

Revised Manuscript [2021-04-19]

NEW PERSPECTIVES ON THE CALCULATION OF BIOACCUMULATION METRICS FOR ACTIVE SUBSTANCES IN LIVING ORGANISMS

Aude Ratier^a, Christelle Lopes^a, Gauthier Multari^a, Vanessa Mazerolles^b, Patrice Carpentier^b,

Sandrine Charles^{1,a}

^aUniversité de Lyon, Université Lyon 1, CNRS UMR5558, Laboratoire de Biométrie et Biologie Evolutive,
69100 Villeurbanne, France.

^bAnses, 14 rue Pierre et Marie Curie, 94701 Maisons-Alfort Cedex, France.

Acknowledgement

The authors are thankful to ANSES for providing the financial support. The MOSAIC_{bioacc} web tool is hosted at the Rhône-Alpes Bioinformatics Center PRABI (PRABI, 2020). This work benefited from the French GDR “Aquatic Ecotoxicology” framework which aims at fostering stimulating scientific discussions and collaborations for more integrative approaches. This work is part of the ANR project APPROVe (ANR-18-CE34-0013) for an integrated approach to propose proteomics for biomonitoring: accumulation, fate and multi-markers (<https://anr.fr/Projet-ANR-18-CE34-0013>). This work was also made under the umbrella of the Graduate School H2O’Lyon (ANR-17-EURE-0018) and “Université de Lyon” (UdL), as part of the program “Investissements d’Avenir” run by “Agence Nationale de la Recherche” (ANR). The authors are truly grateful to the anonymous colleagues who participated in testing MOSAIC_{bioacc} and for giving their feedback. The authors also are grateful for Benoît BRET for creating the logo for MOSAIC_{bioacc}. The authors declare no competing interests.

Data availability

Data are accessible directly within MOSAIC_{bioacc} at <https://mosaic.univ-lyon1.fr/bioacc>.

¹ Corresponding Author

Email address: sandrine.charles@univ-lyon1.fr (Sandrine Charles)

Revised Manuscript [2021-04-19]

NEW PERSPECTIVES ON THE CALCULATION OF BIOACCUMULATION METRICS FOR ACTIVE SUBSTANCES IN LIVING ORGANISMS

ABSTRACT

Today, only few ready-to-use and convenient decision-making tools are available in ecotoxicology concerning accumulation and effects of chemical substances on organisms, accounting for exposure situations that are known to be complex (routes of exposure, metabolism, mixtures, etc.). This paper presents new perspectives on the generic calculation of bioaccumulation metrics via the innovative web tool MOSAIC_{bioacc} (<http://mosaic.univ-lyon1.fr/bioacc>). MOSAIC_{bioacc} provides all kind of bioaccumulation metrics associated with their uncertainty whatever the species-compound combination. MOSAIC_{bioacc} expects accumulation-depuration data as inputs, even with complex exposure and clearance patterns, to quickly perform their relevant analysis. MOSAIC_{bioacc} intends to facilitate the daily work of regulators, or any ecotoxicologist, who will freely benefit from a user-friendly on-line interface that automatically fits toxicokinetic models without needs for users to invest in the technical aspects to get bioaccumulation metrics estimates. MOSAIC_{bioacc} also provides all results in a fully transparent way to ensure reproducibility.

KEYWORDS

Environmental Risk Assessment - Toxicokinetic models - Bayesian inference - Uncertainties – User-friendly web platform

Revised Manuscript [2021-04-19]

INTRODUCTION

Faced with the current environmental challenges, linked in particular to environmental pollution, ecotoxicology must today provide relevant and effective decision-making tools regarding bioaccumulation and effects of chemical substances on living organisms. Such tools must account for various exposure situations, environmentally realistic but complex (*e.g.*, several routes of exposure, metabolism of substances, mixtures, etc.). Among available methods, toxicokinetic-toxicodynamic (TKTD) models are now strongly recommended to describe the link between exposure concentrations and effects on individual life-history traits over time from experimental data collected through toxicity tests, even standard ones (EFSA PPR Panel 2018). More specifically, the TK part of these models is used to relate the exposure concentration to the time course of the internal concentration within organisms, considering various processes such as accumulation, depuration, metabolization and excretion (known as ADME processes). As some recent regulations, the EU regulation No 283/2013 for plant protection products in marketing authorisation applications requires for example a bioaccumulation test on fish according to OECD Test guideline 305 (OECD 2012), which consists in an accumulation phase followed by a depuration phase. During the accumulation phase, fish are exposed to a substance of interest at a range of concentrations, chosen according to the assumed mode of action of the substance. After a certain time period fixed by the experimenter, organisms are transferred to a clean medium for a depuration phase. The concentration of the substance (and of its potential metabolites) is followed within fish at regular time points during both phases leading *in fine* to the estimation of bioaccumulation metrics. In this paper, for the sake of generality, we chose the generic expression “bioaccumulation metrics” to denote either bioconcentration factors (BCF) used when exposure is via water, biota-sediment accumulation factors (BSAF) when exposure is via sediment or biomagnification factors (BMF) when exposure is via food. Bioaccumulation metrics appeared to us as the best compromise regarding the wide diversity of terms used in the scientific

Revised Manuscript [2021-04-19]

literature (for example, USEPA (1994), Gobas et al. (2009) and Burkhard et al. (2012)). Nevertheless, preference is often given to experimentally derived BCF estimates to be used for secondary poisoning assessment under Biocidal Products Regulation (European Commission 2012). Bioaccumulation tests are of course not only limited to fish, even if a test according to OECD Test guideline 305 (OECD 2012) is preferred when experimental information on bioaccumulation is needed for PBT/vPvB assessment under REACH regulation (ECHA 2017; European Commission 2006). Consequently, any other species can be used, such as benthic invertebrates, terrestrial oligochaetes or birds, depending on the substance under consideration. From a regulatory point of view, bioaccumulation metrics are key decision criteria used to evaluate concentrations of active substances in food items of vertebrates (especially piscivorous birds and mammals), making the estimation of these metrics with the most precision as possible a highly crucial methodological challenge.

All bioaccumulation metrics rely on estimates of kinetic parameters as involved in toxicokinetic (TK) models. In the past decades, many types of methods have been proposed to get these estimates from simple TK models, most of them providing BCF estimates separately considering the kinetics for both accumulation and depuration phases as observed in dedicated experiments (OECD 2012). Nevertheless, TK model parameters are known to be highly correlated, so that separating their estimation prevents to account for a mutual influence on their uncertainty. Moreover, kinetic parameter estimates are only rarely provided with their uncertainty, although this is now expected by the regulatory bodies (EFSA Scientific Committee 2018). Consequently, concomitantly to the above-mentioned challenge, environmental risk assessment could be improved if complete tools allowing for a simultaneously estimation of all TK model parameters associated with their uncertainty would be available in support of stakeholders that need to fulfil regulatory expectations. These tools must also be easy to use to overcome the scepticism of regulators who are faced with multiple TK models and implementation methods, while thinking about their standardization at the same

Revised Manuscript [2021-04-19]

time (Tan et al. 2020). An R-code was first proposed in 2016 (Aldenberg 2019) that allowed to analyse data collected only from the OCED test guideline 305. In the same line of thought, a spreadsheet was recently proposed by Gobas et al. (2020). Our paper goes beyond by considering all type of species-compound accumulation-depuration data (not only those of the OECD test guideline 305) which analysis leads to one or several bioaccumulation metrics of interest.

Ratier et al. (2019) recently proposed a full revisit of the TK modelling approach based on a unified inference method to estimate parameters of TK models for both accumulation and depuration phases, simultaneously, automatically associating the uncertainties. This innovative framework has been thought to make it possible to further incorporate the TK part into complete TKTD models. Benefiting from this innovation, we present new perspectives for a facilitated calculation of any type of bioaccumulation metrics (such as BCF/BSAF/BMF) thanks to the new ready-to-use statistical web tool MOSAIC_{bioacc} (<http://mosaic.univ-lyon1.fr/bioacc>). MOSAIC_{bioacc} runs one-compartment TK models that are automatically designed according to the input data. MOSAIC_{bioacc} leads to bioaccumulation metrics associated with their uncertainty propagated from the kinetic parameter estimates, without the need for users to invest underlying technical aspects. MOSAIC_{bioacc} is free of use, fully integrated within the all-in-one facility MOSAIC itself (<http://mosaic.univ-lyon1.fr>). MOSAIC_{bioacc} is regularly updated to always offers the very latest conceptual advances related to TK models. Today, MOSAIC_{bioacc} allows accounting for several exposure routes (water, pore water, sediment and food), for metabolism of chemicals (if the input experimental data include measurements for both the parent chemical and its metabolites), and for potential growth of organisms (if growth measurements are included within the data set). The use of MOSAIC_{bioacc} only requires users to upload their experimental data, collected via standard protocols or from home-made experimental designs. Priority was first given to the calculation of BCF/BSAF/BMF because they are the widely used bioaccumulation metrics in the current regulatory guidelines. MOSAIC_{bioacc} get their estimate

Revised Manuscript [2021-04-19]

as probability distributions that are summarized for users by the median (50th centile of the distribution) and the 95% credible interval (delimited by the 2.5th and 97.5th centiles) quantifying the uncertainty. In addition, the fitting plots and all model parameter estimates are provided, followed by a collection of goodness-of-fit criteria allowing users to check the relevance of the results. All outputs can be downloaded under different formats for further inclusion into any home-made document (in particular the open-source programming code), and a full report can also be directly downloaded gathering everything that is displayed to users on the web page. These two latter features of MOSAIC_{bioacc} guarantee both reproducibility and transparency of underlying calculations.

MATERIALS AND METHODS

MOSAIC_{bioacc} is part of the web platform MOSAIC (<https://mosaic.univ-lyon1.fr/>, Charles et al. 2018). It was developed as a Shiny environment (Chang et al. 2020), available at <https://mosaic.univ-lyon1.fr/bioacc> and hosted at the Rhône-Alpes Bioinformatics Center PRABI (PRABI 2020). A user guide and an explanatory video are immediately available in the introductory section of the application to fully assist users step by step in appropriating the tool and its features. Details on underlying ordinary differential equations and their solving are also provided within a detailed user guide (see supplemental data, **Annex 1**), thus ensuring all required transparency as recommended by EFSA for a good modelling practice (EFSA PPR Panel 2014).

Data uploading

When using MOSAIC_{bioacc}, the first step is to upload input data (**Fig. 1-a**). MOSAIC_{bioacc} expects to receive experimental exposure time-course data, including at least an accumulation phase, as a .txt file or a .csv file (comma, semicolon or tabular separator) with a specific format. Each line of the table corresponds to one time point for a given replicate and a given exposure concentration of the contaminant. The data table must contain at least four columns, with the

Revised Manuscript [2021-04-19]

exact following headers, the order of columns being not important (**Table 1**): ‘time’ (the time point of the measurement at the exposure concentration, in hours, minutes, days or weeks); ‘expw’, ‘exppw’, ‘exps’ or ‘expf’: the exposure concentration of the contaminant in the medium, that is water, pore water, sediment or food, respectively, all expressed in $\mu\text{g.mL}^{-1}$ or in $\mu\text{g.g}^{-1}$; ‘replicate’ (a number or a character that is unique for each replicate, dimensionless); and ‘conc’ (the concentration of the contaminant, and of its potential metabolites, measured within organisms, in $\mu\text{g.g}^{-1}$). According to the experimental design, further columns can be added in the data file: ‘expw’, ‘exppw’, ‘exps’ and/or ‘expf’ if several exposure routes are considered together, ‘concm ℓ ’ (the concentrations of metabolite ℓ derived from the parent compound within the organisms, *e.g.*, concm1, concm2, ...), and ‘growth’ (if growth measurements of the organisms are available).

As shown on **Fig. 1-a**, MOSAIC_{bioacc} users can either upload their own data set with a click on ‘Browse’ (by taking care about the expected format specification) or try MOSAIC_{bioacc} with example data sets (in total six data sets are proposed, each with different characteristics). When the upload is complete, users must manually select the appropriate separator, the time unit and the duration of the accumulation phase; please note that, when using example data sets, these fields are automatically filled in.

MOSAIC_{bioacc} first provides a table with the raw data allowing users to check if the data were correctly entered. Users can also visualize a plot with the raw data (**Fig. 1-b**). In case data were collected for several exposure concentrations and if the users have uploaded all, one must be chosen for the MOSAIC_{bioacc} analysis. Note that only one file at a time can be analysed. Also, when another exposure concentration from the same data file is chosen, the duration of the accumulation phase is reset and framed in orange to invite users for update before to launch new calculations. Example files provided in MOSAIC_{bioacc} are dedicated to assist users in formatting their own data and to appropriate the different MOSAIC_{bioacc} features from the very first step. This paper illustrates this step-by-step process based on a typical data set of a

Revised Manuscript [2021-04-19]

toxicokinetic bioassay where internal concentrations were collected in fathead minnows (*Pimephales promelas*) exposed via contaminated water to a highly hydrophobic chemical ($\log K_{ow} = 9.06$) at an exposure concentration of $0.0044 \mu\text{g.mL}^{-1}$ over 49 days, with one replicate at each time point. After 49 days, minnows were transferred in a clean medium for 98 days more (Crookes and Brooke 2011). The data set (CSV format) and the report with all results (HTML format) can be downloaded directly from MOSAIC_{bioacc}.

Model and parameters

All TK models considered in MOSAIC_{bioacc} describe organisms as single compartments for which a first-order kinetic bioaccumulation model accounting for several exposure routes and elimination processes can be expressed in a generic way as follows (Eqs. (1) to (4)):

$$\begin{cases} \frac{dC_p(t)}{dt} = U - (E + M)C_p(t) \\ \frac{dC_{m_\ell}(t)}{dt} = k_{m_\ell}C_p(t) - k_{e_{m_\ell}}C_{m_\ell}(t) \end{cases} \quad \text{for } 0 \leq t \leq t_c \quad (1) \quad (2)$$

$$\begin{cases} \frac{dC_p(t)}{dt} = -(E + M)C_p(t) \\ \frac{dC_{m_\ell}(t)}{dt} = k_{m_\ell}C_p(t) - k_{e_{m_\ell}}C_{m_\ell}(t) \end{cases} \quad \text{for } t > t_c \quad (3) \quad (4)$$

where $C_p(t)$ is the internal concentration of the parent compound at time t ($\mu\text{g.g}^{-1}$), $C_{m_\ell}(t)$ the internal concentration ($\mu\text{g.g}^{-1}$) of metabolites ($\forall \ell = 1 \dots L$ with L the total number of metabolites) at time t , U the sum of all uptake terms, E the sum of all elimination terms for the parent compound, M the sum of all metabolization terms, k_{m_ℓ} the metabolization rate of metabolite ℓ (time^{-1}) and $k_{e_{m_\ell}}$ the elimination rate of metabolite ℓ (time^{-1}). **Table 2** gives an overview of all parameter and variable meaning. The dynamical system in equations (1) to (4), corresponding to the deterministic part of the model, can explicitly be solved when the exposure concentration is assumed to be constant over time (**Annex 1**). A Gaussian probability distribution was assumed as the stochastic part of the final model, based on the quantitative

Revised Manuscript [2021-04-19]

continuous nature of the concentration variables for both the parent compound and its metabolites within the organisms (Eqs. (5) and (6)):

$$C_{obs,p}(t) \sim \mathcal{N}(C_p(t), \sigma_{C_p}^2) \quad (5)$$

$$C_{obs,m_\ell}(t) \sim \mathcal{N}(C_{m_\ell}(t), \sigma_{m_\ell}^2) \quad (6)$$

Today MOSAIC_{bioacc} proposes data analyses by including until four exposure routes via water, pore water, sediment, and/or food), until three processes of elimination, which are excretion, growth dilution and biotransformation, with a maximum of 15 metabolites directly deriving from the parent compound (*i.e.*, phase I metabolism). For example, if three metabolites are considered ($L = 3$), the number of parameters involved in the most complete TK model equals 19. In total, users have 112 possible models, automatically designed according to their data (see supplemental data, **Annex 2**). For a given data set, the most complete model is built up by default and first proposed to users, from which they can perform the MOSAIC_{bioacc} analysis (**Fig. 1-c**). Users can also deselect some of the parameters (based on biological hypotheses related to the most probable exposure route or by neglecting one elimination process, for example). These choices lead to the automatic building of a nested TK sub-model to fit again on the data.

For clarity reasons, this paper illustrates MOSAIC_{bioacc} features from a simple data set (Pimephales_two.csv file) considering only the water exposure route (parameter k_{uw} for the uptake rate) and the excretion process (parameter k_{ee} for the elimination rate), that is one of the simplest TK models among the 112 possibilities. The corresponding equations for the deterministic part of this TK model are given below (Eqs. (7) and (8)):

$$\begin{cases} \frac{dC_p(t)}{dt} = k_{uw} \times c_w - k_{ee} \times C_p(t) & \text{for } 0 \leq t \leq t_c \end{cases} \quad (7)$$

$$\begin{cases} \frac{dC_p(t)}{dt} = -k_{ee} \times C_p(t) & \text{for } t > t_c \end{cases} \quad (8)$$

Finally, there are only three parameters to estimate: k_{uw} , k_{ee} and parameter σ_{C_p} of the Gaussian distribution (Eq. (5)). An illustration of a more complex data set with

Revised Manuscript [2021-04-19]

biotransformation and growth processes for a benthic invertebrate is given as supplemental material (**Annex 3**), as well as examples for exposure route by sediment (**Annex 4**) or food (**Annex 5**).

Once the Pimephales_two.csv example file has been uploaded, all required fields related to the experimental design are automatically filled in and raw data can be visualized, either as a table (default) or a plot, and the button to launch calculations ('Calculate and Display') is unlocked (**Fig. 1-c**). Once this button clicked, calculations start running with a progress bar informing users about progress. When calculations are finished, results are displayed, as plots or tables first for the bioaccumulation metrics, then for the fitting plots and some relevant goodness-of-fit criteria.

Calculations

Bayesian inference

Computations underlying MOSAIC_{bioacc} results are performed with JAGS (Plummer 2019) and the R software (R Core Team 2020, version 4.0.2) via the `rjags` and `jagsUI` packages (Plummer 2019; Kellner 2019). Models are fitted to bioaccumulation data using Bayesian inference via Monte Carlo Markov Chain (MCMC) sampling. For each model, calculation running starts with a short sampling on three MCMC chains (5,000 iterations after a burn-in phase of 10,000 iterations) using the Raftery and Lewis method (Raftery and Lewis 1992) to set the necessary thinning and the appropriate number of iterations in order to reach a precise and accurate estimation of each model parameter. Thanks to `rjags`, model parameters are retrieved as a joint posterior distribution from the likelihood of the observed data combined with prior information on parameters. All details on this approach can be found in the original research paper (Ratier et al. 2019) but also in many other papers in the field of ecotoxicology (Billoir et al. 2011; EFSA PPR Panel 2018).

Revised Manuscript [2021-04-19]

Choice of prior distributions

For simplicity reasons, we hid the choice of priors to MOSAIC_{bioacc} users; hence, they cannot be changed, except by downloading the open source programming code and handling it directly within the R software. To ensure genericity of priors we chose non-informative (-5, 5) log10-uniform distributions for all uptake and elimination rate constants, and non-informative (0, A) uniform distributions for all standard deviations with a large A, here defined as five times the maximum internal measured concentration, which is then removed from the data set, as usually proceeded (Gelman 2006).

Bioaccumulation metrics

Bioaccumulation metrics are the first outputs delivered by MOSAIC_{bioacc}. From the example chosen for this paper, the data analysis led to both the kinetic bioconcentration factor (BCF_k) and the steady state bioconcentration factor (BCF_{ss}), with the following exact mathematical expressions (Eqs. (9) and (10)):

$$BCF_k = \frac{k_{uw}}{k_{ee}} \quad (9)$$

$$BCF_{ss} = \frac{C_p(t_c)}{c_w} \quad (10)$$

where $C_p(t_c)$ is the internal parent compound concentration (in $\mu\text{g.g}^{-1}$) at the end of the accumulation phase (that is at $t = t_c$, in time) and c_w is the exposure contaminant concentration in water ($\mu\text{g.mL}^{-1}$). More details about calculations of bioaccumulation metrics at steady state are provided in **Annex 6**.

RESULTS AND DISCUSSION

Bioaccumulation metrics

When users click on the ‘Calculate and Display’ button, some results are provided by default. First, the BCF_k is given as a probability distribution (**Fig. 2-a**) and summarized with its median and its 95% uncertainty limits, that is 95% credible interval delimited by the 2.5th and

Revised Manuscript [2021-04-19]

the 97.5th centiles of the posterior probability distribution (**Table 3-a**). If users ask for the BCF_{ss} , its probability distribution is also delivered (**Fig. 2-b**) also summarized with the median and the 95% uncertainty limits (**Table 3-a**). Credible intervals are crucial information to quantify the uncertainty on parameter estimates. If data are available for several exposure routes (according to the experimental design) and uploaded within MOSAIC_{bioacc}, the BCF/BSAF/BMF metrics are displayed in separate tabs (see **Annex 4** for an example).

Predictions

Following bioaccumulation metrics calculations, the fitted curve and its uncertainty band superimposed to the observations is provided (**Fig. 2-c**). From the joint posterior distribution of model parameters, MOSAIC_{bioacc} then provides the marginal posterior distributions for each parameter, which are also summarized with quantiles in a table (**Table 3-b**): medians (for point estimates) and 2.5th and 97.5th centiles (for 95% credible intervals).

Goodness-of-fit criteria

After fitting plots, several goodness-of-fit criteria follow in a prioritized order chosen based on their relevance and their ease of interpretation. The fitting quality of the model can be first checked using the Posterior Predictive Check (PPC) plot: the idea is to compare each observed value to its prediction from the fitted model at the corresponding exposure concentration associated with its 95% credible interval. If the fit is correct, we expect to get 95% of the observed values falling within the 95% credible intervals of their predictions. As shown on **Fig. 3-a**, x-axis locates the observed values, while the y-axis reports their median predictions (black dots) with their 95% credible intervals (vertical segments).

The relevance of the inference process can also be checked using the comparison of prior and posterior distributions for each model parameter. The overall expectation is to get a narrower posterior distribution compared to the prior one for each parameter, reflecting that data contributed enough to precisely estimate parameters (**Fig.3-b**). Users have the possibility to select plots for deterministic (e.g., k_{uw} , k_{ee}) or stochastic (e.g., σ_{Cp}) parameters.

Revised Manuscript [2021-04-19]

Then, MOSAIC_{bioacc} provides a coloured matrix in order to see at a glance the most correlated or anti-correlated parameters, in order to quickly diagnose potential problems of precision due to highly correlated parameters. Moreover, MOSAIC_{bioacc} provides plots to visualize correlations between parameters (**Fig. S1-a**). Such a plot is obtained by projecting the joint posterior distribution as a matrix in planes of parameter pairs where contours have shapes reflecting both the sign and the strength of the correlations (sub-diagonal). The correlation plot also gives marginal posterior distribution of each model parameter (diagonal) and Pearson correlation coefficients (upper diagonal). Correlations between parameters are important to consider in particular when they are high (namely, greater than 0.75) what would mean that one parameter estimate could considerably influence the other, and reciprocally. Users can display the correlation plot for deterministic parameters only or for all parameters.

The convergence of MCMC chains can be checked with the Gelman-Rubin diagnostic (Gelman and Rubin 1992) expressed via the potential scale reduction factor (PSRF) which is expected to be close to 1.00 (**Fig. S1-b**). It can be also visually verified from the MCMC trace plots, which show the time series of the sampling process leading to the posterior distribution for each parameter; it is expected to get overlap of all MCMC chains (**Fig. S1-c**). Users can visualize the MCMC trace plots for deterministic or stochastic parameters.

Finally, the Deviance Information Criterion (DIC) is provided. It is a penalized deviance statistics accounting for the number of parameters that is only useful to compare several models fitted to a same data set. Models with lower DIC values will be preferred. So, the DIC is only useful when several sub-models are compared based on different choices of parameters from all the possible combinations that the users can choose from the beginning according to the uploaded data.

Downloads

At the bottom of the result web page, all outputs, either separately or as a full report, can be downloaded. Users can also download the entire R code corresponding to all calculations

Revised Manuscript [2021-04-19]

and graphs from the uploaded data set. This ensures transparency and reproducibility of all MOSAIC_{bioacc} results. This R code can be used as a steppingstone to change default options or to perform further analyses directly in the R software. For example, users can modify figures at their convenience or make several analyses on several data sets at the same time.

CONCLUSION

Offering MOSAIC_{bioacc} as a new on-line service, free, user-friendly and ready-to-use, raises important methodological issues: (1) automation of the inference process, in particular for Bayesian inference; (2) options and choices in a transparent and facilitated way for users; (3) default outputs whose order has been chosen to facilitate their step-by-step interpretation; (4) easy accessibility to figures, tables and R code to be downloaded under different convenient formats; and (5) a final full report of all MOSAIC_{bioacc} analyses. Besides its user-friendliness, MOSAIC_{bioacc} is free of use while ensuring privacy of the uploaded data as well as transparency and reproducibility of results, together with a short response time. MOSAIC_{bioacc} is particularly useful to estimate parameters of TK models leading to predictions of chemical concentrations bioaccumulated within living organism (whatever the species, aquatic, aerial or terrestrial) from accumulation-depuration data, even standard ones. MOSAIC_{bioacc} could thus be of particular interest for risk assessors and decision makers in their daily work of evaluating dossiers, *e.g.*, for market authorisation of active substances (EFSA PPR Panel 2018). Indeed, all results provided by MOSAIC_{bioacc} account for uncertainty and correlations between parameters, making possible to reproduce any previous analysis that would need to be confirmed. MOSAIC_{bioacc} can also be used for more exploratory research purposes by any environmental scientists or ecotoxicologists when accumulation-depuration data are collected and need to be analysed. MOSAIC_{bioacc} allows analyses for any species-compound combinations under consideration even with biotransformation processes, allowing users to easily perform TK analyses accounting for several exposure routes and phase I metabolites. MOSAIC_{bioacc} is

Revised Manuscript [2021-04-19]

available as a new statistical service dedicated to TK modelling approaches embedded within the MOSAIC platform (Charles et al. 2018). This makes MOSAIC an all-in-one facility for many applications. MOSAIC_{bioacc} will be shortly extended with a prediction tool, in order to help in designing new TK experiments, for example, when a new species-compound combination requires attention or when additional data are needed to get a better precision on bioaccumulation metrics. To gain again in generality, an R-package (rbioacc) is already in progress to include all the functionalities encompassed in MOSAIC_{bioacc}, which will also make it possible to extend the use of TK models when the exposure concentrations vary over time, and when it is necessary to consider non-first order kinetics.

SUPPLEMENTAL DATA

Supplemental data are available on-line.

Revised Manuscript [2021-04-19]

LIST OF FIGURES

Figure 1. Data uploading and user information as required from the MOSAIC_{bioacc} homepage.

Figure 2. Probability distributions of BCF_k (a) and BCF_{ss} (b). The middle-dotted line represents the median value, while left and right dotted lines delimit the 2.5th and 97.5th centiles. (c): Observed (black dots) and predicted contaminant concentrations in the organisms ($\mu\text{g.g}^{-1}$), where the median curve is displayed as the solid orange line and the uncertainty band as the grey zone, delimited by the 2.5th and 97.5th centiles in dotted orange lines.

Figure 3. Some goodness-of-fit criteria as provided by MOSAIC_{bioacc}: (a) Posterior Predictive Check (PPC) where observed values are read on the x -axis, while the y -axis reports median predictions (black dots) and their 95% credible intervals (vertical segments, coloured in green if they contain the observed value, in red otherwise); (b) prior (grey) and posterior (orange) marginal distributions of parameters of the chosen TK model (here the simplest one).

LIST OF TABLES

Table 1. Example of a data set ready to be used in MOSAIC_{bioacc}. The data set must contain four columns whatever their order: ‘time’ (time points of the measurements at the exposure concentration, in time unit: here in days), ‘conc’ (contaminant concentrations measured within organisms, that must be expressed in $\mu\text{g.g}^{-1}$), ‘expw’ (the contaminant concentration in the exposure medium, here water, that must be expressed in $\mu\text{g.mL}^{-1}$), and ‘replicate’ (a number or a character that is unique for each replicate, dimensionless).

Table 2. Meaning of parameters and variables of the TK model used in MOSAIC_{bioacc}.

Table 3. Example of BCF (a) and parameter (b) estimates expressed as medians (50th centile) and 95% credibility intervals (2.5th - 97.5th centiles). Hyphens stand for dimensionless parameters.

Revised Manuscript [2021-04-19]

REFERENCES

- Aldenberg T. 2019. bmcfR: Tools for Modeling Bioaccumulation Potential in Fish. Version 0.4-18 (bcmfR_0.4-18.tar.gz). <https://www.oecd.org/chemicalsafety/testing/section-3-environmental-fate-behaviour-software-tg-305.htm>
- Billoir E., Delhay H., Clément B., Delignette-Muller M.L., Charles S. 2011. Bayesian modelling of daphnid responses to time-varying cadmium exposure in laboratory aquatic microcosms. *Ecotoxicol. Environ. Saf.* 74(4):693-702. doi: 10.1016/j.ecoenv.2010.10.023
- Burkhard L.P., Arnot J.A., Embry M.R., Farley K.J., Hoke R.A., Kitano M., Leslie H.A., Lotufo G.R., Parkerton T.F., Sappington K.G., Tomy G.T., Woodburn K.B. 2012. Comparing laboratory and field measured bioaccumulation endpoints. *Integr Environ Assess Manag.* 8(1):17-31. doi: 10.1002/ieam.260
- Chang S., Cheng W., Allaire J. J., Xie Y., McPherson, J. 2020. Shiny: Web Application Framework for R. Version 1.5.0. <http://shiny.rstudio.com/>
- Charles S., Veber P., Delignette-Muller M.L. 2018. MOSAIC: a web-interface for statistical analyses in ecotoxicology. *Environ. Sci. Pollut. Res.* 25:11295–11302. doi: <https://doi.org/10.1007/s11356-017-9809-4>
- Crookes M.J., Brooke D.N. 2011. Estimation of fish bioconcentration factor (BCF) from depuration data. Environment Agency. Bristol (UK).
- European Commission. 2006. COMMISSION REGULATION (EU) No 1907/2006 of the European Parliament and of the Council of 18 December 2006 concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH), establishing a European Chemicals Agency, amending Directive 1999/45/EC and repealing Council Regulation (EEC) No 793/93 and Commission Regulation (EC) No 1488/94 as well as Council Directive 76/769/EEC and Commission Directives 91/155/EEC, 93/67/EEC, 93/105/EC and 2000/21/EC. Official Journal of the European Union 1–849.

Revised Manuscript [2021-04-19]

European Commission. 2012. REGULATION (EU) No 528/2012 of the European Parliament and of the Council of 22 May 2012 concerning the making available on the market and use of biocidal products. Official Journal of the European Union.

European Commission. 2013. COMMISSION REGULATION (EU) No 283/2013 of 1 March 2013 setting out the data requirements for active substances, in accordance with Regulation (EC) No 1107/2009 of the European Parliament and of the Council concerning the placing of plant protection product. Official Journal of the European Union.

ECHA. [ECHA] European Chemicals Agency. 2017. Guidance on Information Requirements and Chemical Safety Assessment. Chapter R.11: PBT/vPvB Assessment. doi:10.2823/128621

EFSA PPR Panel. [EFSA] European Food Safety Authority. 2014. Scientific Opinion on good modelling practice in the context of mechanistic effect models for risk assessment of plant protection products. EFSA Journal. 12:3589. doi: 10.2903/j.efsa.2014.3589

EFSA Scientific Committee. [EFSA] European Food Safety Authority. 2018. Guidance on Uncertainty Analysis in Scientific Assessments. EFSA Journal. 16:1–39. doi: 10.2903/j.efsa.2018.5123

EFSA PPR Panel (Ockleford C., Adriaanse P., Berny P., Brock T., Duquesne S., Grilli S., Hernandez-Jerez A.F., Bennekou S.H., Klein M., Kuhl T., Laskowski R., Machera K., Pelkonen O., Pieper S., Smith R.H., Stemmer M., Sundh I., Tiktak A., Topping C.J., Wolterink G., Cedergreen N., Charles S., Focks A., Reed M., Arena M., Ippolito A., Byers H., Teodorovic I.). [EFSA] European Food Safety Authority. 2018. Scientific Opinion on the state of the art of Toxicokinetic/Toxicodynamic (TKTD) effect models for regulatory risk assessment of pesticides for aquatic organisms. EFSA Journal. 16(8):5377. doi: 10.2903/j.efsa.2018.5377

Gelman A., Rubin D.B. 1992. Inference from Iterative Simulation Using Multiple Sequences. Stat. Sci. 7:457–472.

Revised Manuscript [2021-04-19]

404 Gelman, A. 2006. Prior Distribution for Variance Parameters in Hierarchical Models.
405 Bayesian Analysis. 3:5901–5906.

406 Gobas, F.A.P.C., de Wolf, W., Burkhard, L.P., Verbruggen, E., Plotzke, K. 2009,
407 Revisiting Bioaccumulation Criteria for POPs and PBT Assessments. Integr Environ Assess
408 Manag. 5: 624-637.

409 Gobas F.A.P.C., Lee Y.S., Lo J.C., Parkerton T.F., Letinski D.J. 2020. A Toxicokinetic
410 Framework and Analysis Tool for Interpreting Organisation for Economic Co-operation and
411 Development Guideline 305 Dietary Bioaccumulation Tests. Environ Toxicol Chem.
412 39(1):171-188. doi: 10.1002/etc.4599.

413 Kellner, K. 2019. jagsUI: A Wrapper Around 'rjags' to Streamline 'JAGS' Analyses.
414 Version 1.5.1. <https://cran.r-project.org/web/packages/jagsUI/index.html>

415 OECD. [OECD] Organisation for Economic Co-operation and Development. 2012. No.
416 305. Bioaccumulation in Fish: Aqueous and Dietary Exposure. Paris (FR): OECD. OECD.
417 doi: 10.1787/9789264185296-en.

418 Plummer M. 2019. rjags: Bayesian Graphical Models using MCMC. Version 4.10.

419 PRABI. 2020. Date Accessed: 2020-08-28. <http://www.prabi.fr>

420 Raftery A.E., Lewis S.M. 1992. [Practical Markov chain Monte Carlo]: Comment: one
421 long run with diagnostics: implementation strategies for Markov chain Monte Carlo. Stat. Sci.
422 7:493–497.

423 Ratier A., Lopes C., Labadie P., Budzinski H., Delorme N., Quéau H., Peluhet L.,
424 Geffard O., Babut M. 2019. A Bayesian framework for estimating parameters of a generic
425 toxicokinetic model for the bioaccumulation of organic chemicals by benthic invertebrates:
426 Proof of concept with PCB153 and two freshwater species. Ecotoxicol. Environ. Saf. 180:33-
427 42. doi:10.1016/j.ecoenv.2019.04.080

428 R Core Team, 2020. R: A Language and Environment for Statistical Computing.
429 Version 4.0.2.

Revised Manuscript [2021-04-19]

430 Tan Y.M., Chan M., Chukwudebe A., Domoradzki J., Fisher J., Hack C.E., Hinderliter
 431 P., Hirasawa K., Leonard J., Lumen A., Paini A., Qian H., Ruiz P., Wambaugh J., Zhang F.,
 432 Embry M. 2020. PBPK model reporting template for chemical risk assessment applications.
 433 Regul. Toxicol. Pharmacol. 115. doi:10.1016/j.yrtph.2020.104691
 434 USEPA. [USEPA] U.S. Environmental Protection Agency. 1994. Great Lake water
 435 quality initiative technical support document for the procedure to determine bioaccumulation
 436 factors 822-R-94-002. Office of water. Off. of Sci. and Tech., Washington, DC, USA.

Revised Manuscript [2021-04-19]

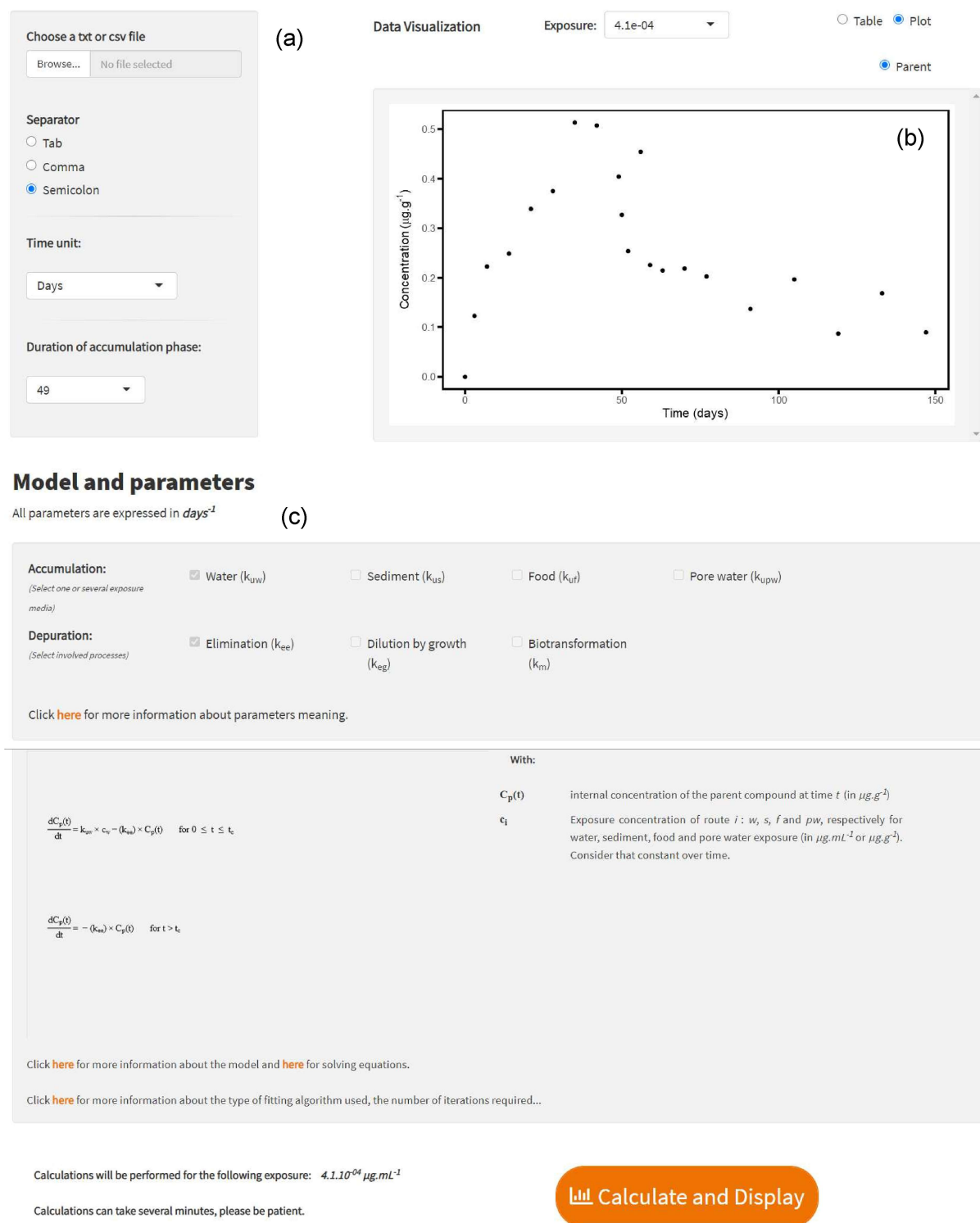


Figure 1. Data uploading and user information as required from the MOSAICbioacc homepage.

Revised Manuscript [2021-04-19]

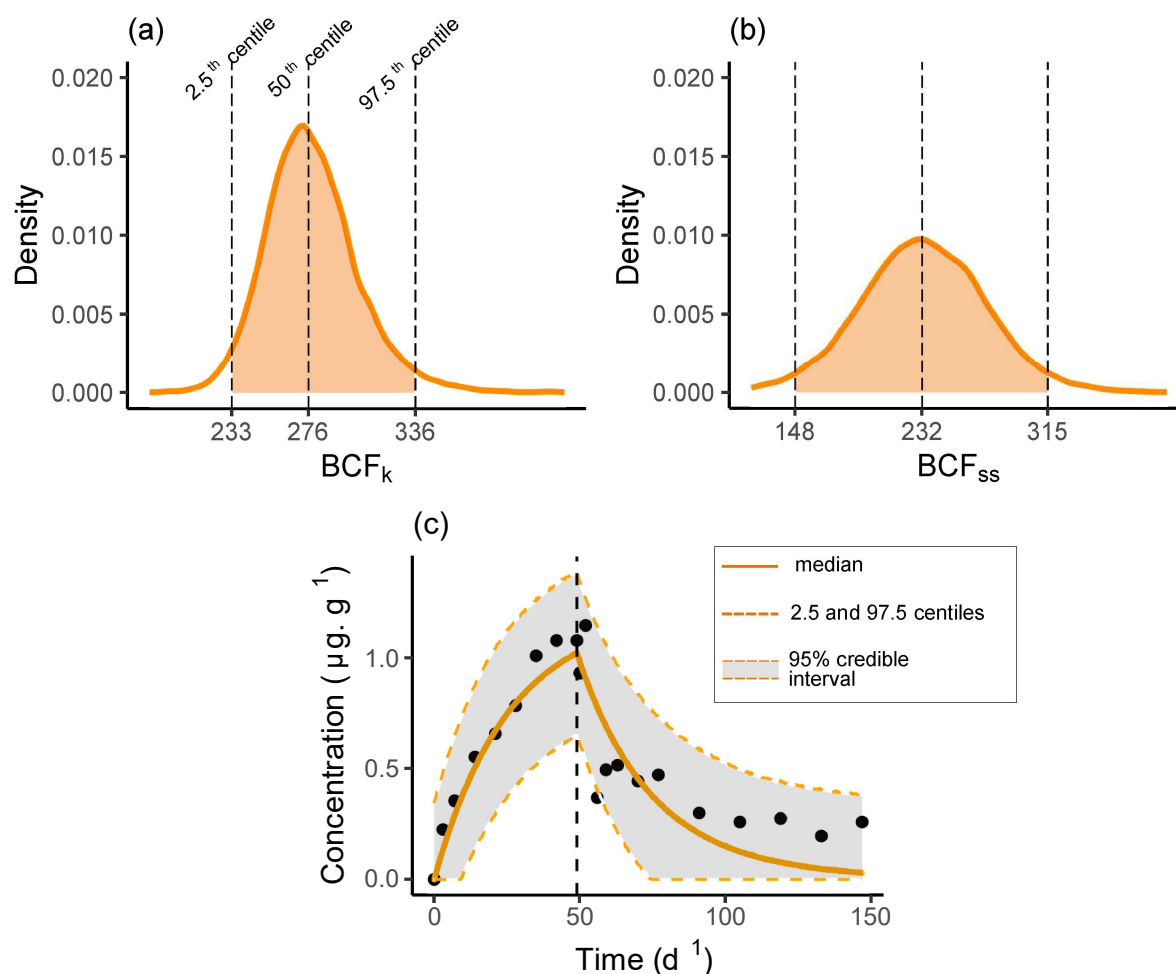


Figure 2. Probability distributions of BCF_k (a) and BCF_{ss} (b). The middle-dotted line represents the median value, while left and right dotted lines delimit the 2.5th and 97.5th centiles. (c): Observed (black dots) and predicted contaminant concentrations in the organisms ($\mu\text{g.g}^{-1}$), where the median curve is displayed as the solid orange line and the uncertainty band as the grey zone, delimited by the 2.5th and 97.5th centiles in dotted orange lines.

Revised Manuscript [2021-04-19]

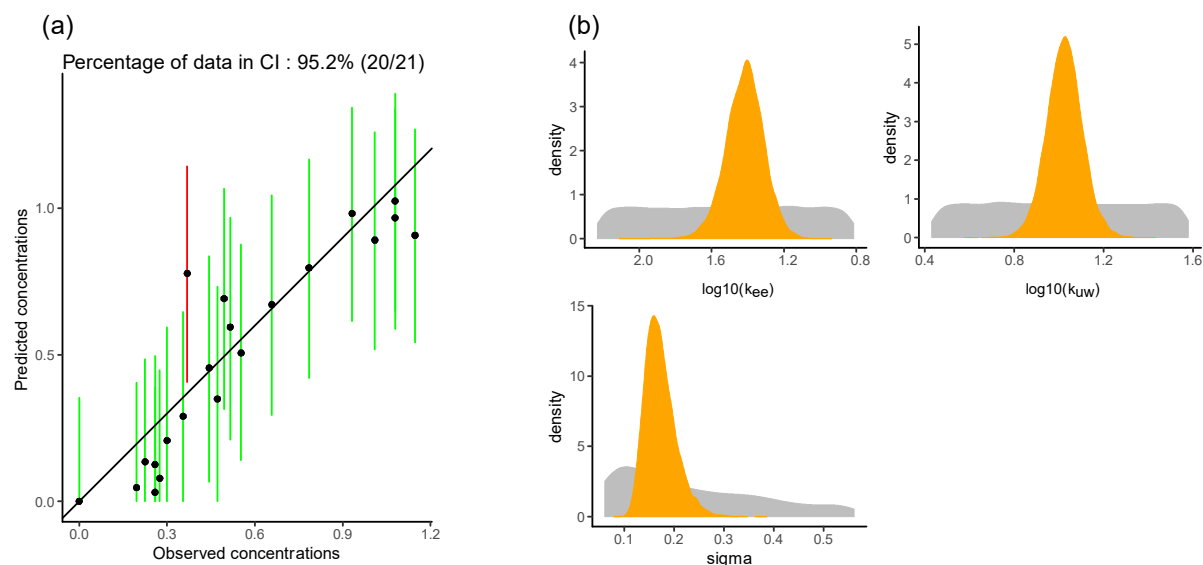


Figure 3. Some goodness-of-fit criteria as provided by MOSAIC_{bioacc}: (a) Posterior Predictive Check (PPC) where observed values are read on the x-axis, while the y-axis reports median predictions (black dots) and their 95% credible intervals (vertical segments, coloured in green if they contain the observed value, in red otherwise); (b) prior (grey) and posterior (orange) marginal distributions of parameters of the chosen TK model (here the simplest one).

Revised Manuscript [2021-04-19]

Table 1. Example of a data set ready to be used in MOSAIC_{bioacc}. The data set must contain four columns whatever their order: ‘time’ (time points of the measurements at the exposure concentration, in time unit: here in days), ‘conc’ (contaminant concentrations measured within organisms, that must be expressed in $\mu\text{g.g}^{-1}$), ‘expw’ (the contaminant concentration in the exposure medium, here water, that must be expressed in $\mu\text{g.mL}^{-1}$), and ‘replicate’ (a number or a character that is unique for each replicate, dimensionless).

time	conc	expw	replicate
0	0.000	0.0044	2
3	0.225	0.0044	2
7	0.355	0.0044	2
14	0.553	0.0044	2
21	0.658	0.0044	2
28	0.785	0.0044	2

Revised Manuscript [2021-04-19]

Table 2. Meaning of parameters and variables of the TK model used in MOSAIC_{bioacc}.

Symbol	Meaning
i	index of exposure routes, $i = 1 \dots I$
j	index of elimination processes, $j = 1 \dots J$
ℓ	index of metabolites, $\ell = 1 \dots L$
I	total number of exposure routes
J	total number of elimination processes
L	total number of metabolites
t	time (in time unit)
c_i	exposure concentration of route i (in $\mu g. mL^{-1}$)
$C_p(t)$	internal concentration of the parent compound at time t (in $\mu g. g^{-1}$)
$C_{m_\ell}(t)$	internal concentration of metabolite ℓ (in $\mu g. g^{-1}$)
k_{u_i}	uptake rate of exposure route i (expressed in $time^{-1}$)
k_{e_j}	elimination rate of elimination process j (expressed in $time^{-1}$)
$k_{e_{m_\ell}}$	elimination rate of metabolite ℓ (expressed in $time^{-1}$)
k_{m_ℓ}	biotransformation rate of metabolite ℓ (expressed in $time^{-1}$)
t_c	duration of the accumulation phase (in time unit)
\mathcal{N}	Gaussian probability distribution
$C_{obs,p}(t)$	internal measured concentration of the parent compound at time t (in $\mu g. g^{-1}$)
$C_{obs,m_\ell}(t)$	internal measured concentration of metabolite ℓ (in $\mu g. g^{-1}$)
σ_{Cp}	standard deviation for internal concentration of the parent compound (in $\mu g. g^{-1}$)
σ_{met_ℓ}	standard deviation for the internal concentration of metabolite ℓ (in $\mu g. g^{-1}$)
$U = \sum_{i=1}^I k_{u_i} c_i$	sum of all uptake terms
$E = \sum_{j=1}^J k_{e_j}$	sum of all elimination terms for the parent compound
$M = \sum_{\ell=1}^L k_{m_\ell}$	sum of all elimination terms for metabolite ℓ

Revised Manuscript [2021-04-19]

Table 3. Example of BCF (a) and parameter (b) estimates expressed as medians (50th centile) and 95% credibility intervals (2.5th - 97.5th centiles). Hyphens stand for dimensionless parameters.

	2.5 th	50 th	97.5 th	Units
(a)				
BCF_k	233	276	338	-
BCF_{ss}	147	233	317	-
(b)				
k_{uw}	7.388	10.59	15.45	d ⁻¹
k_{ee}	0.02307	0.03851	0.06091	d ⁻¹
σ_{Cp}	0.1249	0.1684	0.2445	µg.g ⁻¹