- 1 **Title:** Ecology of active viruses and their bacterial hosts in frozen Arctic peat soil revealed with
- 2 $H_2^{18}O$ stable isotope probing metagenomics
- 3
- 4 **Running title:** SIP-metagenomics of arctic peat viral populations
- 5
- 6 Gareth Trubl¹, Jeffrey A. Kimbrel¹, Jose Liquet-Gonzalez¹, Erin E. Nuccio¹, Peter K. Weber¹,
- 7 Jennifer Pett-Ridge¹, Janet K. Jansson², Mark P. Waldrop³, Steven J. Blazewicz¹
- ⁸ ¹Physical and Life Sciences Directorate, Lawrence Livermore National Laboratory, Livermore,
- 9 CA, USA
- ¹⁰ ²Biological Sciences Division, Pacific Northwest National Laboratory, Richland, WA, USA
- ³U.S. Geological Survey, Geology, Minerals, Energy, and Geophysics Science Center, Menlo
- 12 Park, CA, USA
- 13
- 14 Corresponding authors:
- 15 Gareth Trubl (<u>Trubl1@llnl.gov</u>) and Steve Blazewicz (<u>blazewicz1@llnl.gov</u>)
- 16
- 17 Key Words: soil viruses, bacteriophage, stable isotope probing, permafrost, peat, ¹⁸O-water,
- 18 metagenomics
- 19
- 20
- 21
- 22

23 Abstract

24 Winter carbon loss in northern ecosystems is estimated to be greater than the average growing 25 season carbon uptake. However, most ecosystem carbon measurements neglect winter months 26 since carbon losses (primarily driven by microbial decomposers) are assumed to be negligible at 27 low temperatures. We used stable isotope probing (SIP) targeted metagenomics to reveal the 28 genomic potential of active soil microbial populations under winter conditions, with an emphasis 29 on viruses and virus-host dynamics. Peat soils from the Bonanza Creek LTER site in Alaska were incubated under subzero anoxic conditions with $H_2^{18}O$ for 184 and 370 days. We identified 30 31 46 bacterial populations (MAGs; spanning 9 bacterial phyla) and 243 viral populations (vOTUs) 32 that actively took up ¹⁸O and produced significant CO₂ throughout the incubation. Active hosts, 33 predicted for 33% of the active vOTUs, were some of the most abundant MAGs and capable of 34 fermentation and organic matter degradation. Approximately three-quarters of the active vOTUs 35 carried auxiliary metabolic genes that spanned five functional categories, including carbon 36 utilization, highlighting the potential impact of viruses in this peat soil's microbial 37 biogeochemistry. These results illustrate significant bacterial and viral activity and interactions 38 occur in frozen soils, revealing viruses are a major community-structuring agent throughout 39 winter months.

40

- 41
- 42
- 43

44

45 Introduction

46 Northern peatlands are important terrestrial ecosystems for carbon (C) storage, estimated 47 to contain one-third of soil C (~1,000 gigatons) [1], yet the fate of this C is unknown as these 48 soils are experiencing dramatic changes from anthropogenic climate change [2-4]. While soil-49 warming experiments indicate increased carbon dioxide (CO_2) and methane emissions under 50 global climate change, it is likely that these C losses from northern peatlands are underestimated 51 because virtually all measurements neglect winter processes [5-8]. Recent estimates of winter C 52 loss alone are estimated to be greater than the average growing season C uptake [9]. While the 53 winter months include large temperature fluctuations and extreme temperature minimums [10], 54 the temperatures found in much of the soil profile of permafrost or seasonally frozen bogs can 55 remain stable and just below the freezing point ($\Box 1$ to $\Box 5^{\circ}$ C) [5, 11, 12]. Bacteria have been 56 shown to remain active below the freezing point [13] with both catabolic and anabolic activities 57 observed in frozen soils [14, 15]. Therefore, it is critical to understand the taxonomy, functional 58 capacity, and activities of bacteria and viruses that cause microbial turnover in frozen soils to 59 better predict their biogeochemical implications.

60 In soils, viruses may play a major role in microbial population dynamics and the fate of 61 soil C, similar to the role they play in marine systems, where viruses kill $\sim 40\%$ of bacteria daily 62 and sustain up to 55% of bacterial production via C liberation [16, 17]. Soils hold an enormous 63 viral reservoir, but we know remarkably little about their diversity, activity, host interactions, 64 lysis-induced C-cycling, and persistence as compared to other environments [18, 19]. Viruses 65 can affect soil C cycling by stopping microbial metabolism during lytic infections and releasing 66 cell-derived nutrients into the environment [20]. These nutrients can either fuel other organisms' 67 metabolism or be stabilized via entombing effects [21]. This process can be prolonged by

68 temperate viruses, which remain latent in their host for long periods, before eventually switching 69 into the lytic infection cycle. Viruses can also carry auxiliary metabolic genes (AMGs), host-70 derived genes that can be expressed during infection and typically aid the infection process by 71 overcoming energetic limitations [22, 23]. While viruses and their impacts have been well-72 characterized in marine environments via meta-omic approaches [24] and development of 73 tractable virus-host model systems, these linkages remain enigmatic in soil environments. 74 Growing season studies of viruses in northern peatlands indicate they are largely unrelated to 75 other known viruses, highly diverse, endemic to their habitat, infect dominant microbial 76 populations, and carry AMGs that could impact C cycling [25, 26]. The question of whether 77 viruses are active during the 7-9 months of the year when Arctic peatlands are frozen remains 78 unanswered.

79 Stable isotope probing (SIP) combined with metagenomics is an effective tool for 80 tracking the active microbial taxa in complex communities, linking individuals to specific 81 functions [27-29], or characterizing specific microbes and the viruses that infect them [30]. While many isotope tracing studies use ¹³C-enriched tracers (e.g., ¹³CO₂, ¹³C-plant biomass, or 82 ¹³C-glucose) to identify specific substrate degraders [28] or assess microbial C use efficiency 83 [31], heavy water ($H_2^{18}O$) SIP-metagenomics has the unique benefit of labelling all actively 84 85 growing microbes because water is a universal substrate for nucleic acid synthesis [32-34]. Over 86 adequate time scales, this technique is sensitive enough to label slow-growing or less abundant 87 microbes and identify taxon-specific microbial growth and mortality patterns [35-38], but 88 previously, has not been used to label active viruses.

Here we used heavy-water SIP-metagenomics to label active microbial taxa and viruses
and characterize whether viruses play a role in controlling active microbial population dynamics

91 under subzero and anoxic soil conditions. To our knowledge, viral activities have never been 92 confirmed under such conditions, and understanding their impact on microbial activity and soil C 93 transformation over winter months may be pivotal for understanding the future fate of the vast 94 organic C stocks in Arctic peatlands.

95 Methods

96 Experimental setup

97 We collected soil samples from the Alaska Peatland Experiment (APEX), part of the 98 Bonanza Creek Long Term Ecological Research (LTER) site southwest of Fairbanks, AK (64.70 99 $^{\circ}$ N, \Box 148.3 $^{\circ}$ W), in a zone of discontinuous permafrost (for overview see Fig. S1). Three peat 100 soil cores were collected from the active layer of a Sphagnum-dominated thermokarst bog on 101 June 16, 2013 from the beta site approximately 20 m south of the flux tower. During the week of 102 sampling, midday air temperatures were 27°C, midday surface 10 cm peat temperature was 103 10°C, active layer depth in the bog was 35 cm, and the water table was 3-8 cm below the 104 surface. The pH was approximately 4.80 (1:1 soil:water ratio). Vegetation and other soil and flux 105 data are given in Waldrop et al., [39]. To collect the cores, moss and peat were removed down to 106 the top of the water table using scissors. The top 20 cm below the water table were collected 107 using a 6.3 cm diameter sharpened steel barrel corer attached to an electric drill. Three cores 108 were collected 1 meter apart along a transect. Cores were stored in mason jars filled with 109 porewater collected from the core hole. Sealed jars were immediately shipped on ice to the U.S. 110 Geological Survey (Menlo Park, CA, USA) where they were homogenized in an anaerobic 111 glovebox maintained at 4°C. Anoxic synthetic porewater was created by freeze-drying filtered 112 (0.45 µm PTFE then 0.2 µm nylon filter) soil porewater collected in tandem with the cores, and then rehydrating the remaining particulates with either heavy water (97 atom% $H_2^{18}O$, 113

114 Cambridge Isotope Laboratories) or natural abundance water (control), and then sparging with 115 N₂ to remove O₂. Soil subsamples (2 g of soil wet weight) were collected from each core, 116 pressed to remove porewater (using a 5 ml syringe fitted with a nylon screen and a glass fiber 117 filter), and transferred to Wheaton serum vials (10 ml), creating 12 incubation vessels. Vials 118 were sealed with blue butyl rubber stoppers and the headspace was purged (to remove H_2 from 119 the glovebox atmosphere) by vacuuming and filling with N_2 (10 inches Hg/5 psi) 10 times in a cold room. Anoxic synthetic ¹⁸O-enriched porewater was added (2.5 ml) to half of the incubation 120 121 vials and anoxic synthetic natural abundance porewater was added to the other half using a 5 ml 122 syringe with a 23 G needle that was purged with N_2 . Six parallel samples were set up in a similar 123 manner for headspace gas analysis to quantify CO₂ production, except these used proportionally 124 larger amounts of soil (18.15 g wet soil) and larger incubation vials (100 ml). All the sample 125 vials were submerged in a glycerol bath at 4°C and the temperature was slowly and steadily 126 reduced to □1.5°C, over 48 h. Samples were then continuously maintained at □1.5°C for 184 d 127 (midyear) and 370 d (full year). At each timepoint, biological replicates (n=3) were destructively 128 harvested and snap frozen in liquid N₂ and stored at $\Box 80^{\circ}$ C. Three of the gas production vials 129 were incubated at $\Box 20^{\circ}$ C as a control.

130 *CO*₂ production quantification

Gas samples were collected from gas production vials at 10 timepoints over the 370 d incubation. To prevent oxygen from contaminating the incubation vials, a 5 ml syringe with a stopcock and 23 G needle was cleared 3 times with O_2 -free N_2 . The syringe was then inserted into the vial septa and plunged 3 times to mix the headspace; 2 ml headspace was collected, and the stopcock locked. The 2 ml samples were transferred to 10 ml serum bottles that had been

purged and then filled with N_2 (1 atm). Gas samples were analyzed via gas chromatography (SRI 8610GC, SRI instruments, Torrance, CA) to quantify headspace CO₂.

138 DNA extraction and density gradient SIP

139 DNA was extracted from each replicate using a modified phenol/chloroform protocol 140 [40]. In summary, 0.5 g (+/- 0.01 g) wet soil was added to lysing matrix E tubes, and 100 μ l 1x 141 TE (pH 7.5), 150 µl PO₄ buffer (0.2 M in 1 M NaCl), and 100 µl 10% SDS were added and 142 vortexed. Tubes were bead beat for 30 s at 30 1/s and briefly centrifuged. 0.6 ml 143 phenol:chloroform:isoamyl alcohol (25:24:1) was added, vortexed, and incubated at 65°C for 10 144 min. Tubes were spun for 5 min at 10K x g, and the supernatant was transferred to a new tube. 145 The bead beat tubes were then re-extracted using 220 μ l 1x TE and 80 μ l PO₄ buffers. The 146 supernatant from the first and second extracts were combined in a new 2 ml tube. 550 µl 147 phenol/chloroform/isoamyl alcohol was added, mixed, and centrifuged (10K x g, 5 min). The 148 aqueous top layer was transferred to a new 2 ml tube. 900 µl chloroform: isoamyl alcohol (24:1) 149 was added, mixed, centrifuged (10K x g, 5 min), and the supernatant transferred to a new 2 ml 150 tube. Then, 850 µl chloroform: isoamyl alcohol (24:1) was added, mixed, centrifuged (10K x g, 5 151 min), and the supernatant transferred to a new 1.7 ml tube. RNAase was added (6.44 μ l, 10 152 mg/ml), mixed, and incubated at 50°C for 10 min. 244 µl 10 M NH₄⁺ acetate was added, mixed, 153 and incubated at 4°C for 2 h. Tubes were centrifuged at 16.1K x g for 15 min, and the 154 supernatant transferred to a new tube. Isopropanol (670 µl) was added, mixed, and centrifuged 155 (16.1K x g, 20 min). The supernatant was removed, and the DNA pellet dried in a PCR hood for 156 15 min. 30 µl 1xTE was added, mixed, and the DNA was then stored at \square 80°C.

157 Extracted DNA was fractionated via CsCl density gradient ultracentrifugation to separate
 158 ¹⁸O-enriched DNA as described previously [35]. DNA was binned into 5 fractions based on

density, and the binned DNA from the two heaviest fractions (medium-heavy [MH; 1.717–1.725
g/ml] and heavy [H; 1.725–1.750 g/ml]) were sequenced.

161 *Sequencing and metagenome generation*

162 DNA from the SIP fractions was sent to the Keck Sequencing Facility at Yale University. 163 For each sample, 100 ng of DNA was sheared to 500bp using the Covaris E210 (Covaris, Inc., 164 Woburn, MA, USA), followed by a SPRI bead cleanup using Ampure XP (Beckman Coulter 165 Life Sciences, Brea, CA, USA); DNA quality was checked using a Bioanalyzer chip. The 166 sheared gDNA from the 24 samples was then end-repaired, A-tailed, adapters ligated, and 167 sequenced on an Illumina HiSeq 2500 to generate metagenomes (Table S1). Paired-end 151 nt 168 reads were processed in three steps with bbduk v38 (Bushnell, B.): 1) adapter-trimming (ftl=10 169 ktrim=r k=23 mink=11 hdist=1 tpe tbo minlen=50), 2) PhiX and Illumina adapter/barcode 170 removal (k=31 hdist=1 minlen=50), and 3) quality-trimming (qtrim=r trimq=10 minlen=50). 171 Metagenomes were successfully generated for 23 of the samples (one did not have enough DNA 172 recovered), with 302 Gbp of sequencing data.

173 *Recovery of vOTUs from metagenomes*

174 Virus-specific informatics were used to increase the number of viral sequences detected 175 in these soil datasets [19]. Processed reads were assembled into contigs using SPAdes v3.11.1 (--176 only-assembler --phred-offset 33 --meta -k 25,55,95 -12)[41]. From the 23 SIP-fractionated metagenomes, we assembled 51 487 619 contigs, with 63% of the total reads mapped to the 177 178 contigs. Contigs were processed with VirSorter (virome decontamination mode) [42] and DeepVirFinder (DVF) [43] to detect viral contigs. We retained contigs that were ≥10kb, sorted 179 180 into VirSorter categories 1 and 2, and had a DVF score ≥ 0.9 and p value <0.05. Viral contigs 181 were clustered at 95% average nucleotide identity (ANI) across 85% of the shorter contig [44]

182 using nucmer [45] to generate a nonredundant set of viral populations (vOTUs). vOTU quality 183 was assessed with CheckV (default parameters) [46]. Coverage of the vOTUs was estimated 184 based on post quality-controlled read mapping at \geq 90% ANI and covering \geq 75% of the contig 185 [44] using Bowtie2 [47]. Coverage was then normalized per gigabase-pair of metagenome and 186 by length of the contig [48]. Activity of vOTUs was determined by a vOTU being present in the 187 H₂¹⁸O samples and not present in the paired natural abundance water samples. The vOTUs 188 were annotated using the virus-centric multiPhATE pipeline (default parameters) [49] and the 189 AMG-centric DRAM-v pipeline (with --skip uniref) [50]. We note that DRAM-v provides AMG 190 scores only for vOTUs detected via VirSorter; AMGs predicted from the 208 vOTUs recovered 191 from DVF were manually curated. For this manual inspection, we removed putative AMGs that 192 were at contig ends or those with annotations from multiple functional categories. To determine 193 the proportion of temperate vOTUs, we searched for genes associated with proviruses, such as integrase and parA [51], leveraged classification from VirSorter (categories 4 or 5) and our 194 195 genome-similarity host linkages ($\leq 90\%$ similarity), and used two tools — CheckV [46] and 196 PHASTER [52].

197 MAG curation and host linking

To generate metagenome assembled genomes (MAGs), read-pairs from the biological replicates were grouped for 8 separate co-assemblies (2 timepoints x 2 SIP fractions x 2 O isotopes) with MEGAHIT v1.1.4 (--k-min 27 --k-max 127 --k-step 10) [53]. Contigs greater than 1 KB were separately binned with Concoct v1.0.0 [54], MaxBin v2.2.6 [55] and MetaBAT v2.12.1 [56]. Genome bins from these three binning tools were refined using the bin_refinement

module of MetaWRAP v1.2.1 (-c 50 -x 10) [57] with CheckM v1.0.12 [58]. Only genome bins
with at least 'medium quality' according to MIMAG standards [59] were retained.

205 Two methods were used to define MAGs as active: 1) a read-subtraction approach, and 2) a log-fold-change approach. For the read-subtraction approach, contigs from the ¹⁶O assemblies 206 were used as a reference to subtract the ¹⁸O reads that aligned with the unlabeled dataset using 207 208 bbsplit v38 (maxindel=1) [60]. The isotopically labeled reads that did not align with the 209 unlabeled dataset were considered 'active' reads and were then processed through the same 210 MAG assembly workflow described above (starting with the MEGAHIT assembly through MetaWRAP refinement); this generated genome bins reconstructed from distinct ¹⁸O reads from 211 212 the two SIP fractions at two timepoints.

213 Refined bins from all 12 groups (8 total + 4 active) were dereplicated into a final set of 214 metagenome assembled genomes (MAGs) using dRep v2.2.3 (-comp 50 -con 10 -p 6 -nc 0.6) 215 [61]. MAG taxonomy was determined using GTDB-tk v0.3.2 [62] with the version r89 GTDB 216 database. Structural annotation was done using Patric [63], and general functional annotation 217 with RASTtk [64] for subsystems within Patric v3 (retaining subsystem variants predicted to be 218 active or likely), and KofamScan v1.1.0 [65] with the KEGG [66] 93 database. Per-sample MAG 219 abundances were determined by aligning each sample's filtered reads against the MAG genomes 220 using bbmap v38 [67] to obtain total assigned reads per contig and average fold coverage per 221 contig.

For the log-fold-change approach to define active MAGs, we assessed significant (5% false-discovery rate) and positive log2-fold change in ¹⁸O versus ¹⁶O read abundances within a time point and SIP fraction. Fold changes were determined using wrench-normalized [68] total

assigned reads per MAG with a zero-inflated log-normal model implemented in metagenomeSeq[69].

227 The vOTUs and MAGs were linked via clustered regularly interspaced short palindromic 228 repeats (CRISPR) spacers and shared genomic content as previously described [70]. Briefly, 229 CRISPR regions were detected in the MAGs using MinCED (options -minNR 2 -spacers) [71] 230 and linked to the vOTUs with blastn (options -max_target_seqs 10000000 -dust no -word_size 7) 231 [72]. In addition, BLAST (options -dust no -perc_identity 70) was used to link vOTUs and 232 MAGs based on shared genomic content [73]. Virus-host abundance estimates were made by 233 summing microbial host abundances at the phylum level in each timepoint and linked vOTU 234 abundances.

235 Data availability

The 23 metagenomes were deposited to NCBI under BioProject identifier (ID) PRJNA634918 with BioSample information included in Table S1. Figures were generated with Microsoft Excel and R, using packages Vegan for diversity and pheatmap for heat maps. T-tests and linear regressions were performed using the data analysis package in Excel.

240 **Results**

241 Characterization of viruses

To characterize viral activity in Arctic peat soil under winter conditions, we analyzed viral sequences from heavy-water SIP-targeted metagenomic reads. Using two viral detection methods, we identified 5 737 putative viruses (\geq 5 kb) that clustered into a nonredundant set of 332 vOTUs \geq 10 kb (Table S2) that span 66 viral genera (Fig. S2; Table S3). The size range of these vOTUs was 10 039 – 437 858 bp (average 32 954 bp) with 15 vOTUs \geq 100 kb, including 5

247 so-called 'huge' viruses (i.e., ≥ 200 kb) [74]. The vOTUs were well-covered with an average of 248 17x normalized coverage per metagenome, but with a large range, 3–147x (Fig. 1A). After 249 quality checks, we identified 58 medium-high quality vOTUs, of which four were considered 250 'complete' according to community standards [44]. Genome quality could not be assessed for 93 251 vOTUs because they did not possess a known viral gene and were detected via a reference-free 252 machine learning method [43]. Annotation of the vOTU genomes with the virus-centric 253 Multiphate pipeline predicted 15 772 genes, of which 61% were novel (Table S4). With the 254 AMG-centric pipeline DRAM-v, we predicted 86 putative AMGs (score 1–3) distilled into five 255 functional categories (carbon utilization, energy, organic nitrogen, transporters, and 256 miscellaneous) from 31 vOTUs (Table S5); 21 of the vOTUs were active and carried 63 AMGs 257 (Fig. 2). To identify temperate viruses, we searched the annotations for genes used in the 258 lysogenic infection cycle and predicted nearly half (43%) of our vOTUs were temperate viruses. 259 More than half (59%) of these temperate viruses were active, and the majority (80%) had at least 260 one member of their population integrated at the time of sampling (provirus; Table S6).

261 Significantly more vOTUs were observed in the SIP fractions from heavy-water 262 incubations relative to control (natural abundance water) incubations, confirming that many viruses incorporated ¹⁸O into their DNA and were actively replicating (Fig. S3). A majority 263 (73%) of the vOTUs found in the dense ¹⁸O SIP fractions were active at some point over the 370-264 265 d incubation, with about half active the full year and the other half only active in 184 d or 370 d 266 samples (Fig. 1B). We measured active vOTU diversity to assess changes in the viral community 267 structure from midyear to a full year of incubation. The Shannon's H metric indicates a 268 significantly ($p \le 0.05$) more diverse viral community at 370 d compared to 184 d (Fig. 1C). 269 Shannon's H diversity, which includes both richness and evenness, was driven by an average increase of more than 60% for vOTU richness from 184 to 370 d. Of these, 64 vOTUs became
more abundant from 184 to 370 d, 110 became newly active, and 18 were no longer detected as
labeled at 370 d.

273 Host characterization based on SIP-metagenomics

274 To identify potential viral hosts, we used a suite of metagenome assembly and binning 275 methods which yielded 153 medium to high quality MAGs, spanning 16 bacterial phyla and 1 276 archaeal phylum (Table S7; GTDB taxonomy). The dominant phyla detected were 277 Chloroflexota, Acidobacteria, and Myxococcota (formally Proteobacteria). Incubation with 278 heavy water indicated 30% (46) of these MAGs were actively growing, spanning 9 bacterial 279 phyla, with the most abundant active populations from the Acidobacteria, Bacteroidetes, and 280 Firmicutes. By sequencing both unlabeled and isotopically labeled samples, we gained insight 281 into genetic repertoire from both active and inactive bacteria (Table S8) but focused our efforts 282 on characterizing the active bacterial lineages and metabolisms that viruses may be altering over 283 winter months. We used CO_2 production measurements to confirm microbial activity and 284 positive fluxes occurred continuously throughout the $\Box 1.5^{\circ}$ C incubation, but not from the $\Box 20^{\circ}$ C 285 control incubation (Fig. 3).

286 Linking vOTUs to MAG hosts

To understand the influence of the viruses on microbial dynamics during the winter season, we predicted potential microbial hosts via two different *in silico* approaches based on sequence similarity. First, we identified 11 879 CRISPR spacers in 136 of the MAGs and used these to link 10 vOTUs to 6 MAGs from 4 bacterial phyla (Fig. 4, Table S9). Leveraging the SIP activity patterns, we found most of these identified linkages connected active vOTUs and active MAGs (12 linkages between 8 active vOTUs and 5 active MAGs, 1 linkage between an

293	unlabeled vOTU and an active MAG, and 3 linkages between unlabeled vOTUs and unlabeled
294	MAGs). In a second approach, we used vOTU-MAG genome sequence similarity and recovered
295	798 similarity matches that linked 141 vOTUs to 65 MAGs from 10 bacterial phyla (Fig. 4;
296	Table S10). Combined, the two virus-host linkage approaches indicated 318 unique connections
297	between 145 vOTUs and 65 MAGs spanning 10 bacterial phyla. A significantly higher
298	proportion of these vOTU-host matches were between active populations (Table 1). Notably,
299	four vOTUs (#s 153, 161, 162, and 270) had a broad-host range, and were linked to bacteria
300	from several bacterial phyla (three of these vOTUs infected two phyla and one infecting four
301	phyla; Fig. 4). Two of these broad-host-range vOTUs (153 and 270) were active and linked to
302	only active MAGs from bacterial phyla Bacteriodota and Firmicutes.
303	All 145 of the vOTUs we identified as host-linked may therefore be classified as dsDNA
304	bacteriophages (since the vOTUs were linked to MAGs with a bacterium taxonomic
305	assignment). These represented the majority (88%) of the host-virus matches, and almost all
306	(92%) of the unlabeled vOTUs that were linked to unlabeled MAGs. In our soil incubations,
307	Actinobacteriota (56%), Chloroflexota (24%), and Firmicutes (12%) were the most 'infected'
308	bacterial phyla (i.e., with the most vOTU-MAG linkages). Firmicutes (55%), Bacteroidota
309	(34%), Patescibacteria (9%), and Proteobacteria (2%) were the only phyla that had active MAGs
310	linked to active vOTUs. Of the active populations, 81 (33%) vOTUs and 33 (51%) MAGs were
311	linked, with the top 15% most abundant active vOTUs predicted to infect Firmicutes and
312	Bacteroidota, and the abundances of both these vOTUs and their host populations increased over
313	the year incubation (Fig. 5).

314 **Discussion**

315 Ouantifying and predicting the drivers of C loss in northern latitude peatlands underlain 316 with permafrost is a complex and critical issue, especially as these environments continue to be 317 disproportionately impacted by climate change. Microbes largely govern C release from Arctic 318 soils, and recent work has highlighted the important role viruses could play in microbial C 319 processing and helping their hosts adapt to subzero temperatures [25, 26]. What is not well 320 understood is what proportion of these viruses are active, how the viral community dynamics 321 change over time, and how long viruses persist in frozen anoxic peat soils. To our knowledge, 322 this is the first study to address these concerns, using heavy water SIP-metagenomics to directly 323 identify active microbes and their associated viruses and characterize their dynamics from a half 324 of a year to a full year under subzero and anoxic conditions.

325 Viral activity in cryoecosystems

326 Over the course of a year, we found hundreds of active vOTUs, in stark contrast to the 327 general paradigm that frozen soils have little to no activity [75]. A third of these active vOTUs 328 were linked to active MAGs, highlighting that not only are microbes active in these anoxic 329 subzero temperature soils, but their viruses are also active and likely impacting soil microbial 330 community assembly and ecosystem biogeochemistry, as previously proposed [25, 26, 76, 77]. 331 We observed an increase in vOTU richness and abundance from midyear to a full year of incubation, although vOTU evenness remained the same, likely due to a combination of 332 333 proviruses replicating with active microbes and many generations of virions actively infecting 334 newly active host(s). The few vOTUs that decreased in abundance or were present midyear and 335 not detected after a full year, could have been isotopically enriched then some persisted in the 336 soil as environmental DNA (eDNA), or as inert virions from internal nucleic acid errors (e.g., 337 lethal mutations) or structural damage, or as competent virions that were no longer infective due

to their host evolving resistance (e.g., the 'host-virus arms race' that has been observed in othersystems)[78].

340 While most viruses are host-specific, having a broad host range can influence a virus' 341 ecological significance. Generalist viruses that can infect more than one species of host are 342 thought to have different abundance patterns, infection efficiency, movement within an 343 ecosystem, and other unique attributes [79]. We identified fifteen generalists and four highly 344 promiscuous generalist vOTUs that were predicted to infect hosts from different phyla. In 345 addition, these four generalists were some of the most abundant vOTUs, suggesting that having a 346 very broad host range may offer an advantage in these subzero and anoxic soils. Recently, Malki 347 et al. [80] showed that four viruses from an oligotrophic lake in Michigan could infect bacteria 348 across several phyla including Proteobacteria, Actinobacteria, and Bacteroidetes. Interestingly, 349 the host phyla identified by Malki et al. are the same bacterial phyla that our generalists vOTUs 350 infected, with the addition of Firmicutes and Chloroflexota (also predicted by our network 351 analyses, see Table S11; Fig. S4).

352 *Putative viral influence on winter biogeochemistry*

353 Soil viruses shape the abundance, diversity, and metabolic outputs of their microbial 354 hosts. As recent literature has made clear, viruses may modulate ecosystem biogeochemical 355 processes via an intriguing mechanism—reprogramming host metabolism with AMGs during 356 infection [24]. The metabolic functions and ubiquity of AMGs can vary by environment, with 357 marine viruses carrying AMGs for photosynthesis, central carbon metabolism, sulfur cycling, 358 and nitrogen cycling [24]. Soil viruses, although understudied, appear to carry AMGs for 359 polysaccharide degradation and sporulation [25, 26, 81]. In our cryoecosystem, we identified 360 carbon degradation AMGs (e.g., galactose degradation and Xylan 1,4-beta-xylosidase) that could

361 provide fitness advantages for utilizing much-needed carbon sources, as well as genes for 362 scavenging (e.g., ABC-2 type transport system) that may provide their host an avenue to acquire 363 essential nutrients [82]. Additionally, AMGs such as methionine degradation [83] and Glycosyl 364 transferases [84] were found in high abundance which can help during infection.

365 To better understand the roles viruses play in controlling central biogeochemical 366 processes in Arctic peat under winter conditions, we assessed the active bacterial phyla that were 367 virus-infected. Firmicutes had the largest increase in abundance from midyear to a full year and 368 had the most viral infections (both the number of linked active vOTUs and increase in vOTU 369 abundance), supporting the 'kill-the-winner' theory described in aquatic systems (Fig. 3; Fig. 4) 370 [85]. In this Lotka–Volterra predator-prey framework, the most abundant host would become 371 infected by viruses, leading to a decline in host abundance and an increase in the prevalence of 372 their viruses, at least until microbial hosts with genetic resistance to the currently dominant 373 viruses had time to respond. The dominance of Firmicutes hosts makes sense, since this diverse 374 bacterial phylum is known for adaptations to anoxic conditions, including fermentation and 375 creating endospores [86]. In our samples, all active infected Firmicutes contained fermentation 376 genes for the capacity to produce ethanol, lactate, or both (Table S8). Bacteroidota were also an 377 active and frequently infected bacterial phylum that increased in abundance through time; of this 378 group, all the infected MAGs shared taxonomic affiliation to the order Bacteroidales. Most of 379 these bacteria are obligate anaerobes and known for their diverse arrays of carbohydrate-active 380 enzymes arranged into polysaccharide utilization loci and fermentation [87, 88]. Many active 381 MAGs within this group had the genomic capacity for polysaccharide degradation (e.g., ICE 1 382 encoded pectin degradation protein KdgF) and all encoded genes for lactate fermentation (Table 383 S8). After one year of subzero and anoxic incubation, Bacteroidales had become the most

384 abundant active lineage in our soils and were infected by the most abundant active vOTUs (#s 385 23, 37, 41, 124, and 190), further supporting 'kill-the-winner' theory and highlighting viruses 386 influencing host winter activities. While the abundance of active Bacteroidota and Firmicutes 387 populations were increasing, C mineralization to CO_2 occurred steadily throughout the 388 incubation and vOTUs linked to these populations carried several AMGs that would support 389 central carbon metabolism (Table S5). Together, these results suggest that these two bacterial 390 lineages play an important role in over-winter C loss from these Artic peat soils, and the viruses 391 that infect them likely shape both their population dynamics and functional impact.

392 Another way viruses can impact soil biogeochemistry is by infecting hosts that occupy 393 different metabolic niche spaces. The Patescibacteria we identified (part of the Candidate phyla 394 radiation; CPR) are known for their ultrasmall cell size with reduced genomes and most have a 395 symbiotic relationship with Bacteroidota [89], suggesting direct interactions between these two 396 lineages in our cryoecosystem. Patescibacteria are thought to resist phage infection by lacking or 397 reducing the number of potential phage receptors on their cell membrane [90], but in our study, it 398 was notably one of the most infected lineages and had none of the previously reported phage 399 receptors (Table S8). The prevalence of these infections may be the result of their interactions with Bacteroidota. One member of the Patescibacteria was linked to an unlabeled vOTU with a 400 401 CRISPR spacer match, suggesting this adapted immune system element was successful at 402 protecting the host from infection. Typical CRISPRs are rarely found in CPR bacteria and recent 403 work suggests this may be due to this group using a compact CRISPR-CasY system that is 404 highly divergent to typical CRISPR systems [91]. Another infected and active bacterial phylum 405 we observed was Proteobacteria, with only one active MAG from the class Micavibrionales. 406 Little is known about this clade, and even less about their viruses, because they have an epibiotic

407 lifestyle where they feed on other organisms to survive, making them difficult to culture. These 408 predatory bacteria may be active under winter conditions by either consuming non-active or dead 409 cells, or they may benefit from attaching to a host that can utilize alternative energy sources and 410 recalcitrant organic matter [92]. We identified multiple active vOTUs infecting this lineage, 411 revealing another indirect way that viruses may impact nutrient cycling and microbial diversity, 412 via preving upon bacterial predators.

413 Challenges associated with identifying virus 'activity'

414 Heavy-water SIP has proven to be a robust method for identifying metabolically active 415 microbes in soils [32, 33, 36]. In many ways, the approach is superior to other techniques that 416 track activity such as RNA-to-DNA ratios, rRNA levels, bioorthogonal non-nanonical amino acid tagging (BONCAT) or other stable isotopes (e.g., ¹³C) because active microbes synthesize 417 DNA when their cells divide, incorporating ¹⁸O, and therefore the DNA of all active microbes is 418 419 labeled. Data from RNA studies can be hard to interpret as RNA levels often do not correlate 420 with growth and often have weak or no correlation with proteins levels [93-95]. Compared to other isotopes as tracers, ¹⁸O labeling via heavy-water SIP does not rely on the microbe's carbon 421 422 use efficiency or prior knowledge of the microbe's substrate preference [37].

The application of isotope tracers for direct assessment of activity is not as straightforward for viruses as it is for other microorganisms, and worthy of reasoned consideration. Characterizing activity for viruses is quite different compared to 'free-living' organisms because of their different infection cycles, their lack of metabolism, and the many states in which they can be present [19]. One of the main reasons to identify active entities is to help quantify their impact on their hosts and their environment, with the assumption that inactive entities have less of an impact but are still important [96]. For a microbe, this may be true, but

for a virus there is a range in magnitude of impact for different host metabolic states that depends
heavily on the infection cycle. Assessing the prevalence of an infection cycle, however, is
challenging due to difficulties in quantifying lysogeny or virion abundance, burst size, diversity,
and ecology [51].

434 Arctic soils, that predominantly exist under anoxic and subzero temperature conditions, 435 might be generally considered a harsh environment, and would limit microbial growth. For this 436 reason, we hypothesized most of our viruses would be temperate [51, 97] and they would be 437 detected as proviruses. We did also see a high rate of lysogeny (43%), and in support of our 438 hypotheses, the majority (80%) of temperate viruses were present as proviruses. Further, half of 439 the active vOTUs linked to active MAGs (i.e., bacterial phyla Bacteroidota, Firmicutes, 440 Patescibacteria, and Proteobacteria) were temperate viruses. The number of temperate viruses 441 present in this study was higher than previously reported for desert Antarctic soils (4–20%) and 442 almost identical (44%) to whole soil assays from temperate wetlands soils [98]. We hypothesize 443 the increased incidence of temperate viruses is linked to low host abundances and environmental 444 conditions, therefore increasing the potential for sporadic viral infection.

445 Temperate viruses can undergo lytic infection, where activity is identified by progeny viruses, or lysogenic infection, where activity is difficult to assess. A temperate virus that is 446 447 undergoing lysogenic infection (present as a provirus) during a heavy-water SIP incubation 448 would become isotopically enriched and depend on its host's division rate for abundance. Active 449 proviruses undergoing lysogenic infection need to be distinguished from viruses undergoing lytic 450 infection because the effect of proviruses on the host metabolism (and therefore ecosystem) will 451 not be as pronounced. This is because proviruses do not shut down and redirect host metabolism 452 for progeny production during the lysogenic cycle as compared to the lytic cycle. During

453 lysogeny, viruses still impact host metabolism via host gene regulation and acquisition of new 454 virulence factors, but it is primarily for maintaining the provirus in the host genome [99]. 455 Proviruses may also be labeled, but not currently active if they are proviral remnants of a past 456 infection [100]. These remnants have no negative impact on host metabolism (beyond occupying 457 genome space) but may confer some advantage as a gene transfer agent [51] or by contributing 458 virulence factors which can still impact host physiology and metabolism [99]. SIP-enabled 459 metagenomics alone cannot unequivocally identify a virus' state (e.g., in maintenance mode or 460 remnant), making it currently difficult to fully assess viral activity. Combining SIP-461 metagenomics with other approaches—such as a SIP-virome or induction assays—may identify 462 vOTUs that have undergone lytic infection, and therefore provide a more holistic view of vOTU 463 dynamics.

464 **Conclusions**

465 Winter carbon losses in Arctic peat soils are estimated to be significant and growing, but the mechanisms that drive these losses are poorly understood. Using an ¹⁸O-heavy-water 466 467 incubation under subzero, anoxic conditions, we found that bacteria and their viruses are active 468 over the long winter months in northern peatlands, and these active populations drive significant 469 CO₂ fluxes. Our approach, SIP-targeted metagenomics, allowed us to move beyond a general 470 catalogue of the genetic repertoire of these soil communities, and expose the specific population-471 level dynamics and functional capacities of the activity viral community. Given the high 472 abundance of unlabeled bacteria and their viruses, these bacterial and viral populations would 473 have been traditionally described first and foremost, thus occluding the characterization of the 474 active bacteria and viruses in these soils. Despite the subzero temperatures and lack of oxygen in

475 these peat soils, many active hosts and active viruses (both temperate and lytic) appear to be 476 engaged in a surprisingly high level of biotic interactions and biogeochemical processing.

477 Acknowledgements

478 We would like to thank the Millard lab for curating a list of Genbank bacteriophage genomes 479 used for taxonomic identification of our vOTUs. Thanks to Jack McFarland and Monica Haw 480 (USGS) for help and support with the experimental setup and maintenance. Sample collection 481 and processing was supported by a US Geological Survey Mendenhall Fellowship to S.J.B., the 482 Bonanza Creek LTER Program, jointly funded by NSF (DEB 1026415) and the USDA Forest 483 Service Pacific Northwest Research Station (PNW01-JV112619320-16), and support from the 484 USGS Climate R&D Program and AK Climate Science Center. Analysis of results was 485 supported by a Lawrence Livermore National Laboratory, Laboratory Directed Research & 486 Development grant (18-ERD-041) to S.J.B. and by LLNL's U.S. Department of Energy, Office 487 of Biological and Environmental Research, Genomic Science Program LLNL 'Microbes Persist' 488 Scientific Focus Area (#SCW1632). Work conducted at PNNL, operated for the DOE by Battelle 489 Memorial Institute, was conducted under Contract DE-AC05-76RLO1830. Work conducted at 490 LLNL was conducted under the auspices of the US Department of Energy under Contract DE-491 AC52-07NA27344.

492

493

494

495

496

497

498		
499		
500		
501		
502		
503		
504		
505		
506		
507		
508	1.	Nichols, J.E. and D.M. Peteet, Rapid expansion of northern peatlands and doubled
509		estimate of carbon storage. Nature Geoscience, 2019. 12(11): p. 917-921.
510	2.	Koven, C.D., et al., A simplified, data-constrained approach to estimate the permafrost
511		carbon-climate feedback. Philos Trans A Math Phys Eng Sci, 2015. 373(2054).
512	3.	McGuire, A.D., et al., Dependence of the evolution of carbon dynamics in the northern
513		permafrost region on the trajectory of climate change. Proc Natl Acad Sci U S A, 2018.
514		115(15): p. 3882-3887.
515	4.	Schuur, E.A., et al., Climate change and the permafrost carbon feedback. Nature, 2015.
516		520(7546): p. 171-9.
517	5.	Hodgkins, S.B., et al., Changes in peat chemistry associated with permafrost thaw
518		increase greenhouse gas production. Proc Natl Acad Sci U S A, 2014. 111(16): p. 5819-24.
519	6.	Hopple, A.M., et al., Massive peatland carbon banks vulnerable to rising temperatures.
520		Nat Commun, 2020. 11(1): p. 2373.

- 521 7. Parazoo, N.C., et al., *Detecting regional patterns of changing CO2 flux in Alaska*. Proc
- 522 Natl Acad Sci U S A, 2016. 113(28): p. 7733-8.
- 523 8. Wilson, R.M., et al., *Stability of peatland carbon to rising temperatures*. Nat Commun,
- 524 **2016**. 7: p. 13723.
- 525 9. Natali, S.M., et al., Large loss of CO2 in winter observed across the northern permafrost
- 526 *region.* Nature Climate Change, 2019. 9(11): p. 852-857.
- 527 10. Przybylak, R., Diurnal temperature range in the Arctic and its relation to hemispheric and
- 528 Arctic circulation patterns. International Journal of Climatology, 2000. 20(3 %@ 0899-
- 529 8418): p. 231-253.
- 530 11. Bolduc, B., et al., *The IsoGenie database: an interdisciplinary data management solution*531 *for ecosystems biology and environmental research.* PeerJ, 2020. 8: p. e9467.
- 532 12. Männistö, M.K., et al., Acidobacteriadominate the active bacterial communities of Arctic
- 533 tundra with widely divergent winter-time snow accumulation and soil temperatures.
- 534 FEMS Microbiology Ecology, 2013. 84(1): p. 47-59.
- 535 13. Murray, A.E., et al., *Microbial life at -13 C in the brine of an ice-sealed Antarctic lake*.
- 536 Proceedings of the National Academy of Sciences, 2012. 109(50): p. 20626-20631.
- 537 14. Drotz, S.H., et al., Both catabolic and anabolic heterotrophic microbial activity proceed in
- *frozen soils.* Proceedings of the National Academy of Sciences, 2010. 107(49): p. 2104621051.
- 540 15. Margesin, R. and T. Collins, *Microbial ecology of the cryosphere (glacial and permafrost*
- 541 *habitats): current knowledge.* Applied Microbiology and Biotechnology, 2019. 103(6): p.
- 542 2537-2549.

543	16.	Brum, J.R. and M.B. Sullivan, Rising to the challenge: accelerated pace of discovery
544		transforms marine virology. Nature Reviews Microbiology, 2015. 13(3): p. 147-159.
545	17.	Wilhelm, S.W. and C.A. Suttle, Viruses and Nutrient Cycles in the Sea. BioScience, 1999.
546		49(10): p. 781-788.
547	18.	Suttle, C.A., Marine viruses — major players in the global ecosystem. Nature Reviews
548		Microbiology, 2007. 5(10): p. 801-812.
549	19.	Trubl, G., et al., Coming-of-Age Characterization of Soil Viruses: A User's Guide to Virus
550		Isolation, Detection within Metagenomes, and Viromics. Soil Systems, 2020. 4(2): p. 23.
551	20.	Braga, L.P.P., et al., Impact of phages on soil bacterial communities and nitrogen
552		availability under different assembly scenarios. Microbiome, 2020. 8(1).
553	21.	Liang, C., J.P. Schimel, and J.D. Jastrow, The importance of anabolism in microbial control
554		over soil carbon storage. Nature Microbiology, 2017. 2(8): p. 17105.
555	22.	Rosenwasser, S., et al., Virocell Metabolism: Metabolic Innovations During Host–Virus
556		Interactions in the Ocean. Trends in Microbiology, 2016. 24(10): p. 821-832.
557	23.	Thompson, L.R., et al., Phage auxiliary metabolic genes and the redirection of
558		cyanobacterial host carbon metabolism. Proceedings of the National Academy of
559		Sciences, 2011. 108(39): p. E757-E764.
560	24.	Zimmerman, A.E., et al., Metabolic and biogeochemical consequences of viral infection in
561		aquatic ecosystems. Nature Reviews Microbiology, 2020. 18(1): p. 21-34.
562	25.	Emerson, J.B., et al., Host-linked soil viral ecology along a permafrost thaw gradient.
563		Nature Microbiology, 2018. 3(8): p. 870-880.

- 564 26. Trubl, G., et al., Soil Viruses Are Underexplored Players in Ecosystem Carbon Processing.
 565 mSystems, 2018. 3(5).
- 566 27. Barnett, S.E. and D.H. Buckley, *Simulating metagenomic stable isotope probing datasets*
- 567 *with MetaSIPSim.* BMC Bioinformatics, 2020. 21(1).
- 568 28. Dumont, M.G. and M. Hernández García, *Stable Isotope Probing*. Methods and Protocols
- 569 Totowa, NJ, US: Humana Press, 2019.
- 570 29. Starr, E.P., et al., Stable isotope informed genome-resolved metagenomics reveals that
- 571 Saccharibacteria utilize microbially-processed plant-derived carbon. Microbiome, 2018.
- 572 6(1).
- 573 30. Haig, S.-J., et al., *Stable-isotope probing and metagenomics reveal predation by protozoa* 574 *drives E. coli removal in slow sand filters*. The ISME Journal, 2015. 9(4): p. 797-808.
- 575 31. Gross, A., et al., The role of soil redox conditions in microbial phosphorus cycling in
- 576 *humid tropical forests.* Ecology, 2020. 101(2).
- 577 32. Aanderud, Z.T. and J.T. Lennon, *Validation of Heavy-Water Stable Isotope Probing for*
- 578 the Characterization of Rapidly Responding Soil Bacteria. Applied and Environmental
- 579 Microbiology, 2011. 77(13): p. 4589-4596.
- 580 33. Blazewicz, S.J. and E. Schwartz, *Dynamics of 180 Incorporation from H 2 180 into Soil*581 *Microbial DNA*. Microbial Ecology, 2011. 61(4): p. 911-916.
- 582 34. Schwartz, E., Characterization of Growing Microorganisms in Soil by Stable Isotope
- 583 Probing with H2180. Applied and Environmental Microbiology, 2007. 73(8): p. 2541-
- 584 2546.

- 585 35. Blazewicz, S.J., et al., *Taxon-specific microbial growth and mortality patterns reveal*
- 586 distinct temporal population responses to rewetting in a California grassland soil. The
- 587 ISME Journal, 2020. 14(6): p. 1520-1532.
- 588 36. Koch, B.J., et al., *Estimating taxon-specific population dynamics in diverse microbial*
- 589 *communities.* Ecosphere, 2018. 9(1): p. e02090.
- 590 37. Papp, K., et al., *Quantitative stable isotope probing with H2180 reveals that most*
- 591 bacterial taxa in soil synthesize new ribosomal RNA. The ISME Journal, 2018. 12(12): p.
- *5*92 **3043-3045**.
- 593 38. Schwartz, E., et al., *Characterization of growing bacterial populations in McMurdo Dry*
- 594 Valley soils through stable isotope probing with 180-water. FEMS Microbiology Ecology,
- 595 2014. 89(2): p. 415-425.
- 596 39. Waldrop, M.P., et al., Carbon Fluxes and Microbial Activities from Boreal Peatlands
- 597 *Experiencing Permafrost Thaw.* Journal of Geophysical Research: Biogeosciences, 2021.
- 598 40. Russell, D.W. and J. Sambrook, *Molecular cloning: a laboratory manual*. Vol. 1. 2001:
- 599 Cold Spring Harbor Laboratory Cold Spring Harbor, NY.
- 600 41. Bankevich, A., et al., SPAdes: A New Genome Assembly Algorithm and Its Applications to
- 601 Single-Cell Sequencing. Journal of Computational Biology, 2012. 19(5): p. 455-477.
- 602 42. Roux, S., et al., *VirSorter: mining viral signal from microbial genomic data*. PeerJ, 2015.
- 603 3: p. e985.
- 604 43. Ren, J., et al., *Identifying viruses from metagenomic data using deep learning*.
- 605 Quantitative Biology, 2020. 8(1): p. 64-77.

- 606 44. Roux, S., et al., Minimum Information about an Uncultivated Virus Genome (MIUViG).
- 607 Nature Biotechnology, 2019. 37(1): p. 29-37.
- 608 45. Delcher, A.L., S.L. Salzberg, and A.M. Phillippy, Using MUMmer to identify similar regions
- 609 *in large sequence sets.* Current protocols in bioinformatics, 2003(1): p. 10.3. 1-10.3. 18.
- 610 46. Nayfach, S., et al., *CheckV assesses the quality and completeness of metagenome-*
- 611 *assembled viral genomes.* Nature Biotechnology, 2020.
- 612 47. Langmead, B. and S.L. Salzberg, *Fast gapped-read alignment with Bowtie 2*. Nature
- 613 Methods, 2012. 9(4): p. 357-359.
- 614 48. Roux, S., et al., Benchmarking viromics: an in silico evaluation of metagenome-enabled
- 615 *estimates of viral community composition and diversity.* PeerJ, 2017. 5: p. e3817.
- 616 49. Ecale Zhou, C.L., et al., *multiPhATE: bioinformatics pipeline for functional annotation of*
- 617 *phage isolates.* Bioinformatics, 2019. 35(21): p. 4402-4404.
- 618 50. Shaffer, M., et al., DRAM for distilling microbial metabolism to automate the curation of
- 619 *microbiome function*. Nucleic Acids Research, 2020. 48(16): p. 8883-8900.
- 620 51. Howard-Varona, C., et al., Lysogeny in nature: mechanisms, impact and ecology of
- 621 *temperate phages.* The ISME Journal, 2017. 11(7): p. 1511-1520.
- 622 52. Arndt, D., et al., *PHASTER: a better, faster version of the PHAST phage search tool.*623 Nucleic Acids Research, 2016. 44(W1): p. W16-W21.
- 624 53. Li, D., et al., *MEGAHIT: an ultra-fast single-node solution for large and complex*
- 625 *metagenomics assembly via succinct de Bruijn graph.* Bioinformatics, 2015. 31(10): p.
- 626 1674-1676.

- 627 54. Alneberg, J., et al., Binning metagenomic contigs by coverage and composition. Nature
- 628 Methods, 2014. 11(11): p. 1144-1146.
- 629 55. Wu, Y.-W., B.A. Simmons, and S.W. Singer, MaxBin 2.0: an automated binning algorithm
- 630 to recover genomes from multiple metagenomic datasets. Bioinformatics, 2016. 32(4): p.
- 631 605-607.
- 632 56. Kang, D.D., et al., *MetaBAT, an efficient tool for accurately reconstructing single*
- 633 genomes from complex microbial communities. PeerJ, 2015. 3: p. e1165.
- 634 57. Uritskiy, G.V., J. Diruggiero, and J. Taylor, *MetaWRAP—a flexible pipeline for genome-*
- 635 *resolved metagenomic data analysis.* Microbiome, 2018. 6(1).
- 636 58. Parks, D.H., et al., CheckM: assessing the quality of microbial genomes recovered from
- 637 *isolates, single cells, and metagenomes.* Genome Research, 2015. 25(7): p. 1043-1055.
- 638 59. Bowers, R.M., et al., *Minimum information about a single amplified genome (MISAG)*
- 639 and a metagenome-assembled genome (MIMAG) of bacteria and archaea. Nature
- 640 Biotechnology, 2017. 35(8): p. 725-731.
- 641 60. Bushnell, B., *BBDuk: Adapter.* Quality Trimming and Filtering.
- 642 httpssourceforgenetprojectsbbmap.
- 643 61. Olm, M.R., et al., *dRep: a tool for fast and accurate genomic comparisons that enables*
- 644 *improved genome recovery from metagenomes through de-replication.* The ISME
- 645 Journal, 2017. 11(12): p. 2864-2868.
- 646 62. Chaumeil, P.-A., et al., GTDB-Tk: a toolkit to classify genomes with the Genome
- 647 *Taxonomy Database*. Bioinformatics, 2019.

648	63.	Wattam, A.R., et al., Improvements to PATRIC, the all-bacterial Bioinformatics Database
-----	-----	---

- 649 *and Analysis Resource Center*. Nucleic Acids Research, 2017. 45(D1): p. D535-D542.
- 650 64. Brettin, T., et al., *RASTtk: A modular and extensible implementation of the RAST*
- 651 algorithm for building custom annotation pipelines and annotating batches of genomes.
- 652 Scientific Reports, 2015. 5(1): p. 8365.
- 653 65. Aramaki, T., et al., *KofamKOALA: KEGG Ortholog assignment based on profile HMM and*
- 654 *adaptive score threshold.* Bioinformatics, 2020. 36(7): p. 2251-2252.
- 655 66. Kanehisa, M., et al., *KEGG as a reference resource for gene and protein annotation*.
- 656 Nucleic Acids Research, 2016. 44(D1): p. D457-D462.
- 657 67. Bushnell, B., *BBMap: a fast, accurate, splice-aware aligner*. 2014, Lawrence Berkeley
 658 National Lab.(LBNL), Berkeley, CA (United States).
- 659 68. Kumar, M.S., et al., Analysis and correction of compositional bias in sparse sequencing
 660 count data. BMC Genomics, 2018. 19(1).
- 661 69. Paulson, J.N., et al., Differential abundance analysis for microbial marker-gene surveys.
- 662 Nature Methods, 2013. 10(12): p. 1200-1202.
- 663 70. Andersson, A.F. and J.F. Banfield, Virus Population Dynamics and Acquired Virus
- 664 *Resistance in Natural Microbial Communities.* Science, 2008. 320(5879): p. 1047-1050.
- 665 71. Bland, C., et al., CRISPR Recognition Tool (CRT): a tool for automatic detection of
- *clustered regularly interspaced palindromic repeats.* BMC Bioinformatics, 2007. 8(1): p.
 209.
- 668 72. Altschul, S.F., et al., *Basic local alignment search tool*. Journal of Molecular Biology,
- 669 1990. 215(3): p. 403-410.

670	73.	Edwards, R.A., et al., Computational approaches to predict bacteriophage-host		
671		relationships. FEMS Microbiology Reviews, 2016. 40(2): p. 258-272.		
672	74.	Al-Shayeb, B., et al., Clades of huge phages from across Earth's ecosystems. Nature,		
673		2020. 578(7795): p. 425-431.		
674	75.	Nikrad, M.P., L.J. Kerkhof, and M.M. Häggblom, The subzero microbiome: microbial		
675		activity in frozen and thawing soils. FEMS Microbiology Ecology, 2016. 92(6): p. fiw081.		
676	76.	Adriaenssens, E.M., et al., Environmental drivers of viral community composition in		
677		Antarctic soils identified by viromics. Microbiome, 2017. 5(1).		
678	77.	Zablocki, O., E.M. Adriaenssens, and D. Cowan, Diversity and Ecology of Viruses in		
679		Hyperarid Desert Soils. Applied and Environmental Microbiology, 2016. 82(3): p. 770-		
680		777.		
681	78.	Breitbart, M., et al., Phage puppet masters of the marine microbial realm. Nature		
682		Microbiology, 2018. 3(7): p. 754-766.		
683	79.	Fermin, G., Host Range, Host–Virus Interactions, and Virus Transmission. Viruses, 2018:		
684		p. 101.		
685	80.	Malki, K., et al., Bacteriophages isolated from Lake Michigan demonstrate broad host-		
686		range across several bacterial phyla. Virology Journal, 2015. 12(1).		
687	81.	Van Goethem, M.W., et al., Characteristics of Wetting-Induced Bacteriophage Blooms in		
688		Biological Soil Crust. mBio, 2019. 10(6).		
689	82.	Greiner, T., et al., Genes for membrane transport proteins: Not so rare in viruses.		
690		Viruses, 2018. 10(9): p. 456.		

691	83.	Ankrah, N.Y.D., et al., Phage infection of an environmentally relevant marine bacterium
692		alters host metabolism and lysate composition. The ISME Journal, 2014. 8(5): p. 1089-
693		1100.
694	84.	Markine-Goriaynoff, N., et al., <i>Glycosyltransferases encoded by viruses</i> . Journal of
695		General Virology, 2004. 85(10): p. 2741-2754.
696	85.	Thingstad, T.F., Elements of a theory for the mechanisms controlling abundance,
697		diversity, and biogeochemical role of lytic bacterial viruses in aquatic systems. Limnology
698		and Oceanography, 2000. 45(6): p. 1320-1328.
699	86.	Zeibich, L., O. Schmidt, and H.L. Drake, Dietary polysaccharides: fermentation potentials
700		of a primitive gut ecosystem. Environmental Microbiology, 2019. 21(4): p. 1436-1451.
701	87.	Lapébie, P., et al., Bacteroidetes use thousands of enzyme combinations to break down
702		glycans. Nature Communications, 2019. 10(1).
703	88.	Larsbrink, J. and L.S. McKee, Bacteroidetes bacteria in the soil: Glycan acquisition,
704		enzyme secretion, and gliding motility. Advances in applied microbiology, 2020. 110: p.
705		63-98.
706	89.	Lemos, L.N., et al., Genomic signatures and co-occurrence patterns of the ultra-small
707		Saccharimonadia (phylum CPR/Patescibacteria) suggest a symbiotic lifestyle. Molecular
708		Ecology, 2019. 28(18): p. 4259-4271.
709	90.	Tian, R., et al., Small and mighty: adaptation of superphylum Patescibacteria to
710		groundwater environment drives their genome simplicity. Microbiome, 2020. 8(1).

711	91.	Chen, LX., et al., Candidate Phyla Radiation Roizmanbacteria From Hot Springs Have
712		Novel and Unexpectedly Abundant CRISPR-Cas Systems. Frontiers in Microbiology, 2019.
713		10.
714	92.	Coutinho, F.H., et al., Ecogenomics and metabolic potential of the South Atlantic Ocean
715		microbiome. Science of The Total Environment, 2021. 765: p. 142758.
716	93.	Bashiardes, S., G. Zilberman-Schapira, and E. Elinav, Use of Metatranscriptomics in
717		Microbiome Research. Bioinformatics and Biology Insights, 2016. 10: p. BBI.S34610.
718	94.	Blazewicz, S.J., et al., Evaluating rRNA as an indicator of microbial activity in
719		environmental communities: limitations and uses. The ISME Journal, 2013. 7(11): p.
720		2061-2068.
721	95.	Vogel, C. and E.M. Marcotte, Insights into the regulation of protein abundance from
722		proteomic and transcriptomic analyses. Nature Reviews Genetics, 2012. 13(4): p. 227-
723		232.
724	96.	Joergensen, R.G. and F. Wichern, Alive and kicking: Why dormant soil microorganisms
725		matter. Soil Biology and Biochemistry, 2018. 116: p. 419-430.
726	97.	Stewart, F.M. and B.R. Levin, The population biology of bacterial viruses: Why be
727		temperate. Theoretical Population Biology, 1984. 26(1): p. 93-117.
728	98.	Williamson, K.E., et al., Incidence of lysogeny within temperate and extreme soil
729		environments. Environmental Microbiology, 2007. 9(10): p. 2563-2574.
730	99.	Hernández, S. and M.J. Vives, Phages in Anaerobic Systems. Viruses, 2020. 12(10): p.
731		1091.

732	100.	Casjens, S., Prophages and bacterial genomics: what have we learned so far? Molecular
733		Microbiology, 2003. 49(2): p. 277-300.
734		
735		
736		
737		
738		
739		
740		
741		
742		
743		
744	Table	1. Linkages between ¹⁸ O labeled and unlabeled viruses and bacteria in a SIP
745	incuba	ntion of sub-zero, anoxic Arctic peat soils. Viral operational taxonomic units (vOTUs)
746	were li	nked to metagenome assembled genomes (MAGs) via nucleotide similarity and using
747	CRISP	PR spacers. 'Active' vOTUs and MAGs were defined based on assimilation of ^{18}O
748	enriche	ed water (heavy water) into DNA. In total, there were 814 linkages from active vOTUs to
749	active	MAGs (group 1), active vOTUs to unlabeled MAGs (group 2), unlabeled vOTUs to active
750	MAGs	(group 3) and unlabeled vOTUs to unlabeled MAGs (group 4).

Group	Description	No. of vOTU- MAG matches	Unique vOTU- MAG matches	vOTUs	MAGs
1	Active vOTU-active MAG	164	107	74	37
2	Active vOTU-unlabeled MAG	2	2	2	2
3	Unlabeled vOTU-active MAG	226	77	44	11
4	Unlabeled vOTU-unlabeled MAG	422	132	58	30
	Total	814	318	178	80

752
753
754
755
756
757
758
759
760
761
762
763



Figure 1. Assessment of viral community structure and activity after a ¹⁸O water incubation in Arctic peat soils. (A) Viral 765 sequences were identified from 23 samples grouped by two treatments (natural abundance water " $H_2^{16}O$ ", and heavy water " $H_2^{18}O$ ") 766 767 and two time points: 184 days and 370 days. Number of replicates is indicated in parentheses. Relative abundances of all 332 vOTUs 768 identified in the peat soils, clustered by abundances in each treatment/timepoint. (B) Relative abundances of 243 vOTUs considered 'active' due to DNA ¹⁸O enrichment patterns. Relative abundance for each vOTU (illustrated by blue gradient) was normalized by 769 770 metagenome size (total base pairs) and contig length, and reads were mapped to the contig if they shared \geq 90% average nucleotide identity and had \geq 85% alignment fraction. (C) Diversity metrics for the 243 active vOTUs. Box plots show the median, upper and 771 772 lower quartile range, and the variance among the samples. Asterix denotes significance (p < 0.05).



774 votus
 775 Figure 2. Putative auxiliary metabolic genes associated with peat soil viral genomes.

776 Heatmap of 31 vOTUs carrying confidently predicted auxiliary metabolic genes (AMGs) (Dram-

v score 1–3) and their annotation. The vOTUs are grouped by active or unlabeled (see Table S3

for additional detail) with the sum of AMGs per vOTU indicated in parenthesis and those with an

asterisk are linked to an active MAG. AMGs are grouped by functional category — carbon

- 780 utilization, energy generation, organic nitrogen use, transporters, and miscellaneous.
- 781
- 782
- 783
- 784
- 785





Figure 3. CO₂ production. Cumulative CO₂ production in soil incubated at \Box 1.5°C

(experimental conditions) and $\Box 20^{\circ}$ C (control). Error bars show standard error (n=3) and R² is

- shown for each linear regression.







809





811 **Figure 5. Abundances of active viruses and their predicted bacterial hosts.** Average

812 virus:host abundance ratios for bacterial phyla Bacteroidota (yellow) and Firmicutes (green)

```
813 from heavy-water treatment samples at 184 (n=6) and 370 days (n=6). Host abundance and the
```

814 abundance of viruses for that host were calculated as the mean coverage depth from

- 815 metagenomic read mapping, normalized by the number of reads in the sample.
- 816
- 817
- 818