Boom-bust population dynamics can increase diversity in evolving competitive communities

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November 29, 2020

Running title: Boom-bust dynamics increases ecosystem diversity

Keywords: Competitive exclusion, limiting similarity, adaptive diversification, population fluctuations, ecosystem diversity, niche packing, temporal niche segregation, non-equilibrium population dynamics, overcompensation

The processes and mechanisms underlying the origin and maintenance of biological diversity have long been of central importance in ecology and evolution. The competitive exclusion principle states that the number of coexisting species is limited by the number of resources, or by the species' similarity in resource use. Natural systems such as the extreme diversity of unicellular life in the oceans provide counter examples. It is known that mathematical models incorporating population fluctuations can lead to violations of the exclusion principle. Here we use simple eco-evolutionary models to show that a certain type of population dynamics, boom-bust dynamics, can allow for the evolution of much larger amounts of diversity than would be expected with stable equilibrium dynamics. Boom-bust dynamics are characterized by long periods of almost exponential growth (boom) and a subsequent population crash due to competition (bust). When such ecological dynamics are incorporated into an evolutionary model that allows for adaptive diversification in continuous phenotype spaces, desynchronization of the boom-bust cycles of coexisting species can lead to the maintenance of high levels of diversity.

The amazing diversity of life has sustained the debate about the origins and limits of biodiversity. While random, selectively neutral processes are thought by some to play an important role, e.g. in ecosystem dynamics¹ and in molecular evolution, it seems that a majority of researchers would agree that non-neutral ecological interactions – competition, predation, mutualism – are central to understanding diversity, with competition having received the most attention. Coexistence between competing species requires that intraspecific competition is strong enough relative to interspecific competition. This is captured by the concept of limiting similarity: to coexist, populations must be sufficiently different in their resource use. If populations use the same resource in the same way, they cannot coexist, a phenomenon known as the competitive exclusion principle².

The exclusion principle has faced challenges from many empirical counter examples, in which the number of coexisting and ecologically interacting species was significantly higher than the number of limiting resources. The best known such example is the Paradox of the Plankton³, which is based on a comparison between the relatively small number of biochemical resources essential for plankton growth, and the number of known coexisting plankton species, which is orders of magnitude larger. Different theoretical explanations for conditions that circumvent the exclusion principle have been proposed, and it is known that fluctuating population sizes can lead to violations^{2,4-6}. With fluctuating population sizes, the storage effect^{7,8}, as well as relative non-linearities² can lead to coexistence of more competitor species than essential resource species, e.g. because of cyclical dominance between competitors⁹. Most of these examples involve models with a finite number of distinct resources, but coexistence due to fluctuating population dynamics has also been shown in models with continuous niches. With continuous, externally imposed (seasonal) periodic cycles in population sizes, time essentially becomes an additional niche dimension along which populations

can segregate and coexist¹⁰. In an evolutionary context, it has been shown that limiting similarity in a continuous niche space used by an evolving community whose member species are undergoing externally forced population fluctuations, larger amplitude fluctuations lead to more diversity, and hence effectively to smaller limiting similarity¹¹.

Most previous models used in this context have assumed that the population fluctuations are externally imposed, and that there is a finite number of distinct resources. Here we investigate the questions of evolving diversity and limiting similarity in a setting where population fluctuations are not externally imposed, but are instead due to competitive interactions within and between the evolving species. In addition, we address the question of diversity in continuous phenotype spaces, corresponding to continuously varying resource use. Rather than the 1-dimensional phenotype spaces that are usually assumed with continuous resource distributions^{11–13}, our phenotype spaces are potentially high-dimensional.

We use the models of^{14–18}, which are extensions of classical competition models to high-dimensional continuous phenotype spaces. In previous work, we assumed that the underlying ecological dynamics have a stable equilibrium (the carrying capacity). However, by using difference equations rather than differential equations to describe ecological dynamics, it is straightforward to extend these models to allow for complicated ecological dynamics, which are by definition endogenously generated (i.e., the population fluctuations are a result of the competitive interactions). We show that for certain types of endogenously generated fluctuations, which we term “boom-bust” dynamics, the amount of diversity that evolves and is maintained at evolutionary steady state can be much larger than the diversity maintained without ecological fluctuations.

Boom-bust dynamics consist of long periods of (near-) exponential growth followed by a deep crash, in such a way that the boom-bust cycles of different species become sponta-

neously desynchronized. The amount of excess diversity enabled by boom-bust dynamics increases with the dimension of phenotype space, so that species can be much more tightly packed in high-dimensional spaces, corresponding to a much smaller limiting similarity necessary for coexistence in high-dimensional niche spaces.

Apart from asynchronous boom-bust dynamics, essentially all other types of complex fluctuations, including asynchronous chaotic dynamics not exhibiting the boom-bust fluctuations, do not increase the diversity at evolutionary steady. Our models thus provide a specific and robust mechanism for the evolutionary origin and maintenance of highly diverse competitive communities.

Model and simulation methods

To accommodate various types of ecological dynamics, we consider ecological models given by difference equations, and hence set in discrete time. The basic ecological model we use is a difference equation^{19–21} that links population densities N of two consecutive generations t and $t + 1$,

$$N(t + 1) = F(N(t)) = N(t) \frac{\lambda}{1 + aN(t)^\beta}. \quad (1)$$

where $\lambda > 0$ is the per capita number of offspring, and $a > 0$ and $\beta > 0$ are parameters describing the effect of competition. For $\lambda < 1$, $N(t)$ converges to 0 for any initial condition $N(0) > 0$, and hence extinction is the only possible outcome. We therefore assume $\lambda > 1$ in what follows. In that case, model (1) has a non-zero equilibrium at $K = ((\lambda - 1)/a)^{1/\beta}$,

which is the carrying capacity of the population. It is then convenient to write (1) as:

$$N(t+1) = N(t) \frac{\lambda}{1 + (\lambda - 1)(N(t)/K)^\beta}, \quad (2)$$

as this makes it easy to formulate the model in terms of continuous phenotypes (see below). Model (2) was shown to fit well a wide range of data²⁰, and for $\beta = 1$ can be derived from the logistic differential equation by integration over a finite time interval²².

It is well known that the basic quantity underlying the dynamic behaviour of this model is the derivative dF/dN evaluated at the equilibrium K :

$$\left. \frac{dF}{dN} \right|_{N=K} = 1 - \frac{\lambda - 1}{\lambda} \beta. \quad (3)$$

For $\lambda > 1$, the population dynamics converges to the steady state K if and only if $|1 - \frac{\lambda-1}{\lambda} \beta| < 1$. Thus, for a given λ , for small values of the exponent β the system exhibits stable equilibrium dynamics, and increasing β gives way to a period-doubling route to chaos. Biologically, increasing β can be viewed as reflecting a gradual change from contest to scramble competition²³. This is reflected by the shape of the per capita number of offspring as a function of population size, given by the right hand side of (2) divided by $N(t)$, and viewed as a function of $N(t)$: for any λ , and for high β , the per capita number of offspring is approximately constant until the population size reaches the vicinity of K , but as N increases above K the number of offspring falls rapidly to very low values, essentially generating a population crash as soon as the population size is above K .

Because the difference equation (2) has three rather than two parameters, it can exhibit

certain dynamical properties that better known difference equations, such as the Ricker equation^{20,24}, do not have. For example, for small values of λ , model (2) can exhibit highly chaotic dynamics (as e.g. measured by the Lyapunov exponent) despite the fluctuations in population size being (arbitrarily) small (Figure 1). Importantly, for small values of λ , and for large enough β , model (2) exhibits "boom-bust" dynamics (Figure 1), in which long periods of near-exponential growth (due to small λ) are followed by deep crashes (due to high β) once the population size is above K for the first time after the exponential phase. This cycle repeats itself qualitatively, but the dynamics is in fact chaotic and exhibits sensitive dependence on initial conditions, because the population size after the crash, and hence the length of the subsequent exponential phase, is different in each cycle. We note that such boom-bust dynamics cannot be observed in the Ricker model. Some of the possible dynamic regimes of model (2) and transitions between them are shown in Figure 1.

We note that there are in principle many different models that can exhibit boom-bust dynamics (including models set in continuous time, see Discussion section). We chose model (2) as a generic model with boom-bust dynamics for certain parameter regions, viz. for λ -values close to 1 and large enough β . Rather than being interested in the likelihood of a particular model exhibiting boom-bust dynamics, we are interested in the consequences of such dynamics for the evolution of diversity. Therefore, while pointing out the contrast to the consequences of other types of ecological dynamics, such as cyclic or "regular" chaotic dynamics, delineating the different regions in parameter space generating the different types of dynamics is not relevant for our purposes.

We now consider a generalization of Eq. (2) that includes competition between S phenotypically monomorphic populations. Each population is characterized by its phenotype $x_s = (x_s^1, \dots, x_s^d) \in \mathbb{R}^d$, $s = 1, \dots, S$, where d is the dimension of phenotype space (which

Figure 1: A): Examples of population dynamics; convergence to steady state for $\beta = 2, \lambda = 1.2$ (red line), stable periodic oscillations for $\beta = 16, \lambda = 1.2$ (blue line), chaotic dynamics when $\beta = 20, \lambda = 1.2$ (magenta line) and $\beta = 5, \lambda = 5$ (green line), and boom-bust chaotic dynamics when $\beta = 45, \lambda = 1.2$ (black line with individual generations shown by circles). Note the almost exponential multi-generation growth phases in the boom-bust regime, followed by a single-generation bust. B): bifurcation diagram for Eq. (2) for $\lambda = 1.2$. Both panels computed with $K = 1$.

is assumed to be Euclidean d -space). The population size of phenotype x_s is denoted by N_s . The ecological dynamics of all S clusters are determined by the competition kernel $\alpha(x_s, x_r)$, which measures the competitive impact of phenotype x_r on phenotype x_s , and the carrying capacity $K(x_s)$, which is the equilibrium population size of phenotype x_s in the absence of any other phenotypes. (The competition kernel and the carrying capacity are functions $\alpha : \mathbb{R}^d \times \mathbb{R}^d \rightarrow \mathbb{R}$ and $K : \mathbb{R}^d \rightarrow \mathbb{R}$, respectively.)

The discrete time dynamics of each phenotype in the competitive community is then given by

$$N_s(t+1) = N_s(t) \frac{\lambda}{1 + (\lambda - 1) \left[\sum_{p=1}^S N_p(t) \alpha(x_s, x_p) / K(x_s) \right]^\beta}, \quad (4)$$

$s = 1, \dots, S$. The sum in the denominator on the right hand side of (4) is the effective population size experienced by phenotype x_s . Eq. (4) is a discrete time analog of the continuous-time many-species logistic competition model in multidimensional phenotype space that was used in several previous articles^{16–18}. In contrast to the continuous time models used previously, in the discrete time model populations can undergo ecological fluctuations and sudden collapses not only after a population itself exceeds the carrying capacity, but also when the cumulative competition from other phenotypes is strong enough (i.e, when the effective population size is above K).

For simplicity, and following^{15–17}, we used the following functions for the competition kernel and the carrying capacity,

$$\alpha(x, y) = \exp \left[- \sum_{i=1}^d \frac{(x^i - y^i)^2}{2\sigma_\alpha^2} \right], \quad K(x) = K_0 \exp \left[- \frac{\sum_{i=1}^d (x^i)^4}{4\sigma_K^4} \right]. \quad (5)$$

Thus, competition is symmetric and strongest between phenotypes that are similar, and the carrying capacity has a unique maximum K_0 at 0. We set the scaling parameters K_0 and σ_K to $K_0 = 1$ and $\sigma_K = 1$. (We note that in general, using Gaussian forms for both the competition kernel and the carrying capacity can result in structural instabilities²⁵).

To simulate the evolutionary process, we start with a number S of phenotypes (typically, $S = 1$, and the phenotype is randomly chosen in the vicinity of the maximum of the carrying capacity). We then simulate the ecological dynamics in discrete time, using (4) for each of the phenotypes. In each generation, a new phenotype is generated with a probability μ (typically $\mu = 0.1$). The new phenotype is a mutant of one of the existing phenotypes. Of those, a parental phenotype is chosen with a probability proportional to its population

size, and the offspring phenotype is chosen randomly from a Gaussian distribution with the average centered at the parental phenotype and a small standard deviation Δx (typically $\Delta x = 10^{-2}$).

After addition of the new phenotype, the community now comprises $S + 1$ phenotypes, and the process is repeated for many generations. What one wants to know from this process is how the distribution of phenotypes changes over time. To keep the number of phenotypes from increasing to very large numbers that would render the simulation computationally impossible, we periodically merge phenotypes that are very close together. Specifically, once every t_{merge} generations (typically $t_{merge} = 1000$), phenotypes that are within a distance Δx of each other are merged (preserving their phenotypic center of mass) and their population sizes added. In addition, every generation all clusters with populations densities below a threshold (typically $= 10^{-12}$) are declared extinct and removed from the system. Together, these procedures preserve the phenotypic variance necessary for evolution, but prevent undesirable computational complexity.

To define and count the number of phenotypically distinct species in the community at any given point in time, with each species possibly consisting of a number of similar phenotypes, the phenotypes in the community are clustered with a larger distance $\Delta x_{species}$ (typically $\Delta x_{species} = 10^{-1}$). This phenotypic distance is still significantly smaller than the typical scales of ecological interactions as long as σ_α and σ_K in (5) are of order 1. This implies that the phenotypes within a designated species experience very similar competitive interactions and generally follow the same population dynamics. Note that species designation is only used to gather statistical data from simulations, but not in the actual computational steps of the simulations.

Results

For $\sigma_\alpha < \sigma_K$ in (5), the continuous time analog of the model presented above undergoes adaptive diversification and radiation into a steady state species distribution^{13,17,26}, see also^{27,28}. For example, if we set $\sigma_\alpha = 0.5$ and $\sigma_K = 1$ in (5), then for $\beta = 1$ system (4) is equivalent to the corresponding continuous time system²², and in a 2-dimensional phenotype space undergoes diversification into a stable community of 16 coexisting phenotypic clusters (species) with approximately constant population sizes. This is illustrated in Figure 2a and the corresponding video. As long as $\beta = 1$, the observed diversification is independent of λ and K_0 . The main purpose of this paper is to explore the effect of increasing β to values > 1 , which eventually makes the local dynamics (2) unstable. As the exponent β is increased, stationary populations lose stability and, similarly to the single-species model shown in Fig. 1, the ecological dynamics of populations in an evolving community become first periodic and then chaotic (see below for specific examples of non-equilibrium dynamics).

For low intrinsic growth rates λ this has profound effects on the amount of diversity in the system, as illustrated in Figure 2 (and accompanying videos). For such λ -values, increasing β in the local dynamics (2) has the effect of eventually inducing pronounced boom-bust population dynamics (cf. Figure 1). In an evolving community, increasing β induces boom-bust dynamics in each of the phenotypes present in the community, with each phenotypic cluster (species) undergoing multiple generations of (near-)exponential population growth punctuated by deep crashes. These ecological dynamics unfold in such a way that the dynamics of neighbouring clusters of phenotypes are desynchronized, i.e., such that crashes and subsequent exponential growth phases occur at different time points. With such desynchronized boom-bust ecological dynamics, evolution can generate a drastic increase in diversity com-

Figure 2: Snapshots of cluster distributions for low (left column) and high (right column) β -values after 10^7 generations, for 2-dimensional (top row) and 3-dimensional (bottom row) phenotype spaces. A): 16 species for $\beta = 1, \lambda = 1.2, \sigma_\alpha = 0.5$; B): 47 species for $\beta = 45, \lambda = 1.2, \sigma_\alpha = 0.5$; C): 8 species for $\beta = 1, \lambda = 1.2, \sigma_\alpha = 0.75$; D) 20 species for $\beta = 55, \lambda = 1.2, \sigma_\alpha = 0.75$. The corresponding videos of the evolutionary process that led to these configurations from single ancestors can be found at figshare.com/s/f2d8ecf480fa372319e1. In the top row panels and corresponding videos (2-dimensional case), all phenotypes present in the evolving community are shown (represented as dots). In the bottom row panels and corresponding videos (3-dimensional case), species resulting from clustering of populations of similar phenotypes using a merging distance of $\Delta x_{species} = 10^{-1}$ (see main text) are represented by circles whose size is proportional to a species' population size.

Figure 3: The number of coexisting species as a function of the exponent β for $\lambda = 1.2$ (circles), $\lambda = 1.6$ (squares), $\lambda = 2$ (diamonds), $\lambda = 5$ (triangles). Dimension of phenotype space is $d = 2$, and $\sigma_\alpha = 0.5$. Species are counted after 10^7 generation and after clustering of populations of similar phenotypes using a merging distance of $\Delta x_{species} = 10^{-1}$ (see main text).

pared to that evolving in ecologically stable communities (Figure 2). Increased diversity due to boom-bust ecological dynamics typically occurs as long as the ecological conditions for adaptive diversification due to frequency-dependent competition are met (i.e., as long as $\sigma_\alpha < \sigma_K$ in (5)). While the amount of diversity depends on the exact values of $\sigma_\alpha < \sigma_K$, significantly more diverse communities tend to evolve with boom-bust dynamics than with stable equilibrium ecological dynamics.

Figure 3 shows the number of species coexisting at the evolutionary saturation state as a function of the parameter β for different values of λ . The figure illustrates that λ has to be small enough for a substantial increase in diversity to be observed for high β . Indeed, in the local model (2) boom-bust dynamics can only be observed for small λ , and it is exactly these

kinds of population dynamics that allow for increased diversity. For larger λ , increasing β results in more “traditional” forms of chaotic dynamics with irregular, high frequency oscillations of increasing amplitude. Such local dynamics also lead to chaotic ecological dynamics in populations comprising an evolving community, but they do not generate an increase in the diversity that can evolve and be maintained. In general, as β is increased to very high values, the model becomes less relevant biologically: the population crashes become very severe, which results in extinctions that are frequent enough for diversity not to be able to evolve anymore (see below).

Figure 4 illustrates the desynchronized boom-bust dynamics in an artificial community of 25 species, with each species represented by a single phenotype, and such that the phenotypes are arranged on a regular grid in phenotype space (see inset in Figure 4). This community corresponds to the case indicated by the right-most square with more than 1 species in Figure 3, in which β is large enough for the diversity to increase to 25 coexisting species, rather than the 16 that would evolve for lower β . Figure 4 shows the population dynamics of a subset of 5 phenotypes arranged on a line in the grid, resulting from simulating the ecological dynamics (4) of the whole community of 25 phenotypes. For each phenotype, the dynamics exhibits boom-bust cycles, and neighbouring cycles are desynchronized.

It is worth noting that the increased diversity seen for higher β -values occupies approximately the same phenotypic range as the lower diversity for lower β -values (Figure 2). This implies that with higher diversity, the different species are more closely packed in phenotype space, and hence that, generally speaking, conditions of limiting similarity are relaxed in the boom-bust dynamic regime. Because of lower thresholds for limiting similarity, i.e., denser packing, the increase in diversity at evolutionary stationary state due to boom-bust ecological dynamics becomes more pronounced with higher dimensions of phenotype space, as illus-

Figure 4: Examples of population dynamics of 5 neighbouring phenotypes in a community consisting of the 25 phenotypes indicated by color in the inset. Parameter values are $\beta = 30$, $\lambda = 1.6$, and $\sigma_\alpha = 0.5$, which corresponds to the right-most square with more than 1 species in Figure 3. Dimension of phenotype space is $d = 2$. With equilibrium population dynamics ($\beta = 1$) the 25 phenotypes cannot coexist. Note that the population dynamics of immediate neighbours are desynchronized.

Figure 5: The number of species vs. β for $\lambda = 1.2$, $\sigma_\alpha = 0.5$ and three different dimensions of phenotype space, $d = 1$ (squares), $d = 2$ (circles), and $d = 3$ (diamonds). Species are counted after 10^7 generation and after merging the populations of similar phenotypes using a merging distance of $\Delta x_{species} = 10^{-1}$ (see description in main text). Note that the level of diversity shown for small β -values corresponds to the diversity evolving with stable equilibrium ecological dynamics.

trated in Figure 5. Relaxed limiting similarity conditions require a decrease in competitive pressure that species in neighbouring regions of phenotype space exert on each other. Such a decrease can be achieved if neighbouring populations are fluctuating in opposite phases, as shown for an artificial example in Figure 4. Figure 6 illustrates that in full simulations of the evolutionary process, neighbouring species indeed generally exhibit such an anti-correlation for high β -values. Essentially, the anti-correlation between populations of neighbouring species stems from the asynchrony of their boom-bust cycles. In the Supplementary Material, we show that such desynchronization is expected to emerge spontaneously from an arbitrary small initial difference between populations: in a simple idealized configuration of two competing species with boom-bust dynamics, their population sizes converge to a state

Figure 6: The correlation between population sizes of phenotypes x and y , $C_{xy} \equiv \langle (N_x(t) - \langle N_x \rangle)(N_y(t) - \langle N_y \rangle) \rangle$ as a function of phenotypic distance $|x - y|$ for $\beta = 50$, $\lambda = 1.2$ (black line), $\beta = 20$, $\lambda = 1.2$ (red line) and $\beta = 5$, $\lambda = 5$ (blue line). Dimension of phenotype space is $d = 2$, and $\sigma_\alpha = 0.5$. The correlation was calculated by taking into account all possible pairs of phenotypes over 5×10^6 generations after the steady state level of diversity was reached. Anti-synchronization is only seen for boom-bust dynamics (black) allowing for increased diversity (cf. Figure 3), but not for higher frequency chaotic dynamics with small (red) or large (blue) amplitudes.

of complete anti-synchronization. For smaller values of β or larger values of λ , for which populations do not undergo boom-bust dynamics this anti-correlation is not seen (Figure 6).

The explanation for higher diversity based on anti-correlated boom-bust cycles of phenotypically close species suggests that to make this mechanism work, these cycles should be of sufficient length. This means that the population crashes should be sufficiently severe (large β), and the intrinsic growth rate λ should be sufficiently small. Essentially, the exponential phase should be long enough for robust desynchronization. This effect cannot be achieved with high intrinsic growth parameters λ (Figure 3): the increase in diversity is noticeably di-

minished for $\lambda = 1.6$, and is absent for larger λ . In particular, the type of chaotic population fluctuations induced by high β for λ -values that are significantly larger than 1 do not lead to increased diversity, because for larger λ , complex dynamics are not of the boom-bust type.

To confirm that the boom-bust cycles, rather than chaoticity or other features of the population dynamics defined by Eq. (4), are the essential mechanism for the observed increase in diversity, we stripped the model (4) from all other features except its ability to run boom-bust cycles: We assumed that each phenotypic population grows exponentially with an exponent λ until the effective density experienced by a given phenotype, i.e., the cumulative competitive effect of all phenotypes, given by the term $\sum_{p=1}^S N_p(t)\alpha(x_s, x_p)$ in denominator of (4), becomes greater than the carrying capacity of that phenotype. When that happened, the population of that phenotype was reduced to a small fraction of its population size, simulating a severe crash. The mutation and merging procedures were implemented as in the original model. This modified model shows qualitatively very similar results (not shown) and exhibits significant increases in diversity at a level very similar to the original model, as long as λ -values are close to 1, so that the exponential phase starting from low densities is long and slow, and as long as the population crashes are severe enough. This confirms that the key for the evolution of higher diversity is the existence of pronounced boom-bust dynamics for all phenotypic populations.

An interesting question concerns the effect of the frequency and size of mutations on diversity. These were assumed to be $\mu = 0.1$ and $\Delta x = 10^{-2}$ for the results presented so far (note that it is really only the product of these two parameters that matters). A reduction in μ and/or Δx slows down evolution in general and diversification in particular. This effect is illustrated in Figure S.3A, where we show the number of species vs. time for 4 different mutation frequencies. Smaller mutation rates result in longer times to reach the equilibrium

level of diversity. There is, however, another, less direct effect of mutation rate on diversity. For any non-zero extinction threshold and even moderate β , there is a small but finite probability that all phenotypes of a well-developed cluster, and hence the corresponding species, go extinct during a particularly severe bust. The extinct cluster can eventually get replaced by newly arising mutants, but the time it takes mutations to undergo a sufficient number of phenotypic steps to reach the vacated spot in phenotype space depends on μ and Δx . For any given mutation rate and size, these processes may equilibrate at different levels of diversity. In particular, lower extinction thresholds (making extinction less likely) lead to higher levels of diversity at saturation. This is illustrated in Figure S.3B.

In general, diversity decreases drastically for very high β -values and eventually the system is reduced to just a single phenotypic cluster. This occurs because with large β -values, the crashes due to the effective density experienced being higher than the carrying capacity become progressively more severe and can bring all phenotypes comprising one species below the minimum population threshold, thus rendering the species extinct. Even though diversification is still favoured by selection, the rate of species extinction for high β -values is too high for diversity to evolve. The very dynamic regime of this “competition” between extinction and diversification is illustrated in Fig. S.2.

Discussion

We propose a possible explanation for the emergence and persistence of large amounts of diversity based on competition models for evolving communities with fluctuating population dynamics. When these fluctuations are in the boom-bust regime, in which long periods of exponential growth are followed by deep population crashes, diversity in continuous pheno-

type spaces evolves well beyond what is expected based on limiting similarity with stationary ecological dynamics. The key mechanism that results in higher diversity is the spontaneous de-synchronization of boom-bust cycles between phenotypically similar species, which essentially reduces interspecific competitive impacts and allows for much denser packing of species in niche space.

Population fluctuations have long been considered as a potential mechanism leading to violations of the competitive exclusion principle. For the most part, past studies have either assumed a fixed set of resources^{2,9,29}, or they have assumed externally imposed fluctuations^{10,11}. In such models, the mechanism of ecological fluctuations causing an increase in diversity can be viewed as a form of the temporal storage effect^{7,8}, which intuitively corresponds to temporal segregation in niche space¹⁰. In fact, there have also been models showing that population fluctuations can decrease diversity in an evolutionary context³⁰, but these models appear to allow for jack-of-all-trades mutations on a finite set of resources, which can increase rather than decrease the amount of interspecific competition in the system.

Our models extend previous models for the emergence and maintenance of diversity under stable equilibrium ecological dynamics^{13,17,26–28}. They differ from earlier models such as^{27,28} in key aspects: they consider evolution in high-dimensional phenotypes that characterize continuously variable and multivariate niche use, and persistent ecological fluctuations are intrinsically generated by overcompensating competition. Desynchronized boom-bust cycles provide a robust mechanism for a substantial increase in the diversity that can evolve and be maintained in such models, an effect that increases with increasing dimension of phenotype space. We note that this latter result is not obvious, as with higher phenotypic dimensions the number of phenotypically similar species (nearest neighbours in phenotype space) increases linearly with the dimension, which may be expected to make de-synchronization

of neighbouring boom-bust cycles more difficult due to denser phenotype packing.

This mechanism of “diversification in time” is similar to those previously reported^{10,29}: time acts as additional niche space, and separation along this niche space can alleviate interspecific competition. In the language of³¹, boom-bust desynchronization effectively increases the “environmental dimension”, which is a determinant of the amount of diversity that can be sustained. Again, this is akin to the temporal storage effect^{7,8}, although the latter is mostly invoked for externally generated population fluctuations. The longer the boom-bust cycles, the more temporal separation between similar species is possible. If the population crashes in the boom-bust regime become too severe, they produce frequent extinctions, which eventually leads to a net negative effect on diversity.

In our models, higher diversity can only be observed in the presence of pronounced boom-bust cycles, but not with other types of population fluctuations, such as periodic or chaotic dynamics with high-frequency oscillations. From a modeling perspective, it is worth noting that more standard and more widely used discrete-time models, such as the Ricker equation or the discrete-time logistic model, even in their chaotic regimes cannot exhibit the type of chaotic boom-bust dynamics that model (2) exhibits for low intrinsic growth rates λ and large (overcompensating) β . This reiterates old cautionary notes about the judicious use of discrete maps for modeling ecological dynamics²¹.

Discrete-time models have proved to be very useful for many different purposes in ecology and evolution at least since Ricker’s famous stock and recruitment paper²⁴. However, we note that boom-bust dynamics can also be generated using continuous-time models. To illustrate this, consider a continuous-time analogue of the modified model introduced at the end of the Results section. This continuous-time model has two phases, representing slow exponential growth and fast exponential decline in continuous time. In the first phase, long and

slow exponential growth occurs from low densities for each phenotypic population, while keeping track of the effective density experienced by each phenotype, i.e., of the weighted sum over all phenotypic population sizes, with weights given by the competition kernel (this corresponds to the sum in the denominator of eq. (4)). Once the effective density of a given phenotype reaches the carrying capacity of that phenotype, there is a very fast exponential decline until the phenotype reaches a small fraction of the population size it had before the decline, which corresponds to a severe population crash. Simulations of this simple boom-bust model in continuous time (results not shown) produce qualitatively identical results: the amount of diversity that emerges and is maintained evolutionarily is much larger than the diversity that would evolve with stable equilibrium dynamics (as e.g. reported using a continuous-time logistic model in¹⁷). This again underscores the generality of the effects of boom-bust ecological dynamics on diversity.

We speculate that the mechanisms and results reported here are not limited to competition models, but could also be manifest in communities with other ecological interactions, e.g. in communities with crowding effects³², or in communities containing both predators and prey³³. Whenever the population dynamics exhibit patterns of rapid growth interspersed by crashes (as may e.g. be expected in many predator-prey systems), temporal desynchronization can occur spontaneously and thus lead to increased diversity. Such effects were shown in³³, who reported that "kill-the-winner" mechanism, in which predation generates crashed in the most abundant consumer species, can generate increased levels of diversity. These mechanism differ from the ones reported here in that they are extrinsic to the crashing consumer species (and it is difficult to compare those system to baseline systems with stable ecological dynamics).

There is some empirical support for the effect of boom-bust cycles on ecosystem diver-

sity. For example, such patterns were observed in carefully staged long-running experiments with several plankton species⁶, and the experimental data showed that out-phase oscillations in predator-prey cycles of zooplankton and phytoplankton were important for the maintenance of diversity in this system³⁴. Predation from pathogens have also been reported to induce algal boom-bust cycles³⁵. Generally, boom-bust cycles appear to be common in many marine ecosystems, which are known to be very diverse. For example, it has been suggested to call echinoderms a “boom-bust” phylum³⁶, and recent work shows that in polar plankton communities, which constitute an important ecosystem in the global ocean, phytoplankton dynamics are often categorized by “boom-bust” cycles³⁷.

It is interesting to put our results in the context of observations of “neutral evolution”. For example,³⁸ report neutral taxonomic distributions during early metazoan diversification into relatively empty niche space. In our models, such expansions could be classified as the boom stage, and according to our model assumption would then indeed occur essentially unabated and in the absence of competitive effects. The actual selection only occurs during the bust stage with populations of less adapted species crashing earlier and deeper. Such an application of the model (4) would be rather speculative however, as the unrestricted exponential growth phase would have to last for a very long time and would in any case represent a simplified and unrealistic assumption for such scenarios. We also note that³⁸ consider neutrality based on taxonomic data, not on functional data, whereas our model only considers functional phenotypic data. It has been noted that the distinction between taxonomic and functional data is very important in many microbial ecosystems³⁹, and in particular that functional data can be decidedly non-random even when taxonomic distributions look random.

Overall, we think that our results provide a useful evolutionary perspective for thinking

about diversity in natural ecosystems. Boom-bust population fluctuations are a robust, intuitively appealing and probably under-appreciated potential cause of significantly increased diversity in evolving ecosystems.

Acknowledgement

MD was supported by NSERC Discovery Grant 219930. ECJ was supported funded by the National Agency for Research and Development (ANID) Scholarship 21190785. II acknowledges support from FONDECYT project 1200708.

Supplementary Material

1 Spontaneous desynchronization of coupled boom-bust populations

2 Here we present analytical arguments to illustrate that for large β and populations with
3 sufficiently distinct phenotypes, and thus only moderate mutual competitive effect, the boom-
4 bust cycles inevitably desynchronize.

5 Consider two phenotypes x_1 and x_2 with population sizes N_1 and N_2 whose phenotypes
6 symmetrically diverged from the maximum of the carrying capacity, so that $x_1 = -x_2 \equiv x$.
7 Thus the carrying capacity for each species is equal, $K(x_1) = K(x_2) \equiv K$. The population
8 dynamics of the first species is defined as

$$N_1(t+1) = \frac{\lambda N_1(t)}{1 + (\lambda - 1) \{[N_1(t) + \alpha N_2(t)]/K\}^\beta}, \quad (\text{S1})$$

9 where $\alpha \equiv \alpha(x_1, x_2)$ is the competitive effect that the two species have on each other (recall
10 that α , given by (4) in the main text, is assumed to be symmetric). The population dynamics
11 of the second species is defined analogously.

12 For argument's sake, we assume that initially N_1 and N_2 are small, and the second popu-
13 lation is slightly larger than the first one, $N_1(0) \equiv N$, $N_2(0) = N + \Delta N$. During the boom
14 phase, the effective population size experienced by phenotype x_1 , $N_1(t) + \alpha N_2(t)$, is (much)
15 less than one, so that when elevated to the large power $\beta \gg 1$, this term becomes negligible
16 compared to 1. Then the denominator of (S1) is very close to one. The same is true for N_2 ,
17 so that both populations grow approximately exponentially,

$$N_1(t) \approx \lambda^t N \quad (\text{S2})$$

$$N_2(t) \approx \lambda^t (N + \Delta N).$$

18 We assume that the initial difference between the two population sizes is small enough for
19 both of them to crash in the same time step. The populations of two species immediately
20 after the crash step T are

$$N_1(T+1) = \frac{\lambda^{T+1} N}{1 + (\lambda - 1) \left(\frac{\lambda^T [N + \alpha(N + \Delta N)]}{K} \right)^\beta}, \quad (\text{S3})$$

$$N_2(T+1) = \frac{\lambda^{T+1} (N + \Delta N)}{1 + (\lambda - 1) \left(\frac{\lambda^T [N + \Delta N + \alpha N]}{K} \right)^\beta}.$$

21 We introduce the notation

$$C(t) \equiv \frac{N_1(t) + \alpha N_2(t)}{K}, \quad (\text{S4})$$

22 and ignore the summand 1 in the denominators of (S3), as at the crash step $C(T)$ exceeds
23 1 (for the first time in the cycle), so that for $\beta \gg 1$, the second term in the denominators
24 becomes very large, $C^\beta(T) \gg 1$. Thus:

$$N_1(T+1) \approx \frac{\lambda^{T+1}N}{(\lambda-1)C^\beta(T)}, \quad (\text{S5})$$

$$\begin{aligned} N_2(T+1) &\approx \frac{\lambda^{T+1}(N+\Delta N)}{(\lambda-1)\left[C(T) + \frac{(1-\alpha)\lambda^T\Delta N}{K}\right]^\beta} = \\ &= \frac{\lambda^{T+1}N}{(\lambda-1)C^\beta(T)} \times \frac{1 + \frac{\Delta N}{N}}{\left[1 + \frac{(1-\alpha)\lambda^T\Delta N}{KC(T)}\right]^\beta}. \end{aligned}$$

25 Expanding expression (S5) for $N_2(T+1)$ for small ΔN and observing that the first fraction
26 in this expression is $N_1(T+1)$ and $C(T) \approx \lambda^T N(1+\alpha)/K$, we get

$$N_2(T+1) \approx N_1(T+1) \left\{ 1 - \frac{\Delta N}{N} \left[\beta \frac{1-\alpha}{1+\alpha} - 1 \right] \right\}. \quad (\text{S6})$$

27 For $\beta \gg 1$ and α noticeably less than 1, the factor in square brackets in (S6) is larger than 1,
28 so that the difference between the two populations N_1 and N_2 after the bust will change sign,
29 $N_2(T+1) < N_1(T+1)$, and will grow in absolute value. Thus, after several boom-bust
30 cycles, the difference $|N_1 - N_2|$ becomes large enough for the minority population to avoid
31 crashing at the same time step as the majority population, and hence to delay its crash by at
32 least one step, as it is illustrated in Fig. S.1.

33 In fact, the lag between the two populations increases until reaching approximately half
34 the boom-bust period, thus rendering the two populations maximally anti-correlated. This
35 can be seen by assuming that in each crash the population is reduced to the initial population

Figure S.1: Illustration of the desynchronization of two populations described by (S5) with initially very similar population sizes. Parameter values $\beta = 40$ and $\alpha = 0.7$.

size N . This is an approximation that ignores the chaotic nature of the population dynamics, but serves to demonstrate the general trend. We denote the initial lag between the first and the second population crashes as Δt steps. Since the first population crashed and resumed its growth earlier, its next crash will come no later than that of the second population. It will occur at step T , where T is the smallest integer that satisfies

$$N\lambda^T \left(1 + \frac{\alpha}{\lambda^{\Delta t}}\right) \geq K. \quad (\text{S7})$$

Here we have taken into account that the second population at time T is $N\lambda^{T-\Delta t}$. After the first population crashes down to N , the second population will continue growing for Δs more steps and will crash at the smallest Δs satisfying the condition

$$N\lambda^{T-\Delta t+\Delta s} (1 + \alpha\lambda^{\Delta s}) = N\lambda^{T-\Delta t+\Delta s} \left(1 + \frac{\alpha}{\lambda^{T-\Delta t}}\right) \geq K. \quad (\text{S8})$$

44 If $T - \Delta t > \Delta t$ (which means $\Delta t < T/2$), that is, if the initial lag between crashes is smaller
 45 than half the cycle length, the two conditions (S7, S8) can be simultaneously satisfied only
 46 when $\Delta s > \Delta t$, i.e. when the lag between crashes increases. The lag stops increasing
 47 when $\Delta t = T/2$, i.e., when the boom-bust cycles of the two populations become maximally
 48 desynchronized.

49 In a general sense, the anti-synchronization of boom-bust cycles can be thought of as the
 50 opposite of the well-analyzed phenomenon of synchronization of biological oscillators⁴⁰.
 51 The simplest explanation for this disparity are the opposite convexities of growth functions
 52 used in these models.

53 **Effect of the extinction threshold on diversity**

54 To show that the reduction in diversity for very large $\beta > 60$ is indeed caused by preva-
 55 lence of cluster extinction over diversification, we performed simulations varying the ex-
 56 tinction threshold. Fig. S.2 (and accompanying videos) shows the evolving community for
 57 extinction thresholds 10^{-12} (used for figures in the main text), 10^{-14} and 10^{-10} . Three snap-
 58 shots of the evolving system at their steady state diversity are shown, illustrating that chang-
 59 ing the extinction threshold shifts the balance between extinction and diversification: lower
 60 thresholds increase the level of diversity, and with high enough thresholds (and high enough
 61 β) diversity is reduced to a single species.

Figure S.2: Snapshots of phenotype distributions after 10^7 generations for three values of the extinction threshold: A) extinction threshold 10^{-12} used for figures in main text; B) lower extinction threshold 10^{-14} ; C) higher extinction threshold 10^{-10} . All three scenarios are simulated with very strong overcompensation, $\beta = 110$, and $\lambda = 1.2$. Dimension of phenotype space is $d = 2$, and $\sigma_\alpha = 0.5$. The videos of the evolutionary diversification processes that generated these configurations can be found at figshare.com/s/f2d8ecf480fa372319e1.

62 **Effect of the mutation rate on diversity**

63 Figure S.3A shows the number of species as a function of time (in generations) for differ-
 64 ent mutation rates (other parameters are as for Figure 2B in the main text). The lower the
 65 mutation rate, the longer it takes for diversity to reach saturation levels. Figure S.3B illus-
 66 trates that the extinction threshold has an effect on the saturation levels of diversity: lower
 67 thresholds (fewer extinctions) lead to elevated levels of diversity. For Figure S.3B, the mu-
 68 tation rate was $\mu = 0.005$, so that one mutation of typical size less than 1% of the parental
 69 phenotype occurs every 200 generations in the entire evolving community.

Figure S.3: A: Number of species in the community as a function of time for different mutation rates μ . Species were counted using the clustering algorithm described in the Model and simulation methods section. Other parameter values were as for Fig. 2B in the main text. B: Number of species in the community as a function of time for two different extinction thresholds and for a mutation rate $\mu = 0.005$. Other parameter values were as for Fig. 2B in the main text. For the two scenarios shown in B, videos of the evolutionary process can be found at figshare.com/s/f2d8ecf480fa372319e1.

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