egesting fast-sinking carbon-rich faecal pellets. Direct
173 impacts of fishing include (3) harvesting low-mid trophic level pellet-producing species, (4)
174 removing species living near the seabed where the sink of carbon will be short, and (5)
175 harvesting groundfish disturbing the sediment resuspending carbon which could be
176 remineralised in the water to CO₂, and finally (6) removing resident or migratory
177 mesopelagic species that contribute to the carbon sink. Indirect impacts include (7) causing
178 trophic cascades when removing high trophic level species impacting low trophic level
179 communities that sink carbon, (8) removing prey items for fertilizing species (e.g. mackerel
180 or krill that feed seabirds), (9) killing predators (e.g. seabirds) that may otherwise fertilise
181 the oceans but also help to maintain a balanced food web, and finally (10) the release
182 discards which could cause localized dead zones.

183

184 **Climate change, fishing and the carbon sink**

185 Global carbon export is projected to decline by the end of the century³³, as a result of changes
186 to plankton abundance and composition, and reduced primary production³⁴. There are no
187 forecasts of how climate change impacts to higher trophic levels will affect the future carbon
188 sink. Fishing may further exacerbate the projected climate-driven declines in carbon export,
189 and thus the store of carbon in the deep ocean, by changing the community composition of
190 low trophic levels important in carbon export. For instance 30 years of warming in the Baltic
191 Sea changed the dominant copepod species from the larger *Pseudocalanus acuspes* to the
192 smaller *Acartia spp*, with overfishing of cod amplifying this regime shift³⁵. Climate change
193 will also likely alter the spatial overlap of fishing and carbon export (Fig. 2). Climate-induced
194 spatial shifts have already been observed in fish, including poleward shifts as sea
195 temperatures rise³⁶. As for the carbon sink, projections suggest an expansion of oligotrophic
196 regions where carbon export is currently low (Fig. 1a)³⁷, and increases in carbon export
197 toward the poles. Poleward shifts in both fishing intensity and the carbon sink would result in
198 smaller, more concentrated areas of overlap than today (Fig. 2), with an increased risk of
199 impact.

200 **Conclusions**

201 There is clear spatial overlap between the carbon sink and commercial fishing. Biomass and
202 ecosystem changes caused by fishing could negatively impact carbon sinking and storage
203 throughout the water column and seabed, and therefore atmospheric CO₂ levels. There is an
204 urgent need to clarify through observations and modelling whether and how fisheries reduce
205 the carbon sink, and for policy objectives to include protecting this ecosystem service. These
206 needs are particularly important in the regions where the risk of fishing impacting the carbon
207 sink is high (Northeast Atlantic and Northwest Pacific). Research is also required into the
208 potentially synergistic impacts of fishing on the carbon sink, and climate change on both
209 fishing and the carbon sink. The rebuilding of impacted ecosystems and stocks would help to
210 reverse impacts on the carbon sink. This rebuilding is already an established fisheries
211 management and sustainable development objective^{27,38} but progress towards this goal is
212 extremely limited and up to 63% of monitored stocks remain in need of rebuilding^{27,39}.
213 Recognising that the carbon sink is an additional ecosystem service that requires protection

214 strengthens the case for a holistic approach to managing the oceans^{27,40} and might help to
215 achieve a wider suite of environmental goals. We hope improved understanding of how
216 commercial fisheries disturb the carbon sink will be a step toward realising a sustainable
217 balance of the twin needs for productive fisheries to maintain global food security and strong
218 carbon sinks which play a critical role in climate regulation.

219

220 **Methods**

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222 Our indicator of carbon sink intensity (export) is the critical first step in the carbon sink while
223 our indicator of fishing (effort) is correlated with the main potential route of impact, i.e.
224 biomass removal (see Supplementary Information). Sea surface temperature was downloaded
225 from the NASA ocean colour database (<https://oceancolor.gsfc.nasa.gov>) and primary
226 productivity data from the Ocean Productivity site⁴¹ for the same time frame as availability of
227 fishing data (2012 – 2016), to calculate particulate organic carbon export ($\text{g C m}^{-2} \text{d}^{-1}$) sink of
228 carbon out of the top 100 m of the ocean) using the Henson et al.⁶ algorithm. Carbon sinks
229 through the entire ocean depths, but only is stored and sequestered on long timescales if it
230 reaches the deep sea ($> 1000 \text{ m}$). However, there is not yet a consensus on how to
231 parameterise the transfer efficiency of carbon to the deep due to the many processes which
232 control it, whereas there is a consensus that carbon export out of the upper 100 m is
233 negatively related to temperature^{6,42}. Hence, we use carbon export here as our metric from the
234 global carbon sink. We use data on global fishing intensity (hours fished km^{-2}) taken from all
235 vessels with an automatic identification system (AIS) and published online by the Global
236 Fishing Watch²¹. Only data for the years 2012 – 2016 inclusive have been released so we
237 present the mean annual fishing intensity over this 5-year period. We merged global fishing
238 intensity and export data onto a 1×1 degree resolution grid and identified the areas where
239 both fishing and export were in the top quartile of their respective datasets globally (orange
240 pixels in Fig. 2).

241 We assessed the total export, fishing intensity and dominant fishing method (gear type) for
242 each of the FAO major fishing areas. We obtained gear type data primarily from Tanocet et
243 al.⁴³, which provides total Global Fisheries Landings database⁴⁴ effort by gear type for 2010
244 to 2014. We obtained catch data for the equivalent period from the FAO Global Capture
245 Production database¹³ (Supplementary Table 2). This period overlaps our export and fishing

246 intensity data (Fig. 1a and b) for three years, 2012 – 2014, and fishery catch and effort data
247 are well correlated (Supplementary Fig. 1). We identified those taxa which dominate the
248 catch in each fishing area (i.e. the top ranking taxa in terms of catch weight, which constitute
249 50% or the closest value above 50% of the overall catch) (Supplementary Table 1). We
250 assigned each taxon to one of the following categories: small pelagic fish (SP); groundfish
251 (G); large pelagic fish (LP), deep water fish (DF); unspecified fish (UF), pelagic crustaceans
252 (PC); benthic crustaceans (BC), unspecified crustaceans (UC); squid (S); Unspecified
253 molluscs (UM); and finally bivalves (B). See Supplementary Table 2 for more detail on this
254 classification.

255

256 Gear type data were not available for fishing areas in the Southern Ocean (fishing areas 88,
257 48, 58), nor the Northeast Atlantic (fishing area 27). For the Southern Ocean we were able to
258 characterise our catch data by gear type, providing data that is comparable to the majority of
259 other fishing areas. The dominant Southern Ocean fisheries use either longlines to target
260 toothfish (fishing area 58 & 88) or trawls to target Antarctic krill and mackerel icefish
261 (fishing area 48, Supplementary Table 1). In the case of the Northeast Atlantic, gear type data
262 is presented in terms of percentage of fishing hours rather than percentage of catch⁴³. Our
263 Supplementary Table 1 presents these data, which suggest that trawls are the main fishing
264 gear in the Northeast Atlantic, comprising more than 70% of fishing hours. It is therefore
265 plausible that trawls are also the main fishing gear by catch, although the two metrics are not
266 strictly comparable.

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377

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382

383 **Author Contributions**

384 E.L.C conceived the study and analysed the carbon export and fishing intensity data. E.L.C
385 made the figures. S.H. analysed the catch and gear type data. Both authors contributed
386 equally to the development of ideas and the writing and editing of this manuscript.

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388 **Competing interests**

389 The authors claim no competing interests.

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391 **Additional Information**

392 Supplementary Information is available for this paper. Correspondence and requests for
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