Rapid, but limited, zooplankton adaptation to simultaneous warming and acidification

Hans G. Dam^{1,5*}, James A. deMayo^{1,5,*}, Gihong Park¹, Lydia Norton¹, Xuejia He², Michael B. Finiguerra³, Hannes Baumann¹, Reid S. Brennan⁴, Melissa H. Pespeni⁴.

1. Department of Marine Sciences, University of Connecticut, Groton, CT 06340-6048, USA

2. Research Center for Harmful Algae and Marine Biology, Jinan University, Guangzhou 510632, Guangdong Province, P.R. CHINA

3. Department of Ecology and Evolutionary Biology, University of Connecticut, Groton, CT 06340-6048, USA

4. Department of Biology, the University of Vermont, Burlington, VT 05405, USA

5. Equal contribution

*Communicating authors

Author contact information:

Hans G. Dam: hans.dam@uconn.edu

James A. deMayo: james.demayo@uconn.edu

Gihong Park: gihong.park@uconn.edu

Lydia Norton: lydia.norton@uconn.edu

Xuejia He: hexuejia@sina.com

Michael B. Finiguerra: michael.finiguerra@uconn.edu

Hannes Baumann: hannes.baumann@uconn.edu

Reid S. Brennan: reid.brennan@uvm.edu

Melissa H. Pespeni: melissa.pespeni@uvm.edu

Abstract: Predicting the response of marine metazoans to climate change is hampered by a lack of studies on evolutionary adaptation, particularly to combined ocean warming and acidification (OWA). We provide evidence for rapid adaptation to OWA in the foundational copepod species, *Acartia tonsa*, by assessing changes in population fitness based on a comprehensive suite of life-history traits, using an orthogonal experimental design of nominal temperature (18°C, 22°C) and $p\text{CO}_2$ (400, 2000 μatm) for 25 generations (~1 year). Egg production and hatching success initially decreased under OWA, resulting in a 56% reduction in fitness. However, both traits recovered by the third generation and average fitness was reduced thereafter by only 9%. Antagonistic interactions between warming and acidification in later generations decreased survival, thereby limiting full fitness recovery. Our results suggest such interactions constrain evolutionary rescue and add complexity to predictions of the responses of metazoan populations to climate change.

The continuous increase in atmospheric CO₂ is unparalleled in the last 300 million years¹, causing rapid ocean warming (OW)^{2,3} and ocean acidification (OA)⁴. Warming and acidification in coastal regions are projected to be more extreme^{5,6}. Overall, novel environments may arise as a result of these changes⁷. Predicting how biota respond to this rapid global change is a crucial yet formidable scientific challenge, particularly whether populations can persist at similar levels and phenotypes in the future⁸. Organisms can respond to novel environments through phenotypic plasticity or evolutionary adaptation, both of which can mitigate the deleterious effects of climate change^{8–12}. In particular, long-term experimental evolution studies are a powerful tool to examine the role of adaptation in mitigating the effects climate change on biota and to explore whether populations can evolve fast enough to keep pace¹³.

Organismal performance typically decreases sharply when temperatures increase beyond the optimum¹⁴ leading to disproportional deleterious effects on performance. In addition, marine metazoans require the mobilization of energy-demanding acid-base regulatory processes to counteract internal pH reduction to maintain homeostasis under high CO₂ conditions. This may result in increased metabolic costs at the expense of growth and reproduction, even for non-calcifying metazoans^{15–17}. Although factorial assessments of species sensitivities to warming and acidification have increased rapidly over the past years, multi-generational studies on metazoan populations responding to future simultaneous warming and acidification (OWA) are rare^{18–20}. Moreover, lack of population fitness measures in existing studies preclude considerations of evolutionary rescue – evolution occurring sufficiently fast to allow population recovery before extirpation^{21–23}.

We used an experimental evolution approach to test whether a marine zooplankter, the copepod *Acartia tonsa* (Dana; 1849), can adapt to environments created by OW, OA, and OWA conditions, to identify the functional traits under selection, and to assess evolutionary rescue. As the most abundant metazoans on the planet^{24,25}, copepods link primary producers and other microbes to upper trophic levels, thereby influencing fisheries productivity^{26,27} and mediating marine biogeochemical cycles²⁸. Specifically, *Acartia tonsa* is a dominant copepod in estuarine systems from tropical to temperate regions²⁹ and a main prey item of forage fish³⁰, which makes this species an important zooplankton model. Using both improvements in trait performance and population fitness across generations, we show rapid, yet limited copepod adaptation to combined ocean warming and acidification likely driven by an antagonistic interaction of warming and acidification.

Experimental design

We measured five fitness-relevant life-history traits (survival, egg production rate (EPR), egg hatching success (HS), development time, and sex ratio) across 25 generations in an orthogonal design with two levels of CO₂ and temperature. A population of *Acartia tonsa* was collected from Long Island Sound (41.3°N, 72.0°W; Groton, CT, USA) and kept under standard laboratory conditions (see methods) for at least three generations prior to the experiment. Four lines of the population were established with four replicates of each condition. The target (actual $\pm$ standard deviation) conditions were: ambient temperature = 18°C (18 $\pm$ 0.34, N = 330), ambient *p*CO₂ = 400 μ atm (379 $\pm$ 36, N = 18; pH = 8.26 $\pm$ 0.1, N = 330); high temperature = 22°C (22 $\pm$ 0.81, N = 336), and high *p*CO₂ = 2000 μ atm (2301 $\pm$ 215, N = 18; pH = 7.55 $\pm$ 0.08, N = 330). Ambient target levels represented extant conditions for this species in NE Atlantic

estuaries (see methods for choice of temperature), and high levels corresponded to future conditions based on global projections for the years 2100-2300¹⁻⁴, although *A. tonsa* already periodically experiences high temperature and CO₂ levels in its growth season in NE Atlantic estuaries^{31,32}. Summaries of the temperature, pH, and CO₂ data are shown in Supplementary Tables 1-3. Details of statistical tests and their significance for the transgenerational experiment are in the methods section.

Evidence for adaptation

During the first experimental generation (generation zero), EPR and HS declined in all three future (OW, OA, and OWA) conditions relative to ambient (AM) conditions: OW (EPR: $p < 0.0001$; HS: $p < 0.01$; t-test), OA (EPR: $p < 0.0001$; HS: $p = 0.19$; t-test), and OWA (EPR: $p < 0.0001$; HS: $p < 0.0001$; t-test; Fig. 1), illustrating the ecological effects of these climate change variables. The decrease was strongest under OWA, particularly for HS, indicating synergistic deleterious effects of temperature and CO₂.

Significant interactions of temperature and CO₂ on HS were evident across all generations ($p < 0.0001$; three-way ANOVA). After the first generation, EPR decreased under AM ($p < 0.0001$; GAM ANOVA) and OA conditions ($p < 0.0001$; GAM ANOVA), but partially recovered in later generations. Meanwhile, HS remained stable (AM GAM ANOVA: $p = 0.14$; OA GAM ANOVA: $p = 0.064$; Fig. 1). Under OW, EPR decreased with generation ($p < 0.0001$; GAM ANOVA) but did not recover. Meanwhile, HS increased for OW ($p < 0.0001$; GAM ANOVA; Fig. 1C) suggesting some degree of adaptation. By contrast, under OWA, EPR increased by 50%

($p < 0.02$; GAM ANOVA) and HS doubled ($p < 0.0001$; GAM ANOVA; Fig. 1D) by generation 3, and the improvements were maintained until generation 25. These changes yielded significant effects on population fitness (Table 1) and are consistent with rapid adaptation.

Survival from freshly hatched nauplii to reproductively mature adults was independent of treatment ($p > 0.05$, two-way ANOVA) for the first 12 generations (Fig. 2). By generation 15, however, survival decreased by 30% under OW ($p = 0.02$, t-test) and OWA ($p = 0.01$, t-test) relative to ambient conditions. Acidification alone elicited no decrease in survival ($p > 0.1$, one-way ANOVA). Although survival recovered in the OW treatment by generation 25, survival decreased under OWA by an additional 50% relative to generation 15 for the same treatment ($p < 0.001$, t-test), demonstrating an antagonistic interaction between OW and OA in later generations³³. Traits under selection are expected to increase towards phenotypic optima^{34–36}, thus survival does not appear to be a trait under selection for adaptation to OW, OA or OWA conditions. High temperature resulted in faster development in both the OW and the OWA treatments (Supplementary Fig. 1) with 22–24% shorter development times than at ambient temperature ($p < 0.0001$; two-way ANOVA). High CO₂, in contrast, resulted in 5–6% slower development times than ambient CO₂ across all generations ($p < 0.0001$; two-way ANOVA). Thus, warming and acidification acted antagonistically on development time. Finally, a ~ 1:1 sex ratio remained unchanged for three of the four treatments (Supplementary Fig. 2). Under OWA, however, the proportion of females decreased across generations, with a significantly lower proportion in generation 25 than generation 0 ($p < 0.01$; Tukey HSD).

To understand adaptation not just in terms of individual traits, but with respect to population fitness, we integrated all measured traits (survival, EPR, HS, development time, and sex ratio) to estimate the net reproductive rate, λ ³⁷(λ), which is the fraction of the population replaced in a generation (Fig. 3). In generation 0, OWA conditions resulted in a 56% reduction in λ relative to ambient conditions, while OW and OA resulted in 23% and 13% reductions relative to ambient conditions, respectively ($p < 0.01$, t-test; Fig. 3). However, by generation 3, λ in OWA conditions had improved by 120% relative to generation 0 ($p < 0.0001$; Tukey HSD; Fig. 3) and recovered to levels equal to ambient conditions ($p = 1.0$; Tukey HSD; Fig. 3), mainly driven by improved hatching success (Fig. 1). The λ -frequency distribution shows an inflation of zero values (Supplementary Fig. 3) due to the high abundance of mate pairs with low HS at generation 0, particularly under OWA conditions. In later generations, as HS increased so did the probability of non-zero λ , and this was especially evident under OWA conditions (Supplementary Fig. 4). Thus, as generations progressed, selection under OWA conditions culled off low-fitness individuals in the population. λ remained high in OWA conditions until generation 15, after which there was a 19% reduction ($p < 0.05$, t-test; Fig. 3). This decrease was driven by reduced survival (Fig. 2), which suggests an inability of *Acartia tonsa* to maintain multiple optimal phenotypes, i.e., high hatching success and high survivorship, under OWA conditions. While λ was 22% lower under OWA than ambient conditions by generation 25, it was significantly higher than in generation 0 ($p < 0.0001$, t-test; Fig. 3), still consistent with adaptation over time. In accord, significant effects of generation on λ were evident under OW ($p < 0.04$; ANOVA) and OWA ($p < 0.001$; ANOVA) conditions, but not under ambient ($p = 0.681$; ANOVA) or OA ($p = 0.212$; ANOVA) conditions (Table 1). The latter suggests that OA alone was not a strong selective force in our study.

Adaptation is evident when performance increases towards optimal phenotypes that increase fitness over time^{34–36}. Here, we observed improved performance and fitness under OW and OWA conditions, but with important differences between the treatments: fitness was fully recovered under OW, but not under OWA conditions. *Acartia tonsa* exhibits strong tolerance to high temperatures, consistent with its seasonal dominance in the summer and wide-ranging latitudinal distribution^{29,38}. However, the combination of elevated temperature and CO₂ are known to affect ectotherm resource partitioning and energy distribution¹⁵, which might account for the limited recovery under OWA conditions. During adaptation, multiple phenotypes are expected to reach optimal levels of performance concurrently to yield the maximum possible population fitness for a particular environment. We hypothesize that under OWA conditions copepods could not sustain multiple optimized phenotypes, as evidenced by the observed reduction in survival following HS and EPR recovery to levels equal to or greater than those of ambient conditions. Under OW, copepods improved both EPR and HS (Fig. 1) and maintained high survival (Fig. 2) across generations to yield the highest λ level relative to all conditions by generation 25 (Fig. 3). By contrast, under OWA conditions the improvements in EPR and HS across generations yielded the highest λ values between generations 3 and 12 but decreases in survival reduced λ afterwards. These results are consistent with adaptation via selection, while simultaneously allowing maladaptive traits to persist³⁵ under OWA conditions.

Genetic drift affects global, genome-wide patterns of genetic diversity, while selective processes affect specific regions of the genome^{39,40}. We tested for signatures of genetic drift caused by potential bottlenecks by exploring patterns of nucleotide diversity (π) among treatments using

single nucleotide polymorphisms identified with pooled sequencing of genomic DNA from each replicate of each treatment at generation 25. We found that global levels of nucleotide diversity were equivalent across treatment groups (Supplementary Figure 5; Wilcoxon Rank Sum test, $p > 0.05$; average π across treatments: AM: 0.0135 ± 0.006 ; OA: 0.0130 ± 0.006 ; OW: 0.0134 ± 0.006 ; OWA: 0.0138 ± 0.006), revealing no evidence of genetic bottlenecks or drift after 25 generations. In separate work, we observed divergence in allele frequencies and gene expression at generation 20 between the AM and OWA lines⁴¹. Among the four lineages at generation 25, allele frequency estimates indicate differentiation between the four lines despite the similar levels of nucleotide diversity (our own unpublished observations). Altogether, the improvement in traits and population fitness across generations coupled with the genetic differentiation between the OWA and the AM treatments are consistent with evolutionary adaptation in the OWA treatment.

Traits affecting fitness

To assess the contribution of each life-history trait to adaptation, we quantified the strength of selection via the standardized linear selection coefficient of the fitness landscape⁴². High linear coefficients of selection for a given fitness landscape suggest that a particular trait is under selection, though not necessarily determining fitness. The coefficients are calculated from multiple regression models of relative λ against all changing life-history traits (see methods). We found that HS, but not EPR or survival, was under selection and had a positive impact on fitness. Specifically, relative λ increased as a function of HS with standardized linear selection coefficients (β) of 0.91 (AM), 0.94 (OA), 0.96 (OW), and 0.78 (OWA) at generation 0 ($p < 0.0001$, ANOVA; Fig. 4A). In addition, β_{HS} was significantly different for all treatments between

generations 0 and 25 ($p < 0.0001$, ANOVA), indicating changing degrees of selection on HS for all treatments-between the first and last generations. By contrast, neither EPR or survival were significant at generation 0 (Fig. 4B, C; $p > 0.1$, ANOVA) suggesting that selection acted in favor of higher HS, but not towards higher survival or EPR for the OW and OWA treatments. By generation 25, the impact of HS on relative fitness decreased relative to generation 0 by 88% and 10% for the OW and OWA treatments, respectively. Such decreases corresponded to a shift in phenotype distribution towards higher HS and suggested that selection had relaxed by generation 25 as HS within the population reached the phenotypic optimum (Fig. 4A). Finally, despite the significant decrease in sex ratio over the 25 generations in the OWA treatment (Supplementary Fig. 2), sex ratio had no effect on relative fitness between generations 0 and 25 ($p = 1.0$, ANOVA), suggesting this trait was not under selection.

We also employed path analysis on structural equation models (SEMs) to identify the hierarchical interactions of all life-history traits and the traits' effects on fitness^{43,44}. High correlation coefficients of SEMs indicate causality of a given trait's effect towards fitness. The path analysis revealed that HS had the largest effect on fitness of all life-history traits across treatments, with the exception of the OW treatment at generation 25, and was significant at both generations 0 and 25 (Table 2). Taken together, we conclude that selection on HS both under OW and OWA conditions was the critical factor in adaptation to those environments.

Implications

Our study highlights the need for global change studies to consider population fitness, in addition to individual traits, as an integrative tool for measuring adaptation^{27,45}. For example, in this experiment using survival alone would have led to the erroneous conclusion of no adaptation under any treatment. Similarly, estimating fitness based on EPR alone would have also led to the erroneous conclusion that fitness decreased between generations 3 and 15 for AM. Despite this decrease in EPR, fitness remained unchanged across generations for AM (Fig. 3). Likewise, using EPR and HS alone would have missed the fitness decrease after generation 12 under OWA conditions. At the same time, the tradeoffs of EPR and HS versus survival after generation 12 (Figs. 1 and 2) partly explain the limited evolutionary rescue under OWA conditions (Fig. 3).

The shifting interactive effects of combined ocean warming and ocean acidification through time highlight the need to evaluate long-term evolutionary studies with multiple stressors and underline the complexity that accompanies predicting organism responses to climate change^{33,46}. Previous work has shown that *Acartia tonsa* can adapt to lower CO₂ levels (800 μ atm) than those in our study⁴⁶. Thus, it is conceivable that full evolutionary rescue to OWA can be achieved at lower CO₂ levels. Furthermore, previous research has shown that exposure to OA over multiple generations improves EPR and selects for genes involved in RNA processing and regulation of metabolism in copepods^{11,47}. Forthcoming genomic work from our group will determine if similar genes are also under selection for OW and OWA. Neither deleterious effects of OA nor adaptation to OA alone were observed across generations in our study (Fig. 3). Thus, our results highlight the evolutionary rescue and accompanying limitations that arise from multi-stressor adaptation. Assuming additive effects of warming and acidification in the present study should have led to full evolutionary rescue under OWA conditions, which was not observed (Fig. 3).

Instead, OW and OA had synergistic deleterious effects on fitness on generation 0, but antagonistic effects on generation 25 that counterbalance the individual effects of OW and OA (Supplementary Table 4). We suggest that the limited evolutionary rescue under OWA conditions in our study must have arisen from an antagonistic interaction between warming and acidification, as has been hypothesized earlier in polychaetes^{19,20,48} or for other co-occurring environmental stressors in bivalves⁴⁹. Thus, since OW and OA are occurring simultaneously, even crude predictions of population performance under climate change should consider non-additive effects of temperature and CO₂ interactions on population fitness.

Our study, showing both improved trait performance and population fitness across generations, constitutes the first demonstration of rapid adaptation to global change conditions for a metazoan. We present evidence for adaptation to both warming and combined warming and acidification, but not acidification alone, based on standing genetic variation. Previous work has been limited to phytoplankton³⁹ or documenting marine metazoan traits for limited generations without population fitness estimates^{18–20}. Failing to account for adaptation potential may overestimate future population vulnerability^{45,50}. For example, using the observed deleterious effects at generation 0 to predict future vulnerability would have estimated a fitness reduction for the OWA treatment relative to ambient treatment of 56% (Fig. 3). However, accounting for evolutionary rescue (generations 3–25) resulted in an average fitness reduction of only 9%. At the same time, our results suggest that full evolutionary rescue under OWA conditions was not achieved (Fig. 3). This result is consistent with the hypothesis that adaptation to stress comes at a cost and that full recovery of populations to future climate conditions based on extant genetic variation is limited⁵¹, even though *A. tonsa* periodically experiences the warming and

acidification conditions examined here^{31,32}. The limited evolutionary rescue we observed also has consequences for future oceanic systems under OWA conditions. Because adaptation to OWA conditions appears to be costly even under the food replete conditions of our study, it is unclear to what extent evolutionary rescue would occur under scenarios that predict lower resource availability for zooplankton under climate change⁵⁰. Limited evolutionary rescue, in turn, would reduce prey availability for fish, thereby negatively affecting fish production⁵⁰. Finally, since copepods are a major vector of carbon transfer from the surface waters to the sea floor²⁸, limited evolutionary rescue would also reduce the efficiency of the biological carbon pump, with a concomitant lower drawdown of CO₂ from near surface waters.

Conclusions

We show evidence for metazoan evolutionary adaptation to combined warming and acidification by explicitly assessing changes in population fitness based on a comprehensive suite of life-history traits. Adaptation was evident in the fitness improvements for animals after a few generations, and confirmed by independent evidence of genetic differentiation. The strength of selection coefficients and the path analysis suggested the key trait under selection affecting fitness in our study was hatching success. Evolutionary rescue was limited by the interaction of warming and acidification, which switched from synergistic to antagonistic from the beginning to the end of the experiment, adding complexity to predictions of population adaptation to climate change.

Methods

a) Copepod culturing and maintenance

Copepods were collected in June of 2016 from Esker Point Beach in Groton, CT, USA (41.320725°N, 72.001643°W) and raised for at least three generations as stock cultures prior to the start of transgenerational experiments to limit maternal effects⁵². Stock cultures were split evenly into eight groups of 160 females and 80 males. Four of these eight groups were acclimatized to high temperature at 1 degree Celsius per day and used to seed the two high temperature treatments (OW and OWA). The other four groups remained at ambient temperature and were used to seed the ambient and acidification treatments. After temperature acclimatization, groups of stock cultures seeded the parental (F0) individuals for two days. Stock culture groups yielded an average of 7,173 eggs per group to produce approximately 57,000 parental (F0) eggs. Resulting parental eggs and N1 nauplii were acclimated to one of four experimental treatments over the entire F0 generation: **1)** Ambient control (AM; temperature = 18°C, CO₂ ~400 µatm; pH ~8.2); **2)** Ocean Acidification (OA; ambient temperature, high CO₂ ~2000 µatm CO₂, pH ~7.5); **3)** Ocean Warming (OW; high temperature = 22°C, ambient CO₂); **4)** combined warming and acidification (OWA; high temperature, high CO₂). Treatment replicates were derived from the stock culture groups (i.e. stock culture group 1 after high temperature acclimatization seeded the F0 for OW replicate 1 and OWA replicate 1). The ambient temperature was chosen from a Gaussian fit-model⁵³ for optimal recruitment (egg production × egg hatching success) vs temperature derived from *Acartia tonsa* populations from Casco Bay (Gulf of Maine), Long Island Sound, and Chesapeake Bay (USA), $R = 11.1e^{-0.5\left(\frac{|T-17.7|}{6.4}\right)^2}$ where R is recruitment (egg production rate × hatching success), T is temperature (°C), 17.7 is the optimal temperature and 6.4 is the standard deviation around the

optimal temperature ($N = 54$, $r^2 = 0.42$; our own unpublished data). Each treatment was kept in a separate temperature-controlled incubator (Thermo FisherScientific® Isotemp™; Waltham, MA, USA) and split into four replicate 10L culture containers (Cambro; Huntington Beach, CA, USA). Copepods were fed every 48-72 h at food-replete concentrations ($\geq 800 \mu\text{g Carbon L}^{-1}$) consisting of equal proportions of the phytoplankters *Tetraselmis* sp., *Rhodomonas* sp., and *Thalassiosira weissflogii*, following long-standing copepod culture protocols in our lab⁵⁴. The phytoplankton fed to copepods was deliberately raised under ambient conditions for the entire length of the experiment to avoid confounding effects of possible changes in food quality due to the different temperature and CO₂ among treatments. We minimized the chance of selecting for early developers in the cultures. Based on development time and adult longevity, we allowed copepods to contribute progeny to the next generation for 7-10 days once we observed the first nauplii of a new generation.

Elevated CO₂ levels were achieved with gas proportioners (Cole-Parmer; Vernon Hills, IL, USA) mixing laboratory air with 100% bone dry CO₂ that was delivered continuously to the bottom of each replicate culture. For small volume life-history trait experiments (sections b-d), CO₂-mixed air was fed into custom plexiglass enclosures within each temperature-controlled incubator to allow for passive diffusion of CO₂ into seawater. Target pH values were monitored using a handheld pH probe (Orion Ross Ultra pH/ATC Triode with Orion Star A121 pH Portable Meter (Thermo FisherScientific®; Waltham, MA, USA). The pH probe was calibrated monthly using commercially available NBS pH standards in a three-point calibration (pH 4.01, 7, 10.01; Thermo FisherScientific®; Waltham, MA, USA). Temperature and pH were monitored to ensure that small-volume experiments in the plexiglass enclosures matched those of bulk cultures. To

counteract metabolic CO₂ accumulation, control CO₂ conditions were achieved by forcing compressed ambient air through a series of CO₂-stripping units containing granular soda lime (AirGas®; Waterford, CT, USA) and a particle filter (1 µm), and then to each culture container via airstone. Continuous bubbling maintained dissolved oxygen levels at >8 mg L⁻¹. Temperature, pH, and actual *p*CO₂ were monitored throughout the experiment (Supplementary Table 1). Actual *p*CO₂ conditions were calculated in CO₂SYS⁵⁵ based on measurements of salinity, temperature, pH, and total alkalinity (*A*_T; Supplementary Tables 2-3) with *k*₁/*k*₂ from Lueker, et al. 2000⁵⁶, KHSO₄ from Dickson 1990⁵⁷, total Boron from Uppstrom 1974⁵⁸, and pH based on NBS scale. Total alkalinity was measured in triplicates three times over the course of the experiment using endpoint titration (G20 Potentiometric Titrator; Mettler Toledo®)⁵⁹.

Because treatments were housed in separate incubators, incubator-specific effects are theoretically possible, but unlikely^{60,61} given that incubators were held under identical ambient conditions except for the constant temperature and CO₂ conditions that were meticulously monitored, and verified by independent measures throughout the experiment (Supplementary Table 5). Because of logistical and personnel constraints, egg production rate (EPR), hatching success (HS), survival, and development time (see below) could not be measured at every generation. Instead, measurements were taken at generations 0, 3, 6, 9, 12, 15, and 25.

b) Egg production rate and hatching success

For each replicate culture within a treatment, 10 pairs of newly developed males and females were placed into 20 mL petri dishes for 48 h (*n* = 280 per treatment). The dishes were housed in custom-made, airtight, plexiglass enclosures whose atmosphere was controlled to the appropriate

CO₂ concentration (see *copepod culturing and maintenance* section). There was one enclosure per temperature-controlled incubator. After the 48-h egg laying period, adults were checked for survival and removed from the petri dishes. Eggs were left in the dishes for an additional 72 h to allow for egg hatching and their contents preserved with non-acid Lugol's solution. Dishes with dead males were used for EPR, but not HS, since fertilization could not be assumed. Dishes with dead females were discarded. EPR was calculated as $\frac{E_u + E_h}{t}$ where E_u represents unhatched eggs, E_h represents hatched eggs (nauplii), and t represents egg laying time. Hatching success was calculated as $\frac{E_h}{E_u + E_h}$.

c) *Survival*

Survival was measured from nauplius 1 (N1) stage to copepodid 6 (adult) stage. For a given generation, all adults from the previous generation were removed from the culture and allowed to lay eggs in food-replete media for 48 h. Resulting nauplii were chosen for tracking survival. Unhatched eggs and any nauplii not chosen for survival analysis were returned to their respective replicate cultures for continued population maintenance. To measure survival for all generations where life-history traits were evaluated, three 250-mL beakers for each replicate culture were supplied with 25 randomly chosen N1 nauplii each and housed in the plexiglass enclosure described above (n = 21 per treatment). Copepods were checked every 48-72 h. The number of dead, live, and missing copepods were logged for each beaker along with general stage (i.e. nauplius, copepodite, adult female, or adult male). The fraction of survived individuals (l_x) was calculated as n_x/n_i where n_x represents the number of live individuals on day x , and n_i represents initial individuals. Nauplii were grown with media at levels of 500 µg C L⁻¹ to prevent overgrowth of phytoplankton and allow for adequate nauplii grazing. Following the naupliar

stages, copepods were grown with food-replete ($800 \mu\text{g C L}^{-1}$) media as described earlier. Food was replaced with fresh media on monitoring days. Average survival was calculated per each replicate culture at each generation measured. Differences in day-specific survival between replicates and treatments was assessed using the ‘survival’ package in R⁶².

d) Development time

To calculate development time (time from N1 to adulthood) we recorded the number of days at which adults matured during the survival experiments and averaged the observations.

e) Sex ratio

Sex ratio was calculated based on the number of surviving adult females relative to surviving adult males in survival experiments.

f) Fitness

The population net reproductive rate, λ , was calculated as the dominant eigenvalue of an assembled projected age-structured Leslie Matrix constructed from survival and fecundity data³⁷. Briefly, day-specific probabilities of survival are calculated from day-specific survival as $P_x = \frac{l_x}{l_{x-1}}$ where l_x represents the proportion of individuals on day x and l_{x-1} represents the proportion of individuals on day $x-1$. Probabilities of survival on day 1 are assumed to be 100% or a value of “1.0”. Per capita EPR and HS are calculated as described above, with fecundity rates equaling the product of EPR and HS. Because only females produce offspring, total fecundity rates must be scaled to the sex ratio (proportion of females:males). To account for differences in individual development time for each treatment, fecundity rates are assigned to all days after the first matured adult is observed. We assume that surviving individuals represented by the survival

experiments are equally as likely to experience any of the fecundity values observed in EPR experiments. Therefore, each mate-pair fecundity rate was paired with each survival beaker to construct a matrix. This yields a maximum of 120 matrices per treatment per generation (3 survival beakers $\times$ 4 replicate cultures $\times$ 10 mate pairs).

g) Selection coefficients and path analysis

Linear selection coefficients (β) were evaluated by creating multiple linear regression models of relative fitness ($\lambda/\bar{\lambda}$) against survival, egg production rate, hatching success, and sex ratio at both F0 and F25 for each treatment and calculating the partial regression coefficient estimate for each trait⁶³. The path analysis was completed by creating structural equation models^{43,44} of relative fitness against survival, egg production rate, hatching success, and sex ratio at both F0 and F25 for each treatment using the ‘lavaan’ and ‘semPlot’ packages in R^{64,65}. Development time was omitted from the models because the lack of variance within a replicate violated assumptions of the model.

h) Genetic Diversity

To quantify genetic diversity across the genome, capture probes were designed to target both coding and regulatory regions across the genome. For coding regions, we chose the highest quality probe falling within the region. Regulatory probes were set within 1,000 bp upstream of the transcription start site. Genomic DNA was shipped to Rapid Genomics (Gainesville, FL, USA) for library preparation and was captured with the 32,413 probes (21,311 coding, 11,102 regulatory). Enriched libraries were sequenced on a HiSeq 4000 with 150 bp paired end reads. Raw data were trimmed for quality and adapter contamination with Trimmomatic v0.36⁶⁶. Trimmed reads were mapped to the *A. tonsa* reference genome⁶⁷ with BWA-MEM⁶⁸.

SAMTOOLS⁶⁹ was used to generate a pileup for each sample, from which genetic diversity (π) was estimated with Popoolation⁷⁰. We identified 1,450 100bp windows in 704 unique scaffolds across the genome that were present across all samples. We estimated π in 100bp sliding windows with a 100bp step size. Each position required a minimum coverage of 30×, max coverage of 1000× (to avoid mapping errors), and at least 0.5 of the window meeting these thresholds. Resulting windows were required to be sequenced across all samples. To take into account the independent replicates within each treatment, we used pairwise Wilcoxon Rank Sum tests with a Holm correction for multiple testing. Genomic data is deposited in GenBank: BioProject number PRJNA590963. All statistics were performed in R⁷¹.

i) Statistical analyses

Statistical analyses were computed using the software package R (v 4.0.2)⁷¹. To examine effects of generation on changing life-history traits, we created trait-specific generalized additive models (GAMs) smoothed across generations for each treatment^{72,73}. To evaluate differences between life-history traits, we created separate linear mixed models with replicates as random effects and used post-hoc t-tests and Tukey HSD tests to compare life-history trait values that were significantly different from other treatments at each generation ($\alpha < 0.05$). Analysis of fitness (λ) calculations also included estimations with a zero-inflated generalized linear mixed effects model with generation and treatment as fixed effects and replicates as random effects. Analyzing the data with linear mixed models also allowed us to evaluate the effects of treatment replicates on life-history traits. A low intra-class correlation coefficient suggests no predictive effect of random variables. For the model constructed for fitness over generations, the variance due to replicates within a treatment is very low (ICC = 0.08; Table 1) and does not affect the model results. To estimate the predicted probabilities of $\lambda = 0$ across generations in Supplementary Fig.

4, we converted λ values to either 0 or 1 representing λ values of 0 and >0 , respectively. We used the binomial-converted λ data to fit a linear mixed effects model against generation and treatment with replicates included as random effects. To evaluate individual effects of temperature, pH, or generation on life-history traits, we constructed a third linear model that was tested with a three-way analysis of variance.

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Author Contributions: HGD conceived the project, designed research, aided in data analysis, and wrote the manuscript. JAD conducted experiments, analyzed data, created figures, and wrote the manuscript with HGD. GP, LN, and XH conducted experiments. MBF conceived the project and designed research. HB conceived the project, designed research, and designed the CO₂ delivery system. RSB performed genomic diversity analysis. MHP conceived the project, designed research, and performed genomic analysis. All authors edited and approved the paper.

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469 **Data and Script availability:** The phenotypic data and R-scripts referred in the text have been
470 deposited in zenodo (<https://doi.org/10.5281/zenodo.4477401>). The genetic diversity data are
471 deposited in GenBank: BioProject number PRJNA590963, to be released upon publication of
472 this paper.

References

1. Hönisch, B. *et al.* The Geological Record of Ocean Acidification. *Science* (80-.). **335**, 1058–1063 (2012).
2. Bindoff, N. L. *et al.* Changing Ocean, Marine Ecosystems, and Dependent Communities. *IPCC Spec. Rep. Ocean Cryosph. a Chang. Clim.* 447–588 (2019).
3. Pörtner, H.-O. *et al.* IPCC, 2019. Technical Summary. *IPCC Spec. Rep. Ocean Cryosph. a Chang. Clim.* 35–74 (2019).
4. Caldeira, K. & Wickett, M. E. Anthropogenic carbon and ocean pH. *Nature* **425**, 365 (2003).
5. Cai, W. J. *et al.* Acidification of subsurface coastal waters enhanced by eutrophication. *Nat. Geosci.* **4**, 766–770 (2011).
6. Wallace, R. B., Baumann, H., Grear, J. S., Aller, R. C. & Gobler, C. J. Coastal ocean acidification: The other eutrophication problem. *Estuar. Coast. Shelf Sci.* **148**, 1–13 (2014).
7. Radeloff, V. C. *et al.* The rise of novelty in ecosystems. *Ecol. Appl.* **25**, 2051–2068 (2015).
8. Munday, P. L., Warner, R. R., Monroe, K., Pandolfi, J. M. & Marshall, D. J. Predicting evolutionary responses to climate change in the sea. *Ecol. Lett.* **16**, 1488–1500 (2013).
9. Kelly, M. W. & Hofmann, G. E. Adaptation and the physiology of ocean acidification. *Funct. Ecol.* **27**, 980–990 (2013).
10. Pespeni, M. H. *et al.* Evolutionary change during experimental ocean acidification. *Proc. Natl. Acad. Sci. U. S. A.* **110**, 6937–6942 (2013).
11. Thor, P. & Dupont, S. Transgenerational effects alleviate severe fecundity loss during ocean acidification in a ubiquitous planktonic copepod. *Glob. Chang. Biol.* **21**, 2261–2271 (2015).
12. Donelson, J. M., Salinas, S., Munday, P. L. & Shama, L. N. S. Transgenerational plasticity and climate change experiments: Where do we go from here? *Glob. Chang. Biol.* **24**, 13–34 (2018).
13. Chevin, L. M., Lande, R. & Mace, G. M. Adaptation, plasticity, and extinction in a changing environment: Towards a predictive theory. *PLoS Biol.* **8**, (2010).
14. Angilletta, M. J. *Thermal Adaptation: A Theoretical and Empirical Synthesis*. (2009). doi:10.1093/acprof
15. Byrne, M. Impact of ocean warming and ocean acidification on marine invertebrate life history stages: Vulnerabilities and potential for persistence in a changing ocean. *Oceanogr. Mar. Biol. An Annu. Rev.* **49**, 1–42 (2011).
16. Whiteley, N. M. Physiological and ecological responses of crustaceans to ocean acidification. *Mar. Ecol. Prog. Ser.* **430**, 257–271 (2011).

- 510 17. Cripps, G., Lindeque, P. & Flynn, K. J. Have we been underestimating the effects of
511 ocean acidification in zooplankton? *Glob. Chang. Biol.* **20**, 3377–3385 (2014).
- 512 18. Baumann, H. Experimental assessments of marine species sensitivities to ocean
513 acidification and co-stressors: How far have we come? *Can. J. Zool.* **97**, 399–408 (2019).
- 514 19. Gibbin, E. M. *et al.* Can multi-generational exposure to ocean warming and acidification
515 lead to the adaptation of life history and physiology in a marine metazoan? *J. Exp. Biol.*
516 **220**, 551–563 (2017).
- 517 20. Gibbin, E. M., Massamba N’Siala, G., Chakravarti, L. J., Jarrold, M. D. & Calosi, P. The
518 evolution of phenotypic plasticity under global change. *Sci. Rep.* **7**, 1–8 (2017).
- 519 21. Gonzalez, A., Ophelie, R., Ferriere, R. & Hochberg, M. E. Evolutionary rescue: an
520 emerging focus at the intersection between ecology and evolution. *Philos. Trans. R. Soc. B*
521 *Biol. Sci.* **368**, 1–8 (2012).
- 522 22. Bell, G. & Gonzalez, A. Evolutionary rescue can prevent extinction following
523 environmental change. *Ecol. Lett.* **12**, 942–948 (2009).
- 524 23. Carlson, S. M., Cunningham, C. J. & Westley, P. A. H. Evolutionary rescue in a changing
525 world. *Trends Ecol. Evol.* **29**, 521–530 (2014).
- 526 24. Hardy, A. *The Open Sea: The World of Plankton*. (Collins, 1970).
- 527 25. Huys, R. & Boxshall, G. A. *Copepod Evolution*. (Ray Society, 1991).
- 528 26. Beaugrand, G. & Reid, P. C. Long-term changes in phytoplankton, zooplankton and
529 salmon related to climate. *Glob. Chang. Biol.* **9**, 801–817 (2003).
- 530 27. Möllmann, C., Müller-Karulis, B., Kornilovs, G. & St John, M. A. Effects of climate and
531 overfishing on zooplankton dynamics and ecosystem structure: Regime shifts, trophic
532 cascade, and feedback loops in a simple ecosystem. *ICES J. Mar. Sci.* **65**, 302–310 (2008).
- 533 28. Steinberg, D. K. & Landry, M. R. Zooplankton and the Ocean Carbon Cycle. *Annu. Rev.*
534 *Mar. Sci.* **9**, 413–444 (2017).
- 535 29. Mauchline, J. The Biology of Calanoid Copepods. in *Advances in Marine Biology* 33
536 (1998).
- 537 30. Turner, J. T. *The feeding ecology of some zooplankters that are important prey items of*
538 *larval fish*. (1984).
- 539 31. Rice, E., Dam, H. G. & Stewart, G. Impact of Climate Change on Estuarine Zooplankton:
540 Surface Water Warming in Long Island Sound Is Associated with Changes in Copepod
541 Size and Community Structure. *Estuaries and Coasts* 1–11 (2014). doi:10.1007/s12237-
542 014-9770-0
- 543 32. Gobler, C. J. & Baumann, H. Hypoxia and acidification in marine ecosystems : Coupled
544 dynamics and effects on ocean life Hypoxia and acidification in ocean ecosystems :
545 coupled dynamics and effects on marine life. *Biol. Lett.* **12**, (2016).
- 546 33. Côté, I. M., Darling, E. S. & Brown, C. J. Interactions among ecosystem stressors and

- their importance in conservation. *Proc. R. Soc. B Biol. Sci.* **283**, 1–9 (2016).
34. Burt, A. Perspective: The Evolution of Fitness. *Evolution (N. Y.)* **49**, 1–8 (1995).
35. Hendry, A. P. & Gonzalez, A. Whither adaptation? *Biol. Philos.* **23**, 673–699 (2008).
36. Arnold, S. J., Pfrender, M. E. & Jones, A. G. The adaptive landscape as a conceptual bridge between micro- and macroevolution. *Genetica* **112–113**, 9–32 (2001).
37. Caswell, H. *Matrix Population Models: Construction, Analysis, and Interpretation*. (Sinauer Associates, Inc., 2001).
38. Sasaki, M. C. & Dam, H. G. Integrating patterns of thermal tolerance and phenotypic plasticity with population genetics to improve understanding of vulnerability to warming in a widespread copepod. *Glob. Chang. Biol.* **25**, 4147–4164 (2019).
39. Luikart, G., England, P. R., Tallmon, D., Jordan, S. & Taberlet, P. The power and promise of population genomics: From genotyping to genome typing. *Nat. Rev. Genet.* **4**, 981–994 (2003).
40. Black IV, W. C., Baer, C. F., Antolin, M. F. & DuTeau, N. M. Population genomics: Genome-wide sampling of insect populations. *Annu. Rev. Entomol.* **46**, 441–469 (2001).
41. Brennan, R. *et al.* Loss and recovery of transcriptional plasticity after long-term adaptation to global change conditions in a marine copepod. *bioRxiv* (2021).
doi: <https://doi.org/10.1101/2020.01.29.925396>
42. Kingsolver, J. G. & Pfennig, D. W. Patterns and Power of Phenotypic Selection in Nature. *Bioscience* **57**, 561–572 (2007).
43. Crespi, B. J. & Bookstein, F. L. A Path-Analytic Model for the Measurement of Selection on Morphology. *Evolution (N. Y.)* **43**, 18 (1989).
44. Pigliucci, M. & Kaplan, J. *Making Sense of Evolution*. (2013).
doi: [10.7208/chicago/9780226668352.001.0001](https://doi.org/10.7208/chicago/9780226668352.001.0001)
45. Bush, A. *et al.* Incorporating evolutionary adaptation in species distribution modelling reduces projected vulnerability to climate change. *Ecol. Lett.* **19**, 1468–1478 (2016).
46. Riebesell, U. & Gattuso, J. Lessons learned from ocean acidification research. *Nat. Clim. Chang.* **5**, 2014–2016 (2015).
47. De Wit, P., Dupont, S. & Thor, P. Selection on oxidative phosphorylation and ribosomal structure as a multigenerational response to ocean acidification in the common copepod *Pseudocalanus acuspes*. *Evol. Appl.* **9**, 1112–1123 (2016).
48. Chakravarti, L. J. *et al.* Can trans-generational experiments be used to enhance species resilience to ocean warming and acidification? *Evol. Appl.* **9**, 1133–1146 (2016).
49. Carrier-Belleau, C., Drolet, D., Mckindsey, C. W. & Archambault, P. Environmental stressors , complex interactions and marine benthic communities ’ responses. *Sci. Rep.* 1–14 (2021). **doi:** [10.1038/s41598-021-83533-1](https://doi.org/10.1038/s41598-021-83533-1)
50. Dam, H. G. & Baumann, H. Climate Change, Zooplankton and Fisheries. *Clim. Chang.*

- 584 *Impacts Fish. Aquac.* 851–874 (2017). doi:10.1002/9781119154051.ch25
- 585 51. Bell, G. Evolutionary rescue and the limits of adaptation. *Philos. Trans. R. Soc. B Biol.*
586 *Sci.* **368**, 1–6 (2013).
- 587 52. Falconer, D. S. *Introduction to Quantitative Genetics*. (Longman Scientific and Technical,
588 1989).
- 589 53. Angilletta Jr, M. J. Estimating and comparing thermal performance curves. *J. Therm. Biol.*
590 **31**, 541–545 (2006).
- 591 54. Feinberg, L. R. & Dam, H. G. Effects of diet on dimensions, density and sinking rates of
592 fecal pellets of the copepod *Acartia tonsa*. *Mar. Ecol. Prog. Ser.* **175**, 87–96 (1998).
- 593 55. Pierrot, D., Lewis, E. & Wallace, D. W. R. MS Excel Program Developed for CO2
594 System Calculations. (2006). doi:10.3334/CDIAC/otg.CO2SYS_XLS_CDIAC105a
- 595 56. Lueker, T. J., Dickson, A. G. & Keeling, C. D. Ocean pCO₂ calculated from dissolved
596 inorganic carbon, alkalinity, and equations for K₁ and K₂: validation based on laboratory
597 measurements of CO₂ in gas and seawater at equilibrium. *Mar. Chem.* **70**, 105–119
598 (2000).
- 599 57. Dickson, A. G. Standard potential of the reaction: $\text{AgCl(s)} + 12\text{H}_2\text{(g)} = \text{Ag(s)} + \text{HCl(aq)}$,
600 and the standard acidity constant of the ion HSO_4^- in synthetic sea water from 273.15 to
601 318.15K. *J. Chem. Thermodyn.* **22**, 113–127 (1990).
- 602 58. Uppstrom, L. R. The boron/chlorinity ratio of deep-sea water from the Pacific Ocean.
603 *Deep. Res. Oceanogr. Abstr.* **21**, 161–162 (1974).
- 604 59. Murray, C. S. & Baumann, H. You better repeat it: Complex CO₂ × temperature effects in
605 Atlantic silverside offspring revealed by serial experimentation. *Diversity* **10**, (2018).
- 606 60. Schank, J. C. & Koehnle, T. J. Pseudoreplication is a Pseudoproblem. *J. Comp. Psychol.*
607 **123**, 421–433 (2009).
- 608 61. Oksanen, L. Logic of experiments in ecology: is pseudoreplication a pseudoissue? *Oikos*
609 **94**, 27–38 (2001).
- 610 62. Therneau, T. *_A Package for Survival Analysis in S_*. (2015).
- 611 63. Lande, R. & Arnold, S. J. The Measurement of Selection on Correlated Characters.
612 *Evolution (N. Y.)*. **37**, 1210 (1983).
- 613 64. Rosseel, Y. Lavaan: An R package for structural equation modeling. *J. Stat. Softw.* **48**,
614 (2012).
- 615 65. Epskamp, S., Stuber, S., Nak, J., Veenman, M. & Jorgensen, T. D. semPlot: Path
616 Diagrams and Visual Analysis of Various SEM Packages' Output. (2019).
- 617 66. Bolger, A. M., Lohse, M. & Usadel, B. Trimmomatic: A flexible trimmer for Illumina
618 sequence data. *Bioinformatics* **30**, 2114–2120 (2014).
- 619 67. Jørgensen, T. S. *et al.* The genome and mRNA transcriptome of the cosmopolitan calanoid
620 copepod *acartia tonsa* dana improve the understanding of copepod genome size evolution.

621 *Genome Biol. Evol.* **11**, 1440–1450 (2019).

622 68. Li, H. Aligning sequence reads, clone sequences and assembly contigs with BWA-MEM.
623 **00**, 1–3 (2013).

624 69. Li, H. *et al.* The Sequence Alignment/Map format and SAMtools. *Bioinformatics* **25**,
625 2078–2079 (2009).

626 70. Kofler, R. *et al.* Popoolation: A toolbox for population genetic analysis of next generation
627 sequencing data from pooled individuals. *PLoS One* **6**, (2011).

628 71. R Core Team. R: A language and environment for statistical computing. (2013).

629 72. Wood, S. N. Fast stable restricted maximum likelihood and marginal likelihood estimation
630 of semiparametric generalized linear models. *J. R. Stat. Soc. Ser. B Stat. Methodol.* **73**, 3–
631 36 (2011).

632 73. Simpson, G. L. Modelling palaeoecological time series using generalised additive models.
633 *Front. Ecol. Evol.* **6**, 1–21 (2018).

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Figure Legends:

Figure 1 – Mean egg production and mean hatching success for transgenerational experiment. Egg production rate (EPR, histograms, left Y axis) and egg hatching success (HS, lines, right Y axis) during the transgenerational experiment. **A)** Ambient conditions, AM: Blue (top left) – ambient temperature (18 °C) and CO₂ (400 µatm); **B)** Ocean acidification conditions, OA: Green (top right) – ambient temperature and high CO₂ (2000 µatm); **C)** Ocean warming conditions, OW: Orange (bottom left) – high temperature (22 °C) and ambient CO₂; **D)** Ocean warming and acidification, OWA: Red (bottom right) – high temperature and high CO₂. Error bars represent 95% confidence intervals. Probability (p) values for EPR and HS for each panel are derived from the effect of generation on each trait taken from generalized additive models. Letters represent statistically similar groupings for post-hoc t-test comparison. Box and whiskers plots of the same graphs are displayed on SI Fig. 6.

Figure 2 – Survival for transgenerational study. Mean survival is for nauplius stage 1 to adult stage. Error bars represent 95% confidence intervals. Asterisks indicate significant decrease ($p < 0.001$). Treatment colors are the same as histograms in Fig. 1. Treatment curves are offset for clarity. Box and whiskers plots of the same graphs are displayed on SI Fig. 7.

Figure 3 – Fitness. Mean fitness values, λ , calculated for the transgenerational study. Error bars represent 95% confidence intervals. Treatment colors are the same as histograms in Fig. 1. Treatment curves are offset for clarity. Box and whiskers plots of the same graphs are displayed on SI Fig. 8.

Figure 4 – Fitness landscapes. Frequency distributions (left axes) and relative fitness landscape linear models (right axes) for **A)** Hatching success, **B)** Egg production rate, and **C)** Survival at F0 (black) and F25 (red). Fitness landscapes estimate the correlations between changing quantitative phenotypes and changing relative fitness. Shading around lines represents standard error.

Tables:

Table 1 – Generational effects on fitness. Zero-inflated model results of generation and treatment on fitness, with replicates as random effects. Count model results represents when $\lambda=0$ values are omitted. Control treatment is the reference intercept for the model. Interactive effect of generation on AM conditions is noted as just “Generation” for each model results.

<i>Predictors</i>	<i>Estimates</i>	<i>CI</i>	<i>P</i>
Count model			
(Intercept)	1.19	1.16-1.23	<0.001
Generation	-0.00	-0.00 – -0.00	0.001
OA	0.01	-0.04 – 0.06	0.817
OW	0.04	-0.01 – 0.09	0.115
OWA	0.13	0.07 – 0.18	<0.001
Generation × OA	0.00	-0.00 – 0.00	0.951
Generation × OW	0.00	0.00 – 0.00	0.009
Generation × OWA	-0.01	-0.01 – -0.01	<0.001
Zero-inflated model			
(Intercept)	-1.97	-2.62 – -1.32	<0.001
Generation	-0.01	-0.04 – 0.03	0.681
OA	0.55	-0.36 – 1.46	0.233
OW	1.27	0.39 – 2.16	0.005
OWA	1.54	0.65 – 2.44	0.001
Generation × OA	-0.03	-0.07 – 0.02	0.212
Generation × OW	-0.04	-0.09 – -0.00	0.038
Generation × OWA	-0.19	-0.25 – -0.13	<0.001
Random effects			
σ^2	0.01		
τ_{00} Replicate	0.00		
Intra-correlation coefficient	0.08		
N Replicate	16		
Observations	2867		
Marginal R^2 /Conditional R^2	0.230/0.293		

Table 2 – Path analysis and effects on lambda. Standardized parameter estimates from structural equation models for each trait’s effect on lambda at both F0 (gray bars) and F25 (white bars) for each treatment. The highest parameter estimate for each treatment at each generation is highlighted in bold.

	AM	OA	OW	OWA
<i>Survival</i>	0.12	-0.09	0.03	0.08
	-0.09	0.13	0.72	0.33
<i>EPR</i>	-0.07	0.04	0.03	0.10
	0.34	0.25	0.86	0.23
<i>HS</i>	0.91	0.94	0.96	0.78
	0.53	0.65	0.11	0.70
<i>Sex ratio</i>	0.12	-0.15	0.09	0.21
	-0.39	0.37	0.37	-0.27